



Prepared in cooperation
with the Glen Canyon Dam
Adaptive Management Program

Glen Canyon Dam Adaptive Management Program Triennial Budget and Work Plan— Fiscal Years 2025–2027

Prepared by
Bureau of Reclamation
Upper Colorado Regional Office
and
U.S. Geological Survey
Grand Canyon Monitoring and
Research Center



Cover Background:

Upper Little Colorado River, Navajo Nation. Photo by Jeff Muehlbauer, U.S. Geological Survey.

Inset:

- ~ Glen Canyon Dam. Flows from bypass tubes designed to disrupt establishment of nonnative smallmouth bass. Photo by U.S. Department of the Interior, Bureau of Reclamation.
- ~ Sacred Datura flower (*Datura wrightii*). Photo by Emily Palmquist, U.S. Geological Survey.
- ~ Glen Canyon, Glen Canyon National Recreation Area. Photo by Joel Sankey, U.S. Geological Survey.
- ~ Humpback chub. Photo by David Herasimtschuk, Freshwaters Illustrated.



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Prepared by

U.S. Geological Survey
Southwest Biological Science Center
Grand Canyon Monitoring and Research Center
Flagstaff, Arizona

and

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Upper Colorado Regional Office
Salt Lake City, Utah

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**U.S. Department of the Interior
U.S. Geological Survey**

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Abbreviations

- AGFD: Arizona Game and Fish Department
- AICc: Akaike Information Criterion
- AIS: Aquatic Invasive Species
- AMWG: Adaptive Management Work Group
- AMR: Active Metabolic Rate
- ASU: Arizona State University
- AWS: Amazon Web Services
- AZDEQ: Arizona Department of Environmental Quality
- BACI-based analysis: Before-After-Control-Impact experiment design
- BAHG: Budget Ad Hoc Group
- BO or BiOp: Biological Opinion
- BTELSS: Brown Trout Early Life Stage Survey
- Bug Flows: Macroinvertebrate Production Flows
- CE-QUAL-W2: A water quality and hydrodynamic model
- CHIRP: Frequency-modulated echosounder system
- CHS: Cloud Hosting Solutions
- CMINs: Core monitoring information needs
- CMP: Core Monitoring Plan
- CPUE: Catch-per-unit-effort
- CPUs: Central Processing Units
- CR: Colorado River
- CRe: Colorado River ecosystem
- DASA: Data Acquisition, Storage, and Analysis Program
- DEMs: Digital Elevation Models
- DFCs: Desired Future Conditions
- DOE: Department of Energy
- DOI: Department of Interior
- DSM: Digital surface model
- EarthMAP: Earth Mapping, Analysis, and Processing
- eDNA: Environmental DNA
- EGC: Eastern Grand Canyon
- EGIS: Enterprise GIS
- EIS: Environmental Impact Statement
- EMA: USGS Ecosystem Mission Area
- EMINs: Effects of monitoring information needs
- EPT: Aquatic insect orders Ephemeroptera, Plecoptera, Trichoptera
- EROS: USGS Earth Resources Observation and Science Center
- ESA: Endangered Species Act
- ESRI: Environmental Systems Research Institute

- FACA: Federal Advisory Committee Act
- FCDC: Federal Geographic Data Committee
- FGDC: Federal Geographic Data Committee
- FLAHG: FLOW Ad Hoc Group
- FLIR: Forward looking infrared radar camera
- ft³/s: cubic feet per second (cfs)
- FY: Fiscal year
- GCD: Glen Canyon Dam
- GCDAMP: Glen Canyon Dam Adaptive Management Program
- GCMRC: Grand Canyon Monitoring and Research Center
- GCNP: Grand Canyon National Park
- GCPA: Grand Canyon Protection Act
- GIS: Geographical Information Systems
- GLCA: Glen Canyon National Recreation Area
- GLMMs: Generalized linear mixed models
- GNSS: Global navigation satellite system
- GPP: Gross primary production
- GPS: Geographic Positioning System
- GRAV-D: American Vertical Datum
- DSM: Digital Surface Model
- HFE: High-Flow Experiment
- HPP: Historic Preservation Plan
- ISCO samplers: Water monitoring samplers
- ITAC: Information Technology Advisory Council
- JCM: Juvenile chub monitoring
- LCR: Little Colorado River
- LRRs: Log-response ratios
- LTEMP: Long-term Experimental and Management Plan
- LTEMP EIS: Long-Term Experimental and Management Plan Environmental Impact Statement
- LTEMP ROD: Long-term Experimental and Management Plan Record of Decision
- MNA: Museum of Northern Arizona
- MR: Metabolic Rate
- NAD83: North American Datum of 1983
- NATRF2022: North American Terrestrial Reference Frame of 2022
- NCPN: Northern Colorado Plateau Network Inventory and Monitoring Network
- NGSIDB: National Geodetic Survey integrated database
- NHDPlus: National Hydrography Dataset Plus
- NHPA: National Historic Preservation Act
- NO: Natal origins
- NPS: National Park Service
- NRHP: National Register of Historic Places
- NSSDA: National Standard for Spatial Data Accuracy

- NWIS: National Water Information System
- OEI: USGS Office of Enterprise Information
- OSL: Optically Stimulated Luminescence
- OWI: USGS Office of Water Information
- PA: Programmatic agreement
- PEP: Protocol Evaluation Panel
- PIs: Principle Investigators
- PIT-tag: Passive integrated transponder electronic tag
- RAMPS: Restoration Assessment and Monitoring Program
- RINs: Research information needs
- RM: River mile
- ROD: Record of Decision
- RTELSS: Rainbow trout early life stage studies
- SBSC: Southwest Biological Science Center
- SCADA: Supervisory Control and Data Acquisition
- SEAHG: Socioeconomics Ad Hoc Group
- SEINet: Southwest Environmental Information Network
- SMBI: Sand mass balance index
- SMR: Standard metabolic rate
- SRP: Soluble reactive phosphorous
- SSP: Strategic Science Plan
- SSQs: Strategic science questions
- SUR: Submersible ultrasonic receivers
- TCPs: Traditional Cultural Properties
- TELS: Trout early life stage
- TMFs: Trout management flows
- TRGD: Trout reproductive and growth demographics
- TWG: Technical Work Group
- TWP: Triennial Work Plan
- USFWS: U.S. Fish and Wildlife Service
- USGS: U.S. Geological Survey
- VIE: Visible implant elastomer tags
- YOY: Young-of-year
- WAPA: Western Area Power Administration
- WSC: USGS Water Science Center

Project A: Streamflow, Water Quality, and Sediment Transport and Budgeting in the Colorado River Ecosystem

Investigators

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U.S. Geological Survey, Southwest Biological Science Center, Grand Canyon Monitoring and Research Center

Project Summary and Purpose

The primary linkage between Glen Canyon Dam operations and the characteristics of the physical, biological, and cultural resources of the Colorado River ecosystem (CRE) downstream from Glen Canyon Dam is through the stage, discharge, water quality, and sediment transport of the Colorado River. This project makes and interprets the basic measurements of these parameters at locations throughout the CRE. Project A thus collects the physical data that directly link dam operations to all resources in the downstream CRE. The data collected by this project are used to implement the High-Flow Experiment (HFE) Protocol (i.e., trigger and design HFE hydrographs), to evaluate the segment-scale sand mass-balance response to the HFE Protocol (U.S. Department of the Interior, 2011; Grams and others, 2015), and to evaluate the downstream effects of releases conducted under the Long-Term Experimental and Management Plan (LTEMP) Environmental Impact Statement (EIS) (U.S. Department of the Interior, 2016a, b). Two of the metrics proposed to evaluate LTEMP management for sediment are measured by this project.

Long-Term Experimental and Management Plan (LTEMP) Resource Goals

Project A supports the following ten LTEMP resource goals:

- Archaeological and Cultural Resources
- Natural Processes
- Humpback Chub
- Hydropower and Energy (through ramping-rate effects)
- Other Native Fish
- Recreational Experience
- Sediment
- Rainbow Trout Fishery
- Nonnative Invasive Species
- Riparian Vegetation

The data collected by Project A are used by most of the physical, ecological, and socio-cultural projects funded by the Glen Canyon Dam Adaptive Management Program (GCDAMP). Although a major emphasis of Project A is supporting the LTEMP sediment goal (i.e., “Increase and retain fine sediment volume, area, and distribution in the Glen, Marble, and Grand Canyon reaches above the elevation of the average base flow for ecological, cultural, and recreational purposes”), the basic data collected by this project supports the other nine LTEMP goals listed above. Most of the project funds support basic data collection at U.S. Geological Survey (USGS) gaging stations, with the remainder funding data interpretation. The funds requested under this proposal cover ~80% of the costs required to collect surface-water, water-quality, and sediment data at the network of USGS gaging stations used by this project. An additional approximately \$184,000 to support this network is provided directly to the USGS Arizona Water Science Center from funds appropriated by Congress for the USGS, the Bureau of Reclamation, and from funds provided by the Arizona Department of Water Resources. Project A is designed to provide measurements of stage (i.e., water-surface elevation), discharge (i.e., streamflow), water quality, and suspended sediment at sufficiently high temporal resolutions (~15-minute) to resolve changes in these parameters and to allow accurate determination of suspended-sediment loads for use in sediment budgeting (Grams and others, 2019; Topping and others, 2021; Griffiths and others, 2024). The proposed monitoring under this project is a continuation of that conducted over the last 24 years. Work conducted under the previous work plan indicates that sand storage in the channel and sandbars of the CRe is not likely sustainable unless tributary sand inputs remain well above average and dam releases remain slightly below average (Topping and others, 2021; Griffiths and others, 2024). The work proposed in this current work plan is therefore that required to address this important conclusion.

Background

Systematic measurements of streamflow and water quality, including suspended-sediment concentration, in the CRe began with installation of the Lees Ferry gaging station (USGS gaging station 09380000, Colorado River at Lees Ferry, AZ) in May 1921 (Howard, 1947; Topping and others, 2003). During much of the 20th century, daily measurements of suspended-sediment concentration and water temperature, and episodic measurements of other water-quality parameters, were made by the USGS at multiple gaging stations in the CRe and on key tributaries.

This intensive period of measurements ended in the early 1970s (Andrews, 1991; Topping and others, 2000a). Concern about the effects of the operation of Glen Canyon Dam on the CRe resulted in a new emphasis on scientific measurements and modeling of water quality and sediment transport beginning in the early 1980s (National Research Council, 1996). The results of these studies have been published in numerous USGS reports and journal articles, and ultimately resulted in the current form of Project A.

The operation of Glen Canyon Dam controls the CRe because it is the dominant controller of river stage, discharge, and water quality, and is a primary regulator of sediment transport, erosion, and deposition (Topping and others, 2000a, b, 2003, 2021; Rubin and Topping, 2001, 2008; Gloss and others, 2005; Griffiths and others, 2024). Water temperature, salinity, dissolved oxygen, and water chemistry at the foot of the dam are determined by the physical and chemical characteristics of the reservoir water at the penstock and/or jet-tube elevations on the upstream face of the dam (Vernieu and others, 2005). Because the amount of water supplied by downstream tributaries is small and large floods on these tributaries are infrequent, dam operations largely determine stage, discharge, and key water-quality parameters (water temperature and salinity) throughout the CRe (Wiele and Smith, 1996; Wiele and Griffin, 1998; Voichick and Wright, 2007; Voichick, 2008; Wright and others, 2009; Voichick and Topping, 2010). In addition, because sediment transport in the CRe is controlled by both changes in discharge and changes in bed-sediment grain size (Rubin and Topping, 2001; 2008; Rubin and others, 2020; Topping and others, 2021), and because dam operations control discharge, the operation of Glen Canyon Dam acts as a primary regulator of sediment transport in the CRe. As dam operations regulate the amount of sediment in the water column (i.e., suspended sediment), and because suspended sediment largely determines turbidity (Voichick and Topping, 2014; Voichick and others, 2018), dam operations therefore influence downstream turbidity in the CRe. Finally, because dam operations largely determine water temperature and also influence turbidity, other downstream water-quality parameters regulated by water temperature and turbidity, such as dissolved oxygen, are also affected by dam operations throughout the CRe (Hall and others, 2015; Deemer and others, 2022). Dissolved oxygen is generally negatively related to water temperature, turbidity, and suspended-sediment concentration, e.g., data at https://www.gcmrc.gov/discharge_qw_sediment/. Because water temperature, turbidity, and dissolved oxygen in the CRe are either controlled or regulated by operation of Glen Canyon Dam, dam operations thereby exert a strong control on gross primary productivity throughout the CRe (Deemer and others, 2022).

Suspended sediment is an important water quality parameter in the CRe for several reasons. First, deposition and/or erosion of the eddy sandbars and channel-margin deposits important to many biological, cultural, and recreational resources are directly controlled by the transport of sand (Gloss and others, 2005). The rates of deposition and/or erosion of eddy sandbars and channel-margin deposits are related by mass conservation to spatial gradients in the suspended-sand flux (after Exner, 1920, 1925).

By theory (Grams and others, 2013) and experiments (Schmidt and others, 1993), eddy-sandbar deposition is most efficient when the flux of suspended sand is the highest in the main channel of the river. Because suspended-sand flux is the depth-integrated product of suspended-sand concentration and water discharge, maximum main-channel sand flux occurs when the concentration of suspended sand is the highest (determined largely by the water discharge and bed-sand grain-size distribution, e.g., Topping and others, 2007) and the velocity is the highest (i.e., at higher discharge).

Under these conditions, the convergence (i.e., negative spatial gradient) in the sand flux between the main channel and the riverbank in an eddy is the largest, leading to the greatest sand deposition rates in an eddy (Topping and others, 2010; Grams and others, 2013).

Data collected during HFEs confirmed that eddy sandbars are, in fact, built most efficiently when the discharge is high and the amount of finer sand in a segment is maximized (Topping and others, 2019). Thus, effective management of eddy sandbars and associated resources in different segments of the CRe requires managers to know when finer sand is maximized in those specific segments when designing HFEs (Wright and others, 2005, 2008; Topping and others, 2010).

The second major reason as to why suspended sediment is an important water quality parameter is that it largely determines turbidity, and therefore influences the aquatic and fish ecology of the river (Voichick and others, 2016; Yackulic and others, 2018). The endemic fishes of the CRe evolved in a highly turbid river (Gloss and Coggins, 2005). Turbidity is primarily determined by the concentration of suspended silt and clay and, to a lesser degree, suspended sand (Voichick and Topping, 2014; Voichick and others, 2018). Because closure of Glen Canyon Dam cut off the upstream supply of silt and clay, the post-dam Colorado River in Marble and Grand canyons is much less turbid than ever occurred naturally (Voichick and Topping, 2014). Although on average turbidity increases in a stepwise fashion in the downstream direction (at the mouths of the Paria and Little Colorado rivers), the Colorado River is only highly turbid during periods of tributary flooding (Voichick and Topping, 2014). Through turbidity, suspended sediment controls gross primary productivity throughout the CRe (Deemer and others, 2022).

The transport of suspended sediment in the CRe is controlled by both the discharge released from the dam and the episodic tributary resupply of sand, silt, and clay (Topping and others, 2000b, 2021). The fining of the bed sand in the Colorado River following a tributary flood and the daily increases in discharge caused by dam operations can both cause several orders of magnitude increase in suspended-sand concentration (Topping and others, 2021). This finding – that sand transport in the post-dam Colorado River was essentially co-equally regulated by changes in discharge and changes in bed-sand grain size (Rubin and Topping, 2001; 2008) – refuted key aspects of the 1995 Glen Canyon Dam EIS (U.S. Department of the Interior, 1995, 1996). The 1995 EIS incorrectly assumed that sand transport was regulated only by changes in discharge (Rubin and others, 2002). Rather, the amount of sand that can be stored in each river segment in the CRe is controlled by the interaction of dam releases with the grain size of the sand on the bed (Topping and others, 2021).

Owing to the influence of changing grain size, the residence time of tributary-supplied sand in the CRe is much shorter than assumed in the 1995 EIS (Topping and others, 2000b, 2021; Rubin and others, 2002; Wright and others, 2005). Thus, the 1995 EIS management strategy – of using multi-year accumulation of tributary-supplied sand for sandbar rebuilding in the CRe during relatively rare HFEs – was not valid (Rubin and others, 2002; Wright and others, 2005).

These findings led to the current design of Project A and led to the management strategies for sediment described in the LTEMP EIS, where HFEs are designed based on the availability of tributary-supplied sand (Wright and Kennedy, 2011; Grams and others, 2015).

Sand management in the CRe is challenging because Glen Canyon Dam has cut off almost all the natural sand supply to the CRe, and the remaining tributary sand supply to the CRe has been declining over time. By cutting off most of the sediment formerly supplied to the Colorado River in Marble and Grand canyons (Figure 1), closure of Glen Canyon Dam in 1963 reduced the supply of sand, silt and clay at the upstream boundary of Grand Canyon National Park to ~5% of its pre-dam amount (Topping and others, 2000a). Although other smaller tributaries downstream from the dam do supply sand, silt, and clay to the CRe, the Little Colorado River (LCR) is the largest supplier of silt and clay and the Paria River is by far the largest supplier of sand (Topping and others, 2021). Though they generally supply only a small fraction of the sand supplied by the Paria River, the smaller tributaries can supply greater amounts of sand during rare years (Griffiths and Topping, 2017). For example, during the Low Summer Steady Flow experiment in summer 2000, House Rock Wash supplied more sand to the Colorado River than either the Paria River or LCR (Schmidt and others, 2007; Griffiths and Topping, 2017). Although the LCR was historically the largest sand supplier, non-climatic changes in the LCR basin from water development and biogeomorphic feedbacks, i.e., channel narrowing and floodplain growth associated with vegetation encroachment (Dean and Topping, 2019, 2024), have caused a likely permanent decline in LCR floods and greatly curtailed the delivery of sand to the CRe from the LCR (Dean and Topping, 2019, 2024).

In addition to its sand-supply implications, this progressive decline in geomorphic disturbance in the LCR has likely negative implications for the spawning habitat of the endangered humpback chub (*Gila cypha*) in the lower LCR (Unema and others, 2021). In addition to the changes in the LCR, an apparent regional decline in winter-spring tributary floods has resulted in the summer-fall season now being the only season of dependable larger sand-supplying events (Topping and others, 2021). Thus, sand management in the CRe can typically only utilize sand supplied by the Paria River during summer-fall thunderstorms.

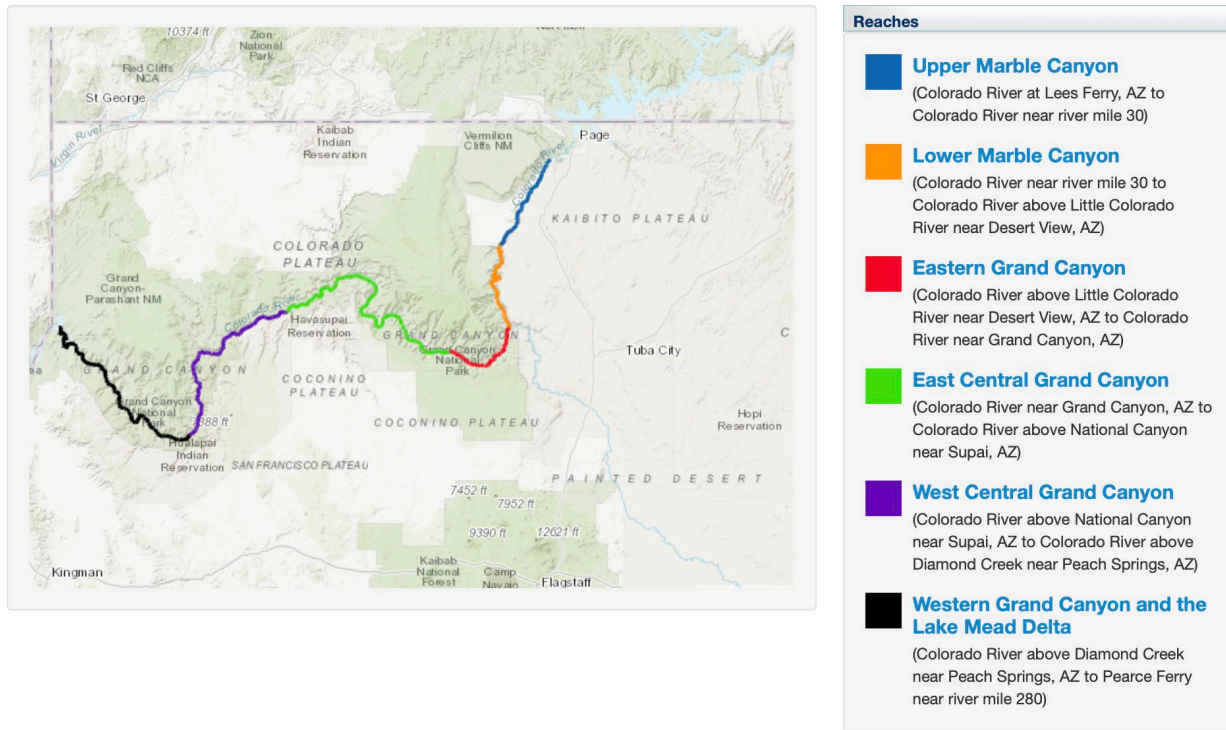


Figure 1. Map showing the extents of the six segments (a.k.a. reaches) of the Colorado River in Marble and Grand canyons in which mass-balance sand budgets are constructed. The seventh monitoring reach, lower Glen Canyon, lies upstream from Upper Marble Canyon. Map taken from Project A's website (U.S. Geological Survey, 2024).

Owing to the dual controls of discharge and grain size on sand transport, the sand supplied to the Colorado River during tributary floods migrates rapidly downstream as an elongating sand wave (Topping and others, 2000b, 2021). The leading edges of these waves migrate downstream at a velocity slightly slower than the velocity of the water. Thus, the leading edges of sand waves exit the CRE within days of a large tributary sand-supplying flood, with this migration rate increasing as a function of discharge. The bed-sand grain-size changes that accompany the migration of these waves cause the sand waves to bifurcate into two packets.

The first leading packet is composed of the finest size classes of sand, is transported as quasi-washload, and is fully transported to Lake Mead within a week of a large tributary flood. The second packet lags the first packet, includes most of the sand supplied during a large tributary flood, and migrates downstream more slowly in the Colorado River, taking several hundreds of days to transit the CRE under most dam operations. For example, following large Paria River floods, the finest part of this second packet takes on average ~63 days to transit Upper Marble Canyon (above river mile 30) and ~144 days to exit Marble Canyon (Topping and others, 2021). This result indicates that the most efficient rebuilding of sandbars in Marble Canyon will occur during HFEs conducted within 60 to 140 days after a large Paria River flood, except when dam releases are abnormally low such as during 2021 through early 2023 when low dam releases allowed for multi-year retention of Paria-supplied sand.

Project A's measurements indicate that these sand-wave dynamics cause large coupled longitudinal gradients in bed-sand grain size and suspended-sand concentration. As expected, based on Exner (1920, 1925) and Smith (1970), these measured gradients cause net sand deposition in some segments while they cause net sand erosion in other segments, with similar longitudinal gradients in bed-sand grain size being associated with larger amounts of either deposition or erosion at higher discharge (Topping and others, 2021). Thus, owing to downstream migration of tributary-generated sand waves in the CRe, the same dam operation will cause erosion in one segment while it causes deposition in another, with higher dam operations simply causing greater amounts of erosion or deposition in these segments.

Moreover, the locations where any dam operation causes erosion or deposition will change with time as the sand wave migrates downstream. Consequently, sand budgets in the CRe do not generally get more positive in the downstream direction; downstream segments in Grand Canyon may erode while upstream segments closer to the dam gain sand and *vice versa* (Topping and others, 2021).

For this reason, and because of a large decline in the LCR sand supply, sand erosion has been recently more prevalent in downstream segments in Grand Canyon than in Marble Canyon. It is thus fundamentally wrong to assume that monitoring in only Marble Canyon is sufficient to know the status of sand throughout the CRe. Thus, effective CRe sand management cannot utilize empirical relations fit to sparse datasets, but rather requires continuous sand-transport monitoring in key tributaries and at a variety of locations along the Colorado River, as is done in Project A (Griffiths and others, 2012).

Sand management in the CRe is therefore difficult because the tributary sand supply is limited, and sand storage in the CRe is self-limited by the combined effects of grain size and discharge. Although tributary floods supply sand to the Colorado River, these floods also cause the bed-sand grain size to decrease, thereby greatly increasing the downstream transport of sand. Owing to this grain-size effect, tributary flooding causes both sand storage and sand export to increase. Because sand transport increases nonlinearly as a function of discharge, this grain-size effect leads to the retention of less sand in the Colorado River at higher discharge. Consequently, multi-year sand accumulation is only possible in the CRe during years of well-above-average tributary sand supply and below-average dam releases (Figure 2; Topping and others, 2021). Sand only accumulates in Marble Canyon during years when the Paria River sand supply is >124% of average and annual-mean dam releases are below $\sim 13,400$ ft³/s (9.7 million acre-feet). Similarly, sand only accumulates in Grand Canyon during years when the combined Paria River and LCR sand supply is >136% of average and annual-mean dam releases are typically well below $\sim 14,100$ ft³/s (10.2 million acre-feet). In lower sand-supply years, sand is eroded from Marble Canyon during years where the annual-mean discharge exceeds $\sim 11,100$ ft³/s (8.0 million acre-feet) and from Grand Canyon when the annual-mean discharge exceeds $\sim 11,300$ ft³/s (8.2 million acre-feet).

Regardless of the magnitude of the Paria River sand supply, sand is eroded from Marble Canyon when the annual-mean discharge exceeds $\sim 13,600 \text{ ft}^3/\text{s}$ (9.8 million acre-feet). Thus, maintaining a level of sand storage sufficient for maintaining sandbars in the CRe may require timing periods of higher and lower dam releases based on the tributary sand-supply conditions. Whether the sand resources of the CRe can be sustainably managed in perpetuity therefore remains an open question.

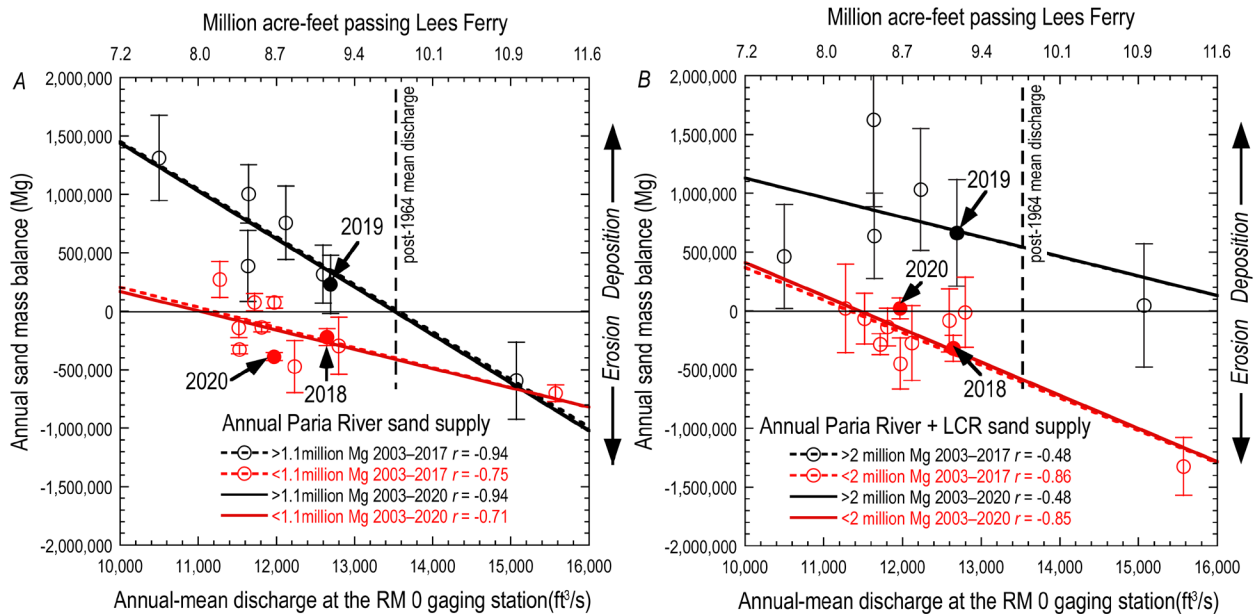


Figure 2. Figure modified from Griffiths and others (2024) Figure 2. Annual sand mass balance plotted as a function of annual-mean discharge and annual dam release (in acre-feet) at the Colorado River at Lees Ferry, AZ (RM 0) gaging station by sediment year (July 1–June 30) for (A) Marble Canyon and (B) Grand Canyon. Data are separated into relatively high (black) and relatively low (red) annual tributary sand supply. The sediment-year 2018–2020 values from Griffiths and others (2024) are depicted as filled circles, with each year labeled; values from previous years reproduced from Topping and others (2021) depicted as unlabeled open circles. The least-squares linear regressions fit to each sand-supply condition for the sediment-year 2003–2017 data from Topping and others (2021) are depicted as dashed lines; regressions fit to each sand-supply condition for all data from sediment years 2003–2020 are depicted as solid lines. Correlation coefficients (r) associated with each regression are shown. Error bars indicate the magnitudes of the uncertainties in annual sand mass balances propagated through equation 5 in Topping and others (2021). The post-1964 mean discharge is the mean discharge at the RM 0 gaging station during sediment years 1965–2020.

There is no “short cut” to sediment monitoring in a river like the Colorado River where large changes in sand transport occur independently of the discharge of water; ongoing continuous suspended-sediment measurements are required. Monitoring the CRe therefore requires a strategy where the CRe is divided into segments based on key tributaries that supply sediment and affect water quality, with continuous monitoring at stations bracketing these reaches (Figure 1). This is the strategy used in Project A since it began in the early 2000s, where the CRe is divided into seven monitoring segments (a.k.a. reaches) bracketed by USGS gaging stations.

At each of these stations, stage, discharge, water temperature, specific conductance, dissolved oxygen, and turbidity are measured continuously at 15-minute intervals. At the downstream five of these stations in Marble and Grand canyons, suspended-silt-and-clay concentration, suspended-sand concentration, and suspended-sand grain size are also measured at 15-minute intervals using the methods of Topping and Wright (2016).

These streamflow and suspended-sand data are used to compute the sand loads that are, in turn, used in the six user-interactive mass-balance sand budgets (Topping and others, 2021) served on Project A's website (Sibley and others, 2015); sand budgets located at https://www.gcmrc.gov/discharge_qw_sediment/reaches/GCDAMP.

These sand budgets (described in detail in Topping and others, 2021) are used to evaluate the near-realtime continuous effects of dam operations, including special LTEMP releases for invertebrates (i.e., bug flows), trout management, and releases for smallmouth bass (*Micropterus dolomieu*) suppression on sand resources throughout the CRe, and used in the design and evaluation of HFEs.

Comparison of our continuous mass-balance sand budgets with topographic-based sand budgets measured by Project B.2 indicates that the mass-balance budgets are accurate at our specified level of uncertainty. Repeat mapping of lower Marble Canyon (river miles 30–61) indicates that 640,000 ±350,000 metric tons of sand were eroded from this segment between May 2009 and May 2012 (Grams and others, 2019). During this same period, our flux-based sand budget indicates that 690,000 ±320,000 metric tons of sand were eroded from lower Marble Canyon. In eastern Grand Canyon (river miles 61–87), repeat mapping indicates that 630,000 ±480,000 metric tons of sand were eroded between April 2011 and May 2014, whereas our flux-based sand budget indicates 740,000 ±610,000 metric tons of sand erosion. These results from two independent comparisons (i.e., different river segments and time periods) indicate that our continuous mass-balance sand budgets are sufficiently accurate to inform managers of the effects of dam operations on CRe sand resources over timescales ranging from sub-hourly to multiple years (Topping and others, 2021). The observed rates of increase in uncertainty over time in the Project A mass-balance sand budgets (Topping and others, 2021) indicate that the repeat mapping of the CRe conducted by Project B.2 should continue at least once-per-decade canyon-wide to verify Project A's evaluation of LTEMP sand management. To meet this canyon-wide requirement, individual mass-balance river segments (Figure 1) should be mapped at least once every three years to ensure the entire CRe gets mapped once per decade.

Only Project A can inform managers of the real-time effect of dam operations on downstream resources in the CRe. As such, the GCDAMP-Technical Work Group approved Project A as Core Monitoring in October 2008. Because we collect and serve data at 15-minute intervals, Project A can inform how dam operations affect stage, discharge, water temperature, salinity, turbidity, dissolved oxygen, sediment transport, and sand erosion and deposition in key reaches throughout the CRe on a 15-minute basis.

Specific to the LTEMP sediment goal, this capability allows ramping rates and daily ranges to be linked to their effects on sand resources under a wide range of sediment-supply conditions (e.g., during periods of sand enrichment after large tributary floods vs. during periods of sand depletion following extended equalization releases).

Proposed Work

The work proposed herein is a continuation of that conducted under the previous work plan, with transfer of some Principal Investigator's time and computer-science support to other projects. Ten pay periods of David Dean's time were transferred to Project C.4 and the computer-science support for Project A's database and website was transferred to Project K. These two transfers reduced the annual gross cost of Project A by roughly \$150,000.

As in all previous work plans, new interpretive products are planned to address the guiding hypotheses of how dam releases directly affect sediment and other resources in the CRe and to build on the conclusions of the work funded during FY 2021–24. The Project A data-collection network was developed and made progressively more efficient over the last 24 years. It is now likely the cheapest, most-efficient monitoring network required to address the LTEMP sediment goal and support nine other LTEMP goals. This network relies extensively on 1) new technologies to automatically monitor streamflow, water quality, and sediment, and 2) cost sharing to reduce costs while not sacrificing the data accuracies required by the LTEMP goals. Collecting the data for Project A using only conventional sampling methods instead of incorporating new technologies would require an estimated annual budget in excess of \$9 million.

Research on the Colorado and on other rivers has shown that, to be meaningful, measurements of stage, discharge, water temperature, specific conductance, turbidity, dissolved oxygen, and suspended sediment must be made at temporal intervals shorter than those over which these parameters vary. Owing to the effects of dam operations and tributary floods, substantial changes in all these parameters occur over timescales less than one hour (Figure 3; Wiele and Smith, 1996; Wiele and Griffin, 1998; Topping and others, 2000b, 2003, 2010; Voichick and Wright, 2007; Voichick, 2008; Wright and others, 2009; Voichick and Topping, 2010, 2014; Grams and others, 2019). Project A was therefore designed to provide measurements of stage, discharge, water quality, and suspended sediment at the required accuracies and sufficiently high temporal resolutions (~15-minutes) to capture the variability in these parameters. Specifically, for suspended sediment, this temporal resolution was chosen to be shorter than the sub-hourly data interval required to know both the sign and magnitude of change in sediment budgets (Grams and others, 2019). Collection of data at 15-minute intervals is the USGS standard. Months to years of data collected at this resolution easily fit on modern dataloggers, result in less processing time in the office, and reduce financial costs to the project. In addition, the efficiencies of such largely automatic data collection require less field time, such that only two river trips are now required annually for this project.



Figure 3. Plots of 15-minute gage height (i.e., stage, water elevation), water discharge, water temperature, specific conductance, turbidity, dissolved oxygen, suspended-silt-and-clay concentration, and suspended-sand concentration for the two-week period 8-2-2015 through 8-16-2015 at the Colorado River above Diamond Creek gaging station. Light blue dots with 95%-confidence-level error bars indicate episodically measured silt and clay concentrations and sand concentrations (from physical suspended-sediment samples) used to verify the two-frequency acoustical suspended-sediment measurements. Variability in each of these parameters over this two-week period arises from the interaction of dam operations with tributary floods. Plots from Project A’s website at https://www.gcmrc.gov/discharge_qw_sediment/ (U.S. Geological Survey, 2024).

A map showing the locations at which data are collected/ utilized by Project A can be viewed at https://www.gcmrc.gov/discharge_qw_sediment/stations/GCDAMP. The gold circles on this

map depict the active stations where data are currently collected; the gray circles depict the inactive stations where data were collected historically. Note that the GCDAMP does not fund the data collection at all stations on this map. The data collected/used by Project A are used to evaluate the near-real-time effects of all LTEMP dam releases on stage, discharge, water quality, sediment transport, and sediment storage in the CRe (U.S. Department of the Interior, 2016a, b). The continuous mass-balance sand budgets provide the measurement-based "ground-truthing" of the Sand Mass Balance Index (SMBI) developed in Appendix E of the LTEMP EIS (U.S. Department of the Interior, 2016a). Higher values of the SMBI in the LTEMP EIS were taken as indicators of increased sand storage in the CRe, with increases in sand storage indicating an increase in the sand available to be deposited in sandbars during HFEs. In addition, the sand-transport data and mass-balance sand budgets from Project A are used to trigger HFEs, design the hydrograph of HFEs, and evaluate the effects of HFEs on sand storage in the CRe, as described by U.S. Department of the Interior (2016a).

All data collected by Project A are served and can be downloaded at our website at https://www.gcmrc.gov/discharge_qw_sediment/. At this website, the user can construct plots in time-series or duration-curve format. In addition, the user can construct interactive plots of the mass-balance sand budgets for the six CRe segments, with user-defined uncertainty.

Hypotheses and Science Questions

There are two key hypotheses that guide the long-term monitoring and research conducted under all elements of Project A. These hypotheses directly address the LTEMP sediment goal and the nine other LTEMP goals listed in the previous section.

- Glen Canyon Dam can be operated such that the sand resources in the CRe are sustainable.
- Glen Canyon Dam can be operated such that the other CRe resources affected by dam operations can be sustainably managed. In this usage, "dam operations" refers to the amount and quality of the water released from the dam, where "amount" refers to stage and streamflow, and "quality" refers to temperature, salinity, turbidity, and dissolved oxygen.

These hypotheses are paraphrased from the LTEMP EIS and from earlier goals, information needs, and strategic science questions formulated by the GCDAMP. The first of these two guiding hypotheses is tested using the continuous mass-balance sand budgets (Project Element A.3) constructed using 15-minute streamflow data (Project Element A.1) and suspended-sand data (Project Element A.3). Annual updates on the status of the segment-by-segment testing of the first hypothesis are provided at each Annual Reporting Meeting. Although the second hypothesis guides data collection in Project A, this hypothesis is tested by the other GCDAMP-funded projects, with annual updates provided by these other projects.

Project Element A.1. Stream Gaging and Hydrologic Analyses (Ongoing Study)

This element partially funds the collection, serving, and interpretation of continuous 15-minute measurements of stage and discharge on the main-stem Colorado River at USGS streamflow gaging stations located at river miles (RM) 0, 30, 61, 87, 166, and 225, and at gaging stations on the major tributaries and in a representative subset of the smaller tributaries (Topping and others, 2021). Eighty percent of the budget for Project Element A.1 funds salary for the field and office time required to operate gaging stations and funds the office time for serving data and working on peer-reviewed interpretive publications.

Although the streamflow data collected under this element support GCMRC projects in every discipline (e.g., Korman and Campana, 2009; Kennedy and others, 2016; Neher and others, 2017; Deemer and others, 2022; Sankey and others, 2023; Butterfield and Palmquist, 2024), the data collected under this element are central to the LTEMP sediment goal and are used to design and evaluate HFEs. Of the gaging stations funded by Project A, only the Little Colorado River above the mouth near Desert View, AZ station (Yackulic and Hull, 2019; Unema and others, 2021) is not used to support the LTEMP sediment goal; this singular and most-expensive gaging station is used almost solely to support the LTEMP humpback chub goal and much of the work of Project G in the lower segment of the LCR critical for humpback chub habitat (Gorman and Stone, 1999; Unema and others, 2021). The suspended-sand flux is the product of the instantaneous water discharge and velocity-weighted suspended-sand concentration (Guy, 1970). This flux is augmented to account for sand bedload (Rubin and others, 2001; Topping and others, 2010, 2021; Ashley and others, 2020; Le Coz and others, 2022) and then integrated over time to calculate the sand load over any given time interval (Topping and others, 2021). These loads on the Colorado River and tributaries are used to construct the continuous mass-balance sand budgets described in Project Element A.3.

It is impossible to construct these budgets without accurate streamflow gaging stations on the Colorado River and its key sand-supplying tributaries. In decreasing order of their sand-supply magnitude, based on measurements since 2010, these tributaries are 1) the Paria River, 2) the Little Colorado River, 3) the combined smaller tributaries in Lower Glen Canyon and Upper Marble Canyon (RM -15 to 30), 4) the combined smaller tributaries in East Central Grand Canyon (RM 87 to 166), 5) the combined smaller tributaries in West Central Grand Canyon (RM 166 to 225), 6) Havasu Creek, 7) Kanab Creek, 8) the combined smaller tributaries in Eastern Grand Canyon (RM 61 to 87), and 9) the combined smaller tributaries in Lower Marble Canyon (RM 30 to 61) (Topping and others, 2021). We have therefore designed the Project A gaging-station network to focus resources as appropriate relative to each tributary's importance as a sand source.

To reduce the number of staff at GCMRC required to operate Project A and thereby reduce costs, we have taken a "burden-sharing" approach to operating the streamflow gaging stations in the CRe owing to an insufficient staffing level at the GCMRC.

For example, three of the Colorado River stations are operated by GCMRC staff and three are operated by AZ Water Science Center (WSC) staff. The FY 2025 gross costs (including overhead) to the GCDAMP for the surface-water record at each gaging station, the USGS science center operating each station, and the main LTEMP sediment-goal purpose of the streamflow data at each station are listed in Table 1.

All gaging stations funded by this element are used to directly address LTEMP goals. Although the streamflow gaging stations on the tributaries do not directly monitor the downstream effects of Glen Canyon Dam operations, these gaging stations are required to monitor the tributary sand supply and to monitor the hydrologic and water-quality conditions in tributaries most important for the aquatic foodbase (Project F.3) and fishes (Yackulic and Hull, 2019). Monitoring the tributary sand supply is required to separate the effects of tributary sand-supply events from the effects of dam operations on the sand resources in the CRE. In addition to the collection and serving of stage and discharge data at gaging stations, a large part of the budget for Project Element A.1 supports hydrologic/geomorphic interpretive work in support of the LTEMP sediment, humpback chub, and natural processes goals as described below.

Table 1.

| Gaging station | USGS lead | Gross cost to GCDAMP | Main LTEMP sediment-goal purpose |
|---|------------------|---------------------------------|--|
| Colorado River at Lees Ferry (RM 0)* | AZ WSC | \$0 | Monitoring sand export from lower Glen Canyon during HFEs and other special LTEMP dam operations |
| Colorado River near river mile 30 (RM 30) | GCMRC | \$15,000 | 15-minute sand loads used in sand budgets |
| Colorado River above LCR (RM 61) | GCMRC | \$15,000 | 15-minute sand loads used in sand budgets |
| Colorado River near Grand Canyon (RM 87)* | AZ WSC | \$0 | 15-minute sand loads used in sand budgets |
| Colorado R. above National Canyon (RM 166) | GCMRC | \$9,000 | 15-minute sand loads used in sand budgets |
| Colorado R. above Diamond Creek (RM 225) | AZ WSC | \$27,900 | 15-minute sand loads used in sand budgets |
| Paria River near Kanab | UT WSC | \$20,700 | Flood warning for sampling fieldwork on Paria River at Lees Ferry |
| Paria River at Lees Ferry * | AZ WSC | \$0 | 15-minute sand loads used in sand budgets |
| Moenkopi Wash at Moenkopi** | AZ WSC | \$11,500 | Flood warning for sampling fieldwork on LCR near Cameron |
| Little Colorado River near Cameron* | AZ WSC | \$0 | 15-minute sand loads used in sand budgets |
| Little Colorado River above the mouth | AZ WSC | \$30,600 | NONE; used mainly by humpback chub studies |
| Kanab Creek above the mouth** | AZ WSC | \$14,500 | Event-based sand loads used to verify sand budgets |
| Havasu Creek above the mouth*** | AZ WSC | \$14,500 | Event-based sand loads used to verify sand budgets |
| Eight low-cost research gages on small tributaries in Lower Glen Canyon, Upper Marble Canyon, and East Central Grand Canyon | GCMRC | \$8,200 (total for all 8 gages) | Event-based sand loads used to verify sand budgets, and design and evaluate HFEs |
| Two low-cost research gages on LCR (Grand Falls) and Moenkopi Wash (Cameron) | GCMRC | \$8,200 (total for both) | Dean work monitoring declining LCR sand supply |

*The surface-water records at these gaging stations are entirely funded by non-GCDAMP sources.

**The surface-water records at these gaging stations are partially funded by the USGS toxics program for uranium monitoring.

***The surface-water record at this gaging station is partially funded by the Arizona Department of Water Resources.

Methods

Data for this element are collected and processed using standard USGS stream-gaging methods (Rantz and others, 1982a, b; Mueller and others, 2013) augmented by other peer-reviewed methods (Griffiths and others, 2012, 2014; Griffiths and Topping, 2017).

Anticipated Use of Data

The stage and discharge data collected by this element address both hypotheses listed in the “Proposed Work” section and are used to inform scientific evaluations of management’s success meeting the LTEMP archaeological and cultural resources, natural processes, humpback chub, hydropower and energy, other native fish, recreational experience, sediment, rainbow trout fishery, nonnative invasive species, and riparian vegetation goals. These evaluations are conducted by projects A, B, C, D, E, F, G, H, I, J, L, and N. The discharge data collected under this element are combined with the sediment-concentration data collected under element A.3 to compute the loads used to design HFES (Grams and others, 2015) and in the mass-balance sand budgets used to evaluate management’s progress meeting the LTEMP sediment goal (e.g., Topping and others, 2021; Griffiths and others, 2024).

Outcomes and Products

Outcomes

- Data used to inform LTEMP sediment, aquatic food base, archaeological and cultural resources, humpback chub, hydropower and energy, invasive fish species, natural processes, rainbow trout fishery, recreational experience, and riparian vegetation goals. Data from this element required to design and evaluate HFES and to evaluate the effects of all LTEMP dam releases on sediment resources in the CRe (including those flows releases to suppress smallmouth bass). In addition, gaging data on the LCR, Bright Angel Creek, and Havasu Creek used to inform National Park Service humpback chub translocation efforts. Data on all tributaries (including the low-cost gages on the smaller tributaries) used to support Department of the Interior uranium monitoring efforts (thus the cost-sharing support of these gages from the USGS toxics program).

Products

- 1) Gage height and discharge data served on the Discharge, Sediment, and Water Quality Monitoring page of the GCMRC website (https://www.gcmrc.gov/discharge_qw_sediment/). These 15-minute data are updated daily at the RM 0, RM 87, and RM 225 gaging stations (Table 1), monthly at the RM 30 and 61 gaging stations, and every 6 months at the RM 166 gaging station. These

data are updated at various intervals at the tributary gaging stations, depending on whether satellite telemetry is present and the frequency of field visits.

- 2) Completion of journal article currently under preparation evaluating hydrologic changes, especially the decline in winter floods in the Paria River since initiation of gaging in 1923 (Topping lead). These analyses are required to understand how seasonal changes in flooding have affected sand delivery to the CRe and the implications for sediment-triggered spring HFEs.
- 3) Discharge measurement download tool on the Discharge, Sediment, and Water Quality Monitoring page of the GCMRC website (https://www.gcmrc.gov/discharge_qw_sediment/).
- 4) Data from this element will be used in at least one presentation given by the scientists funded by Project Element A.1 at professional science meetings each year.

Project Element A.2. Continuous Water-Quality Parameters (Ongoing Study)

This element funds the collection, serving, and interpretation of continuous 15-minute measurements of water temperature, specific conductance (a measure of salinity), turbidity, and dissolved oxygen at the outlet of Glen Canyon Dam and at the above-mentioned six main-stem Colorado River gaging stations. This element also funds episodic measurements of specific conductance associated with suspended-sediment samples collected in tributaries (these measurements are intrinsic to the laboratory methods for processing the suspended-sediment samples and therefore cost nothing). Seventy-three percent of the budget for Project Element A.2 funds salary for the field and office time required for making the water-quality measurements and funds the office time for serving the data.

Under this element 15-minute measurements of water temperature, specific conductance, turbidity, and dissolved oxygen are made using YSI multi-parameter sondes in the Colorado River located at the outlet of Glen Canyon Dam and at the gaging stations located at river miles 0, 30, 61, 87, 166, and 225. See Voichick and Wright (2007), Voichick (2008), and Voichick and Topping (2010, 2014) for detailed descriptions of these sondes and measurements. During high-silt-and-clay-concentration events, the turbidity data are corrected for false low readings (Voichick and others, 2018) and are extended to values higher than the detection limit of the optical turbidity probe using the acoustical suspended-silt-and-clay data collected under Project Element A.3 (Voichick and Topping, 2014). In addition to the water-quality data collected at the gaging stations on the Colorado River, 15-minute measurements of water temperature are made at three additional stations on the Colorado River and at stations near the mouths of the Paria and Little Colorado rivers, and Bright Angel, Kanab, and Havasu creeks.

Data collected under Project Element A.2 are routinely used in publications led by investigators in other GCDAMP-funded projects and investigators of projects external to GCDAMP (e.g., Yard and others, 2011; Hall and others, 2015; Korman and others, 2015; Ward and others, 2015,

2016; Yackulic and others, 2018; Yackulic and Hull, 2019; Dibble and others, 2021; Deemer and others, 2022; Dzul and others, 2023). The seasonal median turbidity in each river segment is the proposed metric used to evaluate LTEMP management for the silt-and-clay component of fine sediment. In addition, the turbidity data collected under this element are to be used in combination with the data and analysis from Project E.1 to estimate phosphorous in the CRe.

Methods

Data for this element are collected and processed using standard USGS water-quality methods (Wagner and others, 2006) augmented by other peer-reviewed methods (Voichick and Wright, 2007; Voichick and Topping, 2010, 2014; Voichick and others, 2018; Griffiths and others, 2012).

Anticipated Use of Data

The water-quality data collected by this element addresses the second of the two hypotheses listed in the “Proposed Work” section and are used to inform scientific evaluations of management’s success meeting the LTEMP natural processes, humpback chub, other native fish, recreational experience, rainbow trout fishery, and nonnative invasive species goals. These evaluations are conducted by projects E, F, G, H, I, J, and N.

Outcomes and Products

Outcomes

- Data used to inform LTEMP aquatic food base, humpback chub, invasive fish species, natural processes, and rainbow trout fishery goals. In addition, water-temperature data on the LCR, Bright Angel Creek, and Havasu Creek used to inform National Park Service humpback chub translocation efforts.

Products

- 1) Water-temperature, specific-conductance, turbidity, and dissolved-oxygen data served on the Discharge, Sediment, and Water Quality Monitoring page of the GCMRC website (https://www.gcmrc.gov/discharge_qw_sediment/). These data are updated daily at the RM 0 gaging station (Table 1), monthly at the dam and the RM 30, 61, and 87 gaging stations, every three months at the RM 225 gaging station, and every 6 months at the RM 166 gaging station.
- 2) The chief employee funded by this element will participate as a junior author on articles/reports published by the other GCDAMP-funded projects that use the data collected under Project Element A.2.

Project Element A.3. Sediment Transport and Budgeting (Ongoing Study)

This element funds the collection, serving, and interpretation of continuous 15-minute measurements and episodic measurements of suspended sediment and bed sediment at the above-mentioned gaging stations on the Colorado River and its tributaries. In addition, this project element funds interpretive work in regard to the sand supply from the Paria and Little Colorado rivers, and interpretive work in regard to the effect of dam operations on the sediment resources in the Colorado River between Glen Canyon Dam and Lake Mead. Examples of the analyses used to inform these interpretations can be found in Topping and others (2010, 2019, 2021) and in the other 103 peer-reviewed articles, reports, and books published by Project A since its inception; publications listed and available at <https://www.usgs.gov/centers/southwest-biological-science-center/science/river-sediment-dynamics>. Eighty-one percent of the budget for Project Element A.3 funds salary for the field, laboratory, and office time required to collect and process sediment data, and also funds the office time for serving data and working on peer-reviewed interpretive publications. In addition, data collected under Project Element A.3 is coordinated with the phosphorous monitoring of Project E.1 in the Paria and Little Colorado rivers. The continuous silt-and-clay data collected under Project Element A.3 and continuous turbidity data collected under Project Element A.2 are to be used in combination with the data and analysis from Project E.1 and the turbidity modeling in Project B.4 to estimate phosphorous in the CRE. The continuous suspended-sediment measurements at the six main-stem Colorado River gaging stations, and the episodic suspended-sediment measurements in the tributaries are all used in the construction and evaluation of mass-balance sand budgets, and are used to trigger, design, and evaluate HFEs (Topping and others, 2010, 2019; Grams and others, 2015).

All measurements funded under Project Element A.3 are made using standard USGS methods (Guy, 1970; Edwards and Glysson, 1999; Porterfield, 1972) augmented by additional peer-reviewed methods (Topping and others, 2010, 2011, 2021; Griffiths and others, 2012, 2014; Sabol and Topping, 2013; Topping and Wright, 2016; Sabol and others, 2022). Under this element, continuous two-frequency acoustical suspended-sediment measurements are made in the Colorado River at the gaging stations located at RM 30, 61, 87, 166, and 225 using the method of Topping and Wright (2016). In addition to informing river management in the GCDAMP, our acoustical method pioneered in the Colorado River is now being used to inform river management across the United States (Dean and others, 2016, 2020, 2022; Topping and others, 2018, in press; AuBuchon and others, 2023) and in Europe in the Isere River at Grenoble, France (Marggraf, 2024). The continuous measurements are used to calculate the sand loads used in sand budgeting and also used to calculate continuous measures of bed-sand grain size using the method of Rubin and Topping (2001, 2008). Because these grain-size values indicate periods of sand enrichment and depletion, they are critical in determining how observed changes in the amount of sand in a segment relate to dam operations. This information allows knowing whether sand erosion or deposition is driven more by dam operations or simply by the longitudinal positions of tributary-generated sand waves in the Colorado River.

In addition to the measurements on the main-stem Colorado River, episodic suspended-sediment measurements are made at the tributary gaging stations funded under Project Element A.1. These measurements are used in conjunction with models (after Topping, 1997) to determine the near-realtime sediment inputs from the Paria and Little Colorado rivers used in sand budgeting (Topping and others, 2021). On the other tributaries, these measurements are used to document the sand, silt, and clay supply from the other major and lesser tributaries and to refine the long-term estimates of the importance of these other tributaries for supplying sediment to the CRE (Griffiths and others, 2014; Griffiths and Topping, 2017; Topping and others, 2021). Most of the sediment work on tributaries utilizes automatic samplers and has a large payoff in information for relatively low cost.

In addition to the collection of the sediment-transport data, this element provides funding for the web-based construction and analysis of continuous mass-balance sand budgets for the CRE using the suspended-sediment measurements on the Colorado River and its tributaries (Sibley and others, 2015; Topping and others, 2021); funds within Project K provide additional funding for the computer-science support for this task. In addition to being used to evaluate the effects of LTEMP dam releases on the CRE, these mass-balance sand budgets are used in collaboration with the Bureau of Reclamation to trigger, plan, and evaluate HFEs (Grams and others, 2015; U.S. Department of the Interior, 2016a). HFEs are triggered and designed based on the Paria-supplied sand that accumulates in Marble Canyon during fall and spring implementation windows (U.S. Department of the Interior, 2016a, Grams and others, 2015). This process involves using many suspended-sediment samples collected in the Paria River (quickly processed through the GCMRC sediment laboratory) in combination with discharge data (funded under Project Element A.1) and initial model estimates (after Topping, 1997) to determine the near-realtime continuous sand supply from the Paria River (Topping and others, 2021). The Bureau of Reclamation and Project B.4 then use this information, along with information on planned dam releases, as input to the sand-routing model of Wright and others (2010) to design HFEs.

As more suspended-sediment measurements get processed through the laboratory (work funded by this element), the uncertainty is reduced in the calculated Paria River sand supply, and additional model runs are made by Project B.4. As time progresses, Project B.4's model-predictions of sand retention in Marble Canyon are compared against the actual measured sand retention in the continuous mass-balance sand budgets funded under this project element. Because the predictions of the sand-routing model of Wright and others (2010) may be off by a factor of 2, this comparison allows reality-based redesign of each planned HFE hydrograph by the Bureau of Reclamation. Finally, after the completion of each HFE, these sand budgets allow quick post-facto evaluation of the longitudinal effects of each HFE on the sand resources in the CRE in support of the LTEMP sediment goal. The annual sand mass balance in each river segment is one of the proposed metrics to evaluate LTEMP management for the sand component of fine sediment.

Methods

Data for this element are collected and processed using standard USGS sediment methods (Guy, 1970; Porterfield, 1972; Edwards and Glysson, 1999) augmented by other peer-reviewed methods (Topping and others, 2010, 2011, 2021; Griffiths and others, 2012, 2014; Sabol and Topping, 2013; Topping and Wright, 2016; Griffiths and Topping, 2017; Sabol and others, 2022).

Anticipated Use of Data

The sediment data collected by this element address both hypotheses listed in the “Proposed Work” section (because sediment affects water quality through turbidity) and are used to inform scientific evaluations of management’s success meeting the LTEMP archaeological and cultural resources, natural processes, humpback chub, other native fish, recreational experience, sediment, nonnative invasive species, and riparian vegetation goals. These evaluations are conducted by projects A, B, C, D, E, F, G, I, J, L, and N. The sediment-concentration data collected under this element are combined with the discharge data collected under element A.1 to compute the loads used to design HFES (Grams and others, 2015) and in the mass-balance sand budgets used to evaluate management’s progress meeting the LTEMP sediment goal (e.g., Topping and others, 2021; Griffiths and others, 2024).

Outcomes and Products

Outcomes

- Data used to inform LTEMP sediment, archaeological and cultural resources, natural processes, and recreational experience goals. Data from this element required to design and evaluate HFES and to evaluate the effects of all LTEMP dam releases on sediment resources in the CRe.

Products

- 1) Sediment data and sand budgets served on the Discharge, Sediment, and Water Quality Monitoring page of the GCMRC website (https://www.gcmrc.gov/discharge_qw_sediment/). The 15-minute data are updated monthly at the RM 30, 61, and 87 gaging stations (Table 1), every three months at the RM 225 gaging station, and every 6 months at the RM 166 gaging station. These data are updated at various intervals at the tributary gaging stations, depending on the frequency of field visits. The sediment processed through the laboratory are uploaded every few weeks.
- 2) Completion of a USGS professional paper describing the laser-diffraction measurements funded by the GCDAMP during FY 2001–10 and completion of a companion journal article focused on silt-and-clay grain size and flocculation

(Topping leads). These papers have been under preparation for 10 years and will be completed during this funding cycle. Their content is central to the biological role silt and clay plays in the CRE.

- 3) Completion of a USGS report describing how all LTEMP dam releases (including possible balancing-tier or equalization releases) conducted during FY 2021–24 have affected the sand resources in the CRE on the segment scale (Griffiths lead). This report will update the results in Griffiths and others (2024), which described how all LTEMP dam releases affected sand in the CRE during FY 2018–20.
- 4) Progress or completion of a journal article or USGS report describing the conditions that lead to hyper concentrated flows (Beverage and Culbertson, 1964; Topping, 1997) in the Paria River during certain floods (Griffiths lead). This work is required to better inform the Bureau of Reclamation in the HFE-planning and design process.
- 5) Progress or completion of a USGS Professional Paper evaluating geomorphic and sediment-transport changes in the Paria River basin since the 1800s, with predictions for the most likely future Paria River sand supply to the CRE (Topping lead). This report may be published during the next work plan; work on this product began in the 1990s (Topping, 1997). All needed historical aerial photography was obtained and orthorectified under the last work plan.
- 5) Data from this element will be used in at least one presentation given by the scientists funded by Project Element A.3 at professional science meetings each year.

Project Element A.4. HFE Analyses (Ongoing Study, Experimental Fund)

This element funds the collection and processing of streamflow and sediment data before, during, and after HFEs in support of the LTEMP sediment goal. Under this element, crews will be deployed to make discharge measurements and collect suspended-sediment samples at the Colorado River at Lees Ferry, Colorado River near Grand Canyon gaging stations and also at either the Colorado River above Little Colorado River or Colorado River above Diamond Creek gaging stations. This work is required to evaluate the effects on sediment of individual HFEs (e.g., Topping and others, 2010, 2019).

Methods

Data for this element are collected and processed using the standard USGS methods and other peer-reviewed methods described above in the “Methods” sections for Elements A.1 and A.3.

Anticipated Use of Data

The data collected by this element addresses the first of the two hypotheses listed in the “Proposed Work” section and are used to inform scientific evaluation of management’s success

meeting the LTEMP sediment goal through use of repeated sand-enriched HFEs to rebuild sandbars. The data collected by this element are analyzed as in Topping and others (2010, 2019) to evaluate the level of sand enrichment in each CRE river segment during individual HFEs and how this level of sand enrichment translates to sandbar deposition.

Outcomes and Products

Outcomes

- Data used to inform LTEMP sediment, archaeological and cultural resources, natural processes, and recreational experience goals. Data from this element required to evaluate HFEs.

Products

- 1) Sediment data and sand budgets served on the Discharge, Sediment, and Water Quality Monitoring page of the GCMRC website (https://www.gcmrc.gov/discharge_qw_sediment/).
- 2) Interpretation of data collected under this element will be included in either the Project Element A.3 USGS report led by Griffiths or a separate report led by Topping.

Project Element A.5. Real-time Water Temperature in Marble Canyon (New Study; Experimental Fund)

This element funds the collection, processing, and hourly serving of 15-minute water temperature measurements at the downstream end of Marble Canyon at the Colorado River above Little Colorado River near Desert View, AZ, gaging station (RM 61). This would provide direct support of the Cool Mix experiment to disadvantage warmwater nonnative fishes as described in the 2024 LTEMP SEIS (U.S. Department of the Interior, 2024).

Methods

Data for this element are collected and processed using the standard USGS methods described above in the “Methods” sections for Elements A.2.

Anticipated Use of Data

The data collected by this element will be used by Reclamation to initiate, modify, and end Cool Mix releases or other temperature-related actions.

Outcomes and Products

Outcomes

- Data used to inform LTEMP Nonnative Invasive Species and Humpback Chub goals.

Products

- Fifteen-minute water temperature data updated every hour on the USGS NWIS website (<https://waterdata.usgs.gov/az/nwis/rt>) and updated daily on the Discharge, Sediment, and Water Quality Monitoring page of the GCMRC website (https://www.gcmrc.gov/discharge_qw_sediment/).

The annual cost of this element is \$20,000 gross that will be suballocated to the USGS Arizona Water Science Center.

Budgets

Project Element Budget

| Fiscal Year 2025 | | | | | | | | | |
|--|------------------|----------------------|-----------------------|-----------------------|---------------------------|--------------------------|-----------------------------|--------------------|------------------------------|
| Project A Streamflow, Water Quality, and Sediment Transport and Budgeting in the Colorado River Ecosystem | Salaries | Travel & Training | Operating Expenses | Logistics Expenses | Cooperative Agreements | To other USGS Centers | AMP Special Burden Rate* | Total | USGS Contributing Funds** |
| | | | | | | | 21.86% | | Est Full Rate = 55% |
| A.1. Stream gaging and hydrologic analyses | \$181,311 | \$6,000 | \$6,000 | \$27,000 | \$0 | \$174,800 | \$48,166 | \$443,277 | |
| A.2. Continuous water-quality parameters | \$114,492 | \$1,000 | \$12,000 | \$26,000 | \$0 | \$29,260 | \$33,558 | \$216,310 | |
| A.3. Sediment transport and budgeting | \$333,120 | \$6,000 | \$38,000 | \$26,200 | \$0 | \$154,360 | \$88,178 | \$645,857 | |
| Total Project A | \$628,922 | \$13,000 | \$56,000 | \$79,200 | \$0 | \$358,420 | \$169,902 | \$1,305,444 | \$81,155 |

| Fiscal Year 2026 | | | | | | | | | |
|--|------------------|----------------------|-----------------------|-----------------------|---------------------------|--------------------------|-----------------------------|--------------------|------------------------------|
| Project A Streamflow, Water Quality, and Sediment Transport and Budgeting in the Colorado River Ecosystem | Salaries | Travel & Training | Operating Expenses | Logistics Expenses | Cooperative Agreements | To other USGS Centers | AMP Special Burden Rate* | Total | USGS Contributing Funds** |
| | | | | | | | 22.60% | | Est Full Rate = 55.5% |
| A.1. Stream gaging and hydrologic analyses | \$160,807 | \$6,000 | \$6,000 | \$28,000 | \$0 | \$180,120 | \$45,382 | \$426,309 | |
| A.2. Continuous water-quality parameters | \$120,217 | \$1,000 | \$12,000 | \$26,800 | \$0 | \$30,080 | \$36,164 | \$226,260 | |
| A.3. Sediment transport and budgeting | \$339,919 | \$6,000 | \$38,000 | \$27,000 | \$0 | \$158,960 | \$92,868 | \$662,747 | |
| Total Project A | \$620,943 | \$13,000 | \$56,000 | \$81,800 | \$0 | \$369,160 | \$174,414 | \$1,315,317 | \$81,780 |

| Fiscal Year 2027 | | | | | | | | | |
|--|------------------|----------------------|-----------------------|-----------------------|---------------------------|--------------------------|-----------------------------|--------------------|------------------------------|
| Project A Streamflow, Water Quality, and Sediment Transport and Budgeting in the Colorado River Ecosystem | Salaries | Travel & Training | Operating Expenses | Logistics Expenses | Cooperative Agreements | To other USGS Centers | AMP Special Burden Rate* | Total | USGS Contributing Funds** |
| | | | | | | | 23.40% | | Est Full Rate = 56% |
| A.1. Stream gaging and hydrologic analyses | \$168,847 | \$6,000 | \$6,000 | \$29,000 | \$0 | \$185,800 | \$49,104 | \$444,752 | |
| A.2. Continuous water-quality parameters | \$130,384 | \$1,000 | \$12,000 | \$27,400 | \$0 | \$30,900 | \$39,963 | \$241,647 | |
| A.3. Sediment transport and budgeting | \$356,915 | \$6,000 | \$38,000 | \$28,000 | \$0 | \$165,900 | \$100,366 | \$695,181 | |
| Total Project A | \$656,146 | \$13,000 | \$56,000 | \$84,400 | \$0 | \$382,600 | \$189,434 | \$1,381,580 | \$87,139 |

*GCDAMP Special Rate: Rate includes current DOI preferred rate (currently 15%; subject to change) and facilities rate (will vary annually). No USGS bureau overhead is charged (unique to the GCDAMP agreement).

**USGS Contributing Funds: The amount of funds required to cover the subsidy created by the reduced burden rate (i.e., GCDAMP Special Rate). As in previous years, SBSC/GCMRC will request these funds of the USGS cost-share program. These funds are not guaranteed.

Experimental Fund

| Fiscal Year 2025 | | | | | | | | |
|--|------------------|----------------------|-----------------------|-----------------------|---------------------------|--------------------------|-----------------------------|------------------|
| Project A Experimental Fund Projects | Salaries | Travel & Training | Operating Expenses | Logistics Expenses | Cooperative Agreements | To other USGS Centers | AMP Special Burden Rate* | Total |
| | | | | | | | 21.86% | |
| A.4. HFE Experimental Fund | \$117,666 | \$2,000 | \$20,000 | \$45,200 | \$0 | \$0 | \$40,417 | \$225,283 |
| A.5. Real-time water temperature in Marble Canyon | \$0 | \$0 | \$0 | \$0 | \$0 | \$20,000 | \$0 | \$20,000 |
| Total Project A | \$117,666 | \$2,000 | \$20,000 | \$45,200 | \$0 | \$20,000 | \$40,417 | \$245,283 |

| Fiscal Year 2026 | | | | | | | | |
|--|------------------|----------------------|-----------------------|-----------------------|---------------------------|--------------------------|-----------------------------|------------------|
| Project A Experimental Fund Projects | Salaries | Travel & Training | Operating Expenses | Logistics Expenses | Cooperative Agreements | To other USGS Centers | AMP Special Burden Rate* | Total |
| | | | | | | | 22.60% | |
| A.4. HFE Experimental Fund | \$123,549 | \$2,050 | \$20,000 | \$46,700 | \$0 | \$0 | \$43,460 | \$235,758 |
| A.5. Real-time water temperature in Marble Canyon | \$0 | \$0 | \$0 | \$0 | \$0 | \$20,000 | \$0 | \$20,000 |
| Total Project A | \$123,549 | \$2,050 | \$20,000 | \$46,700 | \$0 | \$20,000 | \$43,460 | \$255,758 |

| Fiscal Year 2027 | | | | | | | | |
|--|------------------|----------------------|-----------------------|-----------------------|---------------------------|--------------------------|-----------------------------|------------------|
| Project A Experimental Fund Projects | Salaries | Travel & Training | Operating Expenses | Logistics Expenses | Cooperative Agreements | To other USGS Centers | AMP Special Burden Rate* | Total |
| | | | | | | | 23.40% | |
| A.4. HFE Experimental Fund | \$129,726 | \$2,100 | \$20,000 | \$48,100 | \$0 | \$0 | \$46,783 | \$246,709 |
| A.5. Real-time water temperature in Marble Canyon | \$0 | \$0 | \$0 | \$0 | \$0 | \$20,000 | \$0 | \$20,000 |
| Total Project A | \$129,726 | \$2,100 | \$20,000 | \$48,100 | \$0 | \$20,000 | \$46,783 | \$266,709 |

References Cited

- Andrews, E.D., 1991, Sediment transport in the Colorado River Basin, in Committee to Review the Glen Canyon Environmental Studies Water Science and Technology Board, ed., Colorado River ecology and dam management, Washington, D.C., Proceedings: National Academy Press, p. 54-74, <https://nap.nationalacademies.org/catalog/1832/colorado-river-ecology-and-dam-management-proceedings-of-a-symposium>.
- Ashley, T.C., McElroy, B., Buscombe, D., Grams, P.E., and Kaplinski, M., 2020, Estimating bedload from suspended load and water discharge in sand bed rivers: Water Resources Research, v. 56, e2019WR025883, <https://doi.org/10.1029/2019WR025883>.
- AuBuchon, J., Abraham, D., Posner, A., Brown, J., Jackson, T., and Griffiths, R.E., 2023, Acoustic measurements on a shallow, sand-bed river: A case study from the Rio Grande: Proceedings of SEDHYD 2023, the Federal Interagency Sedimentation and Hydrologic Modeling Conference, May 8–12, 2023, St. Louis, Missouri, <https://www.sedhyd.org/2023Program/1/82.pdf>.

- Beverage, J.P., and Culbertson, J.K., 1964, Hyperconcentrations of suspended sediment: *Journal of the Hydraulics Division, Proceedings of the American Society of Civil Engineers*, v. 90, p. 117–128, <https://doi.org/10.1061/JYCEAJ.0001128>.
- Butterfield, B.J., and Palmquist, E.C. 2024, Divergent physiological responses of hydric and mesic riparian plant species to a Colorado River experimental flow: *Plant Ecology*, v. 225, p. 125–133, <https://doi.org/10.1007/s11258-023-01382-6>.
- Dean, D.J., and Topping, D.J., 2019, Geomorphic change and biogeomorphic feedbacks in a dryland river—The Little Colorado River, Arizona, USA: *Geological Society of America Bulletin*, v. 131, no. 11-12, p. 1920-1942, <https://doi.org/10.1130/B35047.1>.
- Dean, D.J., and Topping, D.J., 2024, The effects of vegetative feedbacks on flood shape, sediment transport, and geomorphic change in a dryland river—Moenkopi Wash, AZ: *Geomorphology*, v. 447, 23 p., <https://doi.org/10.1016/j.geomorph.2023.109017>.
- Dean, D.J., Topping, D.J., Buscombe, D.D., Groten, J.T., Ziegeweid, J., Fitzpatrick, F.A., Lund, J.W., and Coenen, E.N., 2022, The use of continuous sediment-transport measurements to improve sand-load estimates in a large sand-bedded river—The lower Chippewa River, Wisconsin: *Earth Surface Processes and Landforms*, v. 47, p. 2006–2023, <https://doi.org/10.1002/esp.5360>.
- Dean, D.J., Topping, D.J., Grams, P.E., Walker, A.E., and Schmidt, J.C., 2020, Does channel narrowing by floodplain growth necessarily indicate sediment surplus? Lessons from sediment transport analyses in the Green and Colorado Rivers, Canyonlands, Utah: *Journal of Geophysical Research: Earth Surface*, v. 125, article e2019JF005414, <https://doi.org/10.1029/2019JF005414>.
- Dean, D.J., Topping, D.J., Schmidt, J.C., Griffiths, R.E., and Sabol, T.A., 2016, Sediment supply versus local hydraulic controls on sediment transport and storage in a river with large sediment loads: *Journal of Geophysical Research: Earth Surface*, v. 121, n. 1, p. 82–110, <https://doi.org/10.1002/2015JF003436>.
- Deemer, B.R., Yackulic, C.B., Hall, R.O., Dodrill, M.J., Kennedy, T.A., Muehlbauer, J., Topping, D.J., Voichick, N., and Yard, M., 2022, Experimental reductions in subdaily flow fluctuations increased gross primary productivity for 425 river kilometers downstream: *Proceedings of the National Academy of Sciences NEXUS*, v. 1, p. 1–12, <https://doi.org/10.1093/pnasnexus/pgac094>.
- Dibble, K.L., Yackulic, C.B., Kennedy, T.A., Bestgen, K.R., and Schmidt, J.C., 2021, Water storage decisions will determine the distribution and persistence of imperiled river fishes: *Ecological Applications*, v. 31, e02279, <https://doi.org/10.1002/eap.2279>.
- Dzul, M., Yackulic, C.B., Giardina, M., Van Haverbeke, D.R., and Yard, M., 2023, Vital rates of a burgeoning population of humpback chub in western Grand Canyon: *Transactions of the American Fisheries Society*, v. 152, p. 443–459, <https://doi.org/10.1002/tafs.10415>.
- Edwards, T.K., and Glysson, G.D., 1999, Field methods for measurement of fluvial sediment: *Techniques of Water-Resources Investigations of the U.S. Geological Survey 03-C2*, 89 p., <https://doi.org/10.3133/twri03C2>.

- Exner, F.M., 1920, Zur physik der dünen: Akademie der Wissenschaften in Wien, Mathematisch-Naturwissenschaftliche Klasse Sitzungsberichte, abt. Ila, v. 129, p. 929-952.
- Exner, F.M., 1925, Über die wechselwirkung zwischen wasser und geschiebe in flüssen: Akademie der Wissenschaften in Wien, Mathematisch-Naturwissenschaftliche Klasse Sitzungsberichte, abt. Ila, v. 134, p. 165-203.
- Gloss, S.P., and Coggins, L.G., 2005, Fishes of Grand Canyon, in Gloss, S.P., Lovich, J.E., and Melis, T.S., eds., The state of the Colorado River ecosystem in Grand Canyon: U.S. Geological Survey Circular 1282, p. 33-68, <https://pubs.usgs.gov/circ/1282/c1282.pdf>.
- Gloss, S.P., Lovich, J.E., and Melis, T.S., eds., 2005, The state of the Colorado River ecosystem in Grand Canyon: U.S. Geological Survey Circular 1282, 220 p., <https://pubs.usgs.gov/circ/1282/c1282.pdf>.
- Gorman, O.T., and Stone, D.M., 1999, Ecology of spawning humpback chub, *Gila cypha*, in the Little Colorado River, near Grand Canyon, Arizona: Environmental Biology of Fishes, v. 55, p. 115-133, <https://doi.org/10.1023/A:1007450826743>.
- Grams, P.E., Buscombe, D., Topping, D.J., Kaplinski, M.A., and Hazel, J.E., Jr., 2019, How many measurements are required to construct an accurate sand budget in a large river? Insights from analyses of signal and noise: Earth Surface Processes and Landforms, v. 44, no. 1, p. 160-178, <https://doi.org/10.1002/esp.4489>.
- Grams, P.E., Schmidt, J.C., Wright, S.A., Topping, D.J., Melis, T.S., and Rubin, D.M., 2015, Building sandbars in the Grand Canyon: EOS, Transactions of the American Geophysical Union, v. 96, no. 11, p. 12-16, <https://doi.org/10.1029/2015EO030349>.
- Grams, P.E., Topping, D.J., Schmidt, J.C., Hazel, J.E., Jr., and Kaplinski, M., 2013, Linking morphodynamic response with sediment mass balance on the Colorado River in Marble Canyon—Issues of scale, geomorphic setting, and sampling design: Journal of Geophysical Research: Earth Surface, v. 118, no. 2, p. 361-381, <https://doi.org/10.1002/jgrf.20050>.
- Griffiths, R.E., and Topping, D.J., 2017, Importance of measuring discharge and sediment transport in lesser tributaries when closing sediment budgets: Geomorphology, v. 296, p. 59-73, <https://doi.org/10.1016/j.geomorph.2017.08.037>.
- Griffiths, R.E., Topping, D.J., Anderson, R.S., Hancock, G.S., and Melis, T.S., 2014, Design of a sediment-monitoring gaging network on ephemeral tributaries of the Colorado River in Glen, Marble, and Grand Canyons, Arizona: U.S. Geological Survey Open-File Report 2014-1137, 21 p., <http://dx.doi.org/10.3133/ofr20141137>.
- Griffiths, R.E., Topping, D.J., Andrews, T., Bennett, G.E., Sabol, T.A., and Melis, T.S., 2012, Design and maintenance of a network for collecting high-resolution suspended-sediment data at remote locations on rivers, with examples from the Colorado River: U.S. Geological Survey Techniques and Methods, Book 8, chapter C2, 44 p., <https://doi.org/10.3133/tm8c2>.
- Griffiths, R.E., Topping, D.J., and Unema, J.A., 2024, Changes in sand storage in the Colorado River in Grand Canyon National Park from July 2017 through June 2020: U.S. Geological Survey Open-File Report 2023-1093, 9 p., <https://doi.org/10.3133/ofr20231093>.

- Guy, H.P., 1970, Fluvial sediment concepts: Techniques of Water-Resources Investigations of the United States Geological Survey, Book 3, chapter C1, 55 p., <https://doi.org/10.3133/twri03C1>.
- Hall, R.O. Jr., Yackulic, C.B., Kennedy, T.A., Yard, M.D., Rosi-Marshall, E.J., Voichick, N., and Behn, K.E., 2015, Turbidity, light, temperature, and hydropeaking control primary productivity in the Colorado River, Grand Canyon: *Limnology and Oceanography*, v. 60, no. 2, p. 512-526, <https://doi.org/10.1002/lno.10031>.
- Howard, C.S., 1947, Suspended sediment in the Colorado River 1925-41: U.S. Geological Survey Water-Supply Paper 998, 165 p., <https://pubs.usgs.gov/wsp/0998/report.pdf>.
- Kennedy, T.A., Muehlbauer, J.D., Yackulic, C.B., Lytle, D.A., Miller, S.A., Dibble, K.L., Kortenhoeven, E.W., Metcalfe, A.N., and Baxter, C.V., 2016, Flow management for hydropower extirpates aquatic insects, undermining river food webs: *BioScience*, v. 66, p. 561–575, <https://doi.org/10.1093/biosci/biw059>.
- Korman, J., and Campana, S.E., 2009, Effects of hydropeaking on nearshore habitat use and growth of age-0 rainbow trout in a large regulated river: *Transactions of the American Fisheries Society*, v. 138, p. 76-87, <https://doi.org/10.1577/T08-026.1>.
- Korman, J., Yard, M.D., and Yackulic, C.B., 2015, Factors controlling the abundance of rainbow trout in the Colorado River in Grand Canyon in a reach utilized by endangered humpback chub: *Canadian Journal of Fisheries and Aquatic Sciences*, v. 73, p. 105–124, <https://doi.org/10.1139/cjfas-2015-0101>.
- Le Coz, J., Perret, E., Camenen, B., Topping, D.J., Buscombe, D.D., Leary, K.C.P., Dramais, G., and Grams, P.E., 2022, Mapping 2-D bedload rates throughout a sand-bed river reach from high-resolution acoustical surveys of migrating bedforms: *Water Resources Research*, v. 58, e2022WR032434, <https://doi.org/10.1029/2022WR032434>.
- Marggraf, J., 2024, Improving methods for the hydroacoustic monitoring of suspended sand concentration and grain size—Application to the Isère River at Grenoble Campus: Lyon, France, University of Lyon, Ph.D. dissertation, 480 p.
- Mueller, D.S., Wagner, C.R., Rehmel, M.S., Oberg, K.A., and Rainville, Francois, 2013, Measuring discharge with acoustic Doppler current profilers from a moving boat (ver. 2.0, December 2013): U.S. Geological Survey Techniques and Methods, book 3, chap. A22, 95 p., <http://dx.doi.org/10.3133/tm3A22>.
- National Research Council, 1996, River resource management in the Grand Canyon: Washington, D.C., National Academy Press, Committee to review the Glen Canyon Environmental Studies, Water and Technology Board, Committee on Geosciences, Environment, and Resources, 226 p., http://www.nap.edu/catalog.php?record_id=5148.
- Neher, C., Duffield, J., Bair, L., Patterson, D., and Neher, N., 2017, Testing the limits of temporal stability—Willingness to pay values among Grand Canyon whitewater boaters across decades: *Water Resources Research*, v. 53, p. 10108–10120, <https://doi.org/10.1002/2017WR020729>.

- Porterfield, G., 1972, Computation of fluvial sediment discharge: Techniques of Water-Resources Investigations of the U.S. Geological Survey 03-C3, 66 p., <https://doi.org/10.3133/twri03C3>.
- Rantz, S.E., and others, 1982a, Measurement and computation of streamflow—volume 1, Measurement of stage and discharge: U.S. Geological Survey Water-Supply Paper 2175, p. 1–284, <https://doi.org/10.3133/wsp2175>.
- Rantz, S.E., and others, 1982b, Measurement and computation of streamflow—volume 2, Computation of discharge: U.S. Geological Survey Water-Supply Paper 2175, p. 285–631, <https://doi.org/10.3133/wsp2175>.
- Rubin, D.M., Buscombe, D., Wright, S.A., Topping, D.J., Grams, P.E., Schmidt, J.C., Hazel, J.E., Jr., Kaplinski, M.A., and Tusso, R., 2020, Causes of variability in suspended-sand concentration evaluated using measurements in the Colorado River in Grand Canyon: *Journal of Geophysical Research: Earth Surface*, v. 125, e2019JF005226, <https://doi.org/10.1029/2019JF005226>.
- Rubin, D.M., Tate, G.M., Topping, D.J., and Anima, R.A., 2001, Use of rotating side-scan sonar to measure bedload: *Proceedings of the 7th Interagency Sedimentation Conference*, March 25–29, Reno, NV, v. 1, p. III-139–III-143,
- Rubin, D.M., and Topping, D.J., 2001, Quantifying the relative importance of flow regulation and grain-size regulation of suspended-sediment transport α , and tracking changes in bed-sediment grain size β : *Water Resources Research*, v. 37, p. 133-146, <https://doi.org/10.1029/2000WR900250>.
<http://www.riversimulator.org/Resources/GCMRC/PhysicalResources2/Rubin2001c.pdf>.
- Rubin, D.M., and Topping, D.J., 2008, Correction to "Quantifying the relative importance of flow regulation and grain-size regulation of suspended-sediment transport α , and tracking changes in bed-sediment grain size β ": *Water Resources Research*, v. 44, no. 9, 5 p., <https://doi.org/10.1029/2008wr006819>.
- Rubin, D.M., Topping, D.J., Schmidt, J.C., Hazel, J., Kaplinski, K., and Melis, T.S., 2002, Recent sediment studies refute Glen Canyon Dam hypothesis: *EOS, Transactions, American Geophysical Union*, v. 83, no. 25, p. 273, 277-278, <https://doi.org/10.1029/2002EO000191>.
- Sabol, T.A., and Topping, D.J., 2013, Evaluation of intake efficiencies and associated sediment-concentration errors in US D-77 bag-type and US D-96-type depth-integrating suspended-sediment samplers: U.S. Geological Survey Scientific Investigations Report 2012-5208, 88 p., <http://dx.doi.org/10.3133/sir20125208>
- Sabol, T.A., Topping, D.J., Griffiths, R.E., and Dramais, G., 2022, Field investigation of sub-isokinetic sampling by the US D-96-type suspended-sediment sampler and its effect on suspended-sediment measurements: U.S. Geological Survey Open-File Report 2022–1077, 14 p., <https://doi.org/10.3133/ofr20221077>.
- Sankey, J.B., East, A., Fairley, H.C., Caster, J., Dierker, J., Brennan, E., Pilkington, L., Bransky, N., and Kasprak, K., 2023, Archaeological sites in Grand Canyon National Park along the Colorado River are eroding owing to six decades of Glen Canyon Dam operations: *Journal of*

- Environmental Management, v. 342, 118036,
<https://doi.org/10.1016/j.jenvman.2023.118036>.
- Schmidt, J.C., Rubin, D.M., and Ikeda, H., 1993, Flume simulation of recirculating flow and sedimentation: *Water Resources Research*, v. 29, no. 8, p. 2925-2939,
<https://doi.org/10.1029/93WR00770>.
- Schmidt, J.C., Topping, D.J., Rubin, D.M., Hazel, J.E., Jr., Kaplinski, M., Wiele, S.M., and Goeking, S.A., 2007, Streamflow and sediment data collected to determine the effects of low summer steady flows and habitual maintenance flows in 2000 on the Colorado River between Lees Ferry and Bright Angel Creek, Arizona: U.S. Geological Survey Open-File Report 2007-1268, 79 p., <https://doi.org/10.3133/ofr20071268>.
- Sibley, D., Topping, D.J., Hines, M., and Garner, B., 2015, User-interactive sediment budgets in a browser—A web application for river science and management, in *Proceedings of the 3rd Joint Federal Interagency Conference (10th Federal Interagency Sedimentation Conference and 5th Federal Interagency Hydrologic Modeling Conference)*, Reno, Nev., April 19-23, 2015: p. 595-605, <https://acwi.gov/sos/pubs/3rdJFIC/Contents/4A-Sibley.pdf>.
- Smith, J.D., 1970, Stability of a sand bed subjected to a shear flow of low Froude number: *Journal of Geophysical Research*, v. 75, 5928–5940,
<https://doi.org/10.1029/JC075i030p05928>.
- Topping, D.J., 1997, Physics of flow, sediment transport, hydraulic geometry, and channel geomorphic adjustment during flash floods in an ephemeral river, the Paria River, Utah and Arizona: Seattle, University of Washington, unpublished Ph.D. dissertation, 406 p.
- Topping, D.J., Grams, P.E., Griffiths, R.E., Dean, D.J., Wright, S.A., and Unema, J.A., 2021, Self-limitation of sand storage in a bedrock-canyon river arising from the interaction of flow and grain size: *Journal of Geophysical Research—Earth Surface*, v. 126, e2020JF005565,
<https://doi.org/10.1029/2020JF005565>.
- Topping, D.J., Grams, P.E., Griffiths, R.E., Hazel, J.E., Kaplinski, M.A., Dean, D.J., Voichick, N., Unema, J.A., and Sabol, T.A., 2019, Optimal timing of high-flow experiments for sandbar deposition, *in* High-flow experiments assessment extended abstracts—Glen Canyon Dam Adaptive Management Program Annual Reporting Meeting presentations, March 12-13, 2019, Phoenix, Ariz.: U.S. Geological Survey, Grand Canyon Monitoring and Research Center, p. 3-9, https://www.usbr.gov/uc/progact/amp/amwg/2019-03-06-amwg-meeting/20190301-HFE_Extended_Abstracts-Combined_FINAL.pdf
- Topping, D.J., Griffiths, R.E., Unema, J.A., and Dean, D.J., in press, Controls on sediment transport and storage in the Little Snake, Yampa, and Green rivers in the vicinities of Dinosaur National Monument and Ouray National Wildlife Refuge, Colorado and Utah, with implications for fish habitat in the middle Green River: U.S. Geological Survey Scientific Investigations Report.
- Topping, D.J., Mueller, E.R., Schmidt, J.C., Griffiths, R.E., Dean, D.J., and Grams, P.E., 2018, Long-term evolution of sand transport through a river network: Relative influences of a dam

- versus natural changes in grain size from sand waves, *Journal of Geophysical Research: Earth Surface*, v. 123, <https://doi.org/10.1029/2017JF004534>.
- Topping, D.J., Rubin, D.M., Grams, P.E., Griffiths, R.E., Sabol, T.A., Voichick, N., Tusso, R.B., Vanaman, K.M., and McDonald, R.R., 2010, Sediment transport during three controlled-flood experiments on the Colorado River downstream from Glen Canyon Dam, with implications for eddy-sandbar deposition in Grand Canyon National Park: U.S. Geological Survey Open-File Report 2010-1128, 111 p, <https://doi.org/10.3133/ofr20101128>.
- Topping, D.J., Rubin, D.M., and Melis, T.S., 2007, Coupled changes in sand grain size and sand transport driven by changes in the upstream supply of sand in the Colorado River—Relative importance of changes in bed-sand grain size and bed-sand area: *Sedimentary Geology*, v. 202, no. 3, p. 538-561, <https://doi.org/10.1016/j.sedgeo.2007.03.016>.
- Topping, D.J., Rubin, D.M., Nelson, J.M., and Kinzel, P.J., III, 2000b, Colorado River sediment transport—2. Systematic bed-elevation and grain-size effects of sand supply limitation: *Water Resources Research*, v. 36, no. 2, p. 543-570, <https://doi.org/10.1029/1999WR900286>.
- Topping, D.J., Rubin, D.M., and Vierra, L.E., Jr., 2000a, Colorado River sediment transport—1. Natural sediment supply limitation and the influence of the Glen Canyon Dam: *Water Resources Research*, v. 36, no. 2, p. 515-542, <https://doi.org/10.1029/1999WR900285>.
- Topping, D.J., Rubin, D.M., Wright, S.A., and Melis, T.S., 2011, Field evaluation of the error arising from inadequate time averaging in the standard use of depth-integrating suspended-sediment samplers: U.S. Geological Survey Professional Paper 1774, 95 p., <http://pubs.usgs.gov/pp/1774/>.
- Topping, D.J., Schmidt, J.C., and Vierra, L.E., 2003, Computation and analysis of the instantaneous-discharge record for the Colorado River at Lees Ferry, Arizona—May 8, 1921, through September 30, 2000: U.S. Geological Survey Professional Paper 1677, 118 p., <http://pubs.water.usgs.gov/pp1677/>.
- Topping, D.J., and Wright, S.A., 2016, Long-term continuous acoustical suspended-sediment measurements in rivers—Theory, application, bias, and error: U.S. Geological Survey Professional Paper 1823, 98 p., <https://doi.org/10.3133/pp1823>.
- U.S. Department of the Interior, 1995, Operation of Glen Canyon Dam, Colorado River storage project, Arizona—Final Environmental Impact Statement: Salt Lake City, Utah, U.S. Department of Interior, Bureau of Reclamation, 337 p. plus appendices, <https://www.usbr.gov/uc/envdocs/eis/gc/gcdOpsFEIS.html>.
- U.S. Department of the Interior, 1996, Record of Decision—Operation of Glen Canyon Dam Final Environmental Impact Statement: Salt Lake City, Utah, U.S. Department of the Interior, Bureau of Reclamation, Glen Canyon Dam Adaptive Management Program, 15 p., https://www.usbr.gov/uc/envdocs/rod/Oct1996_OperationGCD_ROD.pdf.
- U.S. Department of the Interior, 2011, Environmental assessment—Development and implementation of a protocol for high-flow experimental releases from Glen Canyon Dam, Arizona, 2011 through 2020: Salt Lake City, Utah, U.S. Department of the Interior, Bureau

- of Reclamation, Upper Colorado Region, 176 p. plus appendices, <https://www.usbr.gov/uc/envdocs/ea/gc/HFEProtocol/HFE-EA.pdf>.
- U.S. Department of the Interior, 2016a, Glen Canyon Dam Long-term Experimental and Management Plan final Environmental Impact Statement (LTEMP FEIS): U.S. Department of the Interior, Bureau of Reclamation, Upper Colorado Region, National Park Service, Intermountain Region, online, <http://ltempeis.anl.gov/documents/final-eis/>.
- U.S. Department of the Interior, 2016b, Record of Decision for the Glen Canyon Dam Long-Term Experimental and Management Plan final Environmental Impact Statement (LTEMP ROD): Salt Lake City, Utah, U.S. Department of the Interior, Bureau of Reclamation, Upper Colorado Region, National Park Service, Intermountain Region, 22 p. plus appendices, http://ltempeis.anl.gov/documents/docs/LTEMP_ROD.pdf.
- U.S. Geological Survey, 2024, Discharge, sediment, and water quality monitoring: Flagstaff, Ariz., Grand Canyon Monitoring and Research Center, online data accessed on March 10, 2024, https://www.gcmrc.gov/discharge_qw_sediment/.
- U.S. Department of the Interior, 2024, Glen Canyon Dam Long-Term Experimental and Management Plan, Draft Supplemental Environmental Impact Statement: U.S. Department of the Interior, Bureau of Reclamation, Upper Colorado River Basins, Interior Region 7, 374 p., https://www.usbr.gov/uc/DocLibrary/EnvironmentalImpactStatements/GlenCanyonDamLong-TermExperimentalManagementPlan/20240207-Draft-GCDLTEMP-SEIS_508.pdf.
- Unema, J.A., Topping, D.J., Kohl, K.A., Pillow, M.J., and Caster, J.J., 2021, Historical floods and geomorphic change in the lower Little Colorado River during the late 19th to early 21st centuries: U.S. Geological Survey Scientific Investigations Report 2021–5049, 34p., <https://doi.org/10.3133/sir20215049>.
- Vernieu, W.S., Hueftle, S.J., and Gloss, S.P., 2005, Water quality in Lake Powell and the Colorado River, *in* Gloss, S.P., Lovich, J.E., and Melis, T.S., eds., The state of the Colorado River ecosystem in Grand Canyon—A report of the Grand Canyon Monitoring and Research Center 1991-2004: U.S. Geological Survey Circular 1282, 69-85 p., <https://pubs.usgs.gov/circ/1282/>.
- Voichick, N., 2008, Specific conductance in the Colorado River between Glen Canyon Dam and Diamond Creek, northern Arizona, 1988–2007: U.S. Geological Survey Data Series 364, 16 p, <https://pubs.usgs.gov/ds/364/>.
- Voichick, N., Kennedy, T., Topping, D., Griffiths, R., and Fry, K., 2016, Water clarity of the Colorado River—Implications for food webs and fish communities: U.S. Geological Survey Fact Sheet 2016-3053, 4 p., <https://doi.org/10.3133/fs20163053>.
- Voichick, N., and Topping, D.J., 2010, Use of specific conductance in estimating salinity and as a natural tracer of water parcels in the Colorado River between Glen Canyon Dam and Diamond Creek, Northern Arizona, *in* Melis, T.S., Hamill, J.F., Bennett, G.E., Coggins, L.G., Jr., Grams, P.E., Kennedy, T.A., Kubly, D.M., and Ralston, B.E., eds., Proceedings of the Colorado River Basin Science and Resource Management Symposium, November 18-20, 2008, Scottsdale, Arizona: U.S. Geological Survey Scientific Investigations Report 2010-5135, 343-46 p., <https://pubs.usgs.gov/sir/2010/5135/>.

- Voichick, N., and Topping, D.J., 2014, Extending the turbidity record—Making additional use of continuous data from turbidity, acoustic-Doppler, and laser diffraction instruments and suspended-sediment samples in the Colorado River in Grand Canyon: U.S. Geological Survey Scientific Investigations Report 2014-5097, 31 p., <https://doi.org/10.3133/sir20145097>.
- Voichick, N., Topping, D.J., and Griffiths, R.E., 2018, Technical note—False low turbidity readings during high suspended-sediment concentrations: *Hydrology and Earth System Sciences*, v. 22, p. 1767-1773, <https://doi.org/10.5194/hess-22-1767-2018>.
- Voichick, N., and Wright, S.A., 2007, Water-temperature data for the Colorado River and tributaries between Glen Canyon Dam and Spencer Canyon, northern Arizona, 1988–2005: U.S. Geological Survey Data Series 251, 24 p., <https://pubs.usgs.gov/ds/2007/251/>.
- Wagner, R.J., Boulger, R.W., Jr., Oblinger, C.J., and Smith, B.A., 2006, Guidelines and standard procedures for continuous water-quality monitors—Station operation, record computation, and data reporting: U.S. Geological Survey Techniques and Methods, book 1, chapter D3, 51 p. plus 8 attachments, <https://doi.org/10.3133/tm1D3>.
- Ward, D.L., and Morton-Starner, R., 2015, Effects of water temperature and fish size on predation vulnerability of juvenile humpback chub to rainbow trout and brown trout: *Transactions of the American Fisheries Society*, v. 144, p. 1184–1191, <https://doi.org/10.1080/00028487.2015.1077160>.
- Ward, D.L., Morton-Starner, R., and Vaage, B., 2016, Effects of turbidity on predation vulnerability of juvenile humpback chub to rainbow trout and brown trout: *Journal of Fish and Wildlife Management*, v. 7, p. 205–212, <https://doi.org/10.3996/102015-JFWM-101>.
- Wiele, S.M., and Griffin, E.R., 1998, Modifications to a one-dimensional model of unsteady flow in the Colorado River through Grand Canyon, Arizona: U.S. Geological Survey Water-Resources Investigations Report 97-4046, 17 p., <https://doi.org/10.3133/wri974046>.
- Wiele, S.M., and Smith, J.D., 1996, A reach-averaged model of diurnal discharge wave propagation down the Colorado River through the Grand Canyon: *Water Resources Research*, v. 32, no. 5, p. 1375–1386, <https://doi.org/10.1029/96WR00199>.
- Wright, S.A., Anderson, C.R., and Voichick, N., 2009, A simplified water temperature model for the Colorado River below Glen Canyon Dam: *River Research and Applications*, v. 25, no. 6, p. 675-686, <https://doi.org/10.1002/rra.1179>.
- Wright, S.A., Melis, T.S., Topping, D.J., and Rubin, D.M., 2005, Influence of Glen Canyon Dam operations on downstream sand resources of the Colorado River in Grand Canyon, in Gloss, S.P., Lovich, J.E., and Melis, T.S., eds., *The state of the Colorado River ecosystem in Grand Canyon—A report of the Grand Canyon Monitoring and Research Center 1991-2004*: U.S. Geological Survey Circular 1282, p. 17-31, <https://pubs.usgs.gov/circ/1282/>.
- Wright, S.A., Schmidt, J.C., Melis, T.S., Topping, D.J., and Rubin, D.M., 2008, Is there enough sand? Evaluating the fate of Grand Canyon sandbars: *GSA Today*, v. 18, no. 8, p. 4-10, <https://doi.org/10.1130/GSATG12A.1>.

- Wright, S.A., Topping, D.J., Rubin, D.M., and Melis, T.S., 2010, An approach for modeling sediment budgets in supply-limited rivers: *Water Resources Research*, v. 46, no. 10, p. 1-18, <https://doi.org/10.1029/2009WR008600>.
- Yackulic, C.B., Korman, J., Yard, M.D., and Dzul, M., 2018, Inferring species interactions through joint mark–recapture analysis: *Ecology*, v. 99, no. 4, p. 812–821, <https://doi.org/10.1002/ecy.2166>.
- Yackulic, C.B., and Hull, J.B., 2019, Effects of water temperature, turbidity, and rainbow trout on humpback chub population dynamics: U.S. Geological Survey Fact Sheet 2019–3049, 4 p., <https://doi.org/10.3133/fs20193049>.
- Yard, M.D., Coggins, L.G., Baxter, C.V., Bennett, G.E., and Korman, J., 2011, Trout piscivory in the Colorado River, Grand Canyon—Effects of turbidity, temperature, and fish prey availability: *Transactions of the American Fisheries Society* v. 140, no. 2, p. 471–486, <https://doi.org/10.1080/00028487.2011.572011>.

Project B: Sandbar and Sediment Storage Monitoring and Research

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Project Summary and Purpose

The purposes of this project are to a) track the effects of individual High-Flow Experiments (HFEs) on sandbars and campsites, b) monitor the cumulative effect of successive HFEs and intervening operations on sandbars and sand conservation, and c) investigate the interactions between dam operations, sand transport, and eddy sandbar dynamics. These objectives are accomplished by annual measurements at long-term sandbar monitoring sites (B.1), measurements of changes in riverbed sand storage and studies of riverbed dynamics (B.2), maintenance of a geodetic control network (B.3), and development of streamflow, sediment transport, and sandbar response models (B.4). Field activities that would occur for monitoring condition-dependent experimental actions such as HFEs are also described (B.5). Results from the monitoring elements of this project are used to evaluate progress towards meeting the Long-Term Experimental and Management Plan (LTEMP) goal, to “Increase and retain fine sediment volume, area, and distribution in the Glen, Marble, and Grand Canyon reaches above the elevation of the average base flow for ecological, cultural, and recreational purposes.” (U.S. Department of the Interior, 2016). The models developed and maintained in this project are used to plan and design HFEs and to forecast the response of sediment and sandbars to potential flow scenarios for long-term planning.

The implementation of HFEs for building sandbars that is one of the central components of LTEMP is an extension of the program initiated with the Environmental Assessment for Development and Implementation of a Protocol for High-Flow Experimental Releases from Glen Canyon Dam (HFE Protocol; U.S. Department of the Interior, 2011). The HFE Protocol asked the question, "Can sandbar building during HFEs exceed sandbar erosion during periods between HFEs, such that sandbar size can be increased and maintained over several years?" In other words, does the volume of sand aggraded into eddies and onto sandbars during controlled floods exceed the volume eroded from sandbars during intervening dam operations? Expectations of improved deposition on sandbars and conservation of sediment were, therefore, among the criteria used in the selection of the preferred LTEMP alternative.

In addition, condition-dependent experiments were included in the preferred alternative, with objectives related to sandbar building and sediment conservation.

Project B includes elements that are designed to evaluate whether the sediment-related goals of the LTEMP are met, provide the information that is needed to proceed with or abort LTEMP experimental activities, evaluate the effectiveness of implemented experiments, and develop predictive models for future planning efforts. Additionally, these models are needed to evaluate ecological impacts of potential dam operations associated with habitat availability and turbidity.

The sandbar monitoring program described here was outlined in the LTEMP Science Plan (Vanderkooi and others, 2017) and provides the data required to compute the sandbar monitoring metric that will be used answer the fundamental question of the HFE Protocol and the LTEMP by monitoring changes in sandbars over many years, including a period that contains several controlled floods. The program is a continuation of the monitoring implemented in previous work plans and is based on annual measurements of sandbars, using conventional topographic surveys supplemented with daily measurements of sandbar change using ‘remote cameras’ that autonomously and repeatedly take photographs. These annual measurements and daily photographs are included in Project Element B.1. Because these long-term monitoring sites represent only a small proportion of the total number of sandbars in Marble and Grand Canyons, Project Element B.2 includes periodic measurements of a larger sample of sandbars within individual 50 to 130 km sediment budget reaches (see Project A for description of sediment budget reaches).

The other critical information that is needed to evaluate the outcome of the HFE Protocol and the LTEMP is the sand supply metric, which is the change in total sand storage in long river reaches. HFEs build sandbars by redistributing sand from the low-elevation portion of the channel to sandbars in eddies and on the banks. The sand available for deposition is the sand that is in storage on the channel bed, which is the sum of the sand contributed by the most recent tributary inputs, any sand that may have accumulated since Glen Canyon Dam (GCD) was completed, and any sand that remains from the pre-dam era. The goal of the HFE Protocol is to accomplish sandbar building by mobilizing only the quantity of sand most recently contributed by the Paria River, thereby preventing depletion of pre-dam era sand. For this reason, conservation of sand was one of the criteria used to evaluate and select the preferred alternative in the LTEMP ROD. Measured trends in sand storage along the channel bed combined with trends in the volume and area of exposed sandbars will provide the necessary information on which to base future decisions about dam operations and other potential management options. If sand storage is maintained or increased, we expect the response of sandbar deposition to future HFEs to be similar or greater than that observed following recent HFEs. In contrast, depletions of fine sediment in the active channel are potentially irreversible if sand supply from tributaries is consistently less than downstream transport. This situation would threaten the long-term ability to maintain sandbars.

These long-term trends are measured in Project Element B.2, which includes one channel mapping campaign. In this work plan, we will map the West-central Grand Canyon segment between River Mile (RM) 166 and 225 on one motorized trip in 2026.

The Control Network and Survey Support element (B.3) has existed to support the development and maintenance of the geodetic control network in Grand Canyon that is used to reference all ground-based and remotely sensed geospatial measurements made by GCMRC. In the past, this element has supported a full-time geodesist, but that position is currently vacant. Potential paths forward are discussed in the project element description, below.

The Streamflow, Sediment, and Sandbar Modeling Project (B.4) includes maintenance and development of models for predicting sand mass balance, sandbar volumes, turbidity, and streamflow. The sand routing and sandbar models are used to compute the LTEMP performance metrics for sand mass balance and sandbars in a predictive framework and are thereby used for planning management actions such as HFEs and proposed changes in dam operations, including hourly, daily, monthly, and annual releases. If funded, this project would support developing a predictive model for turbidity, which is important for a variety of ecological processes, including gross primary productivity, phosphorous loads, and fish dynamics. This project element would also include development of a 2D streamflow model from central Marble Canyon (RM 30) to the Little Colorado River confluence (RM 61). This model will provide spatially resolved hydraulic information that is important for evaluating habitat availability and larval fish transport under different dam operations, and which will provide the groundwork for future morphodynamic models and coupled streamflow-vegetation modeling.

Several experimental components are included in Project B. These include additional sandbar surveys to evaluate condition-dependent experiments that affect sandbars and sediment resources, including extended-duration HFEs, proactive spring HFEs, variations in HFE downramp rate, and potential nonnative fish management flows. Additional bathymetric surveys will be performed in the western Grand Canyon study reach in the event of an HFE. We have also added an experimental element that would include substrate mapping between Glen Canyon Dam and Lees Ferry in response to a high level of interest in habitat in that reach.

Long-Term Experimental and Management Plan (LTEMP) Resource Goals

Project B supports the following LTEMP Resource Goals:

- Sediment
- Recreational Experience
- Archaeological and Cultural Resources
- Tribal Resources

Sediment goals are addressed in each of the questions and monitoring metrics are described in the descriptions for Project Elements B.1, B.2, and B.4. Recreational goals are addressed in Project Element B.1, which includes measurements of campsite area.

Archaeological and cultural resource goals are addressed more directly in Project D, but measurements and modeling of the area of bare sand available for transport by wind in Project Elements B.1, B.2, and B.4 support that work.

Goals for riparian resources are addressed directly in Project C, but measurements of sandbars in B.2 support vegetation monitoring and modeling in B.4 supports future modeling of riparian vegetation. Monitoring in western Grand Canyon as part of B.2 and B.5.5 contributes to the tribal resources goal by studying how dam operations affect the river channel dynamics in the western Grand Canyon, which is especially important to Hualapai commercial river operations. Ecological goals listed in LTEMP are more directly addressed in other projects, but models for turbidity and spatially resolved streamflow are important for a variety of ecological processes, as described in B.4.

Background

The changes to the flow regime and sediment supply associated with completion of GCD (Topping and others, 2000) caused deep scour and armoring of the riverbed in the 25-km reach between the dam and Lees Ferry (Pemberton, 1976; Williams and Wolman, 1984; Grams and others, 2007). Downstream from Lees Ferry in Marble and Grand Canyons, the debris fans at tributary mouths result in a different channel configuration and different style of response to the upstream dam. The boulder and cobble deposits that form rapids have been largely stable (Magirl and others, 2005), while areas of the bed covered by fine sediment have eroded, and many eddy sandbars are much smaller than before flow regulation (Schmidt, 1990; Wright and others, 2005). Topping and others (2021) estimated a highly approximate lower bound for total sand erosion from Marble and Grand Canyons following the closure of Glen Canyon Dam of at least 28 million metric tons through 2017. Because systematic measurements of fine-sediment thickness have not yet been made, the total volume of fine sediment remaining in Marble and Grand Canyons is not known.

Sandbars are one component of the total sediment budget for the Colorado River. The sediment budget, or sediment mass (or volume) balance, is the accounting by mass of all sediment entering and exiting a given river segment. This budget may be expressed as

$$I - O = \Delta S, \quad (1)$$

where I is the sum of all sediment inputs, O is the sum of all outputs, and ΔS is the net change in the sediment deposits that occurs within the river segment. When inputs exceed outputs, sediment accumulation (deposition) occurs; when outputs exceed inputs, sediment evacuation (erosion) occurs. To provide greater spatial resolution, equation (1) can be partitioned by the elevation zone in which ΔS occurs. Sand stored low in the active channel (ΔS_{low}) is always underwater and sand stored higher in the active channel (ΔS_{high}) is only occasionally inundated. Thus, $\Delta S = \Delta S_{low} + \Delta S_{high}$.

We use *low* to refer to fine-sediment deposits below the stage associated with the 8,000 ft³/s discharge and *high* to refer to fine-sediment deposits above the 8,000 ft³/s stage. The low-elevation deposits are underwater except during the trough of some flow fluctuations and consist of the lower parts of eddy sandbars and patches of sand on the riverbed. These low-elevation deposits determine the physical characteristics of the aquatic environment, such as the characteristics of backwaters that are used by native fishes and are the source for sand remobilized during HFEs. The high-elevation fine-sediment deposits are alternately inundated and exposed, depending on the flow regime. These deposits are used as camping beaches, support riparian vegetation, and support other upland resources.

Annual monitoring of high-elevation deposits has been conducted systematically since 1990. These data clearly demonstrate the role of dam operations, primarily HFEs, in causing changes in sandbar size (Figure 1). Each HFE has resulted in deposition and there has been erosion in each of the periods between HFEs (Hazel and others, 2010; Schmidt and Grams, 2011; Mueller and others, 2014; Grams and others, 2019; Hazel and others, 2022). HFEs conducted with the HFE Protocol since 2012 have resulted in sustained, but not progressive, increases in sandbar volume (Grams and others, 2019; Hazel and others, 2022). Additionally, vegetation has established on portions of sandbars in many parts of the river corridor since the beginning of monitoring (Sankey and others, 2015; Mueller and others, 2018), which may stabilize HFE deposits, but decrease the area of exposed bare sand and campsite area.

Annual monitoring of campsite area at a subset of the high-elevation sandbar monitoring sites has been conducted systematically since 1998 (Kaplinski and others, 2010; Kaplinski and others, 2014a; Hadley and others, 2018). Campsite area monitoring was added to address goals for recreation since most campsites are located on sandbars. Changes to campsite area are primarily linked to changes in sandbar size and riparian vegetation, with minor contributions by gullying from surface runoff (Kaplinski and others, 2014a). These data show campsite area has remained about the same since the implementation of regular HFEs and increases in campsite area associated with HFE deposition are offset by campsite area loss due to riparian vegetation expansion (Hadley and others, 2018). These results highlight the importance of HFEs in meeting program goals for recreation.

Low- and high-elevation deposits are coupled through processes of streamflow erosion and deposition, wind erosion and deposition, and mass failure. This coupling means that changes in ΔS will affect both low- and high-elevation sediment. Although HFEs are scheduled based on the quantity of recent sand inputs from the Paria River, both those inputs and residual sand are mobilized to elevate sand concentrations. Recent investigations of the geochemistry of sand deposited during HFEs indicates that between 60% and 90% of the sand within HFE deposits is likely derived from the Paria River (Chapman and others, 2020), with the remainder composed of pre-dam sediment from the channel and its margins.

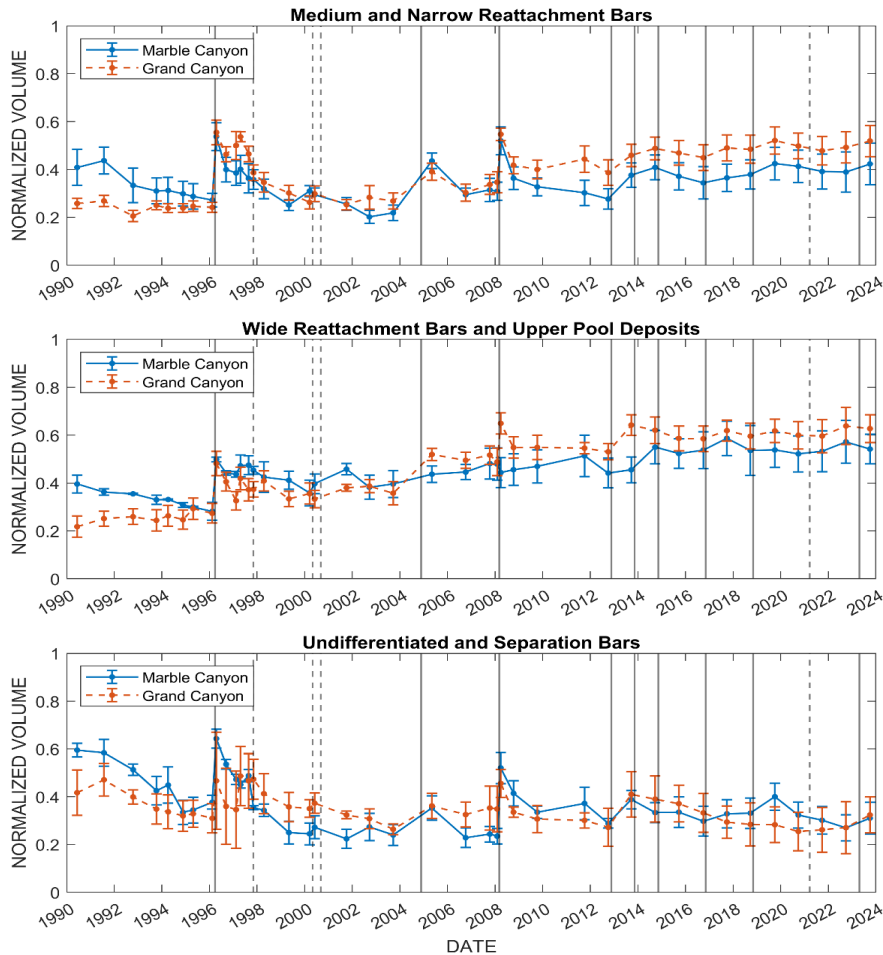


Figure 1. Normalized sandbar volume from 1990 to October 2023. The data are normalized and segregated by bar type, described by Mueller and others (2018). Solid vertical lines indicate high-flow experiments (HFEs) and dashed vertical lines indicate smaller powerplant capacity flows from Glen Canyon Dam.

Paria-derived fractions as low as 25% were observed in some grain size classes for individual samples. Thus, a substantial proportion of the sand deposited during HFEs may be derived from background sand storage – the pre-dam sediment stored in eddies and the riverbed. Because higher concentrations of sand in suspension will result in greater rates of deposition during HFEs (Wiele and others, 1999), decreases in background sand storage – unless they are offset by tributary sediment inputs – will likely lead to diminished capacity to achieve one of the central LTEMP goals of rebuilding and maintaining sandbars using HFEs. Therefore, predictions about the long-term fate of sandbars must be based on understanding long-term trends in ΔS , including both ΔS_{low} and ΔS_{high} . For these reasons, the sandbar research and monitoring is designed around this concept of the sediment budget.

The measurements of suspended sediment made in Project A track the inputs and outputs (I and O in equation (1)) and are used to calculate ΔS for the sediment budget segments.

This approach tracks the accumulation of tributary inputs that is essential for implementation of the HFE Protocol. However, this calculation does not distinguish between low- and high-elevation deposits. Consequently, equation (1) alone cannot be used to evaluate changes in sandbar size, campsite area, sand available for plant colonization, or other changes of recreational or ecological significance. Moreover, uncertainty in the measurements of total sediment flux also accumulate with time (see Project A), limiting the utility of those measurements for tracking long-term trends (Figure 2).

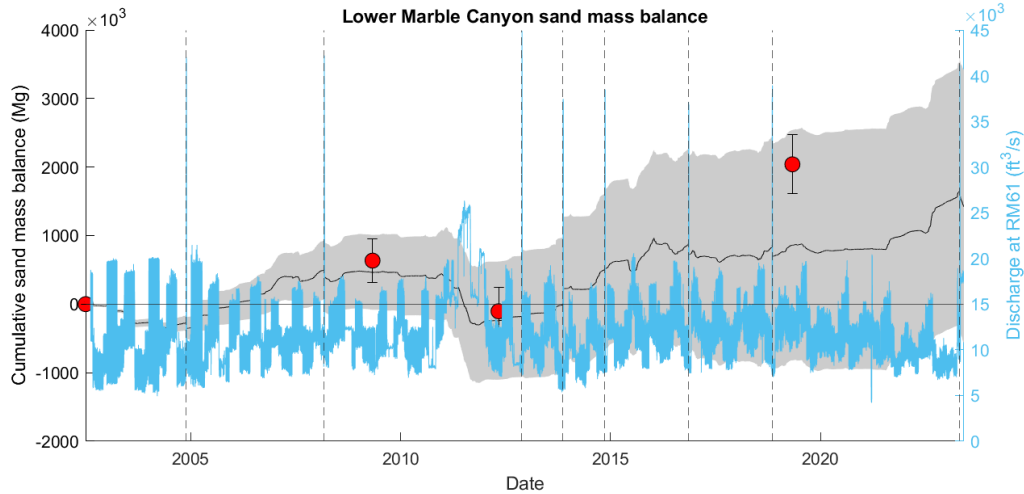


Figure 2. Mass balance sand budget for Lower Marble Canyon from August 2002 to December 2022. The black line shows the mass balance computed from continuous measurements of sediment flux measured by Project A with uncertainty shown by the gray shaded region. The red circles show the mass balance for the same river segment computed from repeat measurements of riverbed elevations measured by Project B.2. These data illustrate the necessity of the Project B.2 measurements of riverbed elevation to compute sediment storage changes over time periods greater than approximately five years. Data from Topping and others (2021), Grams and others (2013), Grams and others (2019), and unpublished data.

Previous studies analyzing repeat topographic measurements of the channel, eddies, and sandbars have found that 90% or more of the changes in sand volume occur at low-elevation, and that high-elevation sandbars comprise only about 10% or less of the fine sediment in the system (Hazel and others, 2006). These studies have also found that ΔS computed for short study reaches yielded different values than ΔS computed as the difference between sand inflows measured at gaging stations using equation (1). This discrepancy stems from the inability to correctly extrapolate measurements from short reaches to larger spatial scales because changes in bed topography are highly localized and spatially variable (Grams and others, 2013; 2019). These findings demonstrate that determining whether sediment storage in each storage environment – at low and high elevations and in the channel and eddy – is increasing, decreasing, or stable requires repeat measurements of sand storage in a large sample of the storage environments within each of the long sediment-budgeting reaches.

Such measurements have been made in some of the sediment-budgeting reaches since 2009, and repeat maps for both lower Marble Canyon and Eastern Grand Canyon capture large spatial variability in erosion and deposition that allow robust calculation of the evacuation of sand that occurred during the period of high releases in summer 2011 (Grams and others, 2019). These measurements also show an overall loss of high-elevation sand in lower Marble Canyon and a slight increase in high-elevation sand in Eastern Grand Canyon. As the period of repeat measurement of the bed and sandbars lengthens, the value of those measurements and the importance of their interpretation will increase.

Figure 3 illustrates how the measurements of bed-sand storage will be interpreted and how they may be used to guide management decisions. This plot shows sand thickness change in bar sand (ΔS_{high}) compared to sand thickness change in the bed sand in eddies and the channel (ΔS_{low}), using data collected in Upper Marble Canyon (UMC), Lower Marble Canyon (LMC) and Eastern Grand Canyon (EGC). The figure also contrasts changes that occur during HFEs with changes that occur over long time periods. The plot is divided into four quadrants based on relative changes in sand on the bed versus sand contained in sandbars, with management implications for each quadrant.

Sandbar deposition accompanied by erosion in the channel (upper left quadrant of Figure 3) is expected to occur over short periods (such as during HFEs) as sand is transferred from low to high elevation. However, this pattern of response over long periods would indicate that sandbars are being built at the expense of progressive depletion of sand from the channel. This occurred in EGC between 2011 and 2014, which included a period of sustained high releases from Glen Canyon Dam. Despite the decrease in supply, sandbars increased because HFEs were implemented in 2012 and 2013. The worst case or “downward spiral” scenario is erosion of high-elevation sandbars accompanied by erosion of low-elevation sand in the channel (lower left quadrant of Figure 3). This occurred in LMC between 2009 and 2012, which included the same period of sustained high releases, but did not include any HFEs. In this case, changes to dam operations such as reducing maximum monthly releases would need to be considered. If changes to operations alone were insufficient, sediment augmentation might become necessary to restore the system to the “sustainable sand management” quadrant (Randle and others, 2007).

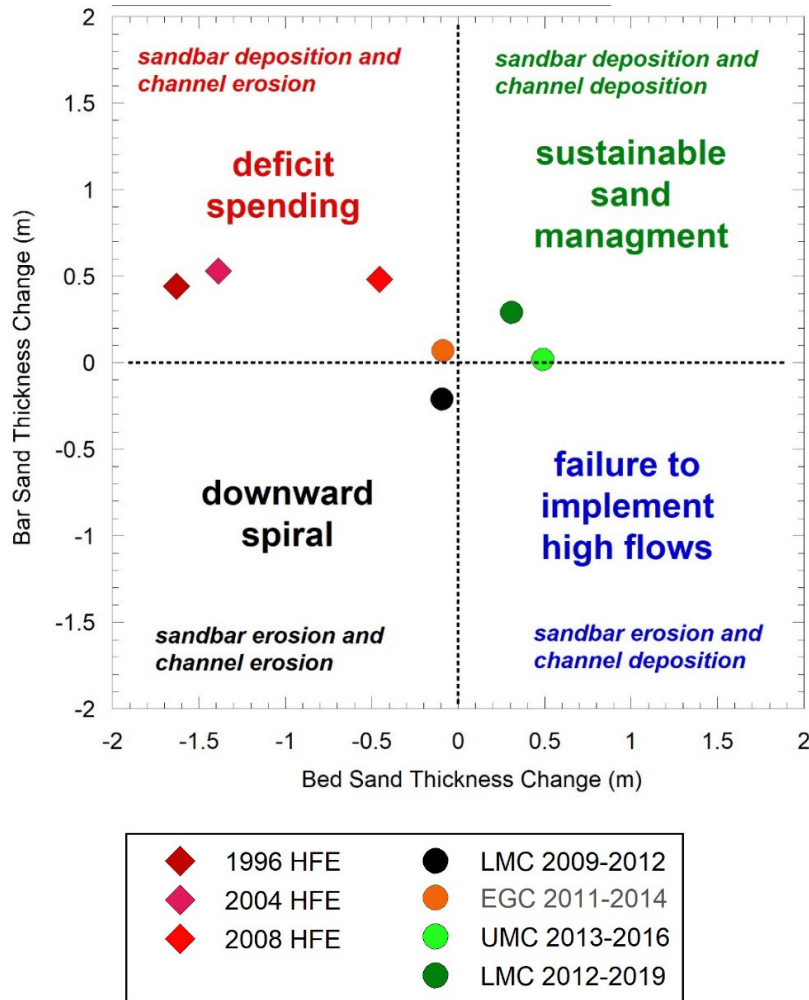


Figure 3. Change in high-elevation sand (sandbars) as a function of change in low-elevation sand (eddy and channel) based on repeat measurements in Upper Marble Canyon (UMC), Lower Marble Canyon (LMC) and Eastern Grand Canyon (EGC). The diamonds show measurements made at individual study sites before and after high-flow experiments (HFEs). The circles show measurements made over the indicated periods for entire ~50-km river segments. Changes are in average thickness of sand. An average thickness change of just 0.5 m over a 50-km reach equates to over 4 million metric tons of sand. Data from Schmidt and Grams (2011), Grams and others (2013), Grams and others (2019), and unpublished data.

The Project B.2 measurements indicate that sustainable sand management (upper right quadrant of Figure 3) has occurred over some periods for some river segments. This occurred in both UMC and LMC for periods that included HFEs and did not include sustained high reservoir releases. One of the objectives of LTEMP is to learn whether this pattern of sustainable sand management can be achieved over long (~20-year) periods. If the measurements made in Project B.2 show this pattern repeatedly, that would indicate that sandbar building and sediment conservation goals are likely being met and, furthermore, that sand supply could support increases in HFE frequency and/or duration.

Finally, sandbar erosion accompanied by sand accumulation in the channel (lower right quadrant of Figure 3) would indicate that despite adequate sand supply, sandbar maintenance goals are not being met. This would suggest that more frequent or longer duration HFEs might be required to achieve different results. In summary, if sandbar maintenance goals are not being met and information on low-elevation sand storage is not available, the cause of sandbar declines will be unknown, and it will not be possible to identify the appropriate management response.

Proposed Work

The sand deposits on the bed and banks of the Colorado River in Glen, Marble, and Grand Canyons are directly affected by the operations of GCD. Depending on the relative magnitudes of dam releases and tributary sediment inputs, sand either accumulates or erodes from the bed of the river. When evaluated over long river segments, sand is evacuated during sustained periods of high dam-releases (Topping and others, 2000; Grams and others, 2015; Grams and others, 2019; Topping and others, 2021) and sand accumulates during periods of average dam-releases and substantial tributary sediment inputs (Grams and others, 2013; Grams and others, 2019).

Sandbars along the riverbanks at elevations above average base flow stage (about 8,000 ft³/s) also change in response to dam operations, but in a different pattern, because they are not always inundated and because they comprise a small fraction of the sand in the system (Hazel and others, 2006; Grams and others, 2013; Hazel and others, 2022).

These deposits aggrade significantly during HFEs that exceed powerplant capacity and, to a lesser extent, during powerplant capacity flows (Schmidt and Grams, 2011; Hazel and others, 2022). These deposits typically erode during normal powerplant operations between HFEs (Hazel and others, 2010; 2022). Efforts by the Glen Canyon Dam Adaptive Management Program (GCDAMP) to manage fine sediment in this context has resulted in the articulation of goals, information needs, and monitoring needs in planning documents, such as the LTEMP EIS (U.S. Department of the Interior, 2016). The science questions articulated within each project element, below, are based on that guidance.

Project Element B.1. Sandbar and Campsite Monitoring with Topographic Surveys and Remote Cameras (Ongoing Study; Partially Funded)

Hypotheses and Science Questions

- Can sandbar building during HFEs exceed sandbar erosion during periods between HFEs, such that sandbar size can be increased and maintained over several years?
- What is the long-term effect of dam operations, including controlled floods, on the size of eddy sandbars at 45 long-term monitoring sites above the 8,000 ft³/s stage?

- How do these changes affect the following recreational and ecosystem resources: camping beaches, substrate for riparian vegetation, and areas of bare sand that are redistributed by wind to upslope locations?

Methods

We will conduct topographic measurements at 45 long-term monitoring sites (Hazel and others, 2022) that will be used to compute sandbar area and volume at each of these sites (Figure 1) and usable campsite area (Hadley and others, 2018) at a subset of 37 sites. The monitoring sites were selected between 1990 and 2002 to represent the range of sandbar types with sites distributed throughout Marble and Grand Canyons such that measurements can be made at all sites on a single annual river trip. Although these study sites comprise less than 10 percent of the subaerially exposed sandbars of similar type in all of Grand Canyon (Hazel and others, 2022), they have been shown to be representative of a larger sample of sandbars of similar type in Lower Marble Canyon and Eastern Grand Canyon (Hazel and others, 2022).

The degree to which these monitoring sites are representative of sandbars in western Grand Canyon where sites are more sparsely distributed is poorly known. Additionally, these sites do not represent the narrow, river-parallel, deposits of sand that occur outside of eddies which are used infrequently for camping.

We will evaluate replacing up to three of the sites that are no longer sensitive to dam operations owing to vegetation expansion with sites that may be of greater interest, such as sites that have active vegetation management programs in place. This project will also include maintenance of remote cameras for daily monitoring at 42 of the 45 long-term monitoring sites (Grams and others, 2018).

The methods for data collection, processing, and analysis are described in detail by Hazel and others (2022). In brief, each site is surveyed for topography and campsite area using conventional survey methods with electronic total stations. Lidar or photogrammetric methods (airborne or ground based) are not practical because they do not reliably penetrate the dense vegetation or measure submerged topography. The submerged topography must be surveyed to consistently measure down to the elevation that is inundated by discharges of 8,000 ft³/s.

The topographic surveys are processed to create digital elevation models (DEMs) for each site for each year. The DEMs are compared among all years to create annual metrics for sand volume and area in two zones: 1) the zone that is inundated by discharges above 8,000 ft³/s and less than 25,000 ft³/s and, 2) the zone that is inundated by discharges above 25,000 ft³/s and up to the elevation that is inundated by dam operations, including HFEs. All analyses are performed in a processing workflow that includes open-source processing scripts and stores the results in a database. The database is linked to a website that serves the sandbar data and images from the remote cameras (U.S. Geological Survey, 2024b, <https://www.usgs.gov/apps/sandbar>).

Data collection for this project will occur on one non-motorized river trip each year that is done in cooperation with Project C to save logistics costs. Also in cooperation with Project C, we will continue to investigate the interactions between sandbars and vegetation by working on a coupled model for sandbar and vegetation change (subject to funding for B.4 and C.4).

This information could be used to understand why some sites are more or less likely to have vegetation expansion and could also be used to choose sites for vegetation management.

This project also includes support (Unfunded) for the Grand Canyon River Guides Adopt-a-Beach program, which provides an assessment of campsite condition from the perspective of river guides based on repeat photographs, which are available on the GCMRC website (<https://www.usgs.gov/apps/sandbar>).

In the third year of this work plan, we will prepare a summary report describing the response of sandbars to dam operations focusing on the period from 2020 to 2026 because the last major report on sandbars included analysis of data through 2020 (Hazel and others, 2022). This summary report will provide necessary data and analysis to inform the LTEMP ten-year review.

If the LTEMP experiment to study an HFE with lower downramp rate (U.S. Department of the Interior, 2016) is implemented, we will investigate the impact of HFE hydrograph shape on sandbar morphology. Because this work is contingent on the implementation of the experiment, funding for logistics and data analysis are in Experimental Project Element B.5.

Anticipated Use of Data

The purpose of this project element is to monitor the annual status and long-term trends of sandbars and campsites in Glen, Marble, and Grand Canyons. The results will be used to evaluate the effects of dam operations, including HFEs, on sandbars and related resources and will include reporting of the sandbar performance metrics for LTEMP.

Outcomes and Products

- Update at each annual reporting meeting on sandbar area and volume and campsite area based on monitoring from the previous year.
- Annual monitoring data made available on website within six months following data collection.
- Remote camera images showing effects of HFEs made available on website within two months following data collection.
- Adopt-a-Beach photography to be served on website on an annual basis (Unfunded).
- Report and/or journal article on response of sandbars to dam operations, 2020 to 2026.

Project Element B.2. Bathymetric and Topographic Mapping for Monitoring Sediment Storage and Riverbed Dynamics (Ongoing Study; Partially Funded)

Hypotheses and Science Questions

- Can the supply of sand that is utilized for sandbar building during HFEs be maintained, such that HFEs can continue to be utilized to rebuild sandbars in the future?
 - What is the long-term effect of dam operations, including controlled floods, on the total amount of fine sediment stored in the active river channel at low and high elevations?
 - What is the long-term effect of dam operations, including controlled floods, on the distribution, abundance, and size of eddy sandbars throughout Grand Canyon above the 8,000 ft³/s stage?
 - How do these changes affect the following recreational and ecosystem resources: in-channel backwater habitat for native fishes, camping beaches, substrate for riparian vegetation, and areas of bare sand that are redistributed by wind to upslope locations?

Methods

The primary purpose of this project element is to track trends in sandbar conditions and sand storage throughout Grand Canyon over the time scale of LTEMP to provide an evaluation of whether the supply of sand (the sum of recent tributary inputs and background storage) necessary for building sandbars is increasing, decreasing, or stable and a robust measure of high-elevation sandbar change. The important distinction between B.1 and B.2 is that the measurements made in Project Element B.2 include a much larger sample of sandbars than included in the B.1 annual sandbar monitoring and that the measurements in B.2 include both exposed sandbars and sand that is submerged in the river channel.

The sampling design used in this project is based on our current understanding of sediment dynamics, the locations of stream-gaging stations, and the timeframe of the LTEMP. The Colorado River ecosystem (CRe) is divided into seven sediment-budget reaches based on the location of the streamflow and sediment gages (Table 1). For each of the five reaches between Lees Ferry and Diamond Creek, flux-based sand budgets are computed at 15-min. intervals (see Project A). In using these same reach boundaries for long-term sandbar and sand-storage monitoring, we are able to correlate and compare the changes that occur in the channel, eddies, and on sandbars with the measurements of sand transport (Figure 2).

Because erosion and deposition are spatially variable (Grams and others, 2013), it is necessary to measure approximately 50% to 70% of the channel and eddies within each of these reaches to ensure that the signal in sand-storage change is greater than the noise caused by that spatial variability (Grams and others, 2019).

Because about 90% of the sand that is available for redistribution by dam operations is submerged (Hazel and others, 2006), the monitoring method must include measurements of sediment on the bed of the river in eddies and pools. Bed sediment data collection will combine multibeam and singlebeam sonars, coupled with conventional topographic surveys for areas above the water surface. All data are referenced to a Grand Canyon geodetic control network and because global navigation satellite system positioning is unreliable in the narrow and deep canyon, positioning is achieved by shore-based range-azimuth robotic tracking instruments (Kaplinski and others, 2017a, b). These methods have been described by Hazel and others (2008, 2010) and Kaplinski and others (2014b; 2017a, b; 2022b). The data will result in high-resolution DEMs of the mapped reaches for each mapping effort (e.g., Kaplinski and others, 2017a, b; 2020; 2022a).

Because uncertainty in the flux-based mass balance increases cause an indeterminate budget in 5 to 10 years (Figure 2), a repeat interval of 10 years or less is required to track the sediment supply. A repeat interval of 5 to 7 years for most segments was proposed in the previous work plan, because that would provide the most certainty in tracking the sediment supply. That schedule would require data collection for at least one of the five segments nearly every year. In this work plan we propose to extend the target repeat interval to ~10 years for each segment to reduce costs (Table 1). A 10-year repeat interval can be accomplished by collecting data in one segment every other year, reducing logistics costs but still meeting the needs for LTEMP monitoring.

For this three-year work plan, we propose to conduct repeat bathymetric and topographic mapping for the segment between RM 166 and 225 on one motorized trip in the three-year work plan (in FY 2026). The next three-year work plan (FY 2028-30) will require two mapping efforts to stay on the 10-year repeat interval schedule (Table 1). In years without data collection, project personnel will focus on data processing, analysis, and reporting. In FY 2025, we will conclude analysis and reporting on data collected in Upper Marble Canyon in 2024 and in FY 2027, we will report on the data collected in 2026. These reports will be valuable for the ten-year LTEMP review, including an evaluation of changes in sand storage and channel morphology in upper and lower Marble Canyons, eastern Grand Canyon, and west-central Grand Canyon (Table 1).

Table 1. Proposed schedule of channel mapping efforts for this work plan through the period of the Long-Term Experimental and Management Plan (LTEMP).

| Segment | River Miles | Completed Surveys | Next Planned Survey | Interval to Next Survey |
|------------------------------|-------------|-------------------|---------------------|-------------------------|
| 1) Glen Canyon | -15 to 0 | 2000, 2015 | research only | -- |
| 2) Upper Marble Canyon | 0 to 30 | 2013, 2016, 2024* | 2034 | 10 yr |
| 3) Lower Marble Canyon | 30 to 61 | 2009, 2012, 2019 | 2028 | 9 yr |
| 4) Eastern Grand Canyon | 61 to 87 | 2011, 2014, 2019 | 2030 | 11 yr |
| 5) East-central Grand Canyon | 87 to 166 | 2022** | 2032 | 10 yr |
| 6) West-central Grand Canyon | 166 to 225 | 2017 | 2026*** | 9 yr |
| 7) Western Grand Canyon | 225 to 280 | none | research only | |

* Was planned for 2023, but funding was not available until 2024.

** Was planned for 2021, be rescheduled for 2022 to avoid overlap with 2021 overflight.

*** Only survey included in this 3-year work plan.

Upon completion of a repeat map of a reach, the DEMs will be compared to compute the net change in the volume of sediment within the reach. These computations will distinguish between fine, coarse, and mixed sediment using recently developed acoustic sediment classification algorithms (Buscombe and others, 2014a, b; 2017), between sediment stored in the channel and eddies, and between sediment at high- and low-elevation. The resulting maps of bed sediment substrates are as highly resolved as the bathymetric maps, and therefore can be used for physical habitat classification efforts in other projects. In the past, these applications have included quantifying the relative proportions of sand and gravel that are substrate for aquatic invertebrates (Kennedy and others, 2014), the extent of submerged aquatic vegetation (Project E), and long-term changes in sand abundance (Kasprak and others, 2018).

Some aspects of B.2 have been proposed as unfunded to reduce the budget and are described in the following paragraph.

Because the overflight is currently proposed as unfunded in Project L, we removed the ground-truth data collection, which would have included a measurement of the water surface and riverbed profile on a motorized trip during the proposed overflight.

We also proposed to map riverbed substrate in the segment between Glen Canyon Dam and Lees Ferry and investigate the potential for detecting and mapping smallmouth bass (*Micropterus dolomieu*) nests in that reach. In addition, we proposed continued evaluation of riverbed response to dam operations in western Grand Canyon that was initiated in the FY 2021-23 work plan (preparation of a report on that work is in progress). This work was proposed to include annual monitoring of the 3-mile study reach that begins at River Mile 273, development of a sediment budget for western Grand Canyon, and monitoring the stability of Pearce Ferry Rapid. Pearce Ferry rapid is a barrier to fish migration and changes could have significant ecosystem implications. Comparison of photographs taken from 2021 to 2024 indicate the rapid has likely eroded, but the magnitude of erosion is unknown. Additional monitoring of this reach may occur as part of Experimental Project Element B.5.5, which includes additional surveys bracketing HFEs.

Anticipated Use of Data

The results will be used to evaluate the outcome of the flow regime adopted in the LTEMP with respect to sandbar building and sand conservation (Figure 3). The monitoring data will also be used to compute LTEMP performance metrics for high-elevation sandbars and sand mass balance by river segment (Figure 2). Changes in sand storage are tracked over long river segments, providing a spatially explicit quantification of changes in the channel, eddies and sandbars (Grams and others, 2013, 2019). These measurements of sand storage in the river channel are critical because that information will be needed to explain the observed trends in sandbar area and volume and whether HFEs should be conducted more frequently or less frequently than prescribed in the LTEMP. This information will also be needed to assess whether the implemented flow regime is able to achieve sediment-related goals with the available sediment supply, or whether additional management actions should be considered. Results from mapping riverbed substrate composition and studies of riverbed dynamics in response to HFEs and dam operations (unfunded) would be used to better understand aquatic habitat and how dam operations affect riverbed conditions in specific river segments.

Outcomes and Products

- Report and maps for (data release) RM 0 to 32 (mapped in 2024).
- Report and maps for (data release) RM 166 to 225 (to be mapped in 2026). Reporting on this element may be delayed owing to reduced budget that may delay data processing.
- Report and/or journal article on geomorphic changes in Upper Marble Canyon and West-Central Grand Canyon describing effects of dam operations on sandbars and sand storage, 2013-2026.

- Journal article in collaboration with Project A comparing the results, with detailed error analysis, from the continuous mass-balance sand budgets under Project A with all topographic-based sand budgets to date measured by Project B.2.
- Fact sheet for distribution to the general public providing a description of the purpose and key results of Project B.

Project Element B.3. Control Network and Survey Support (Ongoing Study; Unfunded)

The purposes of this project element are to establish and maintain the framework for high-accuracy change detection. This project element ensures that geospatial data collected across all projects of the program are accurately referenced, precisely defined, and can be reliably compared with past and future datasets. This project has been included as an element in Project B for the past several work plan cycles because much of the work in expanding the control network was done in cooperation with other Project B elements. Because this expansion of the control network is largely complete and future work will focus on maintenance of the network, documentation, and database management, we are considering migrating this project element to the Geospatial Science Project (Project K) in future work plans.

Knowledge of the accuracy associated with geospatial data sets produced from disparate sources is required for decision making based on long-term resource monitoring. The accuracies of each data set are determined by comparison with independent sources of higher accuracy referencing a common datum (or geospatial reference system). An accurate geodetic control network provides access to the common datum for change detection, the means to validate the accuracy of subsequent data products and ensures that spatially referenced observations are repeatable. For example, the merging of LiDAR (light detection and ranging remote sensing), sonar, digital imagery, and total station measurements along a reach of river requires consistent reference, especially because each are measured from different locations. This project element ensures that these and other data products can be precisely aligned. As such, the ability to provide accurate positions and determine product accuracy benefits several projects including the Streamflow, Water Quality, and Sediment Transport and Budgeting (Project A), Sandbar and Sediment Storage Monitoring (Project B), Riparian Vegetation (Project C), Geomorphic Effects of Dam Operations (Project D), Geospatial Science and Technology (Project K), and Remote Sensing Overflight (Project L).

The control network is a set of monumented and documented reference marks at more than 1200 locations along the river corridor and on the rim together with the collection of observations that determine the relative and absolute positions of those points. These stations serve as the basis for referencing all ground- and air-based monitoring observations. Uncertainties in these positions are particularly critical for the sediment storage project, because small inaccuracies (e.g., centimeters) can lead to very large uncertainties in measured volumetric changes sand storage

when calculated over long reaches. For example, a vertical error of 5 cm distributed over a 30-mile sediment monitoring reach is approximately equivalent to a 250,000 m³ of sand (about 162,500 metric tons). This amount of sand can often be the difference between being able or unable to determine the sign of a sediment budget. Current network accuracy is ~5 cm at 95% confidence, and periodic re-occupations are required to maintain or improve network accuracy. Thus, one of the primary ways in which we can reduce uncertainty in estimates of sand storage change is to increase the accuracy of network positions through repeat observations and to add control where coverage is limited. Currently, nearly all of the river corridor from GCD to Diamond Creek has a sufficient number of control points to support monitoring activities, while control point coverage between Diamond Creek and Pearce Ferry is sparse.

Recently, the United States implemented the North American Terrestrial Reference Frame of 2022 (NATRF2022), replacing the North American Datum of 1983 (NAD83) as the official Federal datum (National Geodetic Survey, 2019). This development requires changing both horizontal and vertical coordinate systems to align with the International Terrestrial Reference Frame. As part of this process, horizontal low-distortion projection systems for the Colorado River through Grand Canyon were developed and published in an open forum so all prospective users will have access. These grid coordinate systems were designed specifically for the region and will allow for more efficient data collection, processing, and analysis of areas and volumes. The new vertical reference system will implement Gravity for the Redefinition of the American Vertical Datum (GRAV-D) data to better realize local height systems, improve geopotential determination of the river system, and monitor geographically dependent changes to the Lake Powell region's gravity over time. Better gravity field models will lead to better understanding of the CRE and more accurate boundary conditions for streamflow modeling. Conversion tools will be implemented to accurately detect geomorphological changes from data referencing the old and new datums (NAD83 and NATRF2022) and new coordinate systems (SPSC1983 and SPSC2022).

Hypotheses and Science Questions

Because project B.3 is a support project, the work is based on goals and objectives rather than hypotheses and science questions. The objectives of the Control Network and Survey Support project element are to maintain the geodetic control network and provide support for geospatial data collection to other projects.

Methods

The Grand Canyon control network was developed and has been managed by a full time Geodesist for the past 26 years, however increased costs and a limited budget has led to that position going unfilled. As an alternative, partial funding (~50% or more) would enable GCMRC to establish a contract or cooperative agreement with a Geodesist.

For this scenario, the objectives for this three-year work plan would be to complete a report summarizing the status of the control network, complete a data release that documents all control points in the current network and their uncertainties (which would inform the LTEMP ten-year review), and develop a plan for migrating to use of the new spatial reference frames (NATRF2022) and maintaining the control network in the future.

Anticipated Use of Data

The control network data are used to establish accurate positions for all geo-spatial measurements that are collected across all projects and to define the uncertainty in those observations.

Outcomes and Products

- Report summarizing the status of the control network.
- Data release that documents all control points in the current network and their uncertainties.
- Plan for migrating to use of the new spatial reference frames (NATRF2022) and maintaining the control network in the future.

Project Element B.4. Streamflow, Sediment, and Sandbar Modeling (Ongoing Study; Unfunded)

The purposes of this project element are to 1) maintain and continue development of the existing sand routing model and sandbar response models, 2) begin development of a new streamflow model for the Colorado River between RM 30 and RM 61, and 3) continue development of a new fine-sediment (silt and clay) routing model. These models will be valuable in identifying and analyzing optimal management actions for sediment resources as part of the LTEMP ten-year review.

The Sand Routing Model (Wright and others, 2010) is used as part of the LTEMP planning process to determine the appropriate HFE duration based on predicted sand mass balance. This modeling has in the past been performed by Reclamation, but since Fall 2022 has been done by GCMRC. Because this is a data-driven empirical model, regular model verification and recalibration are necessary and are provided through this project element. The Sandbar Model (Mueller and Grams, 2021) is a semi-empirical model used to predict sandbar volume through time in response to dam operations and can be used to evaluate the potential outcomes of management actions such as HFEs, as well as proposed changes in dam operations, including hourly, daily, monthly, and annual releases.

We anticipate that the Sandbar Model will be particularly useful for decision makers in GCDAMP if proposed changes to the sediment accounting window under the LTEMP SEIS are implemented, as it can provide quantitative predictions of sandbar volume change associated with fall HFE implementation vs. deferring to spring.

Improvements to the Sandbar Model are needed to realistically predict sandbar responses to different proposed dam operations. In the current version of the Sandbar Model, the deposition rate is based on Andrews and Vincent (2007), and accounts for the impacts of variations in stage, discharge, and sediment concentration on the deposition rate. However, the erosion model is highly simplified, and assumes that the erosion rate is proportional to sandbar volume and is independent of discharge. Prior sandbar surveys have shown that high monthly releases accelerate sandbar erosion (Hazel and others, 2022), and Alvarez and Schmeckle (2012) found that discharge fluctuations accelerate sandbar erosion through seepage erosion. Neither of these effects are included in the current version of the Sandbar Model, making it more difficult to interpret predictions of the model for future flow scenarios. Additionally, the Sandbar Model is currently calibrated based on the group 1a bar type (Mueller and others, 2018), a small subset consisting of nine out of forty-five of the sandbars that are included in the long-term monitoring (Project B.1). In particular, the current Sandbar Model does not include vegetation, and therefore is unsuitable for bar types with high vegetation cover.

New streamflow models are needed because existing models (Wiele and Griffin, 1998; Magirl and others, 2008; Mihalevich and others, 2020) are based on estimated “synthetic” channel geometry, and are limited to a single spatial dimension, which means that while they can reliably predict discharge, they cannot be used to predict water depths, streamflow velocity, or bed shear stress. Predictions of these quantities are necessary for spatially explicit predictions of sediment, nutrient or veliger transport, and quantification of physical habitat for fishes, riparian and in-stream vegetation, and invertebrates (including mussels). We propose to develop and calibrate a two-dimensional, hydraulic model for Marble Canyon and Eastern Grand Canyon, where extensive channel mapping data are available; in this TWP, the initial model construction will occur between RM 30 and RM 61. We will begin with this river segment because it has the best constraint on boundary conditions for both streamflow and sediment transport (U.S. Geological Survey, 2024a; https://www.gcmrc.gov/discharge_qw_sediment). This segment also contains two of the several known aggregations (Valdez and Ryel, 1995) of humpback chub (*Gila cypha*) and is therefore of interest for aquatic habitat. These aggregations are river segments where humpback chub are typically found in greater abundance and extend approximately from RM 30 to RM 36 and RM 57 to RM 77 (Persons and others, 2017). This new streamflow model will provide spatially explicit predictions of water depth, velocity, and shear stress at different discharges and is needed to quantify habitat characteristics for aquatic organisms and larval fish transport. This streamflow model will also provide the groundwork for future morphodynamic models for sediment transport and sandbar dynamics, and modeling interactions between streamflow and riparian vegetation (Project C.4).

The modeling domain from RM 30 to RM 61 complements the previously developed two-dimensional streamflow model for Glen Canyon (Wright and others, 2024).

The third main component of B.4 is the continued development of a silt and clay-routing model for predicting turbidity, which is important for predicting gross primary productivity, phosphorous loads, and fish population dynamics (including nonnative smallmouth bass). Preliminary results from project E show a strong and significant correlation between total phosphorous loads with silt/clay concentrations in the Colorado River mainstem and tributaries. Additionally, turbidity can reduce the vulnerability of juvenile native fishes to predation by nonnative fishes such as brown trout (*Salmo trutta*) and smallmouth bass (Ward and others, 2016; Ward and Vaage, 2019).

Hypotheses and Science Questions

- How will any proposed changes in future dam operations scenarios, including the frequency of HFEs, HFE timing, and changes in monthly or annual release volumes, affect sandbars and sand storage?
- How do current and proposed future dam operations affect turbidity, and can dam operations be modified to change turbidity levels sufficiently to benefit ecological resources?
- How do streamflow patterns associated with dam operations affect the distribution of aquatic habitat and larval fish transport?
- How does the interplay between patterns of streamflow, riparian vegetation, and sediment transport affect sandbar morphology and evolution?

Methods

We will run the Sand Routing and Sandbar models in cooperation with Reclamation as needed for HFE planning and implementation. Each year, model predictions will be validated against sand mass balance predictions generated through Project A.3, and if necessary, we will perform recalibration of the model. We will also make the following modifications and improvements to the Sandbar Model: 1) improve erosion relations used in the sandbar model to better predict the effects of dam operations on sandbar volumes, 2) collaborate with project C to incorporate vegetation effects (that is, increased deposition, and decreased erosion) on sandbars, and 3) calibrate multiple versions of the model to capture the potentially disparate effects of operations on different bar types.

We propose to develop the new RM 30 – RM 61 streamflow model within the International River Interface Cooperative (iRIC) software using either the Flow and Sediment Transport with Morphologic Evolution of Channels (FaSTMECH) solver or the Nays2DH solver which has morphodynamic modeling capabilities.

This is the same modeling platform that was used for the existing Glen Canyon model, and like that model, roughness parameters within the model will either consist of Manning's n , or the z_0 roughness coefficient which is based on the bed-sediment grain size and is not sensitive to fluctuations in discharge like Manning's n . The model will use bathymetry and near-channel topographic data collected during previous channel mapping trips, along with topography above the 8,000 ft^3/s stage collected during the 2021 aerial overflight. Because there are some unsurveyed river sections (i.e., rapids and shallow riffles), we will assess model sensitivity to different approaches for estimating bathymetry in those reaches. The model will be calibrated using surveyed water level profiles at 8,000 ft^3/s along with established discharge/stage relations at sandbar monitoring sites. The effects of vegetation in the Glen Canyon model were not specifically modeled and will be a key component within the new streamflow model between RM 30 and RM 61, such that the effects of vegetation on shoreline flow velocities and shear stress can be determined. These vegetation effects will help guide vegetation management actions within Marble and Grand Canyons. The streamflow/vegetation interactions will be analyzed within Project C.4 (currently unfunded).

We will continue to work on a new fine sediment (silt and clay) routing model that was started in the FY 2021-23 work plan (that work was suspended to provide modeling support for Reclamation compliance efforts). Turbidity in the Colorado River is dominantly controlled by silt and clay, and to a lesser extent fine sand. The preliminary work highlighted the importance of eddies in controlling silt/clay routing following tributary floods, and predicted deposition of a small fraction of the supplied silt/clay within the bed and eddy sandbars, which was gradually released as a function of subsequent dam operations. In this work plan, we will 1) refine calibration parameters to better predict silt, clay, and fine sand routing, 2) develop empirical relationships to translate model predictions of sediment concentration to turbidity, and 3) develop refinements to improve model runtime so that it can be used in future scenario-testing, similar to how the sand routing model was used in the LTEMP SEIS. We will work with Projects E and I to incorporate our results into predictive models for gross primary productivity, phosphorous loads, and fish population dynamics.

Anticipated Use of Data

The sand routing and sandbar models are used each year for planning potential management actions such as HFEs and will also be used to compute the LTEMP performance metrics for sand mass balance and sandbars in a predictive framework to evaluate the potential outcomes of proposed changes in dam operations, including hourly, daily, monthly, and annual releases.

The new streamflow model for the reach between RM 0 and RM 87 will be used to provide the necessary boundary conditions required to run and validate morphodynamic sandbar models which are required to better understand the feedbacks between vegetation encroachment and sandbar dynamics, and to provide flow depth and velocity relations for habitat characterization.

A similar model was developed for the reach between Glen Canyon Dam to Lees Ferry (Wright and others, 2024) that has proved useful for a range of ecosystem studies, such as quantifying available smallmouth bass nesting habitat at different discharges.

The turbidity modeling component of this project element will be valuable for predicting gross primary productivity (Deemer and others, 2022; Project E.2), phosphorous loads (Project E.1), and interpreting fish population dynamics (including smallmouth bass, see Project I.4) (Hansen and others, 2023).

Outcomes and Products

- Communication of Sand Routing Model results to BOR and GCDAMP stakeholders, and/or modeling support for HFE planning provided to BOR. This information would be provided via PowerPoint slides and verbal communication during GCDAMP meetings, and through email communication.
- Communication of Sandbar Model results to GCDAMP stakeholders to help evaluate proposed flow actions, including selecting between fall and spring HFE implementation under a one-year sediment accounting window. This information would be provided via PowerPoint slides and verbal communication during GCDAMP meetings, and through email communication.
- Improvements to the Sandbar model: 1) Improvements to the erosion model: report, journal publication, and/or code release anticipated in FY 2026. 2) Incorporating vegetation effects and multiple bar types: presentation of preliminary results in FY 2027.
- New two-dimensional streamflow model for RM 30 to RM 61. Report, journal publication, and/or code release anticipated in FY 2026.
- Fine Sediment Model (silt/clay and turbidity). Journal publication and code release anticipated in FY 2027.

Project Element B.5. Sandbar and Riverbed Response to Experimental Actions (Experimental Fund)

The LTEMP Environmental Impact Statement and Record of Decision (U.S. Department of the Interior, 2016) included two experimental activities in addition to regular fall and spring sediment-triggered HFEs designed to improve sandbar and sediment resources (extended-duration HFEs, proactive HFEs, and variation in HFE downramp rate) and one experimental activity that may cause increased sandbar erosion (trout management flows). Additionally, although not described specifically in the LTEMP ROD, experiments that involve adjusting the downramp rate of an HFE while maintaining a hydrograph within the parameters of the LTEMP ROD may be conducted.

GCDAMP stakeholders have also proposed evaluation of the effect of HFEs on bed-sediment dynamics in the western Grand Canyon. The purpose of this project element is to collect and analyze field data that will be used to evaluate the effects of any of those flow experiments on sediment resources if those experiments occur. Because the timing of those experiments is condition-dependent, the field components would occur only when the experimental dam operations occur and evaluation of the effects of those flow-release experiments is required. The budgets for these project elements include only the additional costs associated with logistics for field data collection, processing, and analyses of those data.

Project Element B.5.1. Extended-duration HFEs

Hypotheses and Science Questions

- To what extent do extended-duration HFEs result in larger or more numerous sandbars than HFEs less than 96 hours long?

As defined in the LTEMP ROD, extended-duration HFEs are restricted to implementation in the fall sand accounting period and would be triggered according to the same criterion used for other sediment-triggered HFEs—that the sand mass balance for the fall sand accounting-period (July 1 – December 1) remain positive through HFE implementation based on model projections. The original HFE Protocol allows for HFE duration of up to 96 hours with a peak magnitude of 45,000 ft³/s. The extended duration HFEs may be 144, 192, or 250 hours in duration; however, the first test of an extended duration HFE is limited to 192 hours. Extending HFE duration is based on the hypothesis that, under conditions of enriched sand supply, longer duration HFEs will maintain elevated suspended sand concentrations for longer than the duration of a 96-hour HFE, resulting in more deposition and larger sandbars.

Methods

The key information needed to evaluate the effects of extended duration HFEs on sediment resources will be:

- 1) Measurements of suspended sand concentration during each entire HFE,
- 2) Measurements of sandbar size before and after the extended duration HFE, and
- 3) Daily observations of sandbar dynamics during the HFE.

Because the extended duration HFEs are limited to the fall accounting period, data collected in the fall sandbar-monitoring trip, which occurs annually in early October, will be used as the pre-HFE sandbar measurement, which saves logistical costs. One additional sandbar-monitoring trip will be required following the extended duration HFE. The focus of the pre- and post-HFE study will be on deposition above the 8,000 ft³/s stage. Therefore, the surveys will be for sandbar topography only and do not require bathymetry.

Additional information will be gained by conducting daily surveys during the extended duration HFE at two locations. Because the sandbars will be submerged, these surveys will require bathymetry. These surveys will allow for comparison between observed sandbar deposition rates and main-channel suspended sand concentrations. Finally, we will compare observed changes in sandbar volume to predictions based on site-specific sandbar modeling (Project Element B.4) to evaluate the predictive capability of the modeling approach.

Anticipated Use of Data

The measurements of suspended sand concentration will be used to determine if sand concentrations remain elevated throughout the extended HFE or if sand supply becomes depleted and concentrations decline, and these measurements are included in the regular Project A monitoring. The basis for evaluating the effects of extended duration HFEs on sandbar deposition will be by comparison with measurements of deposition for other HFEs. Images from the remote cameras will be used for a qualitative comparison at all sites. The pre- and post-HFE topographic surveys will be used for a quantitative comparison with measurements made before and after the 1996, 2004, and 2008 HFEs at the long-term sandbar monitoring sites (see element B.1).

Project Element B.5.2 and B.5.3. Proactive Spring HFEs

Proactive HFEs are defined in the LTEMP EIS and ROD as releases of up to 45,000 ft³/s and up to 24-hour duration that would occur in spring (April – June) in advance of scheduled equalization flows. The intended purpose of proactive HFEs is to create sand deposits above the expected stage of equalization flows, such that those deposits would not be subject to erosion during the equalization flows.

Hypotheses and Science Questions

- To what extent do proactive spring HFEs provide some mitigation of sandbar erosion in advance of high dam-release water volumes?

Methods

Evaluation of the effectiveness of the proactive HFEs to address the above science question requires surveys of sandbar topography immediately following the proactive HFE and following the equalization flows. Images from remote cameras already in place would be used to monitor the portions of sandbars exposed above water during the equalization flows (see Project Element B.1). The post-equalization flow survey would be accomplished on the annual sandbar-monitoring trip in early October. The post-HFE survey would require one additional survey trip. If river discharge is less than about 16,000 ft³/s during the survey, this could be accomplished with topography only (Experimental Project B.5.2).

If discharge is higher, bathymetric measurements would be required to enable surveying the entire sandbar above the 8,000 ft³/s stage (Experimental Project B.5.3). Surveying the sandbar down to the 8,000 ft³/s stage is required for the purposes of comparison with other surveys.

Anticipated Use of Data

These data will be used to compute sandbar deposition by proactive HFEs and to compute erosion of the deposited sandbars through and immediately following the period of summer equalization flows. This information will be used to evaluate the effectiveness of proactive HFEs.

Project Element B.5.4. Variation in HFE Downramp Rate

One of the challenges faced in implementation of the HFE Protocol is a lack of information for predicting sandbar response to HFEs of different magnitude, duration, or hydrograph shape. Although the LTEMP does not describe specific experiments designed to evaluate if/how variation in HFE magnitude or hydrograph shape (ramp rates) may affect sandbar response, experiments that involve adjusting the hydrograph within the parameters of the LTEMP ROD may be conducted.

HFE magnitude and duration are designed based on estimated mass of accumulated sediment in Marble Canyon and limited by facility and operating constraints. Daily measurements of sandbars during the 2008 HFE indicated that deposition occurred for the entire 60 hours of that event. Measurements during the 1996 HFE, which was not sediment-enriched, indicated that deposition rates decreased, and erosion increased after 3 days at peak discharge. Thus, for short-duration (< 96 hour), sediment-enriched HFEs, sandbar deposition is likely maximized by maximizing the time at peak discharge. To maximize the duration of flow at peak discharge, the HFEs have typically been implemented with the maximum allowed upramp and downramp rates. The maximum allowed upramp rate is 4000 ft³/s per hour and the maximum allowed downramp rate is currently 2500 ft³/s per hour (prior to the 2016 LTEMP ROD, maximum allowed downramp was 1500 ft³/s per hour; U.S. Department of Interior, 2011).

The purpose of experimenting with a lower downramp rate is to allow for sandbar reworking and additional sand deposition to occur as the flow decreases. The expectation is that gradual downramp results in sandbars that have a lower slope on the beach face. A sandbar with gradual slope would likely have less total sand volume than the bar with steeper slope but may have larger area above baseflow (~8,000 ft³/s) discharge. This was observed anecdotally during the 2012 HFE. It is further hypothesized that sandbars with a lower slope will erode at a slower rate and, therefore, persist longer following the HFE. Thus, the hypothesized benefit of decreasing the downramp rate is that the area of usable sandbar above baseflow persists longer, even if the sand volume immediately following the HFE is somewhat less. The purpose of this experimental project element will be to evaluate those hypotheses if a gradual downramp rate is tested.

Because HFE duration is determined based on the total modeled change in sand mass balance including both HFE upramp and downramp, decreasing the downramp rate may mean less time on peak discharge for a given duration. For HFEs where the duration is limited by the limited sand supply, decreasing the time on peak may not be desirable. Supply conditions that provide at least 60 hours on peak and a gradual downramp rate would provide the best test of implementing a gradual downramp rate. This could be achieved either with conditions that allow a 96-hour regular HFE or a longer extended-duration HFE as defined in the LTEMP EIS.

Hypotheses and Science Questions

- To what extent does decreasing the downramp rate of an HFE result in sandbars with lower beach face slopes and are those sandbars more persistent than sandbars deposited during an HFE with a steep downramp rate?

Methods

Addressing these questions will require at least three sets of measurements of sandbar topography at the Project B.1 monitoring sites. The measurements would be collected: 1) immediately following the HFE, 2) approximately 4 months following the HFE, and 3) approximately 10 months following the HFE. Assuming this experiment occurs during a fall HFE, collection of these data would require only two additional sandbar monitoring trips. The third set of data (10 months following the HFE) would be collected as part of the annual Project B.1 sandbar monitoring. The direct measurements of topography would be supplemented with analysis of images from the remote cameras.

Anticipated Use of Data

These surveys would be used to evaluate the slopes of sandbars created by the HFE and to measure post-HFE sandbar area and volume. These data would be used to compute erosion rates and evaluate whether the change in downramp rate resulted in bars with lower erosion rates.

Project Element B.5.5. Channel Response to Flow Pulse in Western Grand Canyon

The purpose of this experimental project element is to collect the field measurements required for the study of channel response to a flow pulse in western Grand Canyon. The Colorado River in western Grand Canyon has significantly different morphology than that of the debris-fan dominated system upstream. Here, the morphology is characterized by tall eroding banks composed of delta and lake deposits from periods of much higher reservoir elevations in Lake Mead, and a shallow sand-bedded channel characterized by migrating sandbars.

Studying how flow pulses affect bed-sediment dynamics in western Grand Canyon is of importance for Hualapai commercial river operations, as shifting sandbars pose navigation challenges in western Grand Canyon. The flow pulse could be a fall or spring HFE or a short-duration pulse of up to 25,000 ft³/s.

Although annual monitoring of this reach is an unfunded component of Project B.2, data collection before and after a HFE will support future efforts to develop a sediment budget for this reach based on measured sand loads from the upstream on the Colorado River at Diamond Creek, estimates of sediment input from channel banks based on boat-based lidar measurements, and bedload measurements inferred from bedform tracking using sonar measurements.

Hypotheses and Science Questions

- How does the river channel in western Grand Canyon change in response to HFEs and other dam operations?

Methods

The required data are repeat measurements of channel bathymetry and bank topography for a 1- to 3-km long study reach near Columbine Falls at roughly RM 274.

Up to five sets of measurements will be collected: 1) before the flow pulse, 2) once during the flow pulse, 3) immediately following the flow pulse, 4) approximately 1 month following the flow pulse, and 5) approximately 4 months following the flow pulse. Each survey will consist of measurements of the channel with multibeam sonar and measurements of the exposed banks with lidar and/or conventional total station. Similar measurements were performed during the 2021 Spring Disturbance Flood and the 2023 Spring HFE, providing robust datasets for comparison.

Anticipated Use of Data

These data will be used to compute changes in riverbed elevation, bank erosion rates and develop sediment budgets for the study reach. Understanding the causes and changes in the sediment budget of this reach will allow us to assess whether navigation challenges in western Grand Canyon are affected by flow pulses.

Project Element B.5.6. Trout Management Flows

Trout Management Flows (TMFs) are described in the LTEMP EIS and ROD as repeated cycles of flow fluctuations between high flows of approximately 20,000 ft³/s and low flows of 8,000 ft³/s or lower. The high-flow component would last between two and seven days with the low-flow component lasting for less than 24 hours. These flows are expected to cause increased rates of sandbar erosion.

The requirement for monitoring sandbar response would depend on the expected number of fluctuation cycles in a given TMF event. If a TMF event consists of only a few cycles, the increased amount of erosion compared to normal fluctuations would likely be small and difficult to measure. Under this scenario, observations from existing remote cameras will be used to determine if sandbar erosion rates are affected by these flows, particularly at the sites where georectification allows for daily to weekly calculation of area and, potentially, bar volume change. In contrast, if a TMF event consists of many fluctuation cycles, the expected additional erosion might require additional sandbar surveys to quantify sandbar change at all sites. Because we do not know the level of effort that will be required, we have not estimated a separate budget for this experiment.

Outcomes and Products for Experimental Project Elements

- Update on results of experimental action at annual reporting meeting.
- Data release for data collected for experimental project.
- Report or journal article describing field data and effects of experimental actions that occur.

Budgets

Project Element Budget

| Fiscal Year 2025 | | | | | | | | | |
|--|------------------|----------------------|-----------------------|-----------------------|---------------------------|--------------------------|-----------------------------|------------------|------------------------------|
| Project B Sandbar and Sediment Storage Monitoring and Research | Salaries | Travel & Training | Operating Expenses | Logistics Expenses | Cooperative Agreements | To other USGS Centers | AMP Special Burden Rate* | Total | USGS Contributing Funds** |
| | | | | | | | 21.86% | | Est Full Rate = 55% |
| B.1. Sandbar and campsite monitoring with topographic surveys and remote cameras | \$224,543 | \$2,000 | \$4,000 | \$27,162 | \$23,500 | \$0 | \$57,047 | \$338,252 | |
| B.2. Bathymetric and topographic mapping for monitoring sediment storage and riverbed dynamics | \$348,531 | \$2,000 | \$2,000 | \$0 | \$18,800 | \$0 | \$77,638 | \$448,969 | |
| Total Project B | \$573,074 | \$4,000 | \$6,000 | \$27,162 | \$42,300 | \$0 | \$134,685 | \$787,221 | \$63,727 |

| Fiscal Year 2026 | | | | | | | | | |
|--|------------------|----------------------|-----------------------|-----------------------|---------------------------|--------------------------|-----------------------------|--------------------|------------------------------|
| Project B Sandbar and Sediment Storage Monitoring and Research | Salaries | Travel & Training | Operating Expenses | Logistics Expenses | Cooperative Agreements | To other USGS Centers | AMP Special Burden Rate* | Total | USGS Contributing Funds** |
| | | | | | | | 22.60% | | Est Full Rate = 55.5% |
| B.1. Sandbar and campsite monitoring with topographic surveys and remote cameras | \$252,959 | \$2,000 | \$4,000 | \$31,921 | \$23,500 | \$0 | \$66,444 | \$380,824 | |
| B.2. Bathymetric and topographic mapping for monitoring sediment storage and riverbed dynamics | \$399,706 | \$2,000 | \$5,000 | \$86,680 | \$37,600 | \$0 | \$112,633 | \$643,619 | |
| Total Project B | \$652,664 | \$4,000 | \$9,000 | \$118,601 | \$61,100 | \$0 | \$179,077 | \$1,024,443 | \$83,107 |

| Fiscal Year 2027 | | | | | | | | | |
|--|------------------|----------------------|-----------------------|-----------------------|---------------------------|--------------------------|-----------------------------|------------------|------------------------------|
| Project B Sandbar and Sediment Storage Monitoring and Research | Salaries | Travel & Training | Operating Expenses | Logistics Expenses | Cooperative Agreements | To other USGS Centers | AMP Special Burden Rate* | Total | USGS Contributing Funds** |
| | | | | | | | 23.40% | | Est Full Rate = 56% |
| B.1. Sandbar and campsite monitoring with topographic surveys and remote cameras | \$260,273 | \$2,000 | \$4,000 | \$32,895 | \$23,500 | \$0 | \$70,710 | \$393,378 | |
| B.2. Bathymetric and topographic mapping for monitoring sediment storage and riverbed dynamics | \$431,301 | \$2,000 | \$2,000 | \$0 | \$18,800 | \$0 | \$102,424 | \$556,526 | |
| Total Project B | \$691,574 | \$4,000 | \$6,000 | \$32,895 | \$42,300 | \$0 | \$173,135 | \$949,904 | \$79,057 |

*GCDAMP Special Rate: Rate includes current DOI preferred rate (currently 15%; subject to change) and facilities rate (will vary annually). No USGS bureau overhead is charged (unique to the GCDAMP agreement).

**USGS Contributing Funds: The amount of funds required to cover the subsidy created by the reduced burden rate (i.e., GCDAMP Special Rate). As in previous years, SBSC/GCMRC will request these funds of the USGS cost-share program. These funds are not guaranteed.

Experimental Fund

| Fiscal Year 2025 | | | | | | | | | |
|---|------------------|----------------------|-----------------------|-----------------------|---------------------------|--------------------------|-----------------------------|------------------|--|
| Project B Experimental Fund Projects | Salaries | Travel & Training | Operating Expenses | Logistics Expenses | Cooperative Agreements | To other USGS Centers | AMP Special Burden Rate* | Total | |
| | | | | | | | 21.86% | | |
| B.5.1. Extended duration Fall HFE (daily surveys during HFE + 1 set sandbar surveys w/o bathymetry) | \$39,793 | \$3,000 | \$1,000 | \$96,787 | \$69,325 | \$0 | \$32,815 | \$242,719 | |
| B.5.2. Proactive Spring HFE (1 set of sandbar surveys w/o bathymetry) | \$36,569 | \$1,500 | \$1,000 | \$55,144 | \$36,425 | \$0 | \$21,691 | \$152,329 | |
| B.5.3. Proactive Spring HFE (1 set of sandbar surveys with bathymetry) | \$36,569 | \$1,500 | \$1,000 | \$61,858 | \$69,325 | \$0 | \$24,145 | \$194,398 | |
| B.5.4. Variation in HFE downramp rate (2 sets of sandbar surveys w/o bathymetry) | \$43,016 | \$3,000 | \$1,000 | \$117,611 | \$69,325 | \$0 | \$38,072 | \$272,025 | |
| B.5.5. Channel response to HFE in western Grand Canyon (4 surveys around fall HFE) | \$33,346 | \$1,000 | \$1,000 | \$4,587 | \$22,325 | \$0 | \$9,400 | \$71,658 | |
| Total Project B | \$189,294 | \$10,000 | \$5,000 | \$335,986 | \$266,725 | \$0 | \$126,123 | \$933,128 | |

| Fiscal Year 2026 | | | | | | | | | |
|---|------------------|----------------------|-----------------------|-----------------------|---------------------------|--------------------------|-----------------------------|------------------|--|
| Project B Experimental Fund Projects | Salaries | Travel & Training | Operating Expenses | Logistics Expenses | Cooperative Agreements | To other USGS Centers | AMP Special Burden Rate* | Total | |
| | | | | | | | 22.60% | | |
| B.5.1. Extended duration Fall HFE (daily surveys during HFE + 1 set sandbar surveys w/o bathymetry) | \$41,783 | \$3,000 | \$1,000 | \$99,980 | \$69,325 | \$0 | \$35,022 | \$250,109 | |
| B.5.2. Proactive Spring HFE (1 set of sandbar surveys w/o bathymetry) | \$38,398 | \$1,500 | \$1,000 | \$56,966 | \$36,425 | \$0 | \$23,210 | \$157,499 | |
| B.5.3. Proactive Spring HFE (1 set of sandbar surveys with bathymetry) | \$38,398 | \$1,500 | \$1,000 | \$63,843 | \$69,325 | \$0 | \$25,751 | \$199,817 | |
| B.5.4. Variation in HFE downramp rate (2 sets of sandbar surveys w/o bathymetry) | \$45,167 | \$3,000 | \$1,000 | \$113,932 | \$69,325 | \$0 | \$38,940 | \$271,365 | |
| B.5.5. Channel response to HFE in western Grand Canyon (4 surveys around fall HFE) | \$35,013 | \$1,000 | \$1,000 | \$4,667 | \$22,325 | \$0 | \$10,089 | \$74,095 | |
| Total Project B | \$198,759 | \$10,000 | \$5,000 | \$339,388 | \$266,725 | \$0 | \$133,013 | \$952,885 | |

| Fiscal Year 2027 | | | | | | | | |
|---|------------------|----------------------|-----------------------|-----------------------|---------------------------|--------------------------|-----------------------------|------------------|
| Project B Experimental Fund Projects | Salaries | Travel & Training | Operating Expenses | Logistics Expenses | Cooperative Agreements | To other USGS Centers | AMP Special Burden Rate* | Total |
| | | | | | | | 23.40% | |
| B.5.1. Extended duration Fall HFE (daily surveys during HFE + 1 set sandbar surveys w/o bathymetry) | \$43,872 | \$3,000 | \$1,000 | \$103,201 | \$69,325 | \$0 | \$37,431 | \$257,829 |
| B.5.2. Proactive Spring HFE (1 set of sandbar surveys w/o bathymetry) | \$40,318 | \$1,500 | \$1,000 | \$58,806 | \$36,425 | \$0 | \$24,873 | \$162,921 |
| B.5.3. Proactive Spring HFE (1 set of sandbar surveys with bathymetry) | \$40,318 | \$1,500 | \$1,000 | \$65,259 | \$69,325 | \$0 | \$27,370 | \$204,772 |
| B.5.4. Variation in HFE downramp rate (2 sets of sandbar surveys w/o bathymetry) | \$47,426 | \$3,000 | \$1,000 | \$117,611 | \$69,325 | \$0 | \$41,634 | \$279,996 |
| B.5.5. Channel response to HFE in western Grand Canyon (4 surveys around fall HFE) | \$36,764 | \$1,000 | \$1,000 | \$4,749 | \$22,325 | \$0 | \$10,852 | \$76,689 |
| Total Project B | \$208,697 | \$10,000 | \$5,000 | \$349,626 | \$266,725 | \$0 | \$142,159 | \$982,206 |

Unfunded Project Element Budget

| Fiscal Year 2025 | | | | | | | | |
|---|------------------|----------------------|-----------------------|-----------------------|---------------------------|--------------------------|-----------------------------|------------------|
| Project B Unfunded Project Elements | Salaries | Travel & Training | Operating Expenses | Logistics Expenses | Cooperative Agreements | To other USGS Centers | AMP Special Burden Rate* | Total |
| | | | | | | | 21.86% | |
| B.3. Control network and survey support | \$129,644 | \$1,500 | \$15,000 | \$0 | \$0 | \$0 | \$31,860 | \$178,004 |
| B.4. Streamflow modeling | \$110,606 | \$1,500 | \$0 | \$0 | \$0 | \$0 | \$24,439 | \$136,545 |
| Total Project B | \$240,250 | \$3,000 | \$15,000 | \$0 | \$0 | \$0 | \$56,299 | \$314,549 |

| Fiscal Year 2026 | | | | | | | | |
|---|------------------|----------------------|-----------------------|-----------------------|---------------------------|--------------------------|-----------------------------|------------------|
| Project B Unfunded Project Elements | Salaries | Travel & Training | Operating Expenses | Logistics Expenses | Cooperative Agreements | To other USGS Centers | AMP Special Burden Rate* | Total |
| | | | | | | | 22.60% | |
| B.3. Control network and survey support | \$137,988 | \$1,500 | \$15,000 | \$0 | \$0 | \$0 | \$34,915 | \$189,403 |
| B.4. Streamflow modeling | \$113,422 | \$1,500 | \$0 | \$0 | \$0 | \$0 | \$25,973 | \$140,895 |
| Total Project B | \$251,410 | \$3,000 | \$15,000 | \$0 | \$0 | \$0 | \$60,888 | \$330,298 |

| Fiscal Year 2027 | | | | | | | | |
|---|------------------|----------------------|-----------------------|-----------------------|---------------------------|--------------------------|-----------------------------|------------------|
| Project B Unfunded Project Elements | Salaries | Travel & Training | Operating Expenses | Logistics Expenses | Cooperative Agreements | To other USGS Centers | AMP Special Burden Rate* | Total |
| | | | | | | | 23.40% | |
| B.3. Control network and survey support | \$146,881 | \$1,500 | \$15,000 | \$0 | \$0 | \$0 | \$38,231 | \$201,612 |
| B.4. Streamflow modeling | \$143,703 | \$1,500 | \$0 | \$0 | \$0 | \$0 | \$33,977 | \$179,180 |
| Total Project B | \$290,584 | \$3,000 | \$15,000 | \$0 | \$0 | \$0 | \$72,208 | \$380,792 |

References Cited

- Alvarez, L.V., and Schmeeckle, M.W., 2012, Erosion of river sandbars by diurnal stage fluctuations in the Colorado River in the Marble and Grand Canyons—Full-scale laboratory experiments: River Research and Applications, v. 29, no. 2, p. 839-954, <https://doi.org/10.1002/rra.2576>.
- Andrews, E.D., and Vincent, K.R., 2007, Sand deposition in shoreline eddies along five wild and scenic rivers, Idaho: River Research and Applications, v. 23, no. 1, p. 7-20, <https://doi.org/10.1002/rra.960>.

- Buscombe, D., Grams, P.E., and Kaplinski, M.A., 2014a, Characterizing riverbed sediment using high-frequency acoustics—1. Spectral properties of scattering: *Journal of Geophysical Research: Earth Surface*, v. 119, no. 12, p. 2674-2691, <https://doi.org/10.1002/2014JF003189>.
- Buscombe, D., Grams, P.E., and Kaplinski, M.A., 2014b, Characterizing riverbed sediment using high-frequency acoustics—2. Scattering signatures of Colorado River bed sediment in Marble and Grand Canyons: *Journal of Geophysical Research: Earth Surface*, v. 119, no. 12, p. 2692-2710, <https://doi.org/10.1002/2014JF003191>.
- Buscombe, D., Grams, P.E., and Kaplinski, M.A., 2017, Compositional signatures in acoustic backscatter over vegetated and unvegetated mixed sand-gravel riverbeds: *Journal of Geophysical Research: Earth Surface*, v. 122, no. 10, p. 1771-1793, <https://doi.org/10.1002/2017JF004302>.
- Chapman, K.A., Best, R.J., Smith, M.E., Mueller, E.R., Grams, P.E., and Parnell, R.A., 2020, Estimating the contribution of tributary sand inputs to controlled flood deposits for sandbar restoration using elemental tracers, Colorado River, Grand Canyon National Park, Arizona: *GSA Bulletin*, v. 133, no. 5-6, p. 1141–1156, <https://doi.org/10.1130/B35642.1>.
- Deemer, B.R., Yackulic, C.B., Hall, R.O., Dodrill, M.J., Kennedy, T.A., Muehlbauer, J.D., Topping, D.J., Voichick, N., Yard, M.D., 2022, Experimental reductions in subdaily flow fluctuations increased gross primary productivity for 425 river kilometers downstream, *PNAS Nexus*, v. 1, no. 3, pgac094, <https://doi.org/10.1093/pnasnexus/pgac094>.
- Grams, P.E., Buscombe, D., Topping, D.J., Kaplinski, M., and Hazel, J.E., 2019, How many measurements are required to construct an accurate sand budget in a large river? Insights from analyses of signal and noise: *Earth Surface Processes and Landforms*, v. 44, no. 1, p. 160–178, <https://doi.org/10.1002/esp.4489>.
- Grams, P.E., Schmidt, J.C., and Topping, D.J., 2007, The rate and pattern of bed incision and bank adjustment on the Colorado River in Glen Canyon downstream from Glen Canyon Dam, 1956-2000: *Geological Society of America Bulletin*, v. 119, no. 5-6, p. 556-575, <https://doi.org/10.1130/B25969.1>.
- Grams, P.E., Schmidt, J.C., Wright, S.A., Topping, D.J., Melis, T.S., and Rubin, D.M., 2015, Building sandbars in the Grand Canyon: *EOS, Transactions of the American Geophysical Union*, v. 96, no. 11, p. 12-16, <https://doi.org/10.1029/2015EO030349>.
- Grams, P.E., Topping, D.J., Schmidt, J.C., Hazel, J.E., Jr., and Kaplinski, M.A., 2013, Linking morphodynamic response with sediment mass balance on the Colorado River in Marble Canyon—Issues of scale, geomorphic setting, and sampling design: *Journal of Geophysical Research: Earth Surface*, v. 118, no. 2, p. 361-381, <https://doi.org/10.1002/jgrf.20050>.
- Grams, P.E., Tusso, R.B., and Buscombe, D., 2018, Automated remote cameras for monitoring alluvial sandbars on the Colorado River in Grand Canyon, Arizona: *U.S. Geological Survey Open-File Report 2018-1019*, 50 p., <https://doi.org/10.3133/ofr20181019>.
- Hadley, D.R., Grams, P.E., and Kaplinski, M.A., 2018, Quantifying geomorphic and vegetation change at sandbar campsites in response to flow regulation and controlled floods, *Grand*

- Canyon National Park, Arizona: River Research and Applications, v. 34, no. 9, p. 1208-1218, <https://doi.org/10.1002/rra.3349>.
- Hansen, L.E., Yackulic, C.B., Dickson, B.G., Deemer, B.R., and Best, R.J., 2023, Linking ecosystem processes to consumer growth rates—Gross primary productivity as a driver of freshwater fish somatic growth in a resource-limited river: *Canadian Journal of Fisheries and Aquatic Sciences*, v. 80, no. 9, p. 1456-1469, <https://doi.org/10.1139/cjfas-2022-0229>.
- Hazel, J.E., Jr., Grams, P.E., Schmidt, J.C., and Kaplinski, M., 2010, Sandbar response following the 2008 high-flow experiment on the Colorado River in Marble and Grand Canyons: U.S. Geological Survey Scientific Investigations Report 2010-5015, 52 p., <https://pubs.usgs.gov/sir/2010/5015/>.
- Hazel, J.E., Kaplinski, M.A., Hamill, D., Buscombe, D., Mueller, E.R., Ross, R.P., Kohl, K., and Grams, P.E., 2022, Multi-decadal sandbar response to flow management downstream from a large dam—The Glen Canyon Dam on the Colorado River in Marble and Grand Canyons, Arizona: U.S. Geological Survey Professional Paper 1873, prepared in cooperation with Northern Arizona University, 104 p., <https://doi.org/10.3133/pp1873>.
- Hazel Jr., J.E., Kaplinski, M., Parnell, R.A., Kohl, K., and Schmidt, J.C., 2008, Monitoring fine-grained sediment in the Colorado River ecosystem, Arizona—Control network and conventional survey techniques: U.S. Geological Survey Open-File Report 2008-1276, 15 p., <https://pubs.usgs.gov/of/2008/1276>.
- Hazel, J.E., Topping, D.J., Schmidt, J.C., and Kaplinski, M., 2006, Influence of a dam on fine-sediment storage in a canyon river: *Journal of Geophysical Research*, v. 111, no. F1, p. 1-16, <https://doi.org/10.1029/2004JF000193>.
- Kaplinski, M., Grams, P.E., Hazel, J.E., Jr., Buscombe, D., Kohl, K., and Hensleigh, J., 2020, Channel mapping of the Colorado River in Grand Canyon National Park, Arizona—May 2012, river miles 29 to 62—Data: U.S. Geological Survey data release, <https://doi.org/10.5066/P9CIMU68>.
- Kaplinski, M.A., Hazel, J.E., Grams, P.E., and Davis, P.A., 2014b, Monitoring fine-sediment volume in the Colorado River ecosystem, Arizona—Construction and analysis of digital elevation models: U.S. Geological Survey Open-File Report 2014-1052, 29 p., <https://doi.org/10.3133/ofr20141052>.
- Kaplinski, M., Hazel, J.E. Jr, Grams, P.E., Gushue, T., Buscombe, D.D., and Kohl, K., 2022a, Channel mapping Glen Canyon Dam to Lees Ferry in Glen Canyon National Recreation Area, Arizona—Data: U.S. Geological Survey data release, <https://doi.org/10.5066/P98GFP93>.
- Kaplinski, M., Hazel, J.E., Jr., Grams, P.E., Gushue, T.M., Buscombe, D., and Kohl, K., 2022b, Channel mapping of the Colorado River from Glen Canyon Dam to Lees Ferry in Glen Canyon National Recreation Area, Arizona: U.S. Geological Survey Open-File Report 2022-1057, 20 p., <https://doi.org/10.3133/ofr20221057>.
- Kaplinski, M., Hazel, J.E., Jr., Grams, P.E., Kohl, K., Buscombe, D.D., and Tusso, R.B., 2017a, Channel Mapping of the Colorado River in Grand Canyon National Park, Arizona – May

- 2009, river miles 29 to 62—Data: U.S. Geological Survey data release, <https://doi.org/10.5066/F7930RCG>.
- Kaplinski, M., Hazel, J.E., Jr., Grams, P.E., Kohl, K., Buscombe, D., and Tusso, R.B., 2017b, Channel mapping river miles 29–62 of the Colorado River in Grand Canyon National Park, Arizona, May 2009: U.S. Geological Survey Open-File Report 2017-1030, 35 p., <https://doi.org/10.3133/ofr20171030>.
- Kaplinski, M., Hazel, J.E., Jr., and Parnell, R., 2010, Colorado River campsite monitoring, 1998–2006, Grand Canyon National Park, Arizona, *in* Melis, T.S., Hamill, J.F., Bennett, G.E., Coggins, L.G., Jr., Grams, P.E., Kennedy, T.A., Kubly, D.M., and Ralston, B.E., eds., Proceedings of the Colorado River Basin Science and Resource Management Symposium, November 18-20, 2008, Scottsdale, Arizona: U.S. Geological Survey Scientific Investigations Report 2010-5135, p. 275-284, <https://pubs.usgs.gov/sir/2010/5135/>.
- Kaplinski, M.A., Hazel, J.E., Parnell, R., Hadley, D.R., and Grams, P.E., 2014a, Colorado River campsite monitoring, 1998-2012, Grand Canyon National Park, Arizona: U.S. Geological Survey Open-File Report 2014-1161, 24 p., <https://doi.org/10.3133/ofr20141161>.
- Kasprak, A., Sankey, J.B., Buscombe, D., Caster, J., East, A.E., and Grams, P.E., 2018, Quantifying and forecasting changes in the areal extent of river valley sediment in response to altered hydrology and land cover: Progress in Physical Geography: Earth and Environment, v. 42, no. 6, p. 739-764, <https://doi.org/10.1177/0309133318795846>.
- Kennedy, T.A., Yackulic, C.B., Cross, W.F., Grams, P.E., Yard, M.D., and Copp, A.J., 2014, The relation between invertebrate drift and two primary controls, discharge and benthic densities, in a large regulated river: Freshwater Biology, v. 59, no. 3, p. 557-572, <https://doi.org/10.1111/fwb.12285>.
- Magirl, C.S., Breedlove, M.J., Webb, R.H., and Griffiths, P.G., 2008, Modeling water-surface elevations and virtual shorelines for the Colorado River in Grand Canyon, Arizona: U.S. Geological Survey Scientific Investigations Report 2008-5075, 32 p., <https://pubs.usgs.gov/sir/2008/5075>.
- Magirl, C.S., Webb, R.H., and Griffiths, P.G., 2005, Changes in the water surface profile of the Colorado River in Grand Canyon, Arizona, between 1923 and 2000: Water Resources Research, v. 41, no. 5, p. 1-10, <https://doi.org/10.1029/2003WR002519>.
- Mihalevich, B.A., Neilson, B.T., Buahin, C.A., Yackulic, C.B., and Schmidt, J.C., 2020, Water temperature controls for regulated canyon-bound rivers: Water Resources Research, v. 56, no. 12, e2020WR027566, <https://doi.org/10.1029/2020WR027566>.
- Mueller, E.R., and Grams, P.E., 2021, A morphodynamic model to evaluate long-term sandbar rebuilding using controlled floods in the Grand Canyon: Geophysical Research Letters, v. 48, no. 9, e2021GL093007, <https://doi.org/10.1029/2021GL093007>.
- Mueller, E.R., Grams, P.E., Hazel, J.E., Jr., and Schmidt, J.C., 2018, Variability in eddy sandbar dynamics during two decades of controlled flooding of the Colorado River in the Grand Canyon: Sedimentary Geology, v. 363, p. 181-199, <https://doi.org/10.1016/j.sedgeo.2017.11.007>.

- Mueller, E.R., Grams, P.E., Schmidt, J.C., Hazel, J.E., Jr., Alexander, J.S., and Kaplinski, M., 2014, The influence of controlled floods on fine sediment storage in debris fan-affected canyons of the Colorado River Basin: *Geomorphology*, v. 226, p. 65-75, <https://doi.org/10.1016/j.geomorph.2014.07.029>.
- National Geodetic Survey, 2019, National Geodetic Survey Strategic Plan—2019-2023: Silver Spring, MD, National Oceanic and Atmospheric Administration and National Geodetic Survey, 50 p., https://geodesy.noaa.gov/web/about_ngs/info/documents/ngs-strategic-plan-2019-2023.pdf.
- Pemberton, E.L., 1976, Channel changes in the Colorado River below Glen Canyon Dam, *in* Third Federal Inter-Agency Sedimentation Conference, Denver, Colo., March 22-25, 1976: Sedimentation Committee, Water Resources Council, 5-61 to 5-73 p.
- Persons, W.R., Van Haverbeke, D.R., and Dodrill, M.J., 2017, Colorado River fish monitoring in Grand Canyon, Arizona—2002–14 humpback chub aggregations: U.S. Geological Survey Open-File Report 2016-1177, 43 p., <https://doi.org/10.3133/ofr20161177>.
- Randle, T.J., Lyons, J.K., Christensen, R.J., and Stephen, R.D., 2007, Colorado River ecosystem sediment augmentation appraisal engineering report: Denver, Colo., Bureau of Reclamation, Technical Service Center, Sedimentation and River Hydraulics Group, 71 p., <https://doi.org/10.3996/102015-JFWM-101.S7>.
- Sankey, J.B., Ralston, B.E., Grams, P.E., Schmidt, J.C., and Cagney, L.E., 2015, Riparian vegetation, Colorado River, and climate—Five decades of spatiotemporal dynamics in the Grand Canyon with river regulation: *Journal of Geophysical Research: Biogeosciences*, v. 120, no. 8, p. 1532-1547, <https://doi.org/10.1002/2015JG002991>.
- Schmidt, J.C., 1990, Recirculating flow and sedimentation in the Colorado River in Grand Canyon, Arizona: *Journal of Geology*, v. 98, no. 5, p. 709-724, <https://doi.org/10.1086/629435>.
- Schmidt, J.C., and Grams, P.E., 2011, The high flows—Physical science results, *in* Melis, T.S., ed., Effects of three high-flow experiments on the Colorado River ecosystem downstream from Glen Canyon Dam, Arizona: U.S. Geological Survey Circular 1366, p. 53-91, <https://pubs.usgs.gov/circ/1366/>.
- Topping, D.J., Grams, P.E., Griffiths, R.E., Dean, D.J., Wright, S.A., and Unema, J.A., 2021, Self-limitation of sand storage in a bedrock-canyon river arising from the interaction of flow and grain size: *Journal of Geophysical Research: Earth Surface*, v. 126, no. 5, e2020JF005565, <https://doi.org/10.1029/2020JF005565>.
- Topping, D.J., Rubin, D.M., and Vierra, L.E., Jr., 2000, Colorado River sediment transport—1. Natural sediment supply limitation and the influence of the Glen Canyon Dam: *Water Resources Research*, v. 36, no. 2, p. 515-542, <https://doi.org/10.1029/1999WR900285>.
- U.S. Department of the Interior, 2011, Environmental assessment—Development and implementation of a protocol for high-flow experimental releases from Glen Canyon Dam, Arizona, 2011 through 2020: Salt Lake City, Utah, U.S. Department of the Interior, Bureau

- of Reclamation, Upper Colorado Region, 176 p. plus appendices, <https://www.usbr.gov/uc/envdocs/ea/gc/HFEProtocol/HFE-EA.pdf>.
- U.S. Department of the Interior, 2016, Glen Canyon Dam Long-Term Experimental and Management Plan final Environmental Impact Statement (LTEMP EIS): U.S. Department of the Interior, Bureau of Reclamation, Upper Colorado Region, National Park Service, Intermountain Region, online, <http://ltempeis.anl.gov/documents/final-eis/>.
- U.S. Geological Survey, 2024a, Discharge, sediment, and water quality monitoring data: Flagstaff, Ariz., U.S. Geological Survey, Southwest Biological Science Center, Grand Canyon Monitoring and Research Center, accessed February 25, 2024, at http://www.gcmrc.gov/discharge_qw_sediment/.
- U.S. Geological Survey, 2024b, Sandbar monitoring data and remote camera images: Flagstaff, Ariz., U.S. Geological Survey, Southwest Biological Science Center, Grand Canyon Monitoring and Research Center, accessed February 25, 2024, at <https://www.usgs.gov/apps/sandbar/>.
- Valdez, R.A., and Ryel, R.J., 1995, Life history and ecology of the humpback chub (*Gila cypha*) in the Colorado River, Grand Canyon, Arizona—final report to Bureau of Reclamation: Salt Lake City, Utah, BIO/WEST, Inc., BIO/WEST report no. TR-250-08, contract no. 0-CS-40-09110, 328 p. plus appendices, <https://prism.lib.asu.edu/items/67282>.
- VanderKooi, S.P., Kennedy, T.A., Topping, D.J., Grams, P.E., Ward, D.L., Fairley, H.C., Bair, L.S., Sankey, J.B., Yackulic, C.B., and Schmidt, J.C., 2017, Scientific monitoring plan in support of the selected alternative of the Glen Canyon Dam Long-Term Experimental and Management Plan: U.S. Geological Survey, Grand Canyon Monitoring and Research Center, U.S. Geological Survey Open-File Report 2017-1006, 18 p., <https://doi.org/10.3133/ofr20171006>.
- Ward, D.L., Morton-Starner, R., and Vaage, B., 2016, Effects of turbidity on predation vulnerability of juvenile humpback chub to rainbow trout and brown trout: *Journal of Fish and Wildlife Management*, v. 7, no. 1, p. 205-212, <https://doi.org/10.3996/102015-JFWM-101>.
- Ward, D.L., and Vaage, B.M., 2019, What environmental conditions reduce predation vulnerability for juvenile Colorado River native fishes?: *Journal of Fish and Wildlife Management*, v. 10, no. 1, p. 196-205, <https://doi.org/10.3996/042018-JFWM-031>.
- Wiele, S.M., Andrews, E.D., and Griffin, E.R., 1999, The effect of sand concentration on depositional rate, magnitude, and location in the Colorado River below the Little Colorado River, in Webb, R.H., Schmidt, J.C., Marzolf, G.R., and Valdez, R.A., eds., *The controlled flood in Grand Canyon*: Washington, D.C., American Geophysical Union, Geophysical Monograph Series, v. 110, p. 131-145.
- Wiele, S.M., and Griffin, E.R., 1998, Modifications to a one-dimensional model of unsteady flow in the Colorado River through Grand Canyon, Arizona: U.S. Geological Survey Water-Resources Investigations Report 97-4046, 17 p., <https://doi.org/10.3133/wri974046>.

- Williams, G.P., and Wolman, M.G., 1984, Downstream effects of dams on alluvial rivers: U.S. Geological Survey Professional Paper 1286, 83 p., <https://doi.org/10.3133/pp1286>.
- Wright, S.A., Kaplinski, M., and Grams, P.E., 2024, Hydrodynamic model of the Colorado River, Glen Canyon Dam to Lees Ferry in Glen Canyon National Recreation Area, Arizona—Tables of model results and accuracy assessment: U.S. Geological Survey data release, <https://doi.org/10.5066/P1QTRNEB>.
- Wright, S.A., Melis, T.S., Topping, D.J., and Rubin, D.M., 2005, Influence of Glen Canyon Dam operations on downstream sand resources of the Colorado River in Grand Canyon, *in* Gloss, S.P., Lovich, J.E., and Melis, T.S., eds., The state of the Colorado River ecosystem in Grand Canyon—A report of the Grand Canyon Monitoring and Research Center 1991-2004: U.S. Geological Survey Circular 1282, p. 17-31, <https://pubs.usgs.gov/circ/1282/>.
- Wright, S.A., Topping, D.J., Rubin, D.M., and Melis, T.S., 2010, An approach for modeling sediment budgets in supply-limited rivers: *Water Resources Research*, v. 46, no. 10, p. 1-18, <https://doi.org/10.1029/2009WR008600>.

Project C: Riparian Vegetation Monitoring and Research

Investigators

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Project Summary and Purpose

This project focuses on riparian plant communities, particularly how dam operations can be used to control plant composition and cover and how plant communities impact other valuable, managed resources. The proposed elements in this project address Goal 11 of Glen Canyon Dam Long-Term Experimental and Management Plan (LTEMP) (U.S. Department of Interior, 2016), LTEMP, which is related to maintaining diverse native riparian plant communities and quality wildlife habitat. Project Elements C.1 and C.4 additionally link to Goals 1 (Archaeological and Cultural Resources), 6 (Recreational Experience) and 7 (Sediment) by evaluating the complex linkages among plants, river flows, sediment, and cultural resources values. Specifically, Project Element C.1 collects and summarizes annual monitoring data that identify if the LTEMP riparian vegetation goal is being met. A newly proposed aspect of this element aims to develop measures of wildlife habitat quality in collaboration with the Navajo Natural Heritage Program (NNHP). Project Element C.2 addresses the LTEMP riparian vegetation goal by experimentally evaluating plant physiological responses to steady vs. daily-fluctuating flows in the context of lower Lake Powell elevations. Project Element C.3 synthesizes data from Project Elements C.1 and C.2, as well as broad-scale regional datasets, to determine which dam-derived flows would be most likely to create plant communities that meet the characteristics desired in Goal 11. Project Element C.4 evaluates the impact of vegetation expansion on Colorado River channel change, including associating plant traits with sediment movement under different hydrological conditions, and the resulting implications for archaeological site preservation potential as well as other cultural resource values, thus linking together a suite of LTEMP goals (Goals 1 Archaeological and Cultural Resources, 2 Natural Processes, 6 Recreation, 7 Sediment, 11 Vegetation). Project Element C.5 provides support for experimental vegetation management actions being implemented by Glen Canyon National Recreation Area and Grand Canyon National Park. Project Element C.6 proposes to measure key physiological responses of plant species of interest during experimental flows. Project Element C.7 proposes to evaluate how plant species with different physical traits alter flow velocity and sediment deposition during experimental flows.

The purpose of this project is to understand how dam operations are shaping riparian plant communities and associated resources (like sand and cultural resources) and determine how dam operations can be modified to align plant communities more closely to conditions listed in the LTEMP Goal 11 desired conditions. The proposed projects are designed to jointly address these topics by 1) assessing how current dam operations are changing plant communities, 2) identifying how the possible loss of daily fluctuating flows will impact plant communities, 3) evaluating what kinds of dam operations promote desired communities, 4) determining if and how increased vegetation cover has changed river channel form, 5) providing research support to experimental management actions, 6) evaluating plant physiological response to experimental flow patterns, and 7) determining the impacts of common species on flow velocity and sediment deposition during flow experiments.

Long-Term Experimental and Management Plan (LTEMP) Resource Goals

- Goal 1. Archaeological and Cultural Resources.
- Goal 2. Natural Processes.
- Goal 6. Recreational Experience.
- Goal 7. Sediment.
- Goal 11. Riparian Vegetation.

Background

Riparian plant communities are an important resource in river ecosystems. Plants alter river geomorphology and sediment dynamics (Butterfield and others, 2020; Dean and Topping, 2024), support migratory and resident animals (Holmes and others, 2005b; Spence, 2006), provide traditional plant resources (Fairley, 2005; Jackson-Kelly and Hubbs, 2007), have positive and negative impacts on recreation (Stewart and others, 2003; Hadley and others, 2018), increase regional biodiversity (Sabo and others, 2005), and mediate resources (like sand) between rivers and uplands (Sankey and others, 2023). Riparian plant communities and the services they provide can vary greatly depending on the flow and climate characteristics that shape the plant species that can establish and grow. In the Colorado River ecosystem between Glen Canyon Dam and Lake Mead (CRe), plant communities are inextricably linked to physical processes (sediment erosion, deposition, transport) and dependent biological communities (for example, birds) (Holmes and others, 2005b; Spence, 2006; Butterfield and others, 2020; Hazel and others, 2022; Sankey and others, 2023). The traits and life histories of the individual plant species that make up riparian plant communities determine the nature of sediment/plant interactions and the services or inconveniences provided to wildlife and recreationists.

Glen Canyon Dam operations control plant community composition and cover along the Colorado River in Glen, Marble, and Grand Canyons (Ralston and others, 2014; Sankey and others, 2015; Palmquist and others, 2023). Many aspects of Glen Canyon Dam derived flow patterns are likely drivers of riparian plant presence and abundance (Figure 1), such as increased base flows, reduction of peak flows (Sankey and others, 2015), daily fluctuating flows (Bejarano and others, 2018a), macroinvertebrate production flows (Gorla and others, 2015), summer and winter high discharge (Butterfield and others, 2023), and high flow experiments (Palmquist and others, 2023; Ralston, 2010). Flow patterns are layered on top of floristic differences along the river, such that plant communities in Glen Canyon, Marble Canyon, eastern Grand Canyon, and western Grand Canyon respond to dam operations in unique ways (Palmquist and others, 2023). Operational and experimental flows in the CRe have been designed for meeting water delivery agreements, producing power, promoting or hindering fish populations, redistributing sediment inputs, and stimulating macroinvertebrate production (Melis and others, 2016). Each of these flow patterns impacts riparian plant community resources by favoring some species over others. As dam operations are adaptively managed, flow patterns could be designed to support plant community goals.

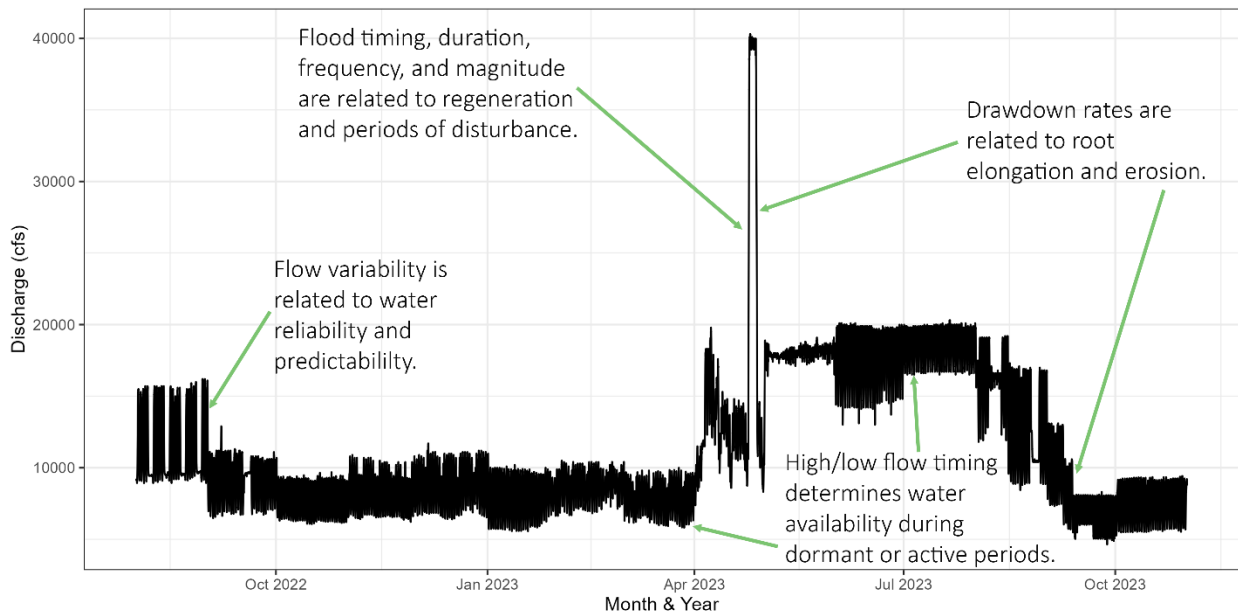


Figure 2. The instantaneous discharge record for water year 2023 (U.S. Geological Survey, 2024) illustrates the Glen Canyon Dam created flow patterns that are influencing riparian plant community cover and composition. Discharge is shown in cubic feet per second (cfs). Relevant aspects of the hydrograph are highlighted with green arrows and descriptive text.

Increases in riparian plant cover since Glen Canyon Dam operations began are very well documented (Turner and Karpiscak, 1980; Webb and others, 2011; Sankey and others, 2015). Impacts of operational adjustments on species composition have also been documented (Stevens and others, 1995; Stevens and Waring, 1986; Kearsley and Ayers, 1996; Ralston, 2010; Durning

and others, 2021; Butterfield and others, 2023; Palmquist and others, 2023). These studies indicate that tamarisk (*Tamarix ramosissima* and hybrids with *T. chinensis*) established quickly after beginning of Glen Canyon Dam operations, followed by *Salix exigua* (coyote willow).

During the period with high fluctuating flows (implementation of hydropower through 1996), riverine marshes formed and supported unique wetland communities (Stevens and others, 1995), which subsequently became drier and transitioned to woody species (Kearsley and Ayers, 1996). Recent plant expansion (< 10 years) has been characterized by several native and nonnative species including arrowweed (*Pluchea sericea*), Emory's baccharis (*Baccharis emoryi*), common reed (*Phragmites australis*), and horsetail (*Equisetum x ferrissii*) (Durning and others, 2021; Palmquist and others, 2023). The current plant communities appear to have been strongly shaped by both patterns of inundation (Butterfield and Palmquist, 2024b) and seasonality of high and low flow periods (Butterfield and others, 2023), such that changes to the frequency and magnitude of high flows and the timing of larger releases will change plant composition along the river. This information is largely based on field observations that are the result of the suite of flow patterns experienced within and across years.

The majority of previous studies providing this information are based on field observations and are limited in their ability to tease apart the impacts of different aspects of the hydrograph and in the ability to forecast outside the range of previously experienced flows (unprecedented conditions).

Riparian plant communities, and the specific species within those communities, are increasingly being recognized as a driving force in sediment transport and geomorphological controls in the CRe (Hadley and others, 2018; Butterfield and others, 2020; Durning and others, 2021; Hazel and others, 2022; Sankey and others, 2023). Vegetation, sediment, and cultural resources are intricately linked through complex interactions and feedback loops (Butterfield and others, 2020; Kasprak and others, 2021; Merritt, 2022; Sankey and others, 2023). Riparian vegetation and cultural resources are dependent on river channel sediment deposits, and the condition of sediment and cultural resources are dependent on riparian vegetation growth and composition. The dam-related changes in plant communities across the CRe are substantive and have altered deposition and erosion in eddy sandbars (Butterfield and others, 2020), reduced windblow sand to uplands (Sankey and others, 2023), and decreased bare sand and campable area (Hadley and others, 2018; Durning and others, 2021; Kasprak and others, 2021). These changes may be large enough to have impacted the width of the Colorado River channel and shoreline complexity, similar to what has been documented in other southwestern river systems (Dean and Schmidt, 2011; Dean and Topping, 2019, 2024).

Climate change and tamarisk defoliation by the introduced tamarisk beetle are changing how riparian plant communities are responding to dam operations. Ongoing aridification is leading to both alterations in Glen Canyon Dam operations and other environmental variables, like air temperature, that shape the riparian communities of the CRe (Overpeck and Udall, 2020; Wheeler and others, 2022).

The previous 60 years of dam operations have illustrated that the riparian plant communities of the CRe are dynamic and capable of rapid, significant change under altered hydrographs. Additionally, the arrival of the tamarisk beetle in 2009 and its subsequent expansion has led to extensive defoliation and death of one of the most abundant species (tamarisk) and primary tree species in the CRe (Bedford and others, 2018; Palmquist and others, 2023). This major change to the plant community will impact the prevalence of associated species. The future conditions of the CRe based on operational changes, climate change, and tamarisk beetle impacts will alter riparian plant communities and have cascading impacts on the CRe. The research presented in this project is aimed at tracking ongoing change, predicting future change, and understanding the impacts to other resources.

Proposed Work

Project Element C.1. Ground-based Riparian Vegetation Monitoring (Modified Study, Partially Funded)

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Monitoring the status and trends of native and nonnative plant species in the CRe provides the data for the LTEMP Goal 11 metrics: native species richness, ratio of native to nonnative species cover, and total native species cover. Annual measurement of plant species cover and composition will be characterized for multiple geomorphic features representative of the CRe and long-term monitoring sandbars and campsites surveyed as a part of Project B.

These data form the basis of plant community status and trends reporting and the underlying data for modeling efforts. Stratified-random sampling of multiple geomorphic features provides a thorough assessment of riparian plant composition, cover, richness, and native to nonnative species dominance on an annual basis throughout the CRe. Collecting data at long-term monitoring sandbars and campsites in conjunction with Project B provides a focused assessment of the impacts of plant communities on recreational resources and an opportunity for integration of vegetation and sediment dynamics.

Hypotheses and Science Questions

- What is the status (composition and cover) of native and nonnative vascular plant species within the riparian zone of the Colorado River from Glen Canyon Dam to the historic high-water line of Lake Mead, approximately 240 river miles downstream of Lees Ferry?

- Can bird habitat quality be reliably assessed by combining plant traits with estimates of plant species composition and cover? Can the data collected by the riparian plant monitoring program be used to assess patterns in riparian-dependent bird communities?

Research Question C.1.1. What is the Status (Composition and Cover) of Native and Nonnative Vascular Plant Species within the Riparian Zone of the Colorado River from Glen Canyon Dam to the Historic High-water Line of Lake Mead, approximately 240 River Miles Downstream of Lees Ferry?

The project aims to answer a suite of research questions that address the many ways dam operations and management actions influence plant communities and how plant communities, in turn, influence valuable resources in the CRe. This project element tracks the composition and cover of plant species growing along the river corridor and provides information on the distribution, frequency, and cover of many plant species.

Methods

The monitoring protocol is described in detail in Palmquist and others (2018b) and monitoring results through 2019 are described in Palmquist and others (2023). The random sampling effort collects data annually at 80-100 sites between Glen Canyon Dam and river mile 240, where the influence of Lake Mead becomes apparent on the shorelines. As part of the experimental design, new sites are selected each year. This data set provides a full representation of the plant communities affected by Glen Canyon Dam. The long-term monitoring sandbar sites are 45 large sandbars that are sampled each year. These sites are not representative of the riparian communities across the entire CRe (Palmquist and others, 2023), but are important recreational sites and the sites used for evaluating the state of sand resources (Hazel and others, 2022).

Data collection at the long-term monitoring sites is conducted in collaboration with Project B.1. The plant survey frames are included in the topographic survey, so that plant data can be closely linked to flow parameters and topographic change. Both survey efforts (random sampling and long-term monitoring sites) use ocular cover estimates of plant species rooted in 1-m² frames stratified by inundation frequency and geomorphic setting. Associated environmental data, such as height above minimum flows, ground cover, etc., are also collected. Previous research and monitoring illustrates that high and low flow patterns designed for other resources or for water delivery can have rapid impacts on plant communities that result in longer term patterns (Melis and others, 2011; Ralston, 2011). For example, the combination of the 2023 high flow experiment and high, consistent summer releases (Figure 1) reversed recent plant cover increases (Figure 2) (Palmquist and others, 2024), but this reversal could lead to a wide variety of plant community outcomes, including compositional shifts, rebounding plant growth, or sustained reduced cover over several years (Stevens and Waring, 1986; Stevens and Ayers, 1994; Melis and others, 2011; Ralston, 2011; Palmquist and others, 2023).

To link specific flow patterns to changes in plant communities, data is needed from before the high flow experiment, shortly after flow patterns return to normal, and at least annually for several years after. Flow experiments and other flow anomalies occur every few years in this system, meaning that at least annual sampling is needed over time. As sampling frequency is reduced, it becomes progressively less clear what aspects of the hydrograph are causing observed change. Thus, annual sampling of both randomly selected sites and the long-term monitoring sites is necessary to evaluate how flow experiments or operational anomalies (like equalization flows or smallmouth bass flows) affect plant communities and to evaluate long-term trends (Palmquist and others, 2023).

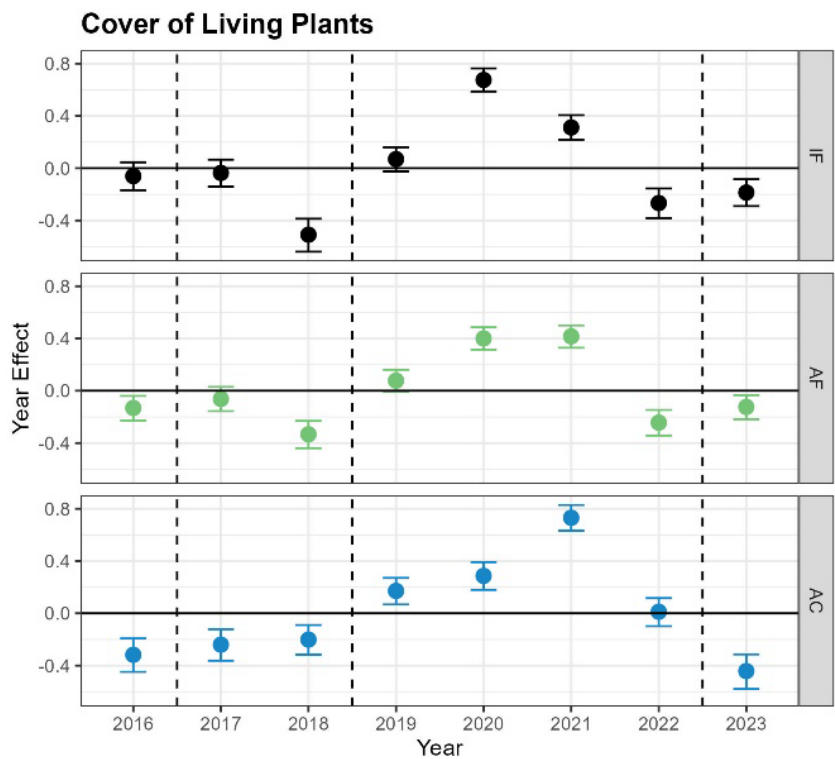


Figure 3. Mean and 95% credible intervals of year effects derived from modeling plant cover with a beta distribution in a Bayesian framework. Error bars that don't cross zero indicate significantly lower or higher values of the metric. Dotted lines indicate high flow experiments. Active channel (AC): area inundated by flows between 8,000 and 25,000 cubic feet per second (cfs). Active floodplain (AF): inundated by flows between 25,000 and 45,000 cfs. Inactive floodplain (IF): inundated by flows over 45,000 cfs. Figure from Palmquist and others (2024).

Anticipated Use of Data

These data are used to track the status of Goal 11 Riparian Vegetation, provide Status and Trends reports, and illustrate changes in riparian plant communities in response to dam operations. The Status and Trends reports and annual trend reporting can include summaries for species of interest and will be responsive to requests from stakeholders (for example, summarizing the occurrence and frequency of sacred datura, *Datura* spp. as requested by Zuni).

Stakeholders will be consulted about species they would like to see reported on in the next Status and Trends report. These data are also used in C.3 and other modeling efforts to evaluate how dam operations could be altered to change plant composition and cover.

Outcomes and Products

- Monitoring data collected annually FY 2025 through FY 2027 for both random sampling and long-term monitoring sandbars, properly archived, and used in other elements.
- Data summaries presented at GCMRC's annual reporting meetings. These can include summaries of species distributions and extents of species of interest along the river corridor, such as sacred datura (*Datura wrightii*), common reed (*Phragmites australis*), coyote willow (*Salix exigua*), arrowweed (*Pluchea sericea*), Emory's baccharis (*Baccharis emoryi*), etc.
- Status and trends presentation(s)/report describing long-term changes in plant communities, including species of interest (e.g., sacred datura, arrowweed).

Research Question C.1.2. Can Bird Habitat Quality be Reliably Assessed by Combining Plant Traits with Estimates of Plant Species Composition and Cover? Can the Data Collected by the Riparian Plant Monitoring Program be used to Assess Patterns in Riparian-dependent Bird Communities? (unfunded)

Maintaining wildlife habitat is explicitly stated in Goal 11, but evaluating wildlife habitat quality requires different vegetation assessments than are needed for determining the diversity and productivity of native plant species. A method for evaluating vegetation as wildlife habitat is needed if this part of the goal is to be assessed. Birds and bird habitat are of interest to Navajo stakeholders (Martin, 2009), is a primary consideration in other southwestern riparian areas (Grand and others, 2024), and has been assessed previously in the CRe (Holmes and others, 2005b).

Methods

As noted in the previous CRe efforts, evaluating vegetation as wildlife habitat and assessing bird use of that habitat requires considerable effort (Holmes and others, 2005b; Spence, 2006). Rather than attempting long-term monitoring using the intensive methods used previously, the work proposed here aims to evaluate if current ongoing monitoring of plant composition and cover can be leveraged to evaluate bird habitat.

Plant guilds based on traits have been used elsewhere to assess how changes to hydrology can impact bird habitat (Merritt and Bateman, 2012) and recent studies have examined links between plant guilds and habitat characteristics (Cubley and others, 2020).

It may be possible to combine the cover and composition data collected by ongoing monitoring efforts with plant trait data (plant height, specific leaf area, tissue density) (Lavorel and Garnier, 2002) to determine habitat suitability for different groups of birds. If these combined data sets (monitoring data and traits) are correlated with standard measures of bird habitat quality, all aspects of Goal 11 could be evaluated with no or very little extra monitoring.

In collaboration with the Navajo Nation Heritage Program (NNHP) Zoologist, the annual vegetation monitoring program will be leveraged to evaluate the quality of bird habitat currently supported by Glen Canyon Dam operations. This project will begin by evaluating methods previously used in the CRE and currently used along the lower Colorado River (Holmes and others, 2005a; Kearsley and others, 2006; Spence, 2006; Grand and others, 2024). Based on these standard methods of measuring bird habitat suitability, we will develop comparable methods for measuring habitat suitability at our randomly selected monitoring sites. In FY 2025 and FY 2026, we will collect habitat structure data along with cover and composition data. The monitoring data from FY 2025 and FY 2026 will be transformed into estimates of suitable bird habitat using the existing plant trait data curated by the GCMRC riparian plant program (Palmquist and others, 2017).

Combining the trait data with the cover and composition data will draw from other studies that use plant trait data in combination with habitat evaluation (Merritt and Bateman, 2012; Bateman and Merritt, 2020; Cubley and others, 2020) and concepts of effect traits used in the broader plant trait literature (Lavorel and Garnier, 2002; Kominoski and others, 2013). These estimates will then be compared to data generated by the traditional measures of habitat suitability. If the existing monitoring data provides similar estimates as the standard measures, bird habitat suitability can be assessed using only the current monitoring protocol and associated trait data (Palmquist and others, 2017; Palmquist and others, 2018b).

As noted above, this effort will be conducted in collaboration with the NNHP. In conjunction with the habitat assessments, NNHP will develop methods to characterize bird communities in the CRE and will be funded through Project 5.P of the FY 2025-27 Reclamation Triennial Work Plan. The habitat and bird monitoring efforts will be coordinated such that the data can be analyzed jointly. The goals and methods of these efforts will be tailored to the needs of stakeholders from traditionally affiliated Tribal Nations, particularly Navajo Nation. There is intentional flexibility built into the above-mentioned data collection and analysis so that the methods can be adjusted and tailored to be responsive to Tribal values.

Anticipated Use of Data

These data will be used to assess the second part of Goal 11 Riparian Vegetation. It will be compared to bird community data collected by NNHP to evaluate bird habitat quality and use. It will also be compared to data collected in C.1.1 to evaluate if ongoing monitoring efforts can be used to evaluate bird habitat quality.

Outcomes and Products

- Assessment of quality of bird habitat and bird use in FY 2025 and FY 2026.
- Data summaries presented at GCMRC's annual reporting meetings.
- Publication on the ability to use ongoing monitoring data and plant trait data to assess wildlife habitat suitability.

Project Element C.2. Mechanistic Experiments with Plant Species of Interest (Modified Study)

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Daily fluctuating flows strongly influence the composition and functioning of riparian plant communities (Bejarano and others, 2018a). These fluctuations have certainly shaped the development of riparian plant communities in the CRe over the past 60 years (Sankey and others, 2015). Moving forward, unprecedented changes to Glen Canyon Dam operations – for example, dropping below power pool – could mean a dramatic shift in flow controls on the riparian plant community and subsequent opportunities to enhance the diversity and functioning of this important ecosystem.

Despite some long-term trends in the CRe, we have little direct evidence for how daily fluctuations have altered the riparian plant community. Most research on the effects of daily fluctuations comes from mesic watersheds in Europe, where contrasts between paired regulated and unregulated rivers have demonstrated general losses in species richness and functional redundancy under regulation (Aguilar and others, 2018; Bejarano and others, 2018b). Less is known about arid rivers. Numerous hydrophilic species, particularly nonnative grasses (for example tall fescue, *Schedonorus arundinaceus*; Bermudagrass, *Cynodon dactylon*), are very abundant in parts of the CRe, and we hypothesize that this is due to pre-adaptations to fluctuating flows. We also hypothesize that facultative riparian species that grow further above the channel, and therefore are not inundated by daily peak flows, such as arrowweed (*Pluchea sericea*), also benefit by accessing soil moisture from those daily peaks (Gill and others, 2018; Palmquist and others, 2022; Butterfield and Palmquist, 2024b).

On the other hand, native riparian trees are largely absent from the CRe riparian zone, which we hypothesize is due to regeneration requirements associated with natural flow regimes and minimal erosion (Mahoney and Rood, 1998; González and others, 2018). While based on multiple lines of indirect evidence, these hypotheses cannot be thoroughly tested without controlled experiments.

We hypothesize that species vary significantly in their responses to hydropeaking, and that this variation is predictable based on plant functional traits related to seed and seedling ecology, clonality, and water use (Baladrón and others, 2023). We propose to test this hypothesis through a comprehensive assessment of plant responses to daily fluctuations across germination, establishment and growth, and including a functionally diverse suite of species.

We have conducted several experiments that inform the development of our daily fluctuation experimental design. First, from the perspective of experimental treatments, we conducted a pilot greenhouse experiment in the summer of 2023 to develop infrastructure for implementing daily fluctuations in a controlled environment. We focused on just three species of interest and simulated (1) daily fluctuation treatments – none (control), daytime peak, and nighttime peak; and (2) hydrological zone treatments – near channel habitat (plant crown inundated at peak) and floodplain habitat (bottom of roots inundated at peak). We quantified three different metrics of plant performance and demonstrated physiologically relevant responses to the experimental treatments. Notably, species differed in their responses to treatments, indicating some species-specificity that could explain their responses to flow conditions in the CRe (Figure 3, Palmquist and Butterfield, 2024).

Two other studies have bearing on this proposal. First, we published a greenhouse experiment focusing on tolerance to persistent drought or flooding, in which we demonstrated significant correlations between species hydrological preferences in the field and their growth and physiological responses in the greenhouse (Butterfield and Palmquist, 2024b). This study demonstrated the value of greenhouse experiments for predicting real-world flow responses and identified new root growth and stomatal conductance as valuable metrics of plant responses. Second, we published results from the spring disturbance flow that, while not representing daily fluctuations, did provide an opportunity to assess plant responses to short-term flow variation in the field (Butterfield and Palmquist, 2024b). We found significant and biologically strong responses of plant water status to these short-term fluctuations that differed between species with different habitat preferences. In particular, we demonstrated the value of plant water potential measurements as a unique dimension of plant responses to short-term fluctuations. These two experiments provide an important foundation for both experimental design and plant responses – including new root growth, stomatal conductance, and stem water potential – as critical indicators of plant responses to short-term fluctuating flows.

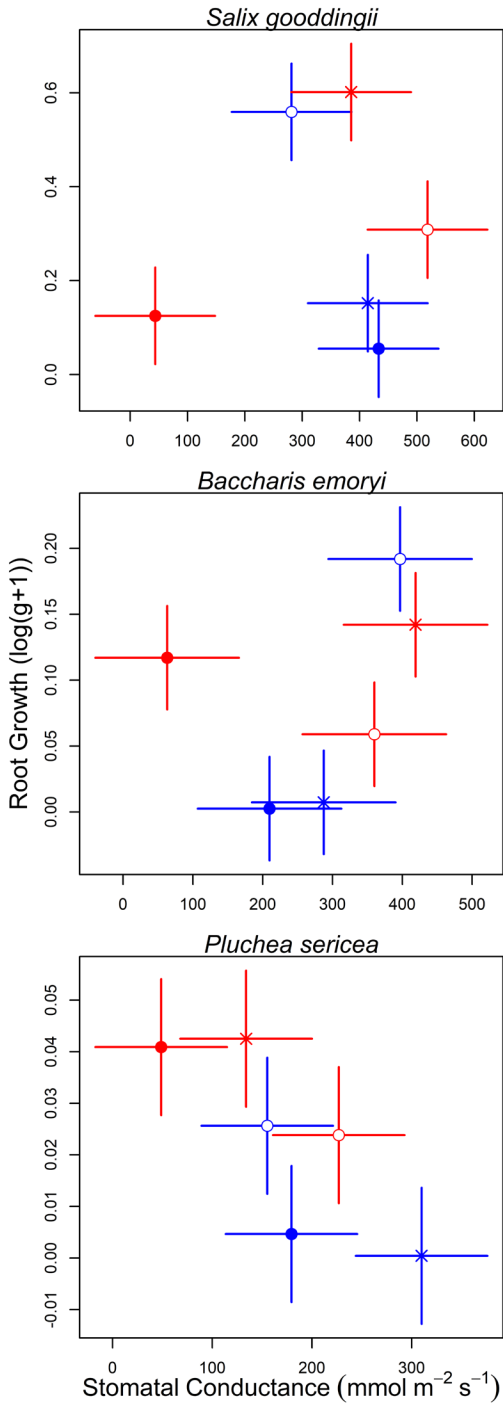


Figure 4. Results of pilot experiment testing the effects of daily fluctuations on performance three important CRe species; Gooding’s willow (*Salix gooddingii*), Emory’s baccharis (*Baccharis emoryi*), and arrowweed (*Plucheia sericea*). Blue symbols indicate plants where the crown was fully submerged during peak “flow”, red symbols indicate plants with just the bottom of the pot submerged during peak flow. These treatments simulate near channel and floodplain habitats, respectively. Points are mean values, lines are 95% confidence intervals. Treatments include controls (x; no fluctuations), nighttime peaks (closed circle) and daytime peaks (open circle). Figure from Palmquist and Butterfield (2024).

Hypotheses and Science Questions

- Which plant species benefit or suffer from daily fluctuations?

Research Question C.2.1. Which Plant Species Benefit or Suffer from Daily Fluctuations?

We intend to expand on the pilot greenhouse experiment in several ways. First, we will increase replication to improve statistical power to identify treatment effects. Second, we will increase the size of our experimental containers to simulate more realistic stage fluctuations. Third, we will assess multiple stages of plant demography, including germination, establishment, and growth. For the latter, we will utilize larger plants that allow us to make destructive measurements of leaf or stem water potentials without harming the plants, which was not feasible with the pilot experiment. In short, we propose to develop larger, more robust infrastructure for conducting daily fluctuation experiments that will be relevant to real-world conditions. We acknowledge that these proposed experiments do not capture effects of variations in velocity, shear stress, sediment transport, or directionality that would occur in a river during flow fluctuations. We hope to combine the results of the proposed, controlled experiments with other research (Butterfield and others, 2020; Project C.4 in this proposal) to gain a more comprehensive picture of the impacts of daily fluctuating flows on riparian ecosystems in the CRe.

Methods

We will establish infrastructure at Northern Arizona University (NAU) Research Greenhouse Complex that will facilitate multiple experiments related to daily fluctuations. The primary components will be,

- 1) Pairs of 1-m tall basins, each with a pond pump, connected by flexible tubing. This is a larger version of the setup we developed in 2023, in which one pump would turn on for a predetermined period of time at night, and the other 12 hours later during the day, in order to create symmetrical night/day peak flow conditions, analogous to the consistent timing of daily peaks and troughs at different locations throughout the CRe by Glen Canyon Dam operations.
- 2) A series of 1-m tall, 25 cm diameter PVC pipes, filled with sand and passively connected to the basins described above. As water levels move up and down in the basins, so too will the soil water levels in the pipes. The height of the containers will allow us to simulate more realistic stage changes than the smaller containers used in the pilot study.
- 3) Pedestals will be used to adjust the base elevations of individual containers (Palmquist and others, 2022), simulating plants growing in near channel versus floodplain habitats.

Following the review of Bejarano and others (2018a), we will test plant responses to fluctuations at multiple stages: reproduction, seedling establishment, and mature plant growth and physiology,

- 1) **Reproduction. (Sexual)** Seeds will be sown on the surface of containers. In treatments where the surface is not inundated by the daily peak, an initial watering event will be implemented to simulate rainfall or a high flow event (HFE). Germination timing and counts will be quantified. **(Asexual)** Plants will be propagated from seeds or cuttings in spring of 2025 in 30 cm tall, 8 cm diameter containers. In summer of 2026, the bottoms of the containers that plants are growing in will be removed, and the containers will be placed on top of the 25 cm diameter pipes filled with sand. Clonal growth will be measured as the number and biomass of new shoots growing out of the sand in the pipe, established via roots from the potted plant. This will simulate the ability of plants to spread clonally from higher elevations into the zone influenced by daily fluctuations near the river channel.
- 2) **Establishment.** Seeds will be germinated in germination chambers in the Butterfield Lab at NAU. Immediately upon root emergence, seedlings will be planted near the surface of containers. Seedling survival will be monitored frequently, and seedling growth, including maximum rooting depth, root biomass and aboveground biomass will be destructively measured after one month.
- 3) **Growth and physiology.** Plants will be propagated from seeds or cuttings in spring of 2025, and up-potted aggressively to maximize plant growth. This was a limitation of our pilot experiment, in that we were not able to establish large enough plants for frequent, semi-destructive measurements of plant water status in the small containers that we used. Plants will be subjected to flow fluctuation treatments in summer of 2027. Leaf and/or stem water potential and stomatal conductance will be monitored weekly, and photosynthetic rate will be monitored bi-weekly. All of the necessary equipment for these measurements is already available from the Butterfield and Palmquist labs. New root growth and plant height growth will be measured at the conclusion of the experiment. Carbon and nitrogen isotope ratios of leaves and roots, indicators of plant water stress, will be quantified at the Colorado Plateau Stable Isotope Laboratory at NAU.

We will use results from a previous study on hydroclimate suitability (Butterfield and others, 2023) to select species for experimental treatments. Three categories of species will be selected: (1) those for which the CRe is predicted to be climatically suitable, and are abundant, (2) those for which the CRe is predicted to be climatically suitable, but are absent or rare, and (3) those for which the CRe is predicted to be climatically unsuitable but are abundant. Category 2 are species that are predicted to be negatively impacted by daily fluctuations, while Category 3 are species that are predicted to benefit from daily fluctuations. Species from each category will be selected among plants with different growth forms (annuals, perennial grasses, sub-shrubs, shrubs, trees).

Effects of experimental treatments will be quantified using general linear modeling approaches and used to quantify effect sizes of experimental treatments (for example, Butterfield and Palmquist, 2024b). Functional traits for each species will be extracted from existing databases (Palmquist and others, 2017) and floras, and simplified via principal coordinates analyses. Trait components and the three species categories delineated above (climatically suitable/abundant, climatically suitable/rare, climatically unsuitable/abundant) will be used to predict treatment effect sizes among species. This framework can then be used to predict how other species not included in the experimental treatments are likely to be influenced by daily fluctuations. These results will then be incorporated into synthesis efforts outlined below.

Anticipated Use of Data

These data will be used to evaluate how plant communities will respond to the loss of daily fluctuating flows if Lake Powell levels drop below hydropower, a major knowledge gap when predicting how future conditions may impact plant communities.

Outcomes and Products

- Presentations on management implications at GCMRC’s annual reporting meetings.
- Journal publication on physiological responses of a suite of riparian plant species to daily fluctuating flows.

Project Element C.3. Predictive Modeling of Vegetation Responses to Dam Operations (Modified Study, Partially Funded)

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This project element will conduct forward-looking modeling efforts to identify the flow conditions that would optimize the vegetation objectives listed in Goal 11.

Lags in vegetation responses to hydrological events and the lack of interannual flow variability in the CRE create the need to use data derived from regional collaborations and manipulative experiments that expand the range of biophysical parameters used to construct vegetation models. This project element will integrate ground-based vegetation monitoring (Project Element C.1), manipulative experiments (Project Element C.2), and existing regional data on riparian vegetation composition and hydrographs using advanced statistical modeling.

Hypotheses and Science Questions

- How do plant species interact with one another, and how are these interactions mediated by flow conditions? What are the implications of these interaction networks for vegetation responses to tamarisk die-off?
- What are the flow scenarios necessary to achieve specific vegetation objectives?

Research Question C.3.1. How do Plant Species Interact with one Another, and How Are These Interactions Mediated by Flow Conditions?

Plants interact with one another, and these interactions affect the relative abundance and dynamics of plant populations (Butterfield, 2009; Cavieres and others, 2014). Plants compete with one another for limiting resources such as water and light but can also facilitate one another through amelioration of environmental stressors like extreme temperatures or flood disturbance (Callaway, 2007). Environmental conditions strongly influence these interaction outcomes (Soliveres and Maestre, 2014), including in riparian zones, where multiple aspects of the flow regime can determine the intensity and outcome of plant-plant interactions. Until now, our riparian vegetation models have assumed that plant species respond to flow conditions independently of one another. This was a simplifying assumption that helped us develop our first set of models, but we know that this ignores potentially important dynamics that influence vegetation status.

We hypothesize that certain species have outsized effects on the rest of the plant community, particularly arrowweed (*Pluchea sericea*) and tamarisk (*Tamarix ramosissima* and hybrids with *T. chinensis*). Tamarisk is particularly relevant, given the spread of the tamarisk beetle: as tamarisk stands die, other species may experience competitive release, while others that benefited from the deep shade and saline litter layer produced by tamarisk may suffer. We also expect that some other species may have strong competitive effects, such as the rapidly growing hydrophilic species like seepwillows (*Baccharis emoryi*, *Baccharis salicifolia*, *Baccharis sarothroides*) and clonal grasses. Less apparent interactions may also be occurring, particularly facilitative effects that buffer disturbances and promote vegetation encroachment. Indeed, we have demonstrated strong, flow-dependent effects of different plant functional groups on sedimentation (Butterfield and others, 2020), which can stabilize sediment that promotes the growth of other species.

We propose to quantify the interaction networks among riparian plant species within the CRe to achieve several inter-related objectives:

- 1) Identify how plant-plant interactions modify predictions of vegetation responses to flow conditions.
- 2) Determine how the balance between competition and facilitation may be influencing vegetation encroachment in different hydrological settings defined by the flow regime.
- 3) Predict the impacts of tamarisk mortality on subordinate species and vegetation metrics.

Methods

We will take advantage of recent developments in joint species distribution modeling (JSDMs; Hui, 2016) to identify associations between species that can be attributed to biotic interactions (competition or facilitation), and the hydrologic settings in which those interactions occur. This modeling effort will leverage the extensive ground-based monitoring data from C.1, an existing trait dataset that has been assembled for the CRe over multiple work plans (McCoy-Sulentic and others, 2017; Palmquist and others, 2017), and hydrological data from Project A.1 (U.S. Geological Survey, 2024).

We will implement JSDMs using the ‘boral’ package (Hui, 2016) in the R statistical platform. We will use elevation above critical hydrological thresholds – minimum of daily fluctuations, maximum of daily fluctuations, and HFEs (if present in a given time period) – as well as botanical region (Palmquist and others, 2018a) as environmental predictors. Functional traits will be used as covariates that mediate plant responses to the environmental variables. This has the effect of improving model predictions for species with lower frequency of occurrence, and improving generalization of model predictions based on trait-environment relationships that can be related to other taxa and communities.

After fitting the JSDMs, we will follow the workflow of D'Amen and others (2018) to identify residual variation in species covariances that can be attributed to competition and facilitation (Zurell and others, 2018). The hydrological preferences (Butterfield and others, 2023) of interacting species will then be used to identify the hydrological conditions that promote different types of interactions outcomes. These interactions will also be compared to the results of Butterfield and others (2020) to explore whether biotic interactions, particularly facilitation, is associated with species belonging to morphological functional groups related to sedimentation and, hence, positive feedbacks on sediment stabilization and plant encroachment. Finally, we will focus particularly on the strength and types of interactions associated with tamarisk and predict how tamarisk mortality is likely to influence vegetation composition in different hydrological settings.

Anticipated Use of Data

These data will be used to provide management guidance on which plant species could be altered, either through removals or dam operations, for greatest effect on the CRe ecosystem.

Outcomes and Products

- Presentations on management implications at GCMRC's annual reporting meetings.
- Journal publication on facilitation and competition among plant species in the CRe.

Research Question C.3.2. What are the Flow Scenarios Necessary to Achieve Specific Vegetation Objectives? (Unfunded)

This research question will conduct forward-looking modeling efforts to identify the flow conditions that would optimize the vegetation objectives listed in Goal 11. The metrics identified for Goal 11 will be used to frame the outcomes of this modeling (native dominance, richness, cover), similar to those used in Yackulic and others (2024). To date, our modeling efforts in the CRe have focused on the responses of riparian vegetation to flow patterns designed for other resources (Butterfield and others, 2023; Yackulic and others, 2024). While important, this approach does not provide a systematic assessment of how vegetation metrics respond to the full range of flow scenarios that could be implemented. Modeling a wider range of flow scenarios will provide a more comprehensive picture of the tradeoffs and compatibilities among different vegetation metrics and may also identify dam operations that can satisfy multiple objectives.

Methods

We will combine our modeling innovations developed for the Colorado River Basin Post-2026 Operations Exploration Tool (WebTool, tool.crbpost2026dmdu.org) with synthetic models of plant responses to multiple hydrological dimensions. We developed an artificial environmental surface for the WebTool that facilitates rapid modeling of thousands of hydrological and climate scenarios while faithfully representing the riparian zone of the CRe (Yackulic and others, 2024). Specific hydrological variables and species-specific responses will include,

- 1) HFE magnitude - species responses will be based on modeled results from our monitoring program, primarily the new JSMD response functions generated from addressing Research Question C.3.1 above.
- 2) HFE seasonality - Previous syntheses (Stromberg and others, 2007; Ralston and others, 2014) and case studies (Amlin and Rood, 2002) will be used to predict species responses to the seasonal timing of HFEs, primarily in the context of regeneration potential. Additional data on plant reproductive phenology will be extracted from floras and herbarium collections to further inform regeneration potential in response to HFE timing.

- 3) Base flow seasonality - Published models based on regional synthesis will be used to predict species responses to shifts in relative monthly volumes (Butterfield and others, 2023).
- 4) Magnitude of daily fluctuations - JSMD response functions described above, existing data on rooting depths (Stromberg, 2013; Palmquist and others, 2017), as well as experimental results from Project Element C.2, will be used to parameterize species responses to daily fluctuations.

Change in species-specific habitat suitability will be modeled in this four-dimensional hydrological space, with subsets selected for representation in 2-D and 3-D response surfaces. Species-specific responses will be aggregated to specific metrics of interest used in previous modeling efforts: total vegetation cover, species richness, and native dominance (Palmquist and others, 2023; Yackulic and others, 2024).

Anticipated Use of Data

These results will be used to determine what modifications can be made to experimental and daily flow patterns to promote desirable qualities of riparian plant communities.

Outcomes and Products

- Presentations on management implications at GCMRC's annual reporting meetings.
- Journal publication on dam operations that would best support plant community metrics for the CRe.

Project Element C.4. Biogeomorphic Linkages between Streamflow, Sediment Transport, and Vegetation Composition (New Study, Partially Funded)

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Traditional models describing drivers of physical river change typically relate the transport capacity of a river's floods to the amount and size of sediment supplied to the river channel. If the floods are able to transport more sediment than is supplied to the channel, erosion will occur; if more sediment is supplied to the river channel than can be transported, deposition will occur (Lane, 1954). However, many studies have shown that those traditional models are overly simplistic because other phenomena, such as vegetation, can substantially modulate flow velocities, shear stress, sediment transport, and flood conveyance, thereby strongly influencing the size and shape of river channels and their floodplains (Figure 4) (Burkham, 1976a; Burkham, 1976b; Vincent and others, 2009; Nepf, 2012; Manners and others, 2014; Gellis and others, 2017; Dean and Topping, 2019, 2024; Walker and others, 2020).

The effects of vegetation on hydraulics and sediment transport have been demonstrated in experimental flume studies (Tal and Paola, 2007; Zong and Nepf, 2010; Le Bouteiller and Venditti, 2014), numerical modeling studies (López and García, 1998; Griffin and others, 2005; Griffin and others, 2014), and large-scale field studies (Burkham, 1976b; Bywater-Reyes and others, 2017; Manners and others, 2013, 2014; Dean and Topping, 2019, 2024). We refer to the effects of vegetation on channel morphology as *biogeomorphic* effects because that term highlights how biology (that is, vegetation) can influence geomorphic process and form.

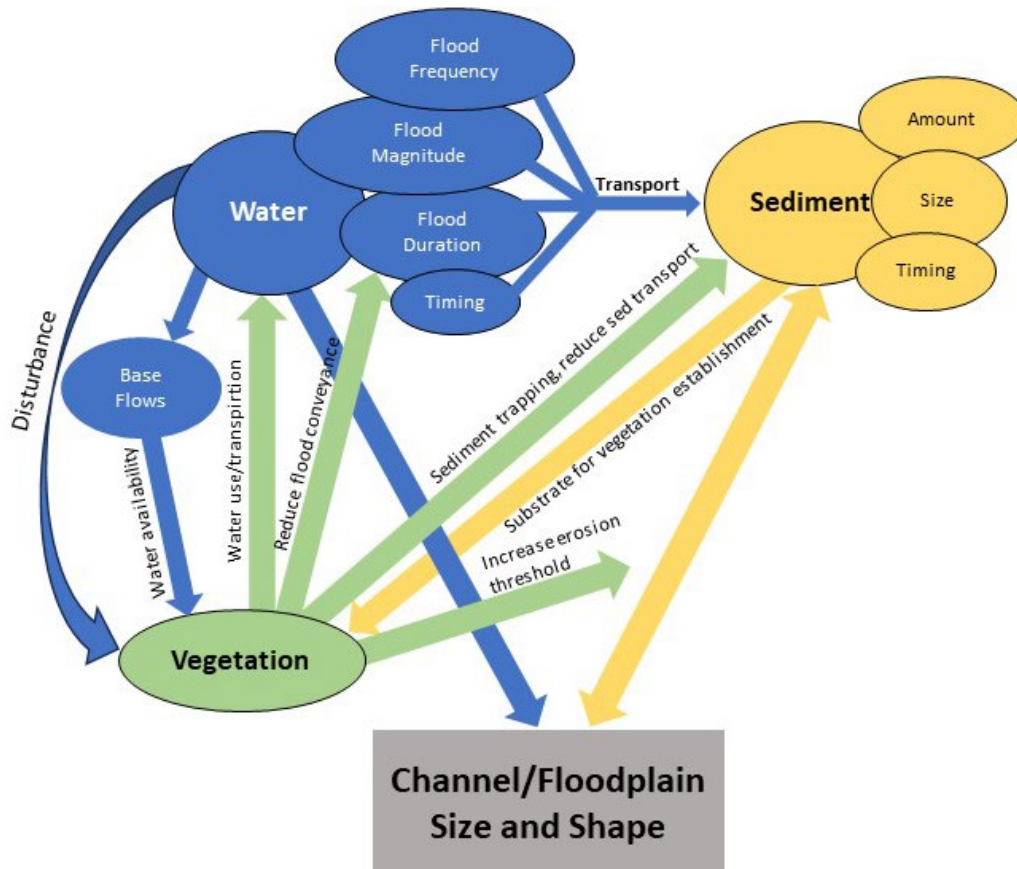


Figure 5. Diagram of the biogeomorphic feedbacks between river flow, sediment (also sed.), and vegetation. The size and shape of river channel and their floodplains is largely determined by the flood hydrology (flood magnitude, frequency, duration, and timing) and sediment supply (amount, size, and timing of delivery to the channel). The flood hydrology and channel morphology determine how much of the supplied sediment may be transported. Vegetation can substantially affect both hydrology and sediment transport because vegetation affects hydraulics, flood and sediment conveyance, and the erosional thresholds of the channel banks. Floods provide disturbance which may limit the establishment of vegetation and control where vegetation can establish. Sediment provides the substrate for vegetation establishment.

Thus, any large-scale change in vegetated area, species, or density may result in corresponding changes to river channels and their floodplains. The magnitude and style of channel and floodplain change driven by vegetation is dependent upon plant size, morphology, and density,

with rigid, dense, multi-stemmed plants exerting a larger influence on geomorphic processes (Diehl and others, 2017; Manners and others, 2015; Bywater-Reyes and others, 2022).

Along the Colorado River in Marble and Grand Canyons, large increases in vegetated area, and large changes in species composition have occurred since the beginning of Glen Canyon Dam operations in 1963 (Sankey and others, 2015; McCoy-Sulentic and others, 2017; Durning and others, 2021) (Figure 5). Sankey and others (2015) and Durning and others (2021) showed that the vegetated area throughout the river corridor expanded by 20% to almost 40%, with much of that expansion occurring through the establishment of relatively rigid, multi-stemmed, nonnative tamarisk and native seepwillow (*Baccharis* spp.). Other plants that were largely responsible for the vegetation expansion were the shrub arrowweed (*Pluchea sericea*), and common reed (*Phragmites australis*). These native shrubs, nonnative tamarisk, and common reed can grow in dense thickets along the channel margin and likely trap sediment. Butterfield and others (2020) showed that deposition of sediment was correlated with vegetation presence and type along the Colorado River, with rhizomatous and herbaceous plants trapping sediment in eddy-separation zones, and tall herbaceous plants and large shrubs trapping sediment in eddy-reattachment zones.



Figure 6. Matched photos from Sankey and others (2023) showing large expansion of riparian vegetation between 1923 (top) and 2019 (bottom) taken just upstream of President Harding Rapid. (Photo Credits. Top: E.C. La Rue, 1923, U.S. Geological Survey Photographic Library, Denver, CO. Bottom: A. H. Fairley, May 2019, U.S. Geological Survey.)

Although Butterfield and others' (2020) work along the Colorado River in Marble and Grand Canyons shows that vegetation has trapped sediment in eddies downstream from debris fans, those results were found at select study sites and there have been no studies investigating whether system-wide changes in the channel of the Colorado River have occurred throughout the river corridor.

We propose to conduct a multi-phased, interdisciplinary study aimed at quantifying large-scale changes to channel morphology relative to vegetation change. The work in Project C.4 will leverage datasets produced by Project Elements L.1, Project B.4, and Project D.2 to evaluate changes in the *biogeomorphic* condition of the river corridor (see Figure 6 for project linkages). Project C.4.1 will build upon Butterfield and others (2020) and Durning and others (2021) by investigating whether large-scale changes in vegetated area and species have had a corresponding impact on river channel form, predominantly channel width. This will be done using a large suite of remote sensing data collected by Project L between 2002 and 2021. These findings will be linked to historical photograph comparisons made in Project Element D.2 that depict changes in the shape of the channel banks, which may indicate whether channel changes visible in the aerial photographs have also corresponded to changes in floodplain elevation, and bank steepness. Project Element C.4.2 will implement field measurements of plant traits within a 2-dimensional hydraulic model to specifically analyze how effective individual vegetation species are at causing channel change by altering channel-margin/floodplain hydraulics and sediment transport. C.4.2 will build upon Butterfield and others (2020) and leverage hydraulic model development conducted in Project B.4. Both phases of work will provide needed information regarding vegetation management actions throughout the river corridor.

Hypotheses and Science Questions

- How has widespread vegetation expansion affected river channel form?
- How effective are individual plant species at altering hydraulics and sediment transport?

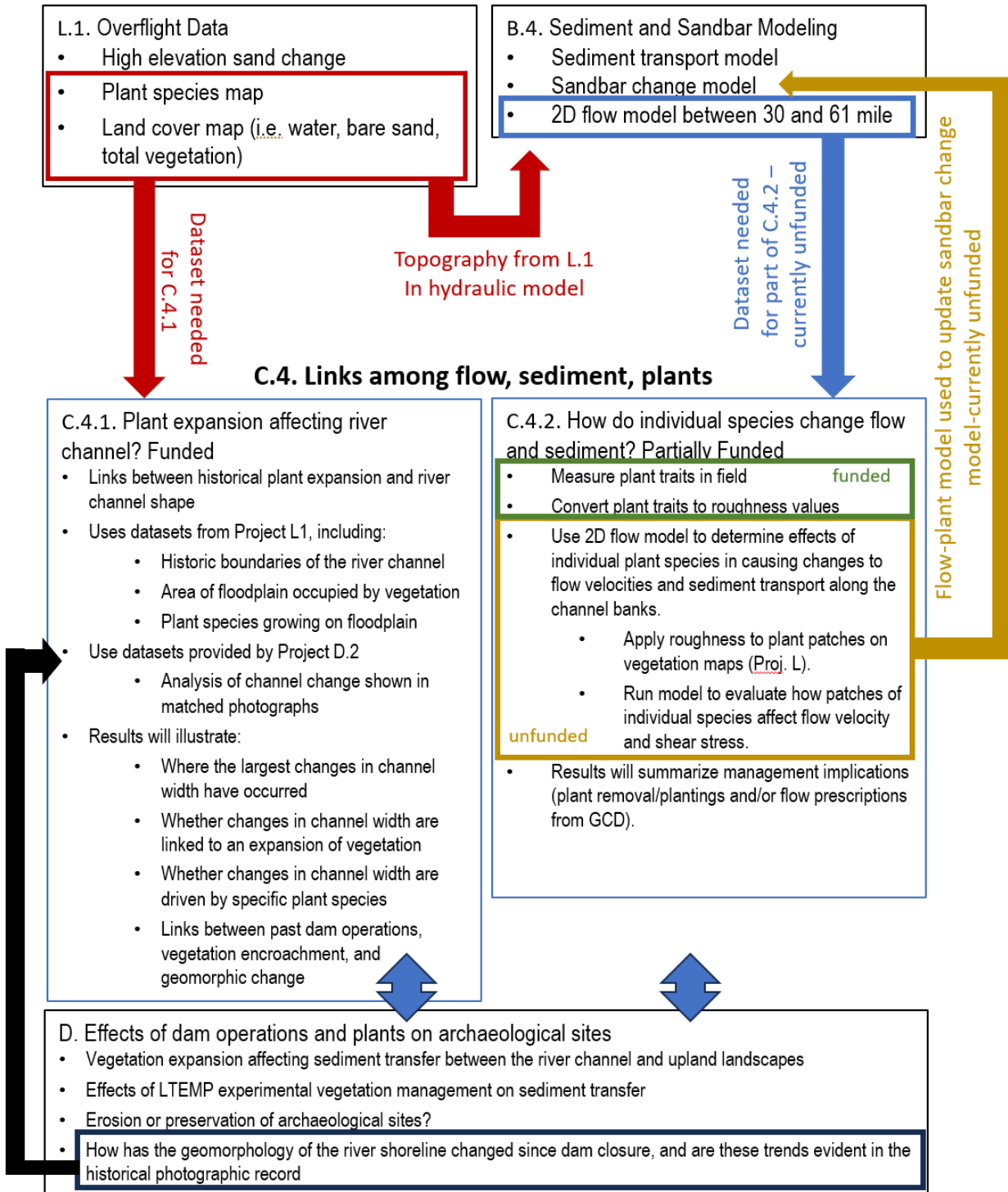


Figure 6. Description of linkages among Projects C.4, L.1, B.4, and D.2. Note that projects produced by Project C.4.2 will feed back into the sandbar response model within Project B.4 when that project element becomes funded.

Research Question C.4.1. How has Widespread Vegetation Expansion Affected River Channel Form?

Methods

C.4.1 is interdisciplinary, bridging work conducted by Projects C, L.1, and D.2. For this project element, we will build on recent studies of vegetation change (Sankey and others, 2015; Durning and others, 2021) and biogeomorphic linkages demonstrated by Butterfield and others (2020) and Sankey and others (2023) by conducting an analysis of historic channel changes relative to vegetation change. This analysis will be conducted for Marble Canyon (from Lees Ferry to the Little Colorado River confluence), Furnace Flats (that is river miles 66-74), and other key areas of interest identified in the historical photograph comparisons. Expansion of the analysis reaches will occur if feasible. The Marble Canyon analysis segment will dovetail with analyses of changes in high elevation sand conducted in Project L.1, which will be conducted over the same segment. We will map the active channel boundaries visible in the 2002, 2009, 2013, and 2021 aerial imagery collected by Project L, and calculate changes in channel width over time; to date, there has been no large-scale analysis of channel width change along the Colorado River in Grand Canyon since the beginning of Glen Canyon Dam operations. Spatial changes in channel width will be correlated to changes in vegetated area and species composition as mapped by Sankey and other (2015) and Durning and others (2021). Historical photograph comparisons made in Project D.2 will further illustrate how changes in channel width measured in this project element correspond to changes in the shape and height of the river shoreline, thus providing an on-the-ground perspective to *biogeomorphic* changes. Additionally, understanding the degree of vegetation change along the channel margin, and the associated change in the availability of bare sand, has implications for the preservation of archaeological sites as further studied in Project D.1. Results of this study will demonstrate where the largest changes in channel width have occurred, whether changes in channel width are linked to an expansion of vegetation area, and whether changes in width are specifically driven by specific vegetation species compositions. Furthermore, this project element may identify potential linkages between past dam operations and the timing of vegetation encroachment and channel change.

Anticipated Use of Data

The results of this study will indicate if ongoing riparian plant expansion is altering the Colorado River channel, and thus its ability to convey water and sediment and support diverse aquatic habitats.

Outcomes and Products

- Maps of channel width change, released as a USGS Data Release.

- Potential linkages between past dam operations and the timing of vegetation encroachment and geomorphic change.
- Summary of management implications presented at GCMRC’s annual reporting meetings.
- Journal publication or USGS publication discussing role of vegetation in causing geomorphic change in the Colorado River in Marble and Grand Canyons.

Research Question C.4.2. How Effective are Individual Plant Species at Altering Hydraulics and Sediment Transport? (Partially Funded)

Methods

The goal of Project Element C.4.2 is to incorporate physical plant traits into a 2-dimensional hydraulic model developed as part of Project Element B.4 to evaluate the effectiveness of different plant species in causing changes to the channel and floodplain. Plants are effective at causing channel/floodplain change because they affect flow velocities, shear stress, and sediment transport. However, Project Element B.4 is currently unfunded, and the hydraulic model will not be developed as part of the FY 2025–27 TWP. Therefore, as part of C.4.2, we will focus on collecting the relevant plant-trait data needed to determine the effects of plants on hydraulics and sediment transport at the plant and patch scale. These data will later be incorporated into the hydraulic model once it is developed to understand the effects of plants over large spatial scales. Collection of the necessary plant trait data will also benefit other aspects of Project C, particularly C.1.2, and be added to the GCMRC plant trait matrix (Palmquist and others, 2017) for future trait-based analyses of plant communities.

The ability of vegetation to alter hydraulics and sediment transport is largely dependent upon plant morphology and flexibility; rigid, multi-stemmed plants have a larger effect than flexible, single-stemmed plants. Thus, to constrain the effects of individual species, we will measure plant traits in the field to determine plant frontal area (that is, the area of a plant that interacts with flow), plant flexibility, and buoyancy. Plant frontal area will be calculated using photographs of the plants and Easy Leaf Area open-source software, plant flexibility measurements will be made using a force gage at multiple heights up the stem, and buoyancy will be calculated using samples of the plant stems. These parameters will be used to develop depth-varied roughness curves for each measured species. A roughness curve indicates the relative degree that a plant affects flow velocity, shear stress, and sediment transport at different flow depths. These roughness curves will provide much needed information on the relative ability of each plant species to affect hydraulics and sediment transport at the plant and patch scale, but will not yet be integrated into a hydraulic model to analyze these phenomena over large spatial scales.

When the hydraulic model to be developed as part of Project Element B.4 is funded and completed at a later date, vegetation maps will be used to determine where patches of individual species exist along the channel margin. The depth-varied roughness curves will be applied to the requisite patches in the hydraulic model. Simulations will be run to evaluate how patches of individual species along the channel margins affect flow velocity and shear stress distributions. We will also attempt to adapt these 2-dimensional models to run morphodynamic simulations to evaluate how these plant patches influence sediment deposition and erosion.

Anticipated Use of Data

The results of this study will indicate if there are some plant species that could be managed, either through removal or flow patterns, to better support sediment goals in the CRe.

Outcomes and Products

- Plant trait data added to published GCMRC plant trait data matrix (Palmquist and others, 2017) and utilized in C.1.2.
- Summary of management implications presented at GCMRC's annual reporting meetings.
- Journal publication or USGS publication discussing hydraulic effects of individual plant species.
- Analyses of C.4.1 with C.4.2 together may help inform how past dam operations have influenced the expansion of different plant species.

Project Element C.5. Experimental Vegetation Treatment Decision Support (Experimental Fund)

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The LTEMP identified the need for experimental vegetation management to be conducted in the CRe (U.S. Department of the Interior, 2016). These efforts are being led by Glen Canyon National Recreation Area and Grand Canyon National Park through Project 4.D of the FY 2025-27 Reclamation Triennial Work Plan (previously Project Elements C.7 and C.8 of the FY 2021-23 Reclamation Triennial Work Plan) and are monitored, in part, by Project D.3 of the FY 2025-27 GCMRC Triennial Work Plan. Thus far, the experimental vegetation treatments have focused on vegetation removal to support recreational and sociocultural resources but plans for revegetation with native species are included.

These projects are collaborative in nature and explicitly state that planning, implementation, and evaluation of these projects will be coordinated among NPS, Tribal partners, and GCMRC.

Hypotheses and Science Questions

- How can experimental vegetation treatments being implemented by National Park Service (NPS) be improved upon?

Methods

This project element supports GCMRC Principal Investigator participation in regular meetings about site selection, monitoring, and experimental treatment plans. It also includes time for site visits, assistance with data analysis. We will collaborate on GCMRC Project D.3 and continue to facilitate NPS work as needed. Aspects of this support include, but are not limited to, assistance designing management experiments, consultation on plant species and collection locations for out-planting, sharing of data and maps, assistance with pre- and post-treatment data collection, facilitating the use of new restoration treatments and newly available research, site visits, input on the selection of work sites, and data analysis.

Anticipated Use of Data

This element dedicates time to assisting NPS partners in interpreting and implementing knowledge gained from other parts of Project C and outside research to experimental management actions.

Outcomes and Products

- Participation in experimental vegetation management planning meetings with partners.
- Contributing data products, analyses, and expertise needed for planning and monitoring purposes.
- Assist with developing or analyzing experiment monitoring data, as needed.
- Assistance with native species plant material decisions and revegetation planning.

Project Element C.6. Plant Physiological Responses to Experimental Flows (Experimental Fund)

Periods of high and low flows are key drivers of riparian plant success (Tabacchi and others, 1998; Greet and others, 2011; Butterfield and others, 2023). Floods provide water to higher elevation plants but can also create plant stress through long periods of submergence.

Periods of low flows can allow for plant establishment or desiccate species that rely on constant water supplies (most riparian species) (Mahoney and Rood, 1998; Stromberg and others, 2007).

The timing, magnitude, and duration of high and low flow anomalies can result in stress-inducing conditions for some species while providing ideal conditions for others (Stromberg and others, 2007; Greet and others, 2013). Flow experiments designed to impact other CRe resources of concern (i.e., sediment deposition or nonnative fish expansion) will impact riparian plants through physiological stress in some species and physiological relief in others.

Current monitoring methods cannot clearly disentangle the impacts of experimental flows from daily operations. The proposed experiments in C.2 provide an understanding of how plants respond to specific flow patterns in a greenhouse setting but are limited by the comparatively small plant size and age of plants used in the experiments and the climate conditions of the greenhouse (Palmquist and others, 2022; Butterfield and Palmquist, 2024b). Collecting physiological measurements on plants before, during, and after planned flow experiments can help connect greenhouse experimental results to observed patterns of plant occurrence and cover in the field (Butterfield and Palmquist, 2024a).

Hypotheses and Science Questions

- What are the short-term physiological responses of common and uncommon riparian plant species to experimental flow patterns?

Methods

Physiological responses of plants during drawdowns and high flow releases can indicate if experimental flows benefit or disadvantage species of interest (Baladrón and others, 2022; Blasini and others, 2022; Butterfield and Palmquist, 2024b; Moran and others, 2023). Measurements of photosynthesis, stomatal conductance, water potential, and leaf water content will be made on plants near Lees Ferry, AZ before, during, and after experimental flows to assess plant reactions to rapid changes in water levels. We conducted similar work during the Spring Disturbance Flow in 2020, resulting in a peer-reviewed publication and new insights into physiological responses of two important species to high and low flow anomalies (Butterfield and Palmquist, 2024b).

Building off the results of this, other previous experiments (Palmquist and others, 2022; Butterfield and Palmquist, 2024b), and Project Elements C.1 and C.2, physiological studies conducted during experimental flows will focus on species expected to be affected by the planned flow experiment and could include both native and nonnative species of interest, such as Emory's baccharis (*Baccharis emoryi*), coyote willow (*Salix exigua*), arrowweed (*Pluchea sericea*), sacred datura (*Datura sp.*), tall fescue (*Schedonorus arundinaceus*), horsetail (*Equisetum xferrissii*), common reed (*Phragmites australis*), honey mesquite (*Prosopis*

glandulosa), tamarisk (*Tamarix* sp.), and species planted at Paria Beach (cottonwood, *Populus fremontii*; Goodding's willow, *Salix gooddingii*).

Anticipated Use of Data

These data will provide information on how experimental flows can be altered to better support riparian plant community goals.

Outcomes and Products

- Summary of management implications presented at GCMRC annual reporting meetings.
- Data on plant species physiological responses to experimental flows, included in other plant response modeling (C.3).

Project Element C.7. Effects of Plants on Flow Velocities and Sediment Transport During Experimental Flows (Experimental Fund)

Changes in river flows can result in a cascade of feedbacks between plant growth/expansion and associated changes in flow velocity and sediment transport. Changes in river flows alter riparian plant growth and expansion through the amount of water available for plant uptake and growth, increases in plant stress through prolonged submergence, seed dispersal, and erosion of suitable habitat. In turn, plant growth adds roughness to the channel banks which affects channel-margin flow velocities and sediment erosion and deposition, while plant loss has different impacts on channel shape and change. In many rivers, the expansion of riparian plants has resulted in large, and often irreversible, changes in river channel form (Vincent and others, 2009; Griffin and others, 2014; Manners and others, 2014; Dean and Topping, 2019, 2024).

In the CRe downstream from Glen Canyon Dam, there has been a widespread expansion of vegetation because of reduced flood disturbance and elevated baseflows. The species responsible for this expansion include nonnative tamarisk (rigid, multi-stemmed), native seepwillow (flexible, multistemmed), native arrowweed (flexible, multi-stemmed), and both native and nonnative common reed (dense, flexible, single-stemmed). Each of these species have complex relationships with flow velocity and sediment deposition that are impacting the CRe (Durning and others, 2021), and other plant species occurring along the river margin likely do, too. The magnitude by which vegetation can affect flow velocity and sediment transport is determined by stem density and morphology, which varies among species. In general, dense, rigid, multi-stemmed plants have a greater effect on flow velocity and sediment transport than sparse, flexible, single-stemmed plants.

Hypotheses and Science Questions

- How do plants with differing physical traits alter flow velocity and sediment deposition during experimental flows?

Methods

We plan to use the experimental fund to directly measure the effects of common plant species on flow velocities and sediment transport during experimental flows. During experimental flows, we will use an acoustic Doppler velocimeter to collect high-resolution 3-dimensional velocity and turbulence data within and adjacent to patches of different plant species to determine the magnitude that different species affect flow velocities. We will collect suspended-sediment samples upstream, within, and downstream of plant patches to determine the effects of plants on sediment transport. We will also try to make measurements in conjunction with experimental vegetation removal areas (Project D.3 and Project 4.D of the FY 2025-27 Reclamation Triennial Work Plan) to evaluate how those removal actions affect flow velocities in those areas.

Additionally, excavations will be made following experimental flows to determine the depth and grain size of sediment deposited within and adjacent to each plant patch.

Anticipated Use of Data

These data will help determine how experimental flows can be altered to better support goals related to sediment deposition/erosion and therefore river channel form.

These data will also help environmental managers plan for restoration actions focused on riparian plant management, including plant removal.

Outcomes and Products

- Summary of management implications presented at GCMRC annual reporting meetings.
- Plant trait data added to published GCMRC plant trait data matrix (Palmquist and others, 2017).
- Data included in future modeling efforts on plant-sand interactions.

Budgets

Project Element Budget

| Fiscal Year 2025 | | | | | | | | | |
|---|------------------|----------------------|-----------------------|-----------------------|---------------------------|--------------------------|-----------------------------|------------------|------------------------------|
| Project C Riparian Vegetation Monitoring and Research | Salaries | Travel & Training | Operating Expenses | Logistics Expenses | Cooperative Agreements | To other USGS Centers | AMP Special Burden Rate* | Total | USGS Contributing Funds** |
| | | | | | | | 21.86% | | Est Full Rate = 55% |
| C.1. Ground-based riparian vegetation monitoring | \$124,827 | \$3,615 | \$3,500 | \$81,775 | \$12,542 | \$0 | \$47,101 | \$273,360 | |
| C.2. Determining hydrological tolerances and management tools for plant species of interest | \$13,445 | \$0 | \$0 | \$0 | \$45,452 | \$0 | \$4,303 | \$63,201 | |
| C.3. Predictive models and synthesis | \$9,845 | \$0 | \$0 | \$0 | \$20,007 | \$0 | \$2,753 | \$32,604 | |
| C.4. Biogeomorphic linkages between streamflow, sediment transport, and vegetation composition | \$105,222 | \$0 | \$0 | \$0 | \$0 | \$0 | \$23,005 | \$128,226 | |
| Total Project C | \$253,339 | \$3,615 | \$3,500 | \$81,775 | \$78,001 | \$0 | \$77,162 | \$497,392 | \$35,739 |

| Fiscal Year 2026 | | | | | | | | | |
|---|------------------|----------------------|-----------------------|-----------------------|---------------------------|--------------------------|-----------------------------|------------------|------------------------------|
| Project C Riparian Vegetation Monitoring and Research | Salaries | Travel & Training | Operating Expenses | Logistics Expenses | Cooperative Agreements | To other USGS Centers | AMP Special Burden Rate* | Total | USGS Contributing Funds** |
| | | | | | | | 22.60% | | Est Full Rate = 55.5% |
| C.1. Ground-based riparian vegetation monitoring | \$131,069 | \$3,615 | \$3,500 | \$85,182 | \$13,327 | \$0 | \$50,880 | \$287,573 | |
| C.2. Determining hydrological tolerances and management tools for plant species of interest | \$14,634 | \$0 | \$0 | \$0 | \$48,681 | \$0 | \$4,768 | \$68,083 | |
| C.3. Predictive models and synthesis | \$10,337 | \$0 | \$0 | \$0 | \$24,423 | \$0 | \$3,069 | \$37,829 | |
| C.4. Biogeomorphic linkages between streamflow, sediment transport, and vegetation composition | \$41,431 | \$0 | \$0 | \$0 | \$0 | \$0 | \$9,363 | \$50,794 | |
| Total Project C | \$197,471 | \$3,615 | \$3,500 | \$85,182 | \$86,431 | \$0 | \$68,081 | \$444,279 | \$30,706 |

| Fiscal Year 2027 | | | | | | | | | |
|---|------------------|----------------------|-----------------------|-----------------------|---------------------------|--------------------------|-----------------------------|------------------|------------------------------|
| Project C Riparian Vegetation Monitoring and Research | Salaries | Travel & Training | Operating Expenses | Logistics Expenses | Cooperative Agreements | To other USGS Centers | AMP Special Burden Rate* | Total | USGS Contributing Funds** |
| | | | | | | | 23.40% | | Est Full Rate = 56% |
| C.1. Ground-based riparian vegetation monitoring | \$141,592 | \$1,615 | \$3,500 | \$88,103 | \$14,112 | \$0 | \$55,369 | \$304,291 | |
| C.2. Determining hydrological tolerances and management tools for plant species of interest | \$11,397 | \$0 | \$0 | \$0 | \$48,820 | \$0 | \$4,131 | \$64,348 | |
| C.3. Predictive models and synthesis | \$10,854 | \$0 | \$0 | \$0 | \$25,156 | \$0 | \$3,294 | \$39,304 | |
| C.4. Biogeomorphic linkages between streamflow, sediment transport, and vegetation composition | \$43,503 | \$0 | \$0 | \$0 | \$0 | \$0 | \$10,180 | \$53,682 | |
| Total Project C | \$207,345 | \$1,615 | \$3,500 | \$88,103 | \$88,088 | \$0 | \$72,974 | \$461,625 | \$32,352 |

*GCDAMP Special Rate: Rate includes current DOI preferred rate (currently 15%; subject to change) and facilities rate (will vary annually). No USGS bureau overhead is charged (unique to the GCDAMP agreement).

**USGS Contributing Funds: The amount of funds required to cover the subsidy created by the reduced burden rate (i.e., GCDAMP Special Rate). As in previous years, SBSC/GCMRC will request these funds of the USGS cost-share program. These funds are not guaranteed.

Experimental Fund

| Fiscal Year 2025 | | | | | | | | |
|---|-----------------|----------------------|-----------------------|-----------------------|---------------------------|--------------------------|-----------------------------|------------------|
| Project C Experimental Fund Projects | Salaries | Travel & Training | Operating Expenses | Logistics Expenses | Cooperative Agreements | To other USGS Centers | AMP Special Burden Rate* | Total |
| | | | | | | | 21.86% | |
| C.5. Experimental Vegetation Treatment Support | \$9,845 | \$0 | \$0 | \$0 | \$4,001 | \$0 | \$2,272 | \$16,119 |
| C.6. Plant physiological responses to experimental flows | \$0 | \$450 | \$400 | \$9,362 | \$0 | \$0 | \$2,233 | \$12,445 |
| C.7. Effects of plants on flow velocity and sand deposition | \$55,869 | \$0 | \$0 | \$14,932 | \$0 | \$0 | \$15,479 | \$86,280 |
| Total Project C | \$65,714 | \$450 | \$400 | \$24,294 | \$4,001 | \$0 | \$19,984 | \$114,843 |

| Fiscal Year 2026 | | | | | | | | |
|---|-----------------|----------------------|-----------------------|-----------------------|---------------------------|--------------------------|-----------------------------|------------------|
| Project C Experimental Fund Projects | Salaries | Travel & Training | Operating Expenses | Logistics Expenses | Cooperative Agreements | To other USGS Centers | AMP Special Burden Rate* | Total |
| | | | | | | | 22.60% | |
| C.5. Experimental Vegetation Treatment Support | \$10,337 | \$0 | \$0 | \$0 | \$4,884 | \$0 | \$2,483 | \$17,704 |
| C.6. Plant physiological responses to experimental flows | \$0 | \$450 | \$400 | \$9,774 | \$0 | \$0 | \$2,401 | \$13,025 |
| C.7. Effects of plants on flow velocity and sand deposition | \$58,663 | \$0 | \$0 | \$15,432 | \$0 | \$0 | \$16,745 | \$90,840 |
| Total Project C | \$69,000 | \$450 | \$400 | \$25,206 | \$4,884 | \$0 | \$21,629 | \$121,568 |

| Fiscal Year 2027 | | | | | | | | |
|---|-----------------|----------------------|-----------------------|-----------------------|---------------------------|--------------------------|-----------------------------|------------------|
| Project C Experimental Fund Projects | Salaries | Travel & Training | Operating Expenses | Logistics Expenses | Cooperative Agreements | To other USGS Centers | AMP Special Burden Rate* | Total |
| | | | | | | | 23.40% | |
| C.5. Experimental Vegetation Treatment Support | \$10,854 | \$0 | \$0 | \$0 | \$5,032 | \$0 | \$2,691 | \$18,576 |
| C.6. Plant physiological responses to experimental flows | \$0 | \$450 | \$400 | \$10,118 | \$0 | \$0 | \$2,567 | \$13,535 |
| C.7. Effects of plants on flow velocity and sand deposition | \$61,596 | \$0 | \$0 | \$15,933 | \$0 | \$0 | \$18,142 | \$95,671 |
| Total Project C | \$72,450 | \$450 | \$400 | \$26,051 | \$5,032 | \$0 | \$23,399 | \$127,782 |

Unfunded Project Element Budget

| Fiscal Year 2025 | | | | | | | | |
|--|-----------------|----------------------|-----------------------|-----------------------|---------------------------|--------------------------|-----------------------------|-----------------|
| Project C Unfunded Project Elements | Salaries | Travel & Training | Operating Expenses | Logistics Expenses | Cooperative Agreements | To other USGS Centers | AMP Special Burden Rate* | Total |
| | | | | | | | 21.86% | |
| C.1.2. Bird habitat collaboration with Navajo Nation | \$15,048 | \$0 | \$300 | \$36,875 | \$0 | \$0 | \$11,385 | \$63,608 |
| C.3.2. Flow scenarios to achieve vegetation objectives | \$0 | \$0 | \$0 | \$0 | \$16,955 | \$0 | \$509 | \$17,464 |
| C.4.2. Biogeomorphic linkages between streamflow, sediment transport, and vegetation composition | \$0 | \$0 | \$0 | \$0 | \$0 | \$0 | \$0 | \$0 |
| Total Project C | \$15,048 | \$0 | \$300 | \$36,875 | \$16,955 | \$0 | \$11,894 | \$81,072 |

| Fiscal Year 2026 | | | | | | | | |
|--|-----------------|----------------------|-----------------------|-----------------------|---------------------------|--------------------------|-----------------------------|-----------------|
| Project C Unfunded Project Elements | Salaries | Travel & Training | Operating Expenses | Logistics Expenses | Cooperative Agreements | To other USGS Centers | AMP Special Burden Rate* | Total |
| | | | | | | | 22.60% | |
| C.1.2. Bird habitat collaboration with Navajo Nation | \$16,102 | \$0 | \$300 | \$38,257 | \$0 | \$0 | \$12,353 | \$67,012 |
| C.3.2. Flow scenarios to achieve vegetation objectives | \$0 | \$0 | \$0 | \$0 | \$20,697 | \$0 | \$621 | \$21,318 |
| C.4.2. Biogeomorphic linkages between streamflow, sediment transport, and vegetation composition | \$8,466 | \$0 | \$0 | \$0 | \$0 | \$0 | \$1,913 | \$10,379 |
| Total Project C | \$24,568 | \$0 | \$300 | \$38,257 | \$20,697 | \$0 | \$14,887 | \$98,709 |

| Fiscal Year 2027 | | | | | | | | |
|--|-----------------|----------------------|-----------------------|-----------------------|---------------------------|--------------------------|-----------------------------|-----------------|
| Project C Unfunded Project Elements | Salaries | Travel & Training | Operating Expenses | Logistics Expenses | Cooperative Agreements | To other USGS Centers | AMP Special Burden Rate* | Total |
| | | | | | | | 23.40% | |
| C.1.2. Bird habitat collaboration with Navajo Nation | \$17,229 | \$0 | \$0 | \$0 | \$0 | \$0 | \$4,032 | \$21,261 |
| C.3.2. Flow scenarios to achieve vegetation objectives | \$0 | \$0 | \$0 | \$0 | \$21,318 | \$0 | \$640 | \$21,958 |
| C.4.2. Biogeomorphic linkages between streamflow, sediment transport, and vegetation composition | \$8,947 | \$0 | \$0 | \$0 | \$0 | \$0 | \$2,094 | \$11,041 |
| Total Project C | \$26,176 | \$0 | \$0 | \$0 | \$21,318 | \$0 | \$6,766 | \$54,260 |

References Cited

- Aguiar, F.C., Segurado, P., Martins, M.J., Bejarano, M.D., Nilsson, C., Portela, M.M., and Merritt, D.M., 2018, The abundance and distribution of guilds of riparian woody plants change in response to land use and flow regulation: *Journal of Applied Ecology*, v. 55, no. 5, p. 2227-2240, <https://doi.org/10.1111/1365-2664.13110>.
- Amlin, N.M., and Rood, S.B., 2002, Comparative tolerances of riparian willows and cottonwoods to water-table decline: *Wetlands*, v. 22, no. 2, p. 338-346, [https://doi.org/10.1672/0277-5212\(2002\)022\[0338:CTORWA\]2.0.CO;2](https://doi.org/10.1672/0277-5212(2002)022[0338:CTORWA]2.0.CO;2).
- Baladrón, A., Bejarano, M.D., and Boavida, I., 2023, Why do plants respond differently to hydropeaking disturbance? A functional approach: *Ecological Indicators*, v. 150, article 110237, <https://doi.org/10.1016/j.ecolind.2023.110237>.
- Baladrón, A., Bejarano, M.D., Sarneel, J.M., and Boavida, I., 2022, Trapped between drowning and desiccation—Riverine plants under hydropeaking: *Science of the Total Environment*, v. 829, article 154451, <https://doi.org/10.1016/j.scitotenv.2022.154451>.
- Batman, H.L., and Merritt, D.M., 2020, Complex riparian habitats predict reptile and amphibian diversity: *Global Ecology and Conservation*, v. 22, e00957, p. 1-10, <https://doi.org/10.1016/j.gecco.2020.e00957>.
- Bedford, A., Sankey, T.T., Sankey, J.B., Durning, L.E., and Ralston, B.E., 2018, Remote sensing of tamarisk beetle (*Diorhabda carinulata*) impacts along 412 km of the Colorado River in the Grand Canyon, Arizona, USA: *Ecological Indicators*, v. 89, p. 365-375, <https://doi.org/10.1016/j.ecolind.2018.02.026>.
- Bejarano, M.D., Jansson, R., and Nilsson, C., 2018a, The effects of hydropeaking on riverine plants—A review: *Biological Reviews*, v. 93, no. 1, p. 658-673, <https://doi.org/10.1111/brv.12362>.
- Bejarano, M.D., Nilsson, C., and Aguiar, F.C., 2018b, Riparian plant guilds become simpler and most likely fewer following flow regulation: *Journal of Applied Ecology*, v. 55, no. 1, p. 365-376, <https://doi.org/10.1111/1365-2664.12949>.
- Blasini, D.E., Koepke, D.F., Bush, S.E., Allan, G.J., Gehring, C.A., Whitham, T.G., Day, T.A., and Hultine, K.R., 2022, Tradeoffs between leaf cooling and hydraulic safety in a dominant arid land riparian tree species: *Plant, Cell & Environment*, v. 45, no. 6, p. 1664-1681, <https://onlinelibrary.wiley.com/doi/abs/10.1111/pce.14292>.

- Burkham, D.E., 1976a, Effects of changes in an alluvial channel on the timing, magnitude, and transformation of flood waves, southeastern Arizona: U.S. Geological Survey Professional Paper 655-K, 25 p., <https://pubs.usgs.gov/publication/pp655K>.
- Burkham, D.E., 1976b, Hydraulic effects of changes in bottom land vegetation on three major floods, Gila River, in southeastern Arizona: U.S. Geological Survey Professional Paper 655J, 14 p., <https://www.usgs.gov/publications/hydraulic-effects-changes-bottom-land-vegetation-three-major-floods-gila-river>.
- Butterfield, B.J., 2009, Effects of facilitation on community stability and dynamics—Synthesis and future directions: *Journal of Ecology*, v. 97, no. 6, p. 1192-1201, <https://doi.org/10.1111/j.1365-2745.2009.01569.x>.
- Butterfield, B.J., Grams, P.E., Durning, L.E., Hazel, J.E., Palmquist, E.C., Ralston, B.E., and Sankey, J.B., 2020, Associations between riparian plant morphological guilds and fluvial sediment dynamics along the regulated Colorado River in Grand Canyon: *River Research and Applications*, v. 36, no. 3, p. 410-421, <https://doi.org/10.1002/rra.3589>.
- Butterfield, B.J., and Palmquist, E.C., 2024a, Inundation tolerance, rather than drought tolerance, predicts riparian plant distributions along a local hydrologic gradient: *Wetlands*, v. 44, no. 6, p. 1-12, <https://doi.org/10.1007/s13157-023-01730-2>.
- Butterfield, B.J., and Palmquist, E.C., 2024b, Divergent physiological responses of hydric and mesic riparian plant species to a Colorado River experimental flow: *Plant Ecology*, v. 225, no. 2, p. 125-133, <https://doi.org/10.1007/s11258-023-01382-6>.
- Butterfield, B.J., Palmquist, E.C., and Yackulic, C.B., 2023, The hydroclimate niche—A tool for predicting and managing riparian plant community responses to streamflow seasonality: *River Research and Applications*, v. 39, no. 1, p. 84-94, <https://doi.org/10.1002/rra.4067>.
- Bywater-Reyes, S., Diehl, R.M., Wilcox, A.C., Stella, J.C., and Kui, L., 2022, A green new balance—Interactions among riparian vegetation plant traits and morphodynamics in alluvial rivers: *Earth Surface Processes and Landforms*, v. 47, no. 10, p. 2410-2436, <https://doi.org/10.1002/esp.5385>.
- Bywater-Reyes, S., Wilcox, A.C., and Diehl, R.M., 2017, Multiscale influence of woody riparian vegetation on fluvial topography quantified with ground-based and airborne lidar: *Journal of Geophysical Research: Earth Surface*, v. 122, no. 6, p. 1218-1235, <https://doi.org/10.1002/2016JF004058>.
- Callaway, R.M., 2007, Positive interactions and interdependence in plant communities: Dordrecht, Netherlands, Springer, <https://doi.org/10.1007/978-1-4020-6224-7>.
- Cavieres, L.A., Brooker, R.W., Butterfield, B.J., Cook, B.J., Kikvidze, Z., Lortie, C.J., Michalet, R., Pugnaire, F.I., Schöb, C., Xiao, S., Anthelme, F., Björk, R.G., Dickinson, K.J.M., Cranston, B.H., Gavián, R., Gutiérrez-Girón, A., Kanka, R., Maalouf, J.-P., Mark, A.F., Noroozi, J., Parajuli, R., Phoenix, G.K., Reid, A.M., Ridenour, W.M., Rixen, C., Wipf, S., Zhao, L., Escudero, A., Zaitchik, B.F., Lingua, E., Aschehoug, E.T., and Callaway, R.M., 2014, Facilitative plant interactions and climate simultaneously drive alpine plant diversity: *Ecology Letters*, v. 17, no. 2, p. 193-202, <https://doi.org/10.1111/ele.12217>.

- Cubley, E.S., Bateman, H.L., Merritt, D.M., and Cooper, D.J., 2020, Using vegetation guilds to predict bird habitat characteristics in riparian areas: *Wetlands*, v. 40, no. 6, p. 1843-1862, <https://doi.org/10.1007/s13157-020-01372-8>.
- D'Amen, M., Mod, H.K., Gotelli, N.J., and Guisan, A., 2018, Disentangling biotic interactions, environmental filters, and dispersal limitation as drivers of species co-occurrence: *Ecography*, v. 41, no. 8, p. 1233-1244, <https://doi.org/10.1111/ecog.03148>.
- Dean, D.J., and Schmidt, J.C., 2011, The role of feedback mechanisms in historic channel changes of the lower Rio Grande in the Big Bend region: *Geomorphology*, v. 126, no. 3-4, p. 333-349, <https://doi.org/10.1016/j.geomorph.2010.03.009>.
- Dean, D.J., and Topping, D.J., 2019, Geomorphic change and biogeomorphic feedbacks in a dryland river—The Little Colorado River, Arizona, USA: *GSA Bulletin*, v. 131, no. 11-12, p. 1920-1942, <https://doi.org/10.1130/B35047.1>.
- Dean, D.J., and Topping, D.J., 2024, The effects of vegetative feedbacks on flood shape, sediment transport, and geomorphic change in a dryland river: Moenkopi Wash, AZ: *Geomorphology*, v. 447, article 109017, <https://doi.org/10.1016/j.geomorph.2023.109017>.
- Diehl, R.M., Wilcox, A.C., Stella, J.C., Kui, L., Sklar, L.S., and Lightbody, A., 2017, Fluvial sediment supply and pioneer woody seedlings as a control on bar-surface topography: *Earth Surface Processes and Landforms*, v. 42, no. 5, p. 724-734, <https://doi.org/10.1002/esp.4017>.
- Durning, L.E., Sankey, J.B., Yackulic, C.B., Grams, P.E., Butterfield, B.J., and Sankey, T.T., 2021, Hydrologic and geomorphic effects on riparian plant species occurrence and encroachment—Remote sensing of 360 km of the Colorado River in Grand Canyon: *Ecohydrology*, v. 14, no. 8, e2344, <https://doi.org/10.1002/eco.2344>.
- Fairley, H.C., 2005, Cultural resources in the Colorado River corridor, in Gloss, S.P., Lovich, J.E., and Melis, T.S., eds., *The state of the Colorado River ecosystem in Grand Canyon—A report of the Grand Canyon Monitoring and Research Center 1991-2004*: U.S. Geological Survey Circular 1282, 177-192 p., <https://pubs.usgs.gov/circ/1282/>.
- Gellis, A.C., Elliott, J.G., and Pavich, M., 2017, Geomorphic processes responsible for decadal-scale arroyo changes, Rio Puerco, New Mexico: *GSA Bulletin*, v. 129, no. 11-12, p. 1660–1680, <https://doi.org/10.1130/B31622.1>.
- Gill, K.M., Goater, L.A., Braatne, J.H., and Rood, S.B., 2018, The irrigation effect—How river regulation can promote some riparian vegetation: *Environmental Management*, v. 61, p. 650–660, <https://doi.org/10.1007/s00267-017-0991-4>.
- González, E., Martínez-Fernández, V., Shafroth, P.B., Sher, A.A., Henry, A.L., Garófano-Gómez, V., and Corenblit, D., 2018, Regeneration of *Salicaceae* riparian forests in the Northern Hemisphere—A new framework and management tool: *Journal of Environmental Management*, v. 218, p. 374-387, <https://doi.org/10.1016/j.jenvman.2018.04.069>.
- Gorla, L., Signarbieux, C., Turberg, P., Buttler, A., and Perona, P., 2015, Effects of hydropeaking waves' offsets on growth performances of juvenile *Salix* species: *Ecological Engineering*, v. 77, p. 297-306, <https://doi.org/10.1016/j.ecoleng.2015.01.019>.

- Grand, J., Meehan, T.D., DeLuca, W.V., Morton, J., Pitt, J., Calvo-Fonseca, A., Dodge, C., Gómez-Sapiens, M., González-Sargas, E., Hinojosa-Huerta, O., Nagler, P., Restrepo-Giraldo, C., Shafroth, P.B., Villagomez-Palma, S., and Wilsey, C.B., 2024, Strategic restoration planning for land birds in the Colorado River Delta, Mexico: *Journal of Environmental Management*, v. 351, article 119755, <https://doi.org/10.1016/j.jenvman.2023.119755>.
- Greet, J., Angus Webb, J., and Cousens, R.D., 2011, The importance of seasonal flow timing for riparian vegetation dynamics—A systematic review using causal criteria analysis: *Freshwater Biology*, v. 56, no. 7, p. 1231-1247, <https://doi.org/10.1111/j.1365-2427.2011.02564.x>.
- Greet, J., Cousens, R.D., and Webb, J.A., 2013, Seasonal timing of inundation affects riparian plant growth and flowering—Implications for riparian vegetation composition: *Plant Ecology*, v. 214, no. 1, p. 87-101, <https://doi.org/10.1007/s11258-012-0148-8>.
- Griffin, E.R., Kean, J.W., Vincent, K.R., Smith, J.D., and Friedman, J.M., 2005, Modeling effects of bank friction and woody bank vegetation on channel flow and boundary shear stress in the Rio Puerco, New Mexico: *Journal of Geophysical Research: Earth Surface*, v. 110, no. F4, <https://doi.org/10.1029/2005JF000322>.
- Griffin, E.R., Perignon, M.C., Friedman, J.M., and Tucker, G.E., 2014, Effects of woody vegetation on overbank sand transport during a large flood, Rio Puerco, New Mexico: *Geomorphology*, v. 207, p. 30-50, <https://doi.org/10.1016/j.geomorph.2013.10.025>.
- Hadley, D.R., Grams, P.E., and Kaplinski, M.A., 2018, Quantifying geomorphic and vegetation change at sandbar campsites in response to flow regulation and controlled floods, Grand Canyon National Park, Arizona: *River Research and Applications*, v. 34, no. 9, p. 1208-1218, <https://doi.org/10.1002/rra.3349>.
- Hazel, J.E., Kaplinski, M.A., Hamill, D., Buscombe, D., Mueller, E.R., Ross, R.P., Kohl, K., and Grams, P.E., 2022, Multi-decadal sandbar response to flow management downstream from a large dam—The Glen Canyon Dam on the Colorado River in Marble and Grand Canyons, Arizona: U.S. Geological Survey Professional Paper 1873, prepared in cooperation with Northern Arizona University, 104 p., <https://doi.org/10.3133/pp1873>.
- Holmes, J.A., Johnson, M.J., and van Riper, C., III, 2005a, Evaluation of the breeding riparian birds monitoring program for the Colorado River ecosystem—final report: Flagstaff, Ariz., U.S. Geological Survey, Colorado Plateau Research Station and Southwest Biological Science Center, submitted to U.S. Department of the Interior, U.S. Geological Survey, Grand Canyon Monitoring and Research Center, cooperative agreement no. 99HQAG0150, 59 p., https://www.researchgate.net/publication/254843216_Evaluation_of_the_Breeding_Riparian_Birds_Monitoring_Program_for_the_Colorado_River_Ecosystem_through_2000.
- Holmes, J.A., Spence, J.R., and Sogge, M.K., 2005b, Birds of the Colorado River in Grand Canyon—A synthesis of status, trends and dam operations effects, *in* Gloss, S.P., Lovich, J.E., and Melis, T.S., eds., *The state of the Colorado River ecosystem in Grand Canyon—A report of the Grand Canyon Monitoring and Research Center 1991-2004*: U.S. Geological Survey Circular 1282, 123-138 p., <https://pubs.usgs.gov/circ/1282/>.

- Hui, F.K.C., 2016, BORAL—Bayesian Ordination and Regression Analysis of multivariate abundance data in R: *Methods in Ecology and Evolution*, v. 7, no. 6, p. 744-750, <https://doi.org/10.1111/2041-210X.12514>.
- Jackson-Kelly, L., and Hubbs, D., 2007, Traditional Hualapai ecological monitoring knowledge monitoring protocols—Adaptive Management Program Technical Working Group meeting: Phoenix, Ariz., April 2-3, 2007, Hualapai Tribe, Department of Cultural Resources, https://www.usbr.gov/uc/progact/amp/twg/2007-04-02-twg-meeting/Attach_13c.pdf.
- Kasprak, A., Sankey, J.B., and Butterfield, B.J., 2021, Future regulated flows of the Colorado River in Grand Canyon foretell decreased areal extent of sediment and increases in riparian vegetation: *Environmental Research Letters*, v. 16, no. 1, p. 1-15, <https://doi.org/10.1088/1748-9326/abc9e4>.
- Kearsley, M.J.C., and Ayers, T.J., 1996, The effects of interim flows from Glen Canyon Dam on riparian vegetation in the Colorado River corridor, Grand Canyon National Park, Arizona—final report: Flagstaff, Northern Arizona University, submitted to Grand Canyon National Park, Grand Canyon Science Center, cooperative agreement no. 8041-8-0002, 702 p., <http://www.riversimulator.org/Resources/GCMRC/Terrestrial/Kearsley1996b.pdf>.
- Kearsley, M.J.C., Cobb, N.S., Yard, H.K., Lightfoot, D.C., Brantley, S.L., Carpenter, G.C., and Frey, J.K., 2006, Inventory and monitoring of terrestrial riparian resources in the Colorado River corridor of Grand Canyon—An integrative approach—final report: Flagstaff, Northern Arizona University, submitted to U.S. Geological Survey, Grand Canyon Monitoring and Research Center, cooperative agreement no. 01-WRAG-0044 (NAU) and 01-WARG-0034 (HYC), 262 p.
- Kominoski, J.S., Shah, J.J.F., Canhoto, C., Fischer, D.G., Giling, D.P., González, E., Griffiths, N.A., Larrañaga, A., LeRoy, C.J., Mineau, M.M., McElarney, Y.R., Shirley, S.M., Swan, C.M., and Tiegs, S.D., 2013, Forecasting functional implications of global changes in riparian plant communities: *Frontiers in Ecology and the Environment*, v. 11, no. 8, p. 423-432, <https://doi.org/10.1890/120056>.
- Lane, E.W., 1954, The importance of fluvial geomorphology in hydraulic engineering—Hydraulic Laboratory Report no. 372: Denver, Colo., U.S. Department of the Interior, Bureau of Reclamation, <https://semspub.epa.gov/work/01/554355.pdf>.
- Lavorel, S., and Garnier, E., 2002, Predicting changes in community composition and ecosystem functioning from plant traits—Revisiting the Holy Grail: *Functional Ecology*, v. 16, no. 5, p. 545-556, <https://doi.org/10.1046/j.1365-2435.2002.00664.x>.
- Le Bouteiller, C., and Venditti, J., 2014, Vegetation-driven morphodynamic adjustments of a sand bed: *Geophysical Research Letters*, v. 41, no. 11, p. 3876-3883, <https://doi.org/10.1002/2014GL060155>.
- López, F., and García, M., 1998, Open-channel flow through simulated vegetation—Suspended sediment transport modeling: *Water Resources Research*, v. 34, no. 9, p. 2341-2352, <https://doi.org/10.1029/98WR01922>.

- Mahoney, J.M., and Rood, S.B., 1998, Streamflow requirements for cottonwood seedling recruitment—An integrative model: *Wetlands*, v. 18, no. 4, p. 634–645, <https://doi.org/10.1007/BF03161678>.
- Manners, R.B., Schmidt, J.C., and Scott, M.L., 2014, Mechanisms of vegetation-induced channel narrowing of an unregulated canyon river—Results from a natural field-scale experiment: *Geomorphology*, v. 211, p. 100-115, <https://doi.org/10.1016/j.geomorph.2013.12.033>.
- Manners, R.B., Schmidt, J.C., and Wheaton, J.M., 2013, Multiscalar model for the determination of spatially explicit riparian vegetation roughness: *Journal of Geophysical Research: Earth Surface*, v. 118, no. 1, p. 65-83, <https://doi.org/10.1029/2011JF002188>.
- Manners, R.B., Wilcox, A.C., Kui, L., Lightbody, A., Stella, J.C., and Sklar, L.S., 2015, When do plants modify fluvial processes? Plant-hydraulic interactions under variable flow and sediment supply rates: *Journal of Geophysical Research: Earth Surface*, v. 120, no. 2, p. 325-345, <https://doi.org/10.1002/2014JF003265>.
- Martin, A.J., 2009, Ayáash Altaas' éi Choo'ínigíí Baa Hane'—The importance of birds to the Navajo: Window Rock, Ariz., Navajo Nation Zoo.
- McCoy-Sulentic, M.E., Kolb, T.E., Merritt, D.M., Palmquist, E., Ralston, B.E., Sarr, D.A., and Shafroth, P.B., 2017, Changes in community-level riparian plant traits over inundation gradients, Colorado River, Grand Canyon: *Wetlands*, v. 37, no. 4, p. 635-646, <https://doi.org/10.1007/s13157-017-0895-3>.
- Melis, T.S., Grams, P.E., Kennedy, T.A., Ralston, B.E., Robinson, C.T., Schmidt, J.C., Schmit, L.M., Valdez, R.A., and Wright, S.A., 2011, Three experimental high-flow releases from Glen Canyon Dam, Arizona—Effects of the downstream Colorado River ecosystem: U.S. Geological Survey Fact Sheet 2011-3012, 4 p., <https://pubs.usgs.gov/fs/2011/3012/>.
- Melis, T.S., Pine, W.E., III, Korman, J., Yard, M.D., Jain, S., and Pulwarty, R.S., 2016, Using large-scale flow experiments to rehabilitate Colorado River ecosystem function in Grand Canyon—Basis for an adaptive climate-resilient strategy, *in* Miller, K.A., Hamlet, A.F., Kenney, D.S., and Redmond, K.T., eds., *Water policy and planning in a variable and changing climate*: Boca Raton, Fla., CRC Press, p. 315–345, <https://pubs.usgs.gov/publication/70178258>.
- Merritt, D.M., 2022, 6.15—Reciprocal relations between riparian vegetation, fluvial landforms and channel processes, *in* Shroder, J.F., ed., *Treatise on Geomorphology (Second Edition)*, v. 6.1: Oxford, Academic Press, p. 269-297, <https://doi.org/10.1016/B978-0-12-818234-5.00001-8>.
- Merritt, D.M., and Bateman, H.L., 2012, Linking stream flow and groundwater to avian habitat in a desert riparian system: *Ecological Applications*, v. 22, no. 7, p. 1973-1988, <https://doi.org/10.1890/12-0303.1>.
- Moran, M.E., Aparecido, L.M., Koepke, D.F., Cooper, H.F., Doughty, C.E., Gehring, C.A., Throop, H.L., Whitham, T.G., Allan, G.J., and Hultine, K.R., 2023, Limits of thermal and hydrological tolerance in a foundation tree species (*Populus fremontii*) in the desert

- southwestern United States: *New Phytologist*, v. 240, no. 6, p. 2298-2311, <https://doi.org/10.1111/nph.19247>.
- Nepf, H.M., 2012, Hydrodynamics of vegetated channels: *Journal of Hydraulic Research*, v. 50, no. 3, p. 262-279, <https://doi.org/10.1080/00221686.2012.696559>.
- Overpeck, J.T., and Udall, B., 2020, Climate change and the aridification of North America: *Proceedings of the National Academy of Sciences*, v. 117, no. 22, p. 11856-11858, <https://doi.org/10.1073/pnas.2006323117>.
- Palmquist, E.C. and Butterfield, B.J., 2024, Getting to know Colorado River riparian plant communities—presentation to the Glen Canyon Adaptive Management Program, Annual Reporting Meeting, January 23, 2024: Phoenix, Ariz., U.S. Geological Survey, Grand Canyon Monitoring and Research Center.
- Palmquist, E.C., Butterfield, B.J., and DiMartini, C., 2024, Riparian plant community metrics—presentation to the Glen Canyon Adaptive Management Program, Annual Reporting Meeting, January 23, 2024: Phoenix, Ariz., U.S. Geological Survey, Grand Canyon Monitoring and Research Center, https://gcdamp.com/images_gcdamp_com/2/2f/MonitoringPoster_20240119_ForBAO_PDF.pdf.
- Palmquist, E.C., Butterfield, B.J., and Ralston, B.E., 2023, Assessment of riparian vegetation patterns and change downstream from Glen Canyon Dam from 2014 to 2019: U.S. Geological Survey Open-File Report 2023–1026, 55 p., <https://doi.org/10.3133/ofr20231026>.
- Palmquist, E.C., Ogle, K., Whitham, T.G., Allan, G.J., Shafroth, P.B., and Butterfield, B.J., 2022, Provenance, genotype, and flooding influence growth and resource acquisition characteristics in a clonal, riparian shrub: *American Journal of Botany*, v. 110, no. 2, e16115, <https://doi.org/10.1002/ajb2.16115>.
- Palmquist, E.C., Ralston, B.E., Merritt, D.M., and Shafroth, P.B., 2018a, Landscape-scale processes influence riparian plant composition along a regulated river: *Journal of Arid Environments*, v. 148, p. 54-64, <https://doi.org/10.1016/j.jaridenv.2017.10.001>.
- Palmquist, E.C., Ralston, B.E., Sarr, D.A., and Johnson, T.C., 2018b, Monitoring riparian-vegetation composition and cover along the Colorado River downstream of Glen Canyon Dam, Arizona: U.S. Geological Survey Techniques and Methods, book 2, chap. A14, 65 p., <https://doi.org/10.3133/tm2A14>.
- Palmquist, E.C., Ralston, B.E., Sarr, D.A., Merritt, D.M., Shafroth, P.B., and Scott, J.A., 2017, Southwestern riparian plant trait matrix, Colorado River, Grand Canyon, Arizona (ver. 2.0, January 2022): U.S. Geological Survey data release, <https://doi.org/10.5066/P974VCDK>.
- Ralston, B.E., 2010, Riparian vegetation response to the March 2008 short-duration, high-flow experiment—Implications of timing and frequency of flood disturbance on nonnative plant establishment along the Colorado River below Glen Canyon Dam: U.S. Geological Survey Open-File Report 2010-1022, 30 p., <https://pubs.usgs.gov/of/2010/1022/>.
- Ralston, B.E., 2011, Summary report of responses of key resources to the 2000 low steady summer flow experiment, along the Colorado River downstream from Glen Canyon Dam,

- Arizona: U.S. Geological Survey Open-File Report 2011-1220, 129 p.,
<https://pubs.usgs.gov/of/2011/1220/of2011-1220.pdf>.
- Ralston, B.E., Starfield, A.M., Black, R.S., and Van Lonkhuizen, R.A., 2014, State-and-transition prototype model of riparian vegetation downstream of Glen Canyon Dam, Arizona: U.S. Geological Survey Open-File Report 2014-1095, 26 p.,
<https://doi.org/10.3133/ofr20141095>.
- Sabo, J.L., Sponseller, R.A., Dixon, M., Gade, K., Harms, T., Heffernan, J., Jani, A., Katz, G., Soykan, C., Watts, J., and Welter, J., 2005, Riparian zones increase regional species richness by harboring different, not more, species: *Ecology*, v. 86, no. 1, p. 56-62,
<https://doi.org/10.1890/04-0668>.
- Sankey, J.B., East, A., Fairley, H.C., Caster, J., Dierker, J., Brennan, E., Pilkington, L., Bransky, N., and Kasprak, A., 2023, Archaeological sites in Grand Canyon National Park along the Colorado River are eroding owing to six decades of Glen Canyon Dam operations: *Journal of Environmental Management*, v. 342, article 118036, p. 1-17,
<https://doi.org/10.1016/j.jenvman.2023.118036>.
- Sankey, J.B., Ralston, B.E., Grams, P.E., Schmidt, J.C., and Cagney, L.E., 2015, Riparian vegetation, Colorado River, and climate—Five decades of spatiotemporal dynamics in the Grand Canyon with river regulation: *Journal of Geophysical Research—Biogeosciences*, v. 120, no. 8, p. 1532-1547, <https://doi.org/10.1002/2015JG002991>.
- Soliveres, S., and Maestre, F.T., 2014, Plant–plant interactions, environmental gradients and plant diversity—A global synthesis of community-level studies: *Perspectives in Plant Ecology, Evolution and Systematics*, v. 16, no. 4, p. 154-163,
<https://doi.org/10.1016/j.ppees.2014.04.001>.
- Spence, J.R., 2006, The riparian and aquatic bird communities along the Colorado River from Glen Canyon Dam to Lake Mead, 1996-2000—final report: Flagstaff, Ariz., National Park Service, Resource Management Division, Glen Canyon National Recreation Area, submitted to U.S. Geological Survey, Grand Canyon Monitoring and Research Center, 283 p.,
<http://www.riversimulator.org/Resources/GCMRC/Terrestrial/Spence2006.pdf>.
- Stevens, L.E., and Ayers, T.J., 1994, The effects of interim flows from Glen Canyon Dam on riparian vegetation along the Colorado River in Grand Canyon National Park, Arizona—draft 1993 annual technical and administrative report: Flagstaff, Ariz., U.S. Department of the Interior, National Biological Service, Colorado Plateau Research Station, submitted to National Park Service, Cooperative Studies Unit, and Northern Arizona University, cooperative agreement no. CA 8021-8-0002, 41 p.
- Stevens, L.E., Schmidt, J.C., Ayers, T.J., and Brown, B.T., 1995, Flow regulation, geomorphology, and Colorado River marsh development in the Grand Canyon, Arizona: *Ecological Applications*, v. 5, no. 4, p. 1025-1039, <https://doi.org/10.2307/2269352>.
- Stevens, L.E., and Waring, G., 1986, Effects of post-dam flooding on riparian substrates, vegetation, and invertebrate populations in the Colorado River corridor in Grand Canyon, Arizona—Terrestrial biology of the Glen Canyon Environmental Studies: Flagstaff, Ariz.,

- Bureau of Reclamation, Glen Canyon Environmental Studies, contract no. IA4-AA-40-01930, GCES 19/87, 175 p.
- Stewart, W., Larkin, K., Orland, B., and Anderson, D., 2003, Boater preferences for beach characteristics downstream from Glen Canyon Dam, Arizona: *Journal of Environmental Management*, v. 69, no. 2, p. 201-211, <https://doi.org/10.1016/j.jenvman.2003.08.001>.
- Stromberg, J.C., 2013, Root patterns and hydrogeomorphic niches of riparian plants in the American southwest: *Journal of Arid Environments*, v. 94, p. 1-9, <https://doi.org/10.1016/j.jaridenv.2013.02.004>.
- Stromberg, J.C., Beauchamp, V.C., Dixon, M.D., Lite, S.J., and Paradzick, C., 2007, Importance of low-flow and high-flow characteristics to restoration of riparian vegetation along rivers in arid south-western United States: *Freshwater Biology*, v. 52, no. 4, p. 651-679, <https://doi.org/10.1111/j.1365-2427.2006.01713.x>.
- Tabacchi, E., Correll, D.L., Hauer, R., Pinay, G., Planty-Tabacchi, A.-M., and Wissmar, R.C., 1998, Development, maintenance and role of riparian vegetation in the river landscape: *Freshwater Biology*, v. 40, no. 3, p. 497-516, <https://doi.org/10.1046/j.1365-2427.1998.00381.x>.
- Tal, M., and Paola, C., 2007, Dynamic single-thread channels maintained by the interaction of flow and vegetation: *Geology*, v. 35, no. 4, p. 347-350, <https://doi.org/10.1130/G23260A.1>.
- Turner, R.M., and Karpiscak, M.M., 1980, Recent vegetation changes along the Colorado River between Glen Canyon Dam and Lake Mead, Arizona: U.S. Geological Survey Professional Paper 1132, 125 p., <https://doi.org/10.3133/pp1132>.
- U.S. Department of the Interior, 2016, Record of Decision for the Glen Canyon Dam Long-Term Experimental and Management Plan final Environmental Impact Statement (LTEMP ROD): Salt Lake City, Utah, U.S. Department of the Interior, Bureau of Reclamation, Upper Colorado Region, National Park Service, Intermountain Region, 196 p., online, http://ltempeis.anl.gov/documents/docs/LTEMP_ROD.pdf.
- U.S. Geological Survey, 2024, Discharge, sediment, and water quality monitoring: Flagstaff, Ariz., Grand Canyon Monitoring and Research Center, online data, https://www.gcmrc.gov/discharge_qw_sediment/.
- Vincent, K., Friedman, J., and Griffin, E., 2009, Erosional consequence of saltcedar control: *Environmental Management*, v. 44, no. 2, p. 218-227, <https://doi.org/10.1007/s00267-009-9314-8>.
- Walker, A.E., Moore, J.N., Grams, P.E., Dean, D.J., and Schmidt, J.C., 2020, Channel narrowing by inset floodplain formation of the lower Green River in the Canyonlands region, Utah: *GSA Bulletin*, v. 132, no. 11-12, p. 2333-2352, <https://doi.org/10.1130/B35233.1>.
- Webb, R.H., Belnap, J., Scott, M.L., and Esque, T.C., 2011, Long-term change in perennial vegetation along the Colorado River in Grand Canyon National Park (1889-2010): *Park Science*, v. 28, no. 2, p. 83-87, <https://irma.nps.gov/DataStore/Reference/Profile/2201709>.

- Wheeler, K.G., Udall, B., Wang, J., Kuhn, E., Salehabadi, H., and Schmidt, J.C., 2022, What will it take to stabilize the Colorado River?: *Science*, v. 377, no. 6604, p. 373-375, <https://doi.org/10.1126/science.abo4452>.
- Yackulic, C.B., Bair, L.S., Eppehimer, D.E., Salter, G.L., Deemer, B.R., Butterfield, B.J., Kasprak, A., Caster, J.J., Fairley, H.C., Grams, P.E., Mihalevich, B.A., Palmquist, E.C., and Sankey, J.B., 2024, Modeling the impacts of Glen Canyon Dam operations on Colorado River resources: Phoenix, Ariz., U.S. Department of the Interior, Bureau of Reclamation, cooperator publication prepared by U.S. Geological Survey, Southwest Biological Science Center, Grand Canyon Monitoring and Research Center, Flagstaff, Ariz., April 2024, 133 p., <https://pubs.usgs.gov/publication/70252976>.
- Zong, L., and Nepf, H., 2010, Flow and deposition in and around a finite patch of vegetation: *Geomorphology*, v. 116, no. 3-4, p. 363-372, <https://doi.org/10.1016/j.geomorph.2009.11.020>.
- Zurell, D., Pollock, L.J., and Thuiller, W., 2018, Do joint species distribution models reliably detect interspecific interactions from co-occurrence data in homogenous environments?: *Ecography*, v. 41, no. 11, p. 1812-1819, <https://doi.org/10.1111/ecog.03315>.

Project D: Effects of Dam Operations and Experimental Vegetation Management for Archaeological Sites

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Project Summary and Purpose

The Colorado River corridor through Glen, Marble and Grand Canyon is both an ecological and cultural landscape reflecting thousands of years of human use and occupation by people of diverse cultural backgrounds. Traces of past human activities are embedded throughout this landscape. These traces are often referred to as archaeological sites or historic properties. For the modern Indigenous people of the Grand Canyon region, however, these sites are much more than historical relics of the past; they are considered to be ancestral homes, places of profound spiritual and cultural significance, tangible footprints of their ancestors that document their deep and continuing spiritual and cultural connections to this landscape, and much more.

Past inventories of the river corridor conducted by the National Park Service in the early 1990s recorded over 475 archaeological sites between Glen Canyon Dam and Separation Canyon 255 miles downstream (Fairley and others, 1994). Over 400 of these sites are considered by the National Park Service and the Bureau of Reclamation to fall within the area of potential effects from dam operations. In Grand Canyon National Park (GRCA), three-quarters of the 362 river-corridor archaeological sites depend on river-derived sand for their geomorphic context; this context, in turn supports native plant communities traditionally used by Indigenous peoples. The vast majority of these cultural sites and associated contributing elements are now deprived of sand resupply in the modern, dam-controlled river system.

Sankey and others (2023) reported that most archaeological sites in GRCA along the Colorado River are eroding and are at increased environmental risk of erosion from six decades of operations of Glen Canyon Dam. The proportion of sites affected by gully processes controlled by the base-level of the Colorado River in Grand Canyon has increased since 2000 (Sankey and others, 2023). The wind-driven supply of river-derived sand, essential for protecting sites, maintaining their geomorphic context, and helping to offset gully erosion has decreased for most archaeological sites since 1973 owing to effects of long-term dam operations on river sediment supply and invasive riparian vegetation expansion on sandbars (Sankey and others, 2023).

These fundamental changes to landscape processes affecting archaeological site context and integrity, limit the ability of the National Park Service (NPS) to achieve environmental management goals to maintain or improve site integrity in-situ and threaten to expose human burials and other culturally significant remains that currently remain buried in the landscape.

Archaeological site monitoring results illustrate some of the negative impacts of human river management and associated gully erosion on site condition and the physical integrity of prehistoric and historic archaeological sites. However, monitoring and research also demonstrate that windblown river sand can help to offset erosion impacts on archaeological site condition (Sankey and others, 2023). Targeted riparian vegetation removal on sandbars may provide an environmental management opportunity to increase windblown sand supply from sandbars to archaeological sites, and thus increase in-situ preservation potential on a site-specific basis, while also helping to maintain the historic geomorphic context of these cultural resources and improve conditions for native, psammophilous (i.e., sand-loving) species. Effectiveness of vegetation management might theoretically be increased when coupled with HFEs to rebuild sandbars, or with periodic low river flows to expose sandbars, which in both cases are the sources of windblown sediment supply. In 2023, for the first time, experimental vegetation management and an HFE were implemented in combination in the same year in Grand Canyon. Without environmental management actions to increase in-situ preservation potential, sites along the Colorado River will likely continue to erode (Figure 1), leaving excavations or other mitigation of individual affected sites (e.g., Thorne, 1991), as the only options for preserving archaeological site information before it is lost, although this approach falls short of stated environmental-management goals and also conflicts with some tribal values.

Long-Term Experimental and Management Plan (LTEMP) Resource Goals

This project addresses the following LTEMP resource goal:

- **Archaeological and Cultural Resources.** Maintain the integrity of potentially affected NRHP-eligible or listed historic properties in place, where possible, with preservation methods employed on a site-specific basis.

The Long-Term Experimental and Management Plan (LTEMP) (U.S. Department of the Interior, 2016a) goal for Archaeological and Cultural Resources is to maintain the integrity of potentially affected National Register of Historic Places (NRHP)-eligible or listed historic properties in place, where possible, with preservation methods employed on a site-specific basis. In addition to this goal, the LTEMP acknowledges that the region “should not be conceptualized merely as multiple discrete or detached archeological sites, traditional cultural properties, and/or sacred places; but rather viewed as interconnected, culturally symbiotic areas of traditional religious and cultural value” (U.S. Department of the Interior, 2016a; page 1).

Project D addresses the LTEMP goal by quantifying changes in the physical condition of river corridor archaeological sites in Grand Canyon and the associated surrounding landscape as a function of 1) dam operations, 2) experimental vegetation management, and 3) interacting natural processes and visitor impacts. While the dam and its operation are not the only sources of change affecting the Colorado River ecosystem (CRe) and associated cultural sites, this project focuses on studying and monitoring dam effects and associated flow and non-flow experimental actions, in keeping with the mandates of the 1992 Grand Canyon Protection Act (GCPA; U.S. Department of the Interior, 1992) and consistent with the monitoring plan for cultural resources developed by GCMRC in 2015 and Reclamation’s 2018 Historic Preservation Plan (Bureau of Reclamation, 2018). The ongoing and experimental dam operations and vegetation management of interest are those that are undertaken under the Record of Decision (ROD) for the Glen Canyon Dam LTEMP final Environmental Impact Statement EIS through 2036 (U.S. Department of the Interior, 2016b). The experimental dam operations and experimental vegetation management are LTEMP flow actions and LTEMP non-flow actions, respectively, with the potential to influence the condition of the cultural landscape.

Throughout this project description, we refer to archaeological sites, cultural landscapes, and cultural resources to maintain consistency with the language used in the LTEMP EIS and the LTEMP goal for “Archaeological Sites and Cultural Resources.” It is important to note, however, that for the Tribes who are traditionally affiliated with the Grand Canyon, these places and the Grand Canyon as a whole have much deeper meaning than the terms “archaeological sites” or “historic properties” convey. In many cases, they are ancestral homes, ancestral burial sites, locations of ceremonial significance, and portals to other worlds. It is recognized that the terminology used in this project description does not do justice to the deep significance these places hold for the indigenous Tribes in the Grand Canyon region. Use of terms common to federal bureaucratic discourse should not be construed as negating or being disrespectful of the views of the various Tribal Nations who attach great significance to these places.

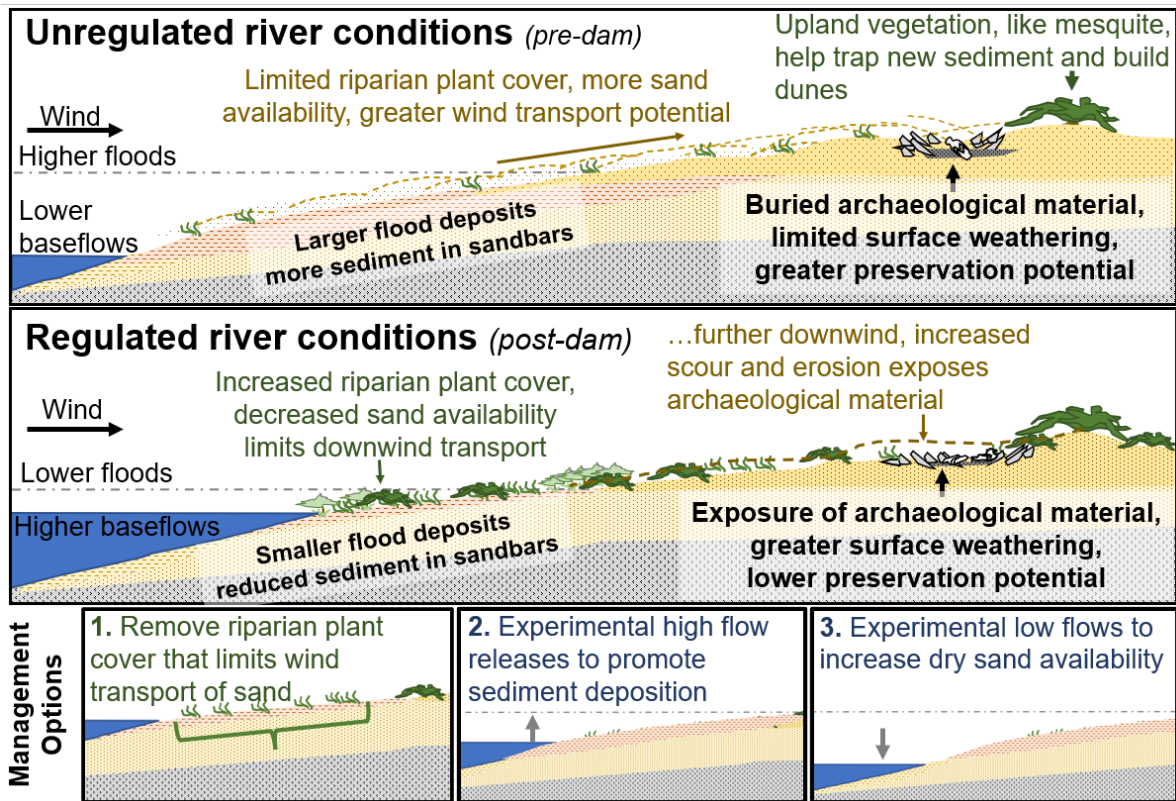


Figure 1. Illustration of effects of river regulation on archaeological sites that are located along the Colorado River downstream of Glen Canyon Dam in Grand Canyon National Park.

Background

Significance and Justification

This project is designed to provide quantifiable information about the effects of Glen Canyon Dam on archaeological sites and other types of cultural resources embedded in the CRe’s sediment-dependent riverine landscape. It will also help to inform decisions that may arise in the future as specific actions are proposed or implemented to protect and maintain cultural resources or modify dam operations. According to the LTEMP ROD (U.S. Department of the Interior, 2016b), the goal for archaeological sites and cultural resources is to “[m]aintain the integrity of potentially affected National Register of Historic Places (NRHP)-eligible or listed historic properties in place, where possible, with preservation methods employed on a site-specific basis.” Additionally, there are other resource goals described in the LTEMP ROD that are directly tied to the goal for cultural resources, such as goals for tribal resources and sediment. For example, the goal for tribal resources is to “[m]aintain the diverse values and resources of traditionally associated Tribes along the Colorado River corridor through Glen, Marble, and Grand Canyon,” while for sediment, the goal is to “[i]ncrease and retain fine sediment volume, area, and distribution in the Glen, Marble and Grand Canyon reaches above the elevation of the

average base flow for ecological, cultural, and recreational purposes.” This project is designed to inform progress towards meeting each of these goals, as well as evaluating predictions about the anticipated effects of the preferred flow regime and other management actions, such as vegetation management, selected through the LTEMP EIS process. For example, the LTEMP ROD states that for cultural resources, the selected alternative (Alternative D) “will result in indirect potential benefits for archaeological sites in the Grand Canyon due to an increase in the availability of sand that will protect site stability...”. Project D is designed to quantitatively evaluate that predicted outcome. Moreover, the LTEMP ROD recommends to “[e]xplore vegetation management to benefit high value recreational beaches and protect vulnerable archaeological sites.” Project D is designed to quantitatively evaluate the outcome of ongoing vegetation management for archaeological sites.

In addition to being responsive to LTEMP goals and predictions, this project is responsive to multiple legal and regulatory mandates. The Grand Canyon Protection Act (U.S. Department of the Interior, 1992) specifically identifies cultural resources as one of the key resource categories that the law is intended to protect. Under GCPA, research and monitoring are required to determine whether the goals of protection, improvement, and/or effective mitigation of adverse effects from Glen Canyon Dam operations are being achieved. The National Historic Preservation Act (NHPA; U.S. Congress, 1966) has somewhat similar obligations as GCPA (Bureau of Reclamation, 2017).

To fulfill its compliance obligations under the GCPA and specifically the NHPA, Reclamation has developed a Programmatic Agreement and a Historic Preservation Plan (HPP; Bureau of Reclamation, 2018). The HPP is intended to guide future monitoring and mitigation activities, thereby fulfilling Reclamation’s Section 106 compliance obligations related to the operation of Glen Canyon Dam and implementation of LTEMP. Among the commitments described in the HPP is an obligation to monitor dam effects using a variety of protocols, including the protocols described in the monitoring plan developed at the request of Reclamation by GCMRC in 2016 and implemented through Project D (described below). Furthermore, as specified in the HPP, results from the GCMRC monitoring project will inform prioritization of future mitigation actions to be carried out under the HPP.

Science and Management

More than two decades of research and monitoring in the CRe have demonstrated that throughout Grand Canyon, numerous archaeological sites and other cultural resources are subject to degradation from erosion processes and visitor impacts (Leap and others, 2000; East and others, 2016; U.S. Department of the Interior, 2016a; Cook and others, 2019; Sankey and others, 2023). Many of these sites occur in landforms, such as fluvial terraces, debris fans, and dunefields, that are located above the elevations inundated by the contemporary dam-regulated river, yet the effects of dam operations have nonetheless accelerated and exacerbated rates of erosion affecting many of these sites (East and others, 2016; U.S. Department of the Interior,

2016a; East and others, 2017; Sankey and others, 2023). Landforms containing cultural resource sites have become disconnected (i.e., no longer receive sediment) from the active river channel downstream of the dam due to the combination of reduced sediment supply in the river, riparian vegetation encroachment, and alterations in flow, which historically supplied sediment during floods but also exposed that sediment for transport by wind during low flows (Sankey and others, 2015; East and others, 2016; Kasprak and others, 2018; Sankey and others, 2018b).

In Grand Canyon National Park (GRCA), three-quarters of the 362 river-corridor archaeological sites depend on river-derived sand for their geomorphic context, and the vast majority of those sites are now deprived of sand resupply in the modern, dam-controlled river system. Sankey and others (2023) showed that most archaeological sites in GRCA along the Colorado River are eroding, and at increased environmental risk of erosion, from six decades of operations of Glen Canyon Dam. The wind-driven supply of river-derived sand, essential for maintaining site geomorphic context by partially re-covering sites and offsetting gully erosion, has decreased for most archaeological sites since 1973 owing to effects of long-term dam operations on river sediment supply and riparian vegetation expansion on sandbars (Sankey and others, 2023). The proportion of sites affected by gulying processes controlled by the local base-level of the Colorado River in Grand Canyon has increased since 2000 (Sankey and others, 2023). These fundamental changes to landscape processes affecting archaeological site context and integrity limit the ability of the National Park Service (NPS) to achieve environmental management goals to maintain or improve site integrity in-situ and threaten to expose human burials and other culturally significant remains that currently remain buried in the landscape.

Archaeological site monitoring results illustrate some of the negative impacts of human river management and gully erosion on site condition and the physical integrity of prehistoric and historic archaeological sites (e.g., Sankey and others, 2023). However, targeted riparian vegetation removal on sandbars implemented by the NPS may provide an environmental management opportunity to increase windblown sand supply from sandbars to archaeological sites, and thus increase in-situ preservation potential on a site-specific basis (Pilkington and others, 2022; Sankey and others, 2023). The effectiveness of vegetation management can theoretically be increased when coupled with HFEs to rebuild sandbars, or with periodic low river flows to expose sandbars, which in both cases are sources of windblown sediment supply (Sankey and others, 2018a; Sankey and others, 2022).

In 2023, for the first time, experimental vegetation management and an HFE were implemented in combination in the same year in Grand Canyon National Park. Without environmental management actions such as these to increase in-situ preservation potential, cultural sites along the Colorado River will likely continue to erode, leaving excavations or other mitigation of individual affected sites (e.g., Thorne, 1991), as the only options for preserving archaeological site information before it is lost, although this approach falls short of stated environmental-management goals.

In Grand Canyon National Park and in Glen Canyon National Recreation Area, terraces and other Holocene fluvial sediment deposits are a substantial component of sediment resources in the ecosystem. They additionally contain widespread evidence of past human activity (e.g., archaeological sites, pollen from cultigens) (Fairley and others, 1994; Hereford and others, 1996; Fairley, 2003; Damp and others, 2007; U.S. Department of the Interior, 2016a). Thus, the ongoing loss of these sediment deposits is contributing to the loss of all sediment-dependent resources in the CRe, including cultural sites and other evidence of human activities (Collins and others, 2016; U.S. Department of the Interior, 2016a). In some places, for example at the large terraces in the Glen Canyon reach of the Colorado River, sediment transfer between the active river channel and upland areas occurs primarily through fluvial erosion and mass failure processes (East and others, 2017). In these areas HFEs have resulted in the erosion of terraces, mainly from the change in pore pressure gradient after flood water recession exposes saturated terrace banks, which then shed material into the river channel (Grams and others, 2007; U.S. Department of the Interior, 2016a). In other areas, sediment connectivity results from aeolian transport of sand from sandbars to dunefields located on terraces or debris fans (Draut and others, 2008; Draut and Rubin, 2008; Draut, 2012; East and others, 2016; U.S. Department of the Interior, 2016a; Sankey and others, 2018a, b). In both situations, the deposition or erosion of sediment can have direct impacts on buried or exposed archaeological sites situated in or on these deposits and can also have indirect impacts such as offsetting rates of erosion from natural processes in the surrounding landscape (Sankey and Draut, 2014; Collins and others, 2016; U.S. Department of the Interior, 2016a; Sankey and others, 2018a). These impacts are in turn interpreted by NPS, tribal resource managers, and LTEMP Cultural Programmatic Agreement (PA) signatories as being either beneficial or deleterious to the cultural resources in question.

The LTEMP EIS (U.S. Department of the Interior, 2016a) relied on a series of conceptual and numerical models to evaluate the likely responses of resources to a suite of proposed alternatives for operating GCD through 2036. The models incorporated past scientific learning and produced generalized predictions about how resource conditions would potentially change under each alternative. The model-based analyses predicted that Alternative D, the alternative ultimately selected for implementation in the LTEMP ROD (U.S. Department of the Interior, 2016b), would result in modest benefits for cultural resources by improving sediment conditions that help to stabilize and preserve archaeological sites *in situ*, while also benefiting natural processes, campsites, riparian vegetation, hydropower, endangered fish, and other resources valued by society.

Over the past decade, GCMRC scientists have developed and refined methods for tracking trends and quantifying rates, amounts, and sources of geomorphic change affecting cultural resources in the CRe (Collins and others, 2008, 2009, 2012, 2014, 2016; East and others, 2016, 2017; Kasprak and others, 2017, 2018, 2021; Sankey and others, 2018a, b; Yackulic and others, 2024). These methods are well suited to evaluating whether the predictions of resource improvement in the LTEMP occur through 2036 (U.S. Department of the Interior, 2016a, b).

Specifically, the methods can be used to evaluate whether changes in operations improve sediment supply to archaeological sites and the associated landforms in which these sites are embedded, and whether such changes in turn result in a reduction of erosion rates and improved preservation of the physical attributes that are necessary to maintain site integrity under the NHPA.

The LTEMP EIS (U.S. Department of the Interior, 2016a) identifies river terraces, specifically in the Glen Canyon National Recreation Area (GLCA) reach, as being vulnerable to erosion and degradation from HFEs which are otherwise intended to distribute sediment throughout the Colorado River downstream of the Paria River (see also Grams and others, 2007). DOI agencies and some tribal resource managers have identified a need for quantifying the effects of dam operations on the erosion of terraces and other river sediment deposits in Glen, Marble, and Grand Canyons by determining erosion rates during the approximately two decades since the implementation of the previous ROD (U.S. Department of the Interior, 1996) with a river flow regime of episodic controlled floods and restricted hydropeaking (U.S. Department of the Interior, 2016b).

The LTEMP ROD (U.S. Department of the Interior, 2016b, subsection 6.4.) also identifies vegetation management as a non-flow action to assist with cultural site protection. Accordingly, GCMRC is providing science information used by NPS and some tribal partners to manage woody riparian vegetation at individual sandbars in order to increase campsite area and also to increase the amount of river sand that is transported by wind and deposited on adjacent dunefields and archaeological sites. GCMRC's ongoing program for monitoring the effects of dam operations on the geomorphic condition of archaeological sites is well-suited for monitoring the vegetation management experiments and for quantifying the effectiveness of the treatments.

Proposed Work

Project Element D.1. Monitoring the Effects of Dam Operations on Archaeological Sites (Modified Study)

During FY 2025-27, GCMRC will continue long-term monitoring of archaeological sites using lidar to report on proposed *LTEMP Cultural Resources Performance Metric 1.2 Lidar Topographic Change Detection*. The purpose of this monitoring is to quantify the effects of dam operations and other factors on the geomorphic condition of a sample of archaeological sites in the Colorado River corridor in Grand Canyon National Park and Glen Canyon National Recreation Area that are within the area of potential effect of Glen Canyon Dam operations. Geomorphic changes are determined from ground-based lidar topographic surveys (also termed terrestrial laser scanning, TLS).

Science Questions

- How do dam operations, HFEs and other LTEMP flow actions impact the LTEMP resource goal for archaeological and cultural resources of preservation in place?

Hypotheses

- HFEs increase the resupply of river sand to archaeological sites in the river corridor and offset erosion, thus increasing the probability of achieving the LTEMP resource goal of preservation in place.
- Vegetation and biological soil crust cover within archaeological sites that are not resupplied with sediment from HFEs help to reduce erosion and increase the probability of achieving the LTEMP resource goal of preservation in place.

Methods

Lidar surveys use multiple scan positions placed outside of culturally sensitive areas to collect hundreds of millions of topographic measurements within and surrounding archaeological sites. These surveys are georeferenced to the established U.S. Geological Survey geodesic network and repeated at individual sites according to an established three-year rotational schedule (Collins and others, 2008, 2009, 2012, 2014, 2016; East and others, 2016, 2017; Sankey and others, 2018a, b, 2023; Caster and others, 2022). These repeated measurements are then used to calculate topographic changes resulting from erosion and deposition of river-sourced sediment (Figure 2).

The total sample size is 40 archaeological sites, with 12-14 sites monitored each year of the three-year work plan. The sample of 40 sites represents approximately 8% of the entire known population of river corridor sites in the CRe. We selected the sample of 40 sites for lidar measurements from the entire population of river corridor sites using two site classification systems (East and others, 2016, 2017) that characterize the extent to which each site is 1) degraded by gully erosion, and 2) positioned within the landscape to be resupplied with sand transferred by wind from adjacent sandbars. During the FY 2025-27 TWP, we will revisit all 40 sites at least once, conduct lidar surveys, quantify changes in geomorphic condition, and relate any changes that are detected to dam operations, and specifically, to the occurrence and timing of HFEs.

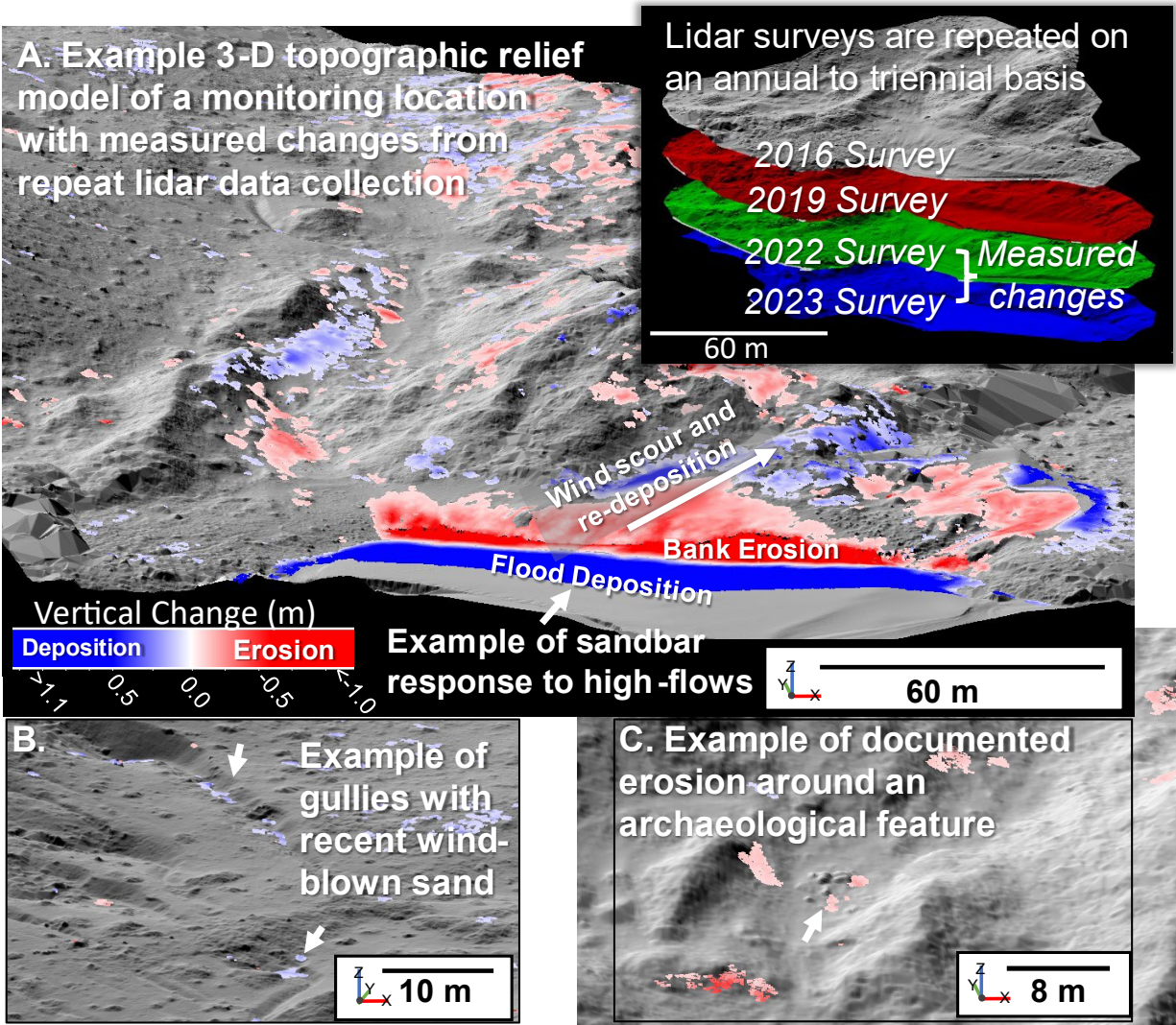


Figure 2. Example of repeat ground-based lidar and geomorphic change detection at a Grand Canyon National Park River Corridor monitoring location. A. 5-cm resolution topographic relief model with measured geomorphic changes (red = erosion; blue=deposition; gray = no significant change) between ground-based lidar surveys collected in April 2023 and May 2022. Labeled flood deposition and bank erosion represent significant geomorphic changes from the April 2023 High-Flow Experiment (HFE) conducted days before the 2023 survey. B. A subset of the 5-cm resolution topographic relief model with gully development demonstrating how wind-blown sand originating from upwind sandbars is partially ameliorating gully erosion through sand deposition. C. A subset of the 5-cm resolution topographic relief model with an archaeological feature. Ground-based lidar provides detailed measurements of archaeological components and repeat surveys permit assessment of changes in site and feature conditions.

Anticipated Use of Data

These monitoring data will be used to quantify how dam operations, HFEs and other LTEMP flow actions impact the geomorphic condition of archaeological sites and impact the LTEMP resource goal for archaeological and cultural resources of preservation in place.

These monitoring data will also be leveraged, as described in Project Element D.3 below, to evaluate effects of LTEMP non-flow actions, such as the experimental vegetation management implemented by NPS, as well as alternative vegetation management strategies, such as Hopi traditional dryland farming knowledge and soil management practices for achieving cultural resource preservation goals.

There are three additional collaborative activities that data from this ongoing project element will contribute to.

The first collaborative activity is that GCMRC and NPS will conduct an analysis of archaeological site monitoring data acquired by both agencies for sites within the area of potential effect of Glen Canyon Dam operations. NPS reports on proposed *LTEMP Cultural Resources Goal Performance Metric 1.1 Site Integrity* through their projects 5.c and 5.d in the BOR TWP titled NPS Cultural Resource Monitoring – Grand Canyon and Glen Canyon, respectively. USGS reports on proposed *Performance Metric 1.2 Lidar Topographic Change Detection* and *Performance Metric 1.3 Fluvial Sediment Connectivity and Drainage Classifications*. Analysis will identify what the relationships are between the monitoring information acquired by NPS to determine site integrity and the data acquired by GCMRC to measure geomorphic and site classification changes.

The second collaborative activity is that an interdisciplinary component of Project Element L.1, with Projects B and D, will explore utility of existing remotely sensed data for measuring system-wide changes in high-elevation sand deposits. That interdisciplinary work will leverage the long-term lidar monitoring topographic change detection results acquired by Project Element D.1 as a basis for checking the accuracy of high elevation sand estimates using photogrammetric data.

The third collaborative activity is that modeling efforts funded externally (i.e., not funded by the GCDAMP) through the USGS Mendenhall Post-doctoral Fellow program will model river sand transport and high-elevation sand deposition with specific consideration to archaeological site preservation potential. The Mendenhall program modeling project will leverage long-term monitoring data acquired by Project Element D.1.

Finally, during FY 2025-27, GCMRC Project Element D.1 will not monitor changes to archaeological site classifications to report on proposed *LTEMP Cultural Resources Performance Metric 1.3 Fluvial Sediment Connectivity and Drainage Classifications*; that metric is reported on an approximately decadal time interval and will be completed in a subsequent work plan.

Outcomes and Products

- Annual presentations and reports on the status of the monitoring conducted under Project Element D.1. Submit to the GCDAMP at the end of Fiscal Years 2025, 2026, and 2027.

Project Element D.2. Monitoring Landscape-Scale Ecosystem Change with Repeat Photography (Ongoing Study)

Repeat photography provides a powerful means of documenting and communicating ecological and geomorphic changes to landscapes over decadal timescales. It has been used for this purpose by scientists from a variety of disciplines for more than a century (Webb and others, 2010). In FY 2015-17, GCMRC initiated a pilot effort to monitor vegetation and geomorphic changes in the riparian zone using repeat photography. The initial results of this pilot photo-matching effort proved to be highly informative and useful for a variety of GCMRC projects: not only do the matched images visually document and illustrate dramatic changes in river corridor vegetation, they also document the ongoing loss of open sand areas throughout the river corridor as well as geomorphic changes to shorelines, campsites, and the river corridor as a whole (Figure 3; 4). This information is useful for reconstructing the pre-dam conditions under which archaeological sites and cultural landscapes existed prior to emplacement of Glen Canyon Dam. Furthermore, supplementary data that is being collected about the specific plant species within the modern photo views, compared to the historical views, provides a site-specific inventory of the plant species growing in the vicinity of archaeological sites today as compared to the past, which in turn informs us about the changing attributes of the cultural landscape in which the sites are embedded. This information also informs on the distribution and abundance of native plant species that are of traditional importance to the Indigenous peoples of the Colorado Plateau. Therefore, GCMRC has continued to acquire high-quality, high-resolution matches of historical imagery during subsequent years, relying heavily on volunteer labor and leveraging logistical support from Project Element D.1 to accomplish the field work, with post-field work photo-processing accomplished exclusively by volunteer labor.

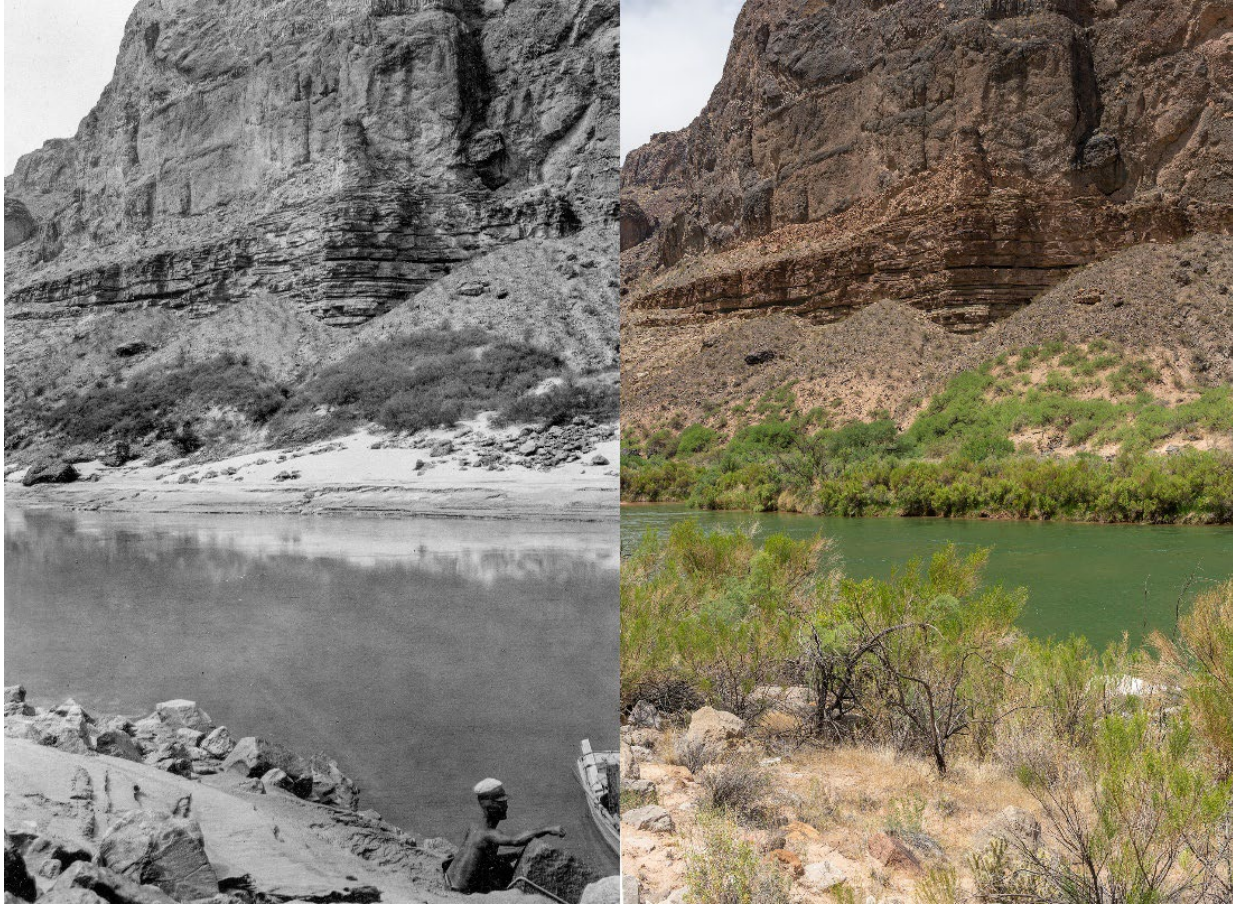


Figure 3. Example of a photo-match from 1923 (left) compared to 2021 (right). Note the wide, gently sloping, unvegetated sandy shoreline in 1923 compared to the nearly vertical, heavily vegetated shoreline in 2021. Also note significant thinning of the mesquite thicket above the pre-dam old high-water line. Stake No. 1157, River mile 179.8, taken from left bank, looking across and downstream to opposite shoreline. Left photo taken by E. C. LaRue September 18, 1923; right photo by A. H. Fairley, May 19, 2021.

Science Questions

- How has riparian vegetation encroachment since dam closure affected the availability of open sand source areas that formerly served to cover and protect archaeological sites in the CRe?
- Does pre-dam riparian vegetation cover within the old high-water zone vary through time? Specifically, do historical photos taken during drought periods characterized by lower annual flows show more riparian cover compared with photographs taken during periods characterized by wetter conditions and higher average annual flows?
- How has the distribution and abundance of plants of traditional importance to Native peoples of the Colorado Plateau changed over time, and specifically, how have they changed as a result of dam operations?

- How has the composition and density of riparian vegetation cover changed during the 50+ years since dam closure?
- Are specific patterns of vegetation encroachment evident in the historical photographic record, and if so, are they indicative of natural successional processes or are they more reflective of changes in dam-controlled flow regimes?
- How has the geomorphology of the river shoreline changed during the 50+ years since dam closure, and are discernible patterns or trends in geomorphic change evident from the historical photographic record?

Methods

During FY 2025-27, GCMRC will continue compiling a photographic record and associated database of ecological and geomorphic changes affecting the landscape and archaeological sites within the river corridor using well-established repeat photography methods (see Turner and Karpisak, 1980; Webb, 1996; Webb and others, 2010, 2011; Scott and others, 2018). To date, photo-matching efforts have focused on replicating photographs taken in 1923 during the USGS Birdseye Expedition (Boyer and Webb, 2007) and images taken 50 years later (in 1973) during an NPS-sponsored campsite inventory project (Weeden and others, 1975). We have also re-matched numerous photographs taken during the 1889-1890 Stanton expedition. Over the next three years, we plan to fill in temporal gaps in the photographic record, focusing initially on matching approximately 50 black-and-white images taken by Barry Goldwater during his 1940 river trip through Grand Canyon.

Another important photograph collection that we plan to match was compiled by Bill Belknap during a 1964 low-water trip through Grand Canyon. The Goldwater images will be valuable to match because they were taken by a highly skilled photographer following a period of lower flows and lower magnitude floods than the earlier Birdseye and Stanton images; therefore, they can provide a useful comparison of pre-dam riparian vegetation conditions closer to the time of dam construction and under somewhat different flow conditions compared to photographs from 1890 and 1923. The Belknap photos, also taken by a skilled photographer, provide a visual record of the river corridor immediately after the dam was completed, during exceptionally low water conditions. Neither the Goldwater images nor the Belknap images have been previously matched, to our knowledge.



Figure 4. Examples of vegetation encroachment on former campsites between 1973 and 2022. Top photos taken by unknown photographer in July 1973, bottom photos taken by A.H. Fairley in May 2022. Left photographs: Stake No. 5623, River Mile 170.5, left bank, looking upstream. Right photographs: Stake No. 5631, River Mile 185.1, left bank, looking upstream.

In FY 2025-27, we will continue to create high quality, accurate matches of these and other historical images to provide a high resolution, detailed visual record of decadal-scale ecosystem changes that can be used and analyzed by a variety of monitoring projects for years into the future. In addition to matching images, we will continue to collect detailed information on species-level vegetation change and geomorphic changes within each matched view.

As in the past, matching of these images will occur in conjunction with previously scheduled GCMRC research and monitoring trips (e.g., element D.1) and will be heavily reliant on volunteer labor to minimize project costs.

Anticipated Use of Data

Matches of historical photographs acquired during previous years are already being used to illustrate journal articles and reports related to projects A, B, C and D. They are also being used in presentations and for general educational purposes to illustrate how Glen Canyon Dam has altered environmental conditions in the downstream Colorado River corridor.

In the future, after the image matches and vegetation and geomorphic data have been acquired, we will qualitatively and quantitatively analyze and ultimately publish the results of this work in one or more peer-reviewed publications. We anticipate that some of the qualitative analysis of geomorphic change evident in the photo-matches will be incorporated into Project C.4.1. We will also make the imagery available to researchers, stakeholders, and the public through GCMRC's website.

Outcomes and Products

- Annual summary of photographs matched each year.

Project Element D.3. Evaluating Effects of LTEMP Non-Flow Actions and other Experimental Vegetation Management on Archaeological Sites (Modified Study)

GCMRC will collaborate with NPS, the Hopi Tribe, and any other parties that express interest, to study effects of experimental vegetation management, with a specific focus on LTEMP non-flow actions that may affect archaeological site preservation. There are three components of this collaborative work; these components are described in more detail below, under the heading “Anticipated Uses of Data”.

Science Questions

- How do LTEMP non-flow actions implemented by the NPS – such as the experimental reduction of riparian vegetation growing on sandbars that blocks windblown transport of river sand – affect archaeological site preservation?
- Can improved experimental management actions not previously implemented in the current suite of LTEMP non-flow actions increase archaeological site preservation? For example, can actions based on different vegetation removal strategies, strategic plantings, or sediment capture using minimally invasive methods or traditional dryland farming knowledge and soil management practices of Hopi or other Indigenous people of the Grand Canyon region, improve archaeological site preservation?

Hypothesis

- Sediment transfer at the experiment sites is greater under the combined effects of vegetation management actions followed by an annual HFE.

Methods

There are three components of this collaborative work. GCMRC will evaluate each component using ground-based lidar measurements from Project Element D.1. These measurements not only include detailed observations of the physical landscape, but also the biological landscape.

We use iterative height filtering (Caster and others, 2022, 2024) to separate ground and vegetation measurements that permit evaluation of NPS treatments on the physical and biological environments (Figure 5; 6).

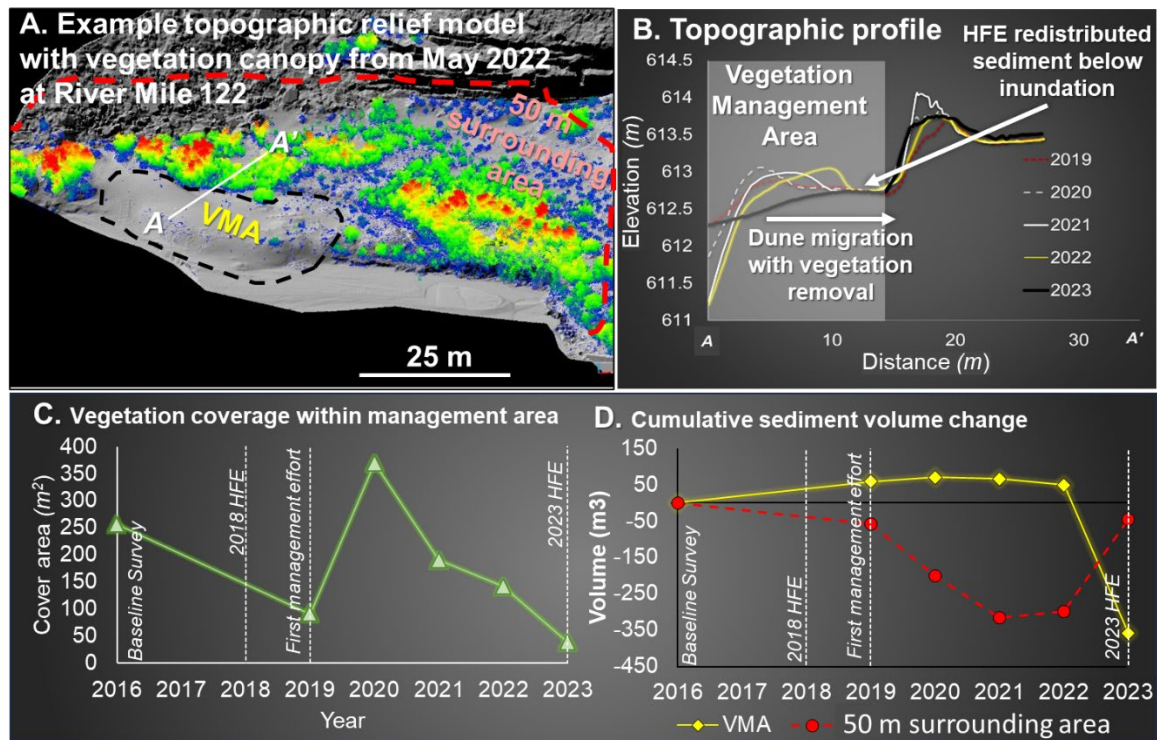


Figure 5. Example of a vegetation management site near river mile 122. A. 5-cm resolution topographic relief model (gray) with vegetation canopy height model (blue = <0.1m; green = <1m; red = >5m). The vegetation management area (VMA) and a 50 m buffer used for summarizing results (Panels C; D) are provided for reference. B. 2-D profile of line A-A' demonstrating dune crest movement during management actions. The topographic trends changed below the area of inundation with the most recent High-Flow Experiment (HFE; April 2023). C. Plot of changes in vegetation area within the VMA between 2016 (baseline year) and 2023 (most recent survey). Vegetation management efforts have been carried out almost annually since 2019. D. Plot of cumulative changes in sediment storage (net volume change) since 2016 for the VMA and the surrounding area, representing direct and downwind effects of vegetation management. Note that at this location, the most significant changes co-occurred with the 2023 high-flow experiment.

Anticipated Use of Data

There are three components of this collaborative work. The first component will continue evaluating effects of ongoing NPS experimental management at six pilot study sites where NPS removes invasive vegetation annually on river sandbars as an experimental LTEMP non-flow

action (Pilkington and others, 2022; Sankey and others, 2023). The Lees Ferry Paria Beach restoration project and associated downwind archaeological sites will be added as a seventh monitoring location in FY 2025-27.

The specific research question that the data will be used to evaluate is whether removal or reduction of riparian vegetation barriers located between river sandbars and archaeological sites can measurably increase the resupply of aeolian sediment to archaeological sites.

The second component will apply lessons learned during FY 2021-24 to propose and evaluate improved experimental management actions based on different vegetation removal strategies, strategic plantings, and sediment capture using minimally invasive methods. GCMRC seeks to collaborate with NPS, the Hopi Tribe, and other interested parties on this work. Specifically, through collaboration with the Hopi Tribe, we will explore the applicability of traditional dryland farming knowledge and soil management practices for achieving cultural resource preservation goals.

The third component will explore site restoration potential relative to degree of sandbar vegetation encroachment. The pilot sites evaluated (described in the first component above) are all associated with sandbars in early stages of vegetation encroachment. In FY 2025-27, data will be used to evaluate additional study sites exhibiting much later vegetation encroachment stages to determine whether they can be effectively restored using similar experimental management techniques or whether different approaches to vegetation management may be warranted.

Note that in the FY 2021-24 work plan, the precedent to the collaborative work proposed in this element was supported by GCMRC Project D.1 to fund USGS efforts, and BOR Project C.7 to fund NPS efforts. For FY 2025-27, a similar funding strategy is proposed for USGS and NPS, but in this work plan, we (GCMRC) have broken out the vegetation management experimental study (this Project Element D.3) from the long-term monitoring Project Element (D.1) to clarify the different objectives involved with each study. For FY 2025-27 the related BOR TWP project is titled Project 4.d Experimental Vegetation Management – Grand Canyon.

Outcomes and Products

- Annual presentations and reports on the status of the monitoring conducted under Project Element D.3. Submit to the GCDAMP at the end of Fiscal Years 2025, 2026, and 2027.

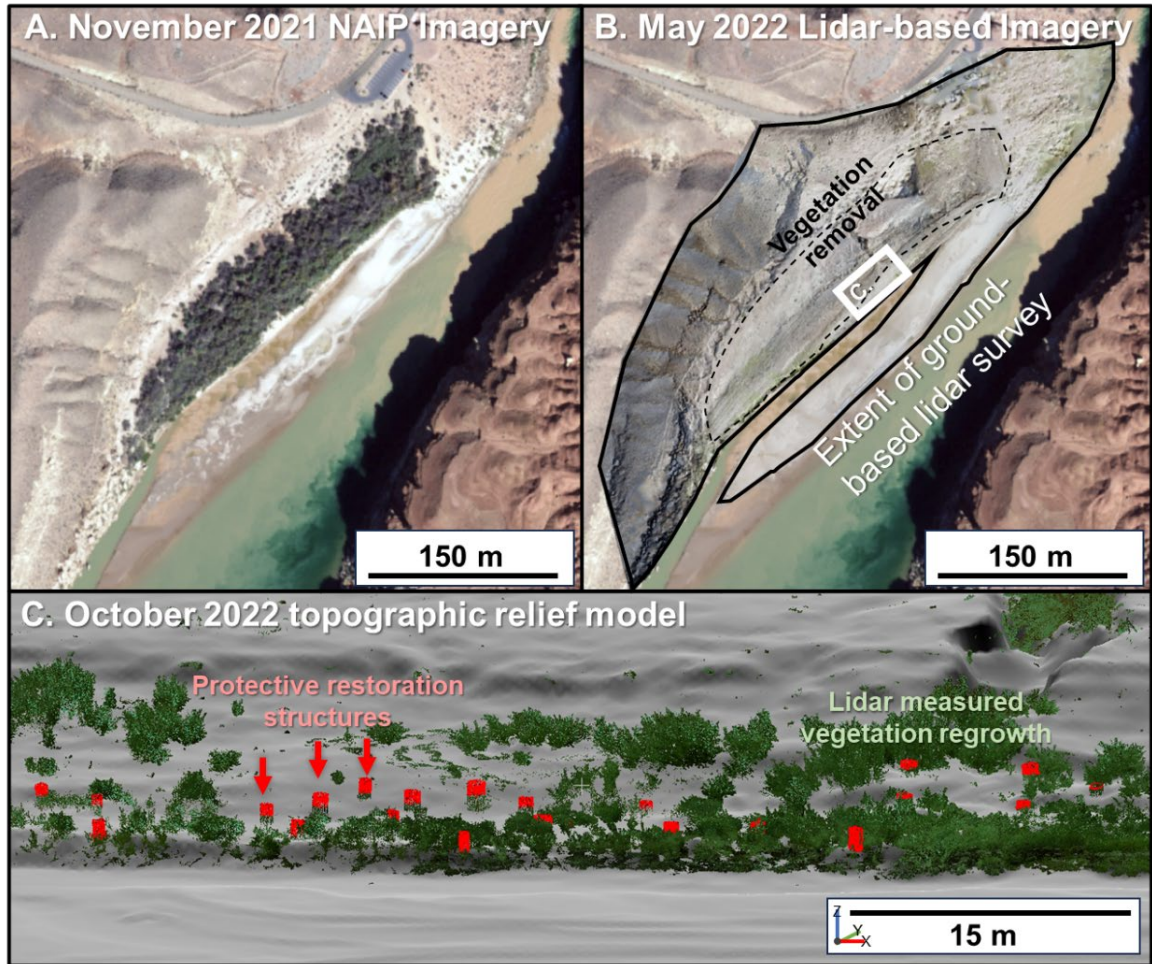


Figure 6. The Paria Beach restoration area. A. National agricultural imagery program (NAIP) aerial photography collected in November 2021 (courtesy of USGS). Note the dense woody vegetation near the center of the imagery. B. A partial “pseudo aerial imagery” dataset developed from photo colored ground-based lidar data collection overlaying the NAIP 2021 imagery. Note that during the May 2022 lidar survey, much of the woody vegetation had been removed. C. A portion of a 50-cm resolution topographic relief model from ground-based lidar collected in October 2022. Note that lidar measurements captured fencing installed to protect native plant restoration efforts as well as additional native and invasive plant re-establishment.

Project Element D.4. Pilot Study to Evaluate Potential to Extract Cultural and Ecological Information from Colorado River Deposits using eDNA, Phytoliths, and Pollen (New Study; Unfunded)

In the GCDAMP, past studies of the sediment resource have mainly focused on the physical dynamics of particles ranging in size from clay to boulders, with an emphasis on tracking the storage and redistribution of fine sediment (e.g., sand, silt, and clay) within and immediately adjacent to the active river channel; however, the sediment resource is much more than just a physical attribute of the CRE.

It also serves as a vehicle for nutrient cycling, provides a substrate for plants, influences the types and distribution of vegetation and wildlife habitat in the river corridor, and serves as a matrix for preserving archaeological sites *in situ*.

In addition, the pre-dam Colorado River sediments are an untapped and largely unexplored repository of information about the pre-dam ecology and cultural history of the river corridor. Information embedded in pre-dam sedimentary deposits has the potential to inform on the characteristics and attributes of the prehistoric cultural landscapes which formed the original context for the hundreds of archaeological sites evident in the river corridor today. These cultural landscapes remain largely unstudied and undocumented, especially in terms of the prehistoric cultural activities and attributes that shaped the river corridor environment through time, including the agricultural fields and associated features that are believed to have been central aspects of the prehistoric cultural landscape in eastern and central Grand Canyon between ca. 3000 and 800 years ago.

Science Questions

- Can ancient eDNA be successfully extracted from pre-dam sedimentary deposits, and if so, what new or corroborating information can eDNA provide about the prehistoric and historic cultural landscape and terrestrial ecology of the pre-dam Colorado River corridor at various points in the past?
- How do the results of eDNA analysis of past environments in the CRe compare with, complement, or contradict information derived from analyzing just pollen or phytoliths from pre-dam sedimentary deposits of the same age?

Methods

In the past two decades, eDNA has become an increasingly important scientific tool for eliciting information about past and present environments and their associated biodiversity; indeed, according to some researchers, it has “revolutionized our knowledge of biogeography” (Pedersen and others, 2015). While the methods and applications of eDNA research are continuing to evolve, ancient eDNA has already proved useful for documenting large-scale ecological changes associated with the Pleistocene-Holocene transition, including extending the persistence of certain species such as woolly mammoth (*Mammuthus primigenius*) in Alaska by several thousand years; documenting vegetation changes spanning 50,000 years; identifying the presence of rare, and in some cases, extinct species; and shedding additional light on many other topics pertaining to ancient environments and the archaeological record (Pedersen and others, 2015).

In FY 2025-26, we propose to collaborate with the NPS and interested Tribes to undertake a pilot study to examine the types of environmental information that potentially can be extracted from pre-dam sedimentary deposits, with a focus on extracting ancient eDNA for the purpose of characterizing the prehistoric vegetation community and cultural landscape at various points in

the past. In addition, we intend to use this pilot study to determine whether eDNA can be used to document the presence of ancient horticulture fields and identify specific cultigens that were grown in the river corridor in the past.

Previous studies using standard palynological methods have produced evidence of maize agriculture possibly dating back more than 3,000 years ago; if this evidence can be substantiated through independent methods, such as eDNA, it would demonstrate that the Colorado River corridor in Grand Canyon contains some of the earliest evidence for maize horticulture anywhere on the Colorado Plateau. Other previous studies have produced evidence of cotton cultivation in association with ~1,000-year-old habitation sites, indicating that Grand Canyon may also have been one of the earliest sources for locally-grown cotton on the Colorado Plateau. Despite the pollen evidence, however, current knowledge of the extent and locations of past horticultural activities in the CRe is essentially non-existent.

This study has the potential to illuminate several aspects of the archaeological and ecological record embedded within the pre-dam sedimentary deposits in Grand Canyon that have never been previously studied. To accomplish this pilot study, we propose to sample pre-dam river sediments at three locations in eastern Grand Canyon. These locations will be selected based on having appropriate characteristics for horticulture in the past and being most likely to preserve ancient eDNA, e.g., very fine-grained, water-laid sediments deposited preferably in an anaerobic environment without evidence of post-depositional disturbance.

We propose to vertically core the selected deposits or if the deposits are exposed in profile, we will core them horizontally, following established protocols to minimize the possibility of contamination with modern DNA. We will collect several samples from each core.

The cores will be sampled for eDNA as well as pollen and phytoliths, so that the results of all three methods can be compared and to ensure that we can recover environmental information from pollen and phytoliths, even if eDNA is not well-preserved. In addition, we will date the samples using either a radiocarbon dating method (if organic material is present in the sediment) or Optically Stimulated Luminescence (OSL).

Anticipated Use of Data

This study will identify the types of environmental information that potentially can be extracted from pre-dam sedimentary deposits, with a focus on characterizing the prehistoric vegetation community and cultural landscape at various points in the past. The study will also help determine whether eDNA can be used to document the presence of ancient horticulture fields and identify specific cultigens that were grown in the river corridor in the past. The results are anticipated to illuminate aspects of the archaeological and ecological record embedded within the pre-dam sedimentary deposits in Grand Canyon that have never been previously studied.

Outcomes and Products

- FY 2027: Presentation and report/journal article summarizing results of pilot project.

Project Element D.5. Monitoring Petroglyphs and Pictographs with Photogrammetry and Lidar (New Study; Unfunded)

The project element *Monitoring Petroglyphs and Pictographs with Photogrammetry and Lidar* is in the unique position of being recommended for funding by all five of the tribes (at time of this draft). We learned of Pueblo of Zuni's support most recently in a letter sent to GCMRC. We knew of Hopi, Hualapai, Navajo, and Southern Paiute Consortium's support prior to the July 2024 Technical Work Group Meeting.

In response to concerns by tribal members of the GCDAMP and at the request of the LTEMP Cultural Programmatic Agreement (PA) signatories, GCMRC will continue work to evaluate potential hazards to petroglyph and pictograph sites and collect data to monitor potential changes. This work will expand from an FY 2023 pilot study – which was unfunded, but GCMRC undertook at the request of the LTEMP Cultural PA in response to a recommendation from Zuni and NPS – which demonstrated the utility of photogrammetry and lidar for these purposes at a single petroglyph site (C:06:0005; “Supai Man”).

Science Questions

- Can repeat LiDAR scanning or photogrammetry at petroglyph or pictograph locations quantify changes in the surfaces owing to natural or other causes?
- Can applications of the LiDAR or photogrammetry methods help to address concerns of individual Tribes and other signatories to the LTEMP Cultural Programmatic Agreement (PA) about the occurrence, rates, and causes of degradation of petroglyph or pictograph sites?

Hypotheses

- Ground-based lidar surveys combined with photogrammetry provide sufficient detail to detect sub-centimeter changes in petroglyph and pictograph panel surfaces.
- Ground-based lidar, photogrammetry, or both improve documentation and analysis of pictograph panels, preserving baseline data for current and future monitoring programs.
- Pictograph form and compositional characteristics are detectable from ground-based lidar near-infrared reflectance.

Methods

Like Project Element D.1, this work uses ground-based lidar but collects measurements at a finer resolution allowing for mm-scale characterization of the panel (Figure 7). Additionally, we use a Digital Single Lens Reflex (DSLR) camera to develop a true photogrammetric surface model (Figure 7) using Agisoft's Metashape software. There are numerous sites and structural remains for which these monitoring procedures could be applied throughout the Colorado River corridor.

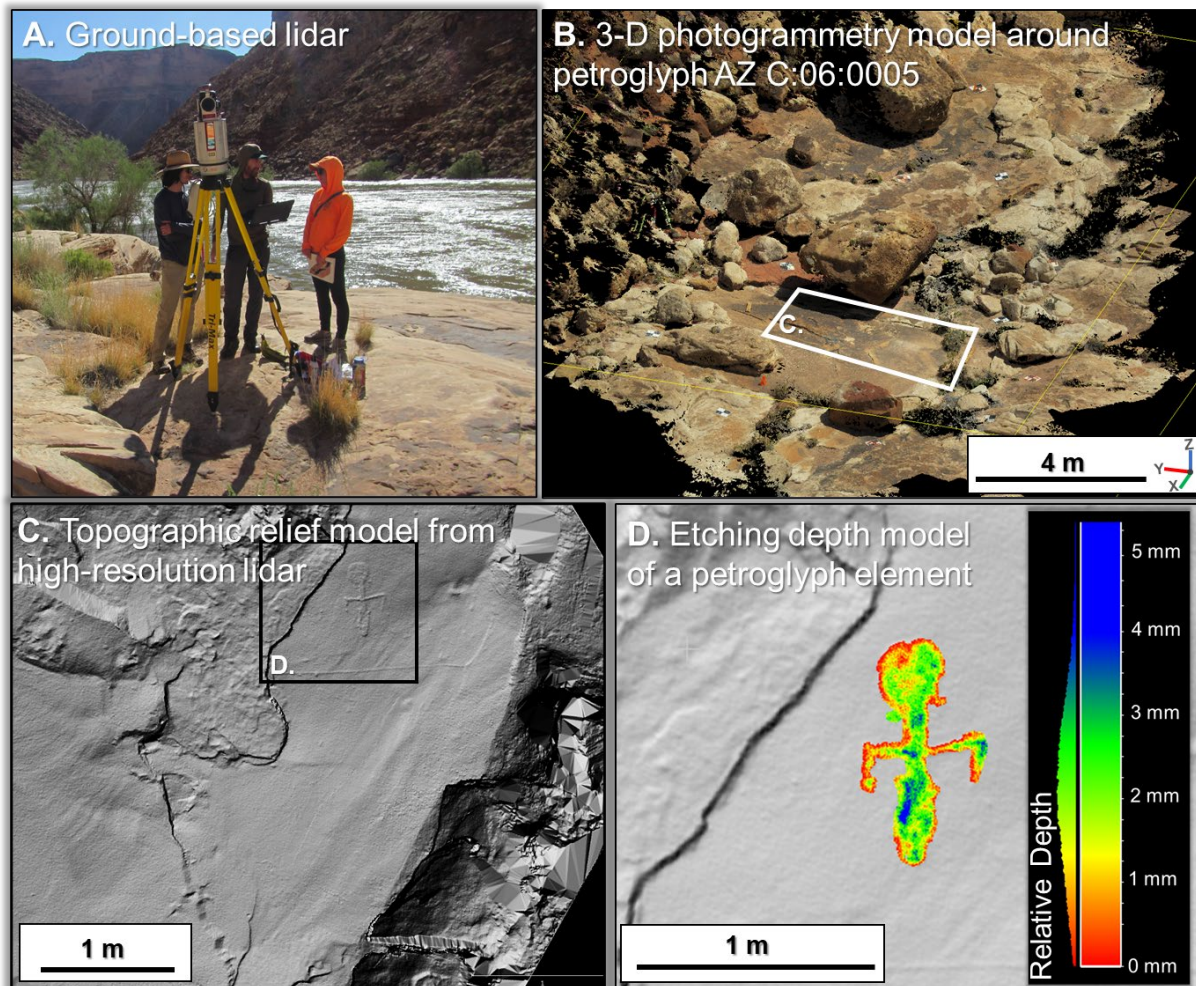


Figure 7. Pilot study data collected at the AZ C:06:0005 petroglyph site. A. Photograph of ground-based lidar collection by USGS scientists. B. Example of a true-color photogrammetry model that includes the petroglyph panel and surrounding area. C. 5-mm resolution topographic relief model of the petroglyph panel. D. Example of calculating estimated etching depth for a portion of the petroglyph panel derived from ground-based lidar.

Anticipated Use of Data

To best meet the needs of the GCDAMP and the LTEMP Cultural PA, this project element is designed to be responsive to identified priorities during FY 2025-27.

During FY 2025, GCMRC will write a report detailing the 2023 findings from lidar monitoring at the petroglyph site C:06:0005. The initial results of this study demonstrated that such sites are vulnerable to a variety of hazards, including potential for direct effects of dam operations from inundation as well as indirect effects associated with visitor access and biochemical weathering.

Collection of detailed monitoring records, as with the ground-based lidar and photogrammetry in this study, provides an important foundation for identifying hazards, providing a baseline condition assessment, and monitoring future changes in site condition.

During FY 2026, these detailed monitoring efforts will either be repeated at this site or conducted at other locations identified as priorities by LTEMP Cultural PA signatories.

For example, the NPS has identified several other petroglyph and pictograph sites in the river corridor that could benefit from more intensive monitoring using the above methods. In addition, several Tribes have expressed interest in possibly applying photogrammetry at other monitoring sites and learning more about the technical aspects of using these methods for monitoring purposes. At the LTEMP Cultural PA Meeting in March 2024 or at the TWG meeting in July 2024, representatives from several tribes, including Hopi, Hualapai, Navajo, Southern Paiute Consortium, and Zuni, stated different specific interests in creating learning opportunities for tribal members. Some representatives requested demonstrations of the photogrammetry and lidar monitoring methods at one or more petroglyph or pictograph sites of their choosing on their lands. As such, GCMRC proposes to work with each individual interested tribe to develop such a demonstration or workshop for tribal members to participate in and learn about collecting these types of data where there is interest in applying these methods within tribal monitoring programs.

Outcomes and Products

- FY 2025: Presentation and report on “Supai Man” (C:06:0005) pilot project results.
- FY 2026 and FY 2027: Hold a demonstration or workshop with each individual interested tribe on the photogrammetry and lidar methods for monitoring petroglyphs and pictographs.

Project Element D.6. Post-HFE Surveys (Experimental Fund)

In each year that an HFE occurs, we will send one technician on a summer trip to do additional post-HFE surveys. The technician will then process those data after returning from the trip, and the data will be added to the monitoring and research results reported by D.1 and D.3. This model for acquiring HFE-specific data that would otherwise not be acquired by Project D was successfully implemented in June 2023 after the spring HFE that year.

Budgets

Project Element Budget

| Fiscal Year 2025 | | | | | | | | | |
|--|------------------|----------------------|-----------------------|-----------------------|---------------------------|--------------------------|-----------------------------|------------------|------------------------------|
| Project D Effects of Dam Operations and Vegetation Management for Archaeological Sites | Salaries | Travel & Training | Operating Expenses | Logistics Expenses | Cooperative Agreements | To other USGS Centers | AMP Special Burden Rate* | Total | USGS Contributing Funds** |
| | | | | | | | 21.86% | | Est Full Rate = 55% |
| D.1. Monitoring the effects of dam operations on archaeological sites | \$166,947 | \$10,000 | \$20,000 | \$34,015 | \$0 | \$0 | \$50,495 | \$281,457 | |
| D.2. Monitoring landscape-scale ecosystem change with repeat photography | \$70,048 | \$3,000 | \$2,500 | \$0 | \$0 | \$0 | \$16,517 | \$92,065 | |
| D.3. Evaluating effects of LTEMP non-flow actions and other experimental vegetation management on archaeological sites | \$19,605 | \$9,000 | \$6,000 | \$0 | \$0 | \$0 | \$7,566 | \$42,170 | |
| Total Project D | \$256,599 | \$22,000 | \$28,500 | \$34,015 | \$0 | \$0 | \$74,578 | \$415,692 | \$35,623 |

| Fiscal Year 2026 | | | | | | | | | |
|--|------------------|----------------------|-----------------------|-----------------------|---------------------------|--------------------------|-----------------------------|------------------|------------------------------|
| Project D Effects of Dam Operations and Vegetation Management for Archaeological Sites | Salaries | Travel & Training | Operating Expenses | Logistics Expenses | Cooperative Agreements | To other USGS Centers | AMP Special Burden Rate* | Total | USGS Contributing Funds** |
| | | | | | | | 22.60% | | Est Full Rate = 55.5% |
| D.1. Monitoring the effects of dam operations on archaeological sites | \$236,012 | \$9,000 | \$15,000 | \$35,129 | \$0 | \$0 | \$66,702 | \$361,843 | |
| D.2. Monitoring landscape-scale ecosystem change with repeat photography | \$58,840 | \$3,000 | \$1,000 | \$0 | \$0 | \$0 | \$14,202 | \$77,042 | |
| D.3. Evaluating effects of LTEMP non-flow actions and other experimental vegetation management on archaeological sites | \$77,940 | \$8,000 | \$2,500 | \$12,699 | \$0 | \$0 | \$22,857 | \$123,996 | |
| Total Project D | \$372,792 | \$20,000 | \$18,500 | \$47,828 | \$0 | \$0 | \$103,761 | \$562,881 | \$48,652 |

| Fiscal Year 2027 | | | | | | | | | |
|--|------------------|----------------------|-----------------------|-----------------------|---------------------------|--------------------------|-----------------------------|------------------|------------------------------|
| Project D Effects of Dam Operations and Vegetation Management for Archaeological Sites | Salaries | Travel & Training | Operating Expenses | Logistics Expenses | Cooperative Agreements | To other USGS Centers | AMP Special Burden Rate* | Total | USGS Contributing Funds** |
| | | | | | | | 23.40% | | Est Full Rate = 56% |
| D.1. Monitoring the effects of dam operations on archaeological sites | \$247,813 | \$9,000 | \$15,000 | \$36,257 | \$0 | \$0 | \$72,088 | \$380,158 | |
| D.2. Monitoring landscape-scale ecosystem change with repeat photography | \$61,782 | \$2,000 | \$1,000 | \$0 | \$0 | \$0 | \$15,159 | \$79,941 | |
| D.3. Evaluating effects of LTEMP non-flow actions and other experimental vegetation management on archaeological sites | \$81,837 | \$7,500 | \$2,500 | \$0 | \$0 | \$0 | \$21,490 | \$113,326 | |
| Total Project D | \$391,431 | \$18,500 | \$18,500 | \$36,257 | \$0 | \$0 | \$108,737 | \$573,425 | \$50,018 |

*GCDAMP Special Rate: Rate includes current DOI preferred rate (currently 15%; subject to change) and facilities rate (will vary annually). No USGS bureau overhead is charged (unique to the GCDAMP agreement).

**USGS Contributing Funds: The amount of funds required to cover the subsidy created by the reduced burden rate (i.e., GCDAMP Special Rate). As in previous years, SBSC/GCMRC will request these funds of the USGS cost-share program. These funds are not guaranteed.

Experimental Fund

| Fiscal Year 2025 | | | | | | | | |
|---|-----------------|----------------------|-----------------------|-----------------------|---------------------------|--------------------------|-----------------------------|-----------------|
| Project D Experimental Fund Projects | Salaries | Travel & Training | Operating Expenses | Logistics Expenses | Cooperative Agreements | To other USGS Centers | AMP Special Burden Rate* | Total |
| | | | | | | | 21.86% | |
| D.6. Post-HFE surveys | \$29,541 | \$250 | \$0 | \$0 | \$0 | \$0 | \$6,513 | \$36,304 |
| Total Project D | \$29,541 | \$250 | \$0 | \$0 | \$0 | \$0 | \$6,513 | \$36,304 |

| Fiscal Year 2026 | | | | | | | | |
|---|-----------------|----------------------|-----------------------|-----------------------|---------------------------|--------------------------|-----------------------------|-----------------|
| Project D Experimental Fund Projects | Salaries | Travel & Training | Operating Expenses | Logistics Expenses | Cooperative Agreements | To other USGS Centers | AMP Special Burden Rate* | Total |
| | | | | | | | 22.60% | |
| D.6. Post-HFE surveys | \$31,018 | \$250 | \$0 | \$0 | \$0 | \$0 | \$7,067 | \$38,335 |
| Total Project D | \$31,018 | \$250 | \$0 | \$0 | \$0 | \$0 | \$7,067 | \$38,335 |

| Fiscal Year 2027 | | | | | | | | |
|---|-----------------|----------------------|-----------------------|-----------------------|---------------------------|--------------------------|-----------------------------|-----------------|
| Project D Experimental Fund Projects | Salaries | Travel & Training | Operating Expenses | Logistics Expenses | Cooperative Agreements | To other USGS Centers | AMP Special Burden Rate* | Total |
| | | | | | | | 23.40% | |
| D.6. Post-HFE surveys | \$32,569 | \$250 | \$0 | \$0 | \$0 | \$0 | \$7,680 | \$40,499 |
| Total Project D | \$32,569 | \$250 | \$0 | \$0 | \$0 | \$0 | \$7,680 | \$40,499 |

Unfunded Project Element Budget

| Fiscal Year 2025 | | | | | | | | |
|--|-----------------|----------------------|-----------------------|-----------------------|---------------------------|--------------------------|-----------------------------|-----------------|
| Project D Unfunded Project Elements | Salaries | Travel & Training | Operating Expenses | Logistics Expenses | Cooperative Agreements | To other USGS Centers | AMP Special Burden Rate* | Total |
| | | | | | | | 21.86% | |
| ‡ D.4. Pilot study to evaluate potential to extract cultural and ecological information from Colorado River deposits using eDNA and pollen | \$21,014 | \$5,000 | \$20,000 | \$0 | \$0 | \$0 | \$10,060 | \$56,074 |
| ‡, ‡‡ D.5. Monitoring petroglyphs and pictographs with photogrammetry and lidar | \$0 | \$2,000 | \$2,000 | \$0 | \$0 | \$0 | \$875 | \$4,875 |
| Total Project D | \$21,014 | \$7,000 | \$22,000 | \$0 | \$0 | \$0 | \$10,935 | \$60,949 |

| Fiscal Year 2026 | | | | | | | | |
|--|-----------------|----------------------|-----------------------|-----------------------|---------------------------|--------------------------|-----------------------------|-----------------|
| Project D Unfunded Project Elements | Salaries | Travel & Training | Operating Expenses | Logistics Expenses | Cooperative Agreements | To other USGS Centers | AMP Special Burden Rate* | Total |
| | | | | | | | 22.60% | |
| ‡ D.4. Pilot study to evaluate potential to extract cultural and ecological information from Colorado River deposits using eDNA and pollen | \$23,083 | \$5,000 | \$20,000 | \$0 | \$0 | \$0 | \$10,867 | \$58,950 |
| ‡, ‡‡ D.5. Monitoring petroglyphs and pictographs with photogrammetry and lidar | \$15,728 | \$2,000 | \$2,000 | \$0 | \$0 | \$0 | \$4,458 | \$24,186 |
| Total Project D | \$38,811 | \$7,000 | \$22,000 | \$0 | \$0 | \$0 | \$15,325 | \$83,136 |

| Fiscal Year 2027 | | | | | | | | |
|--|-----------------|----------------------|-----------------------|-----------------------|---------------------------|--------------------------|-----------------------------|-----------------|
| Project D Unfunded Project Elements | Salaries | Travel & Training | Operating Expenses | Logistics Expenses | Cooperative Agreements | To other USGS Centers | AMP Special Burden Rate* | Total |
| | | | | | | | 23.40% | |
| ‡ D.4. Pilot study to evaluate potential to extract cultural and ecological information from Colorado River deposits using eDNA and pollen | \$24,237 | \$5,000 | \$20,000 | \$0 | \$0 | \$0 | \$11,521 | \$60,758 |
| ‡, ‡‡ D.5. Monitoring petroglyphs and pictographs with photogrammetry and lidar | \$16,515 | \$2,000 | \$2,000 | \$0 | \$0 | \$0 | \$4,800 | \$25,315 |
| Total Project D | \$40,752 | \$7,000 | \$22,000 | \$0 | \$0 | \$0 | \$16,321 | \$86,073 |

‡ Contingent on not descoping other elements.

‡‡ D5 was originally projected to cost \$187,000 gross over three years (FY2025-27); however, it could be funded for \$54,000 by redirecting pay periods of key project staff time from other elements (D.1, D.2, and D.3) to D.5. Project D is in the position to do this by reducing operating expenses and leveraging recent adjustments in labor costs/coverage. Please see KEY POST DRAFT 3 ADJUSTMENTS FOR GCMRC'S FY 2025-27 TWP BUDGET shared with the Glen Canyon Adaptive Management Program (GCDAMP) on 7/25/2024.

References Cited

- Boyer, D.E. and Webb, R.H., 2007, Damming Grand Canyon: The 1923 USGS Colorado River Expedition: Logan, Utah, Utah State University Press, 280 p.
- Bureau of Reclamation, 2017, LTEMP Programmatic agreement among the U.S. Department of the Interior Bureau of Reclamation and National Park Service, Western Area Power Administration, Advisory Council on Historic Preservation, Hualapai Tribal Historic Preservation Officer, Navajo Nation Tribal Historic Preservation Officer, Hopi Tribe, Kaibab Band of Paiute Indians, Paiute Indian Tribe of Utah, Pueblo of Zuni, and the Arizona State Historic Preservation Officer regarding the Glen Canyon Dam operations and non-flow actions identified in the Long Term Experimental and Management Plan Environmental Impact Statement and Record of Decision: Salt Lake City, Utah, Bureau of Reclamation, Upper Colorado Regional Office, 79 p.,
https://gcdamp.com/images_gcdamp_com/d/da/2017_LTEMP_Final_PA.pdf.
- Bureau of Reclamation, 2018, Historic preservation plan for the Glen Canyon Dam operations and non-flow actions identified in the Long-term Experimental and Management Plan Environmental Impact Statement and Record of Decision: Salt Lake City, Utah, U.S. Department of the Interior, Bureau of Reclamation, Upper Colorado Region, 247 p.,
https://gcdamp.com/images_gcdamp_com/8/86/2018_LTEMP_HPP.pdf.
- Caster, J., Sankey, J.B., Fairley, H., and Kasprak, A., 2022, Terrestrial lidar monitoring of the effects of Glen Canyon Dam operations on the geomorphic condition of archaeological sites in Grand Canyon National Park, 2010–2020: U.S. Geological Survey Open-File Report 2022–1097, 100 p., <https://doi.org/10.3133/ofr20221097>.
- Caster, J., Sankey, J.B., Sankey, T.T., Kasprak, A., Bowker, M.A., and Joyal, T., 2024, Do topographic changes tell us about variability in aeolian sediment transport and dune mobility? Analysis of monthly to decadal surface changes in a partially vegetated and biocrust covered dunefield: *Geomorphology*, v. 447, article 109021, p. 1-16,
<https://doi.org/10.1016/j.geomorph.2023.109021>.
- Collins, B.D., Bedford, D.R., Corbett, S.C., Cronkite-Ratcliff, C., and Fairley, H., 2016, Relations between rainfall-runoff-induced erosion and aeolian deposition at archaeological sites in a semi-arid dam-controlled river corridor: *Earth Surface Processes and Landforms*, v. 41, no. 7, p. 899-917, <https://doi.org/10.1002/esp.3874>.

- Collins, B.D., Brown, K.M., and Fairley, H.C., 2008, Evaluation of terrestrial lidar for monitoring geomorphic change at archeological sites in Grand Canyon National Park, Arizona: U.S. Geological Survey Open-File Report 2008-1384, 60 p., <https://pubs.usgs.gov/of/2008/1384/>.
- Collins, B.D., Corbett, S.C., Fairley, H.C., Minasian, D., Kayen, R., Dealy, T.P., and Bedford, D.R., 2012, Topographic change detection at select archeological sites in Grand Canyon National Park, Arizona, 2007–2010: U.S. Geological Survey Scientific Investigations Report 2012–5133, 77 p., <http://pubs.usgs.gov/sir/2012/5133/>.
- Collins, B.D., Corbett, S.C., Sankey, J.B., and Fairley, H.C., 2014, High-resolution topography and geomorphology of select archeological sites in Glen Canyon National Recreation Area, Arizona: U.S. Geological Survey Scientific Investigations Report 2014-5126, 31 p., <https://doi.org/10.3133/sir20145126>.
- Collins, B.D., Minasian, D., and Kayen, R., 2009, Topographic change detection at select archaeological sites in Grand Canyon National Park, Arizona, 2006-2007: U.S. Geological Survey Scientific Investigations Report 2009-5116, 97 p., <https://pubs.usgs.gov/sir/2009/5116/>.
- Cook, T., East, A.E., Fairley, H., and Sankey, J.B., 2019, Managing sand along the Colorado River to protect cultural sites downstream of Glen Canyon Dam: U.S. Geological Survey Fact Sheet 2019-3054, 6 p., <https://doi.org/10.3133/fs20193054>.
- Damp, J., Pederson, J.L., and O'Brian, G., 2007, Geoarchaeological investigations and an archaeological treatment plan for 151 sites in the Grand Canyon, Arizona—unpublished report: Salt Lake City, Utah, submitted to Bureau of Reclamation, Upper Colorado Regional Office, 501 p.
- Draut, A.E., 2012, Effects of river regulation on aeolian landscapes, Colorado River, southwestern USA: *Journal of Geophysical Research: Earth Surface*, v. 117, no. F2, p. 1-22, <https://doi.org/10.1029/2011JF002329>.
- Draut, A.E., and Rubin, D.M., 2008, The role of eolian sediment in the preservation of archeologic sites along the Colorado River corridor in Grand Canyon National Park, Arizona: U.S. Geological Survey Professional Paper 1756, 71 p., <https://pubs.usgs.gov/pp/1756>.
- Draut, A.E., Rubin, D.M., Dierker, J.L., Fairley, H.C., Griffiths, R.E., Hazel, J.E., Jr., Hunter, R.E., Kohl, K., Leap, L.M., Nials, F.L., Topping, D.J., and Yeatts, M., 2008, Application of sedimentary-structure interpretation to geoarchaeological investigations in the Colorado River corridor, Grand Canyon, Arizona: *Geomorphology*, v. 101, no. 3, p. 497-509, <https://doi.org/10.1016/j.geomorph.2007.04.032>.
- East, A.E., Collins, B.D., Sankey, J.B., Corbett, S.C., Fairley, H.C., and Caster, J.J., 2016, Conditions and processes affecting sand resources at archeological sites in the Colorado River corridor below Glen Canyon Dam, Arizona: U.S. Geological Survey Professional Paper 1825, 104 p., <https://doi.org/10.3133/pp1825>.
- East, A.E., Sankey, J.B., Fairley, H.C., Caster, J.J., and Kasprak, A., 2017, Modern landscape processes affecting archaeological sites along the Colorado River corridor downstream of

- Glen Canyon Dam, Glen Canyon National Recreation Area, Arizona: U.S. Geological Survey Scientific Investigations Report 2017-5082, 22 p., <https://doi.org/10.3133/sir20175082>.
- Fairley, H.C., 2003, Changing river—Time, culture, and the transformation of landscape in Grand Canyon—A regional research design for the study of cultural resources along the Colorado River in lower Glen Canyon and Grand Canyon National Park, Arizona: Tucson, Ariz., Statistical Research, Inc., SRI Press, Technical series 79, 179 p., <https://archive.org/details/changingrivertim0000fair>.
- Fairley, H.C., Bungart, P.W., Coder, C.M., Huffman, J., Samples, T.L., and Balsom, J.R., 1994, The Grand Canyon river corridor survey project—Archaeological survey along the Colorado River between Glen Canyon Dam and Separation Rapid: Flagstaff, Ariz., Bureau of Reclamation, Glen Canyon Environmental Studies and Grand Canyon National Park, cooperative agreement no. 9AA-40-07920, 276 p., <https://doi.org/10.6067/XCV8427915>.
- Grams, P.E., Schmidt, J.C., and Topping, D.J., 2007, The rate and pattern of bed incision and bank adjustment on the Colorado River in Glen Canyon downstream from Glen Canyon Dam, 1956-2000: Geological Society of America Bulletin, v. 119, no. 5-6, p. 556-575, <https://doi.org/10.1130/B25969.1>.
- Hereford, R., Thompson, K.S., Burke, K.J., and Fairley, H.C., 1996, Tributary debris fans and the late Holocene alluvial chronology of the Colorado River, eastern Grand Canyon, Arizona: Geological Society of America Bulletin, v. 108, no. 1, p. 3 -19, [https://doi.org/10.1130/0016-7606\(1996\)108<0003:TDFATL>2.3.CO;2](https://doi.org/10.1130/0016-7606(1996)108<0003:TDFATL>2.3.CO;2).
- Kasprak, A., Caster, J.J., and Bangen, S.G., 2017, Geomorphic process from topographic form—Automating the interpretation of repeat survey data in river valleys: Earth Surface Processes and Landforms, v. 42, no. 12, p. 1872-1883, <https://doi.org/10.1002/esp.4143>.
- Kasprak, A., Sankey, J.B., Buscombe, D., Caster, J., East, A.E., and Grams, P.E., 2018, Quantifying and forecasting changes in the areal extent of river valley sediment in response to altered hydrology and land cover: Progress in Physical Geography: Earth and Environment, v. 42, no. 6, p. 739-764, <https://doi.org/10.1177/0309133318795846>.
- Kasprak, A., Sankey, J.B., and Butterfield, B.J., 2021, Future regulated flows of the Colorado River in Grand Canyon foretell decreased areal extent of sediment and increases in riparian vegetation: Environmental Research Letters, v. 16, no. 1, p. 1-15, <https://doi.org/10.1088/1748-9326/abc9e4>.
- Leap, L.M., Kunde, J.L., Hubbard, D.C., Andrews, N.B., Downum, C.E., Miller, A., and Balsom, J.R., 2000, Grand Canyon monitoring project 1992-1999—Synthesis and annual monitoring, report FY99: Salt Lake City, Utah, Grand Canyon National Park, River Corridor Monitoring, submitted to Bureau of Reclamation, Upper Colorado Region, 285 p. plus appendices.
- Pedersen, M.W., Overballe-Petersen, S., Ermini, L., Sarkissian, C.D., Haile, J., Hellstrom, M., Spens, J., Thomsen, P.F., Bohmann, K., Cappellini, E., Schnell, I.B., Wales, N.A., Carøe, C., Campos, P.F., Schmidt, A.M.Z., Gilbert, M.T.P., Hansen, A.J., Orlando, L., and Willerslev,

- E., 2015, Ancient and modern environmental DNA: *Philosophical Transactions of the Royal Society B: Biological Sciences*, v. 370, no. 1660, article 20130383, p. 1-11, <https://doi.org/10.1098/rstb.2013.0383>.
- Pilkington, L.H., Sankey, J.B., Boughter, D.L., Preston, T.N., and Prophet, C.C., 2022, Parks look for ways to alleviate Glen Canyon Dam's downstream impacts: *Park Science Magazine*, National Park Service, Glen Canyon National Recreation Area, Grand Canyon National Park, v. 36, no. 1, Summer 2022, <https://www.nps.gov/articles/000/parks-look-for-ways-to-alleviate-glen-canyon-dams-downstream-impacts.htm>.
- Sankey, J.B., Caster, J.J., Kasprak, A., and East, A.E., 2018a, The response of source-bordering aeolian dunefields to sediment-supply changes 2—Controlled floods of the Colorado River in Grand Canyon, Arizona, USA: *Aeolian Research*, v. 32, p. 154-169, <https://doi.org/10.1016/j.aeolia.2018.02.004>.
- Sankey, J.B., Caster, J., Kasprak, A., and Fairley, H., 2022, The influence of drying on the aeolian transport of river-sourced sand: *Journal of Geophysical Research: Earth Surface*, v. 127, no. 12, e2022JF006816, p. 1-24, <https://doi.org/10.1029/2022JF006816>.
- Sankey, J.B., and Draut, A.E., 2014, Gully annealing by aeolian sediment—Field and remote-sensing investigation of aeolian-hillslope-fluvial interactions, Colorado River corridor, Arizona, USA: *Geomorphology*, v. 220, p. 68-80, <https://doi.org/10.1016/j.geomorph.2014.05.028>.
- Sankey, J.B., East, A., Fairley, H.C., Caster, J., Dierker, J., Brennan, E., Pilkington, L., Bransky, N.D., and Kasprak, A., 2023, Archaeological sites in Grand Canyon National Park along the Colorado River are eroding owing to six decades of Glen Canyon Dam operations: *Journal of Environmental Management*, v. 342, article 118036, p. 1-17, <https://doi.org/10.1016/j.jenvman.2023.118036>.
- Sankey, J.B., Kasprak, A., Caster, J.J., East, A.E., and Fairley, H., 2018b, The response of source-bordering aeolian dunefields to sediment-supply changes 1—Effects of wind variability and river-valley morphodynamics: *Aeolian Research*, v. 32, p. 228 -245, <https://doi.org/10.1016/j.aeolia.2018.02.005>.
- Sankey, J.B., Ralston, B.E., Grams, P.E., Schmidt, J.C., and Cagney, L.E., 2015, Riparian vegetation, Colorado River, and climate—Five decades of spatiotemporal dynamics in the Grand Canyon with river regulation: *Journal of Geophysical Research: Biogeosciences*, v. 120, no. 8, p. 1532-1547, <https://doi.org/10.1002/2015JG002991>.
- Scott, M.L., Webb, R.H., Johnson, R.R., Turner, R.M., Friedman, J.M., and Fairley, H.C., 2018, Evaluating riparian vegetation change in canyon-bound reaches of the Colorado River using spatially extensive matched photo sets, *in* Johnson, R.R., Carothers, S.W., Finch, D.M., and Kingsley, K.J., eds., *Riparian ecology—Past, present, future*: Fort Collins, Colo., U.S. Department of Agriculture, Forest Service, Rocky Mountain Research Station, General technical report RMRS-GTR-377, 148-173 p., <https://doi.org/10.2737/RMRS-GTR-377-CHAP9>.

- Thorne, R.M., 1991, Intentional site burial—A technique to protect against natural or mechanical loss: U.S. Department of the Interior, National Park Service, Archeological Assistance Program Technical Brief no. 5, 8 p., https://home.nps.gov/subjects/archeology/upload/TchBrf5_508.pdf.
- Turner, R.M., and Karpiscak, M.M., 1980, Recent vegetation changes along the Colorado River between Glen Canyon Dam and Lake Mead, Arizona: U.S. Geological Survey Professional Paper 1132, 125 p., <https://pubs.usgs.gov/pp/1132/report.pdf>.
- U.S. Congress, 1966, National Historic Preservation Act—Section 1 as amended through December 16, 2016 and codified in Title 54 of the United States code: Washington, D.C., U.S. Congress, Public Law no. 89-665, as amended by Public Law no. 96-515, 52 p., <https://www.achp.gov/sites/default/files/2018-06/nhpa.pdf>.
- U.S. Department of the Interior, 1992, Grand Canyon Protection Act of 1992—Reclamation projects authorization and adjustment act of 1992—Title XVIII-Grand Canyon protection—Section 1801: Bureau of Reclamation, <https://www.usbr.gov/uc/legal/gcpa1992.pdf>.
- U.S. Department of the Interior, 1996, Record of Decision—Operation of Glen Canyon Dam Final Environmental Impact Statement: Salt Lake City, Utah, U.S. Department of the Interior, Bureau of Reclamation, Glen Canyon Dam Adaptive Management Program, 15 p., https://www.usbr.gov/uc/envdocs/rod/Oct1996_OperationGCD_ROD.pdf.
- U.S. Department of the Interior, 2016a, Glen Canyon Dam Long-Term Experimental and Management Plan Final Environmental Impact Statement (LTEMP EIS): U.S. Department of the Interior, Bureau of Reclamation, Upper Colorado Region, National Park Service, Intermountain Region, online, <http://ltempeis.anl.gov/documents/final-eis/>.
- U.S. Department of the Interior, 2016b, Record of Decision for the Glen Canyon Dam Long-Term Experimental and Management Plan Final Environmental Impact Statement (LTEMP ROD): Salt Lake City, Utah, U.S. Department of the Interior, Bureau of Reclamation, Upper Colorado Region, National Park Service, Intermountain Region, 196 p., http://ltempeis.anl.gov/documents/docs/LTEMP_ROD.pdf.
- Webb, R.H., 1996, Grand Canyon—A century of change—Rephotography of the 1889-1890 Stanton expedition: Tucson, University of Arizona Press, <https://uapress.arizona.edu/book/grand-canyon-a-century-of-change>.
- Webb, R.H., Belnap, J., Scott, M.L., and Esque, T.C., 2011, Long-term change in perennial vegetation along the Colorado River in Grand Canyon National Park (1889-2010): Park Science, v. 28, no. 2, p. 83-87, <https://irma.nps.gov/DataStore/Reference/Profile/2201709>.
- Webb, R.H., Boyer, D.E., and Turner, R.M., eds., 2010, Repeat photography—Methods and applications in the natural sciences: Washington, D.C., Island Press, <https://islandpress.org/books/repeat-photography#desc>.
- Weeden, H., Borden, F., Turner, B., Thompson, O.N., Strauss, C., and Johnson, R.R., 1975, Grand Canyon National park campsite inventory—unpublished Colorado River research report: University Park, Pennsylvania State University, prepared for the National Park

Service, contract no. CX 001-3-0061, 72 p. (Copy on file, Grand Canyon National Park Museum Collections Archives, Grand Canyon, Arizona).

Yackulic, C.B., Bair, L.S., Eppehimer, D.E., Salter, G.L., Deemer, B.R., Butterfield, B.J., Kasprak, A., Caster, J.J., Fairley, H.C., Grams, P.E., Mihalevich, B.A., Palmquist, E.C., and Sankey, J.B., 2024, Modeling the impacts of Glen Canyon Dam operations on Colorado River resources: Phoenix, Ariz., U.S. Department of the Interior, Bureau of Reclamation, cooperator publication prepared by U.S. Geological Survey, Southwest Biological Science Center, Grand Canyon Monitoring and Research Center, Flagstaff, Ariz., April 2024, 133 p., <https://pubs.usgs.gov/publication/70252976>.

Project E: Controls on Ecosystem Productivity: Nutrients, Flow, and Temperature

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Project Summary and Purpose

Ecosystem metabolism, or the rate of organic material fixation (accumulation) or depletion from an ecosystem, is key to understanding the energetic basis of food webs (Bernhardt and others, 2018; Rüegg and others, 2020). The two key components to riverine ecosystem metabolism are aquatic primary production and ecosystem respiration. Aquatic primary production, or the “green” food web, converts sunlight, carbon dioxide and water into simple carbohydrates via photosynthesis. Ecosystem respiration, or the “brown” food web, represents the decomposition and/or consumption of organic material by microbes, aquatic insects, fishes, or other animals. In the Colorado River downstream of Glen Canyon Dam (GCD), fishes are food limited (Cross and others, 2011) and energy (carbon) produced by autotrophs within the river is an important food source (Wellard Kelly and others, 2013). Aquatic primary production, and the aquatic insect community this production supports, is the main source of fish production in Glen Canyon throughout the year (Cross and others, 2011). Primary producers (specifically diatoms) are also an important food source downstream, although the role of non-algal (tributary/terrestrial) “brown” carbon sources can also be an important driver of the food availability near tributary junctions and during flood pulses that occur during monsoon season (Cross and others, 2011; Wellard Kelly and others, 2013; Sabo and others, 2018). Rates of primary production (green food web) are an important control on flannelmouth sucker (*Catostomus latipinnis*) growth rates in Marble and Grand Canyon (Hansen and others, 2023a).

Phosphorus (P) is a key nutrient that can limit organismal growth and metabolic rates. Lake Powell retains disproportionately more P than other key nutrients like nitrogen and silica (Kelly, 2001), creating ecosystem P limitation (Gloss, 1977; Deemer and others, 2023b; Yard and others, 2023).

We now have strong evidence of food web P limitation in Glen Canyon (downstream of Lake Powell). P predicts rainbow trout (*Oncorhynchus mykiss*) recruitment better than flow-based metrics used to predict recruitment for the Long-Term Experimental and Management Plan (LTEMP) Environmental Impact Statement (EIS) (U.S. Department of the Interior, 2016; Yackulic 2020). The density of aquatic insects in Glen Canyon are positively correlated with P availability at the annual scale (Korman and others, 2021), and bioenergetic modeling demonstrates that the Glen Canyon rainbow trout population is strongly regulated by rates of aquatic insect prey production, which is in turn controlled by P concentrations being released from the reservoir (Yard and others, 2023). In Marble and Grand Canyons, the relationship between P and ecosystem productivity is still emerging. Outside of periods when tributaries are flooding for extended periods, the availability of aquatic insect drift and the condition of native fishes are positively related to seasonal rates of gross primary production (GPP) near the Little Colorado River, highlighting the important role for aquatic primary production even 120 km downstream of the dam (Deemer, 2020). In dark sediment incubations conducted in both Glen Canyon and Grand Canyon, pH-mediated P release from sediments consistently led to increases in total protein production (Deemer and others, 2023b), suggesting that the “brown” food web is also limited by P.

Understanding the controls on Colorado River ecosystem production is an important step towards better managing the aquatic food base. For example, the canyon-wide increase in GPP due to springtime low and steady weekend bug flows (Deemer and others, 2022a) was estimated to increase rates of flannelmouth sucker growth by 1.6 mm per month, or approximately the same effect as warming the river by 1.1 °C (Hansen and others, 2023a), demonstrating a measurable link between a dam management strategy and native fish populations. During 2022 and 2023, the river has experienced unprecedented water quality conditions (U.S. Geological Survey, 2024), further underscoring the need to understand how abiotic factors interact to affect the base of the food web. The role of temperature versus other abiotic factors as controls on ecosystem metabolic rates is a key question in ecology (Bernhardt and others, 2018; Battin and others, 2023) and considerable uncertainty exists regarding how the green and brown food web in Grand Canyon will respond to temperature change. In Glen Canyon, P may limit decomposition rates even as waters warm (Scholl and others, 2024). While rates of primary production during clear water conditions appear to scale with river water temperatures at some locations and times, this relationship sometimes falls apart suggesting other factors may control primary productivity, with P limitation a leading candidate (U.S. Geological Survey, 2024). Dramatic temperature changes may also cause shifts in structure of primary producer communities (e.g., Hansson and others, 2020), changing the availability of the most preferential species and the persistence of nuisance taxa. These changes in the food web are expected to interact with direct impacts of temperature, and other environmental drivers, on fish metabolism and other aspects of fish biology to determine the overall impact of environmental drivers on fish population dynamics.

Overall, Project E is designed to capture and link changes in primary productivity, decomposition, and the community composition of primary producers to changes in bottom-up drivers such as light, temperature, flow, and nutrients and to further develop links between these bottom-up drivers and higher trophic levels.

Long-Term Experimental and Management Plan (LTEMP) Resource Goals

Project E directly addresses the natural processes LTEMP goal. Specific project elements also address the Humpback Chub (E4), Other Native Fish (E4), Rainbow Trout Fishery (E2), and Nonnative Aquatic Species (E2). In addition, project elements E2 and E3 may align with the “health of ecosystem” tribal resource goal outlined by Runge and others (2015). This is because river metabolism measurements represent a relatively holistic approach, measuring the way the river breathes. This is in alignment with the Colorado River as a sentient being, although we have not yet discussed this alignment with the tribes.

Project E supports the following resource goals:

- Natural Processes
- Humpback Chub
- Other Native Fish
- Rainbow Trout Fishery
- Nonnative Aquatic Species

Background

Given several challenges associated with quantifying primary production directly below dams, the majority of what we currently know about controls on riverine primary production comes from unregulated rivers and streams or reaches downstream from dams (Bernhardt and others, 2018). In these systems light and disturbance are key factors determining the timing and overall rate of primary production (Bernhardt and others, 2022).

Glen Canyon Dam has fundamentally altered the light and disturbance regimes in the downstream Colorado River. Since damming, the river has experienced a 95% decline in the amount of fine sediment delivered to the upstream boundary of Grand Canyon National Park (Topping and others, 2000) resulting in much less turbid conditions during much of the year through Marble and Grand Canyon (e.g., downstream of the Paria River) and extremely clear water conditions year-round in Glen Canyon (e.g. < 2 Nephelometric Turbidity Units, with the exception of a rare turbidity interflow event from the dam (Wildman and Vernieu, 2017). Disturbance regimes have also shifted, with periodic high flow experiments representing the largest disturbances to the ecosystem, albeit generally much smaller floods than experienced under the natural hydrograph (Melis and others, 2012).

Damming has shifted the hydrograph from large seasonally driven fluctuations in discharge to large sub-daily variation and muted seasonal variation in discharge related to hydropower production. Consistent with work in unregulated rivers, previous modeling work in the Colorado River upstream of Diamond Creek showed that riverine primary production is strongly light limited and that higher diel discharge fluctuations lead to somewhat less primary production (Hall and others, 2015). This finding holds true throughout the Colorado River in Grand Canyon, where low and steady flows during the 2018 and 2019 “Bug Flow” experiment led to significant increases in reach-scale rates of primary production (Deemer and others, 2022a). In Glen Canyon, GPP may be less sensitive to changes in discharge because variable flow upstream of the Paria is not associated with increased turbidity. Across 143 rivers, the presence of an upstream dam was associated with slower recovery of primary producer biomass following disturbance (Lowman and others, 2024), although this dataset was generally only able to examine trends outside the immediate tailwaters (where oxygen disequilibrium prohibits one station modeling approaches). Preliminary examination of GPP time series in the Colorado River suggest that recovery of GPP following high flow experiments may depend on the timing of the experiment (Bishop and others, 2024). Downstream, smaller tributary floods may have an equal or greater effect on GPP than do high flow experiments since these floods carry limiting P into the Colorado River and reduce ecosystem pH, making the P more biologically available (Deemer and others, 2023b).

Changes to flow are not the only dam management lever on downstream ecosystem metabolism. Lower Lake Powell elevations are likely having unprecedented impacts on the primary producer communities in the CRE, both through increased entrainment of plankton and through shifts in water quality (Figure 1, H7). Specifically, lower lake elevations are leading to warmer water temperatures (Dibble and others, 2021) and lower but more variable phosphorus concentrations (Deemer and others, 2023b). The lower lake elevations are also affecting the entrainment of biological organisms. While recently the GCDAMP has largely been focused on entrainment of smallmouth bass (*Micropterus dolomieu*) (Eppheimer and others, 2024), increased entrainment of plankton is also a likely effect of lower reservoir elevations and evidence for this can be observed in the long-term Lake Powell water quality monitoring dataset (Andrews and Deemer, 2022).

Water Quality Conditions Immediately Below Glen Canyon Dam

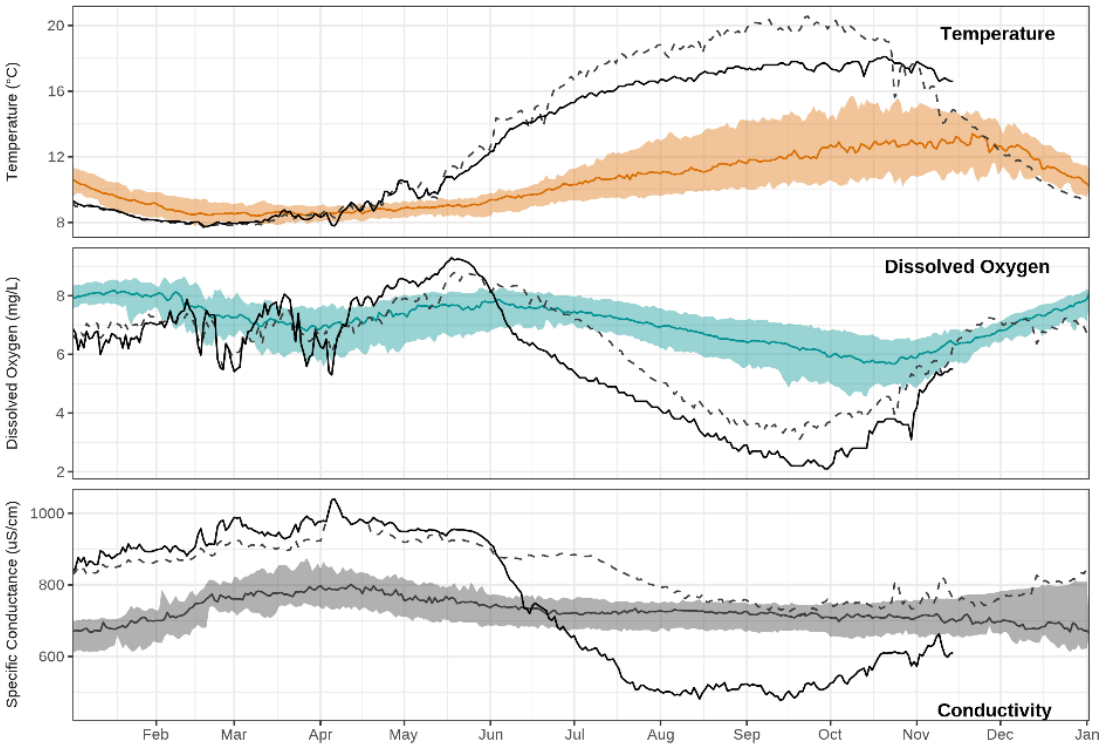


Figure 1. Water quality record from Glen Canyon Dam near Page, AZ (gaging station #09379901*) for temperature (A) dissolved oxygen (B) and specific conductance (C). Dashed and solid lines show daily median values for calendar years 2022 and 2023 respectively (medians are from the continuous data record, logging at 15-minute increments). Colored lines show the long-term median value for each parameter and the orange, blue, and brown color bands represent the daily 10th and 90th quantiles of temperature, dissolved oxygen and specific conductance, respectively. The water quality record represented in this figure contains 12 years of data for temperature and dissolved oxygen, and nine years of data for specific conductance. *Data from this site are currently posted and available through June 2023 at https://www.gcmrc.gov/discharge_qw_sediment/station/GCDAMP/09379901; more recent data are filtered raw values that are awaiting further quality assurance before being posted online. Figure adapted from U.S. Geological Survey (2024).

In 2022, Lake Powell experienced unprecedented warm temperatures, approximately 5 °C warmer than the warmest post-dam recorded temperature in the past 50 years (Figure 1). This allows us to test the role of temperature as a driver of metabolic rates. While warmer temperatures are positively correlated with GPP in some Grand Canyon reaches, in other locations GPP appears to be controlled by other factors (Hansen and others, 2023b). In Glen Canyon, warm temperatures in 2022 and 2023 led to elevated GPP and ecosystem respiration (ER), but not to the extent that would be expected based on lab studies (Bishop and others, 2024, Scholl and others, 2024). Experimental work in other systems has demonstrated the capacity for grazers (Kazanjan and others, 2018) and nutrient limitation (Corman and others, 2016) to limit or even reduce primary producer biomass and production under warming temperatures.

In other cases, increased nutrient use efficiency with warming water supports substantial increases in stream primary producer biomass and overall production (Hood and others, 2018). In a Glen Canyon decomposition study with three litter sources, two of the litter types showed relatively minor responses to the 2022 warming event (+10°C), whereas the litter type with elevated nutrient content was decomposed at a more elevated rate, suggesting the P limitation regulates the response of the brown food web to temperature increases (Scholl and others, 2024). Together, these results illustrate that direct effects of warming on GPP and decomposition may be modified by a suite of environmental and biological factors, however it remains unclear how temperature, GPP (including nuisance producers), grazers (including nonnative consumers), and ER interact in Glen Canyon and throughout the CRe.

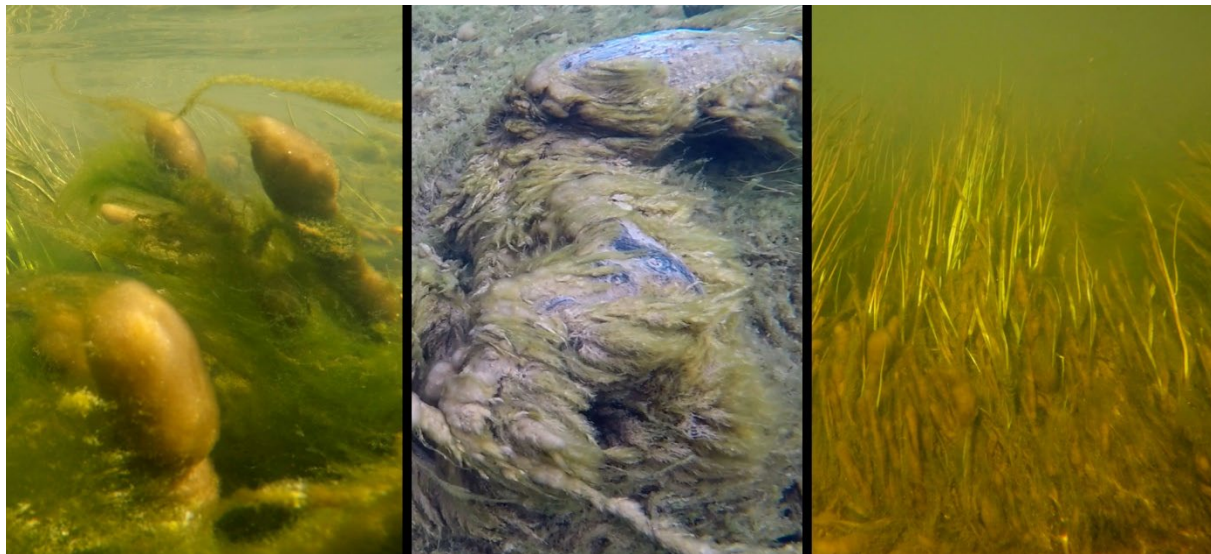


Figure 2. Images of nuisance stalked benthic diatom species, taken in the Colorado River near Lees Ferry in 2021. Photo credit: S. Spaulding, USGS.

Distinct producer communities vary in their responses to dam operations (Benenati and others, 1998) and changing environmental conditions upstream (autecological preference for increasing temperature; Blinn and others, 1989; and phosphorus; Bothwell and others, 2014), and thus in their contribution to secondary production (Krist and Charles, 2012). In one example, plausible decreases in P and increases in temperature downstream of GCD may result in the development of nuisance diatom blooms. These taxa are already present throughout the Colorado river, in Glen and Grand Canyons (Bishop and Spaulding, personal observations, Figure 2), and the former commonly thrives in regulated rivers under severe P limitation ($P < 5 \mu\text{g/L}$; Bothwell and others, 2014). The “rock snot” that these blooming taxa produce can reshape macroinvertebrate composition (Larson and Carreiro, 2008; Gillis and Chalifour, 2010) and is also a relatively poor food source which can negatively impact consumer biomass (Furey and others, 2014).

In another example, shifting Lake Powell elevations may be affecting the movement of lake phytoplankton (another group of primary producers) and zooplankton downstream. Understanding how producer communities may respond to changes in dam operation (through both changes in flow and changes in water quality) has important implications for the food base and for recreational experience (e.g., Beville and others, 2012).

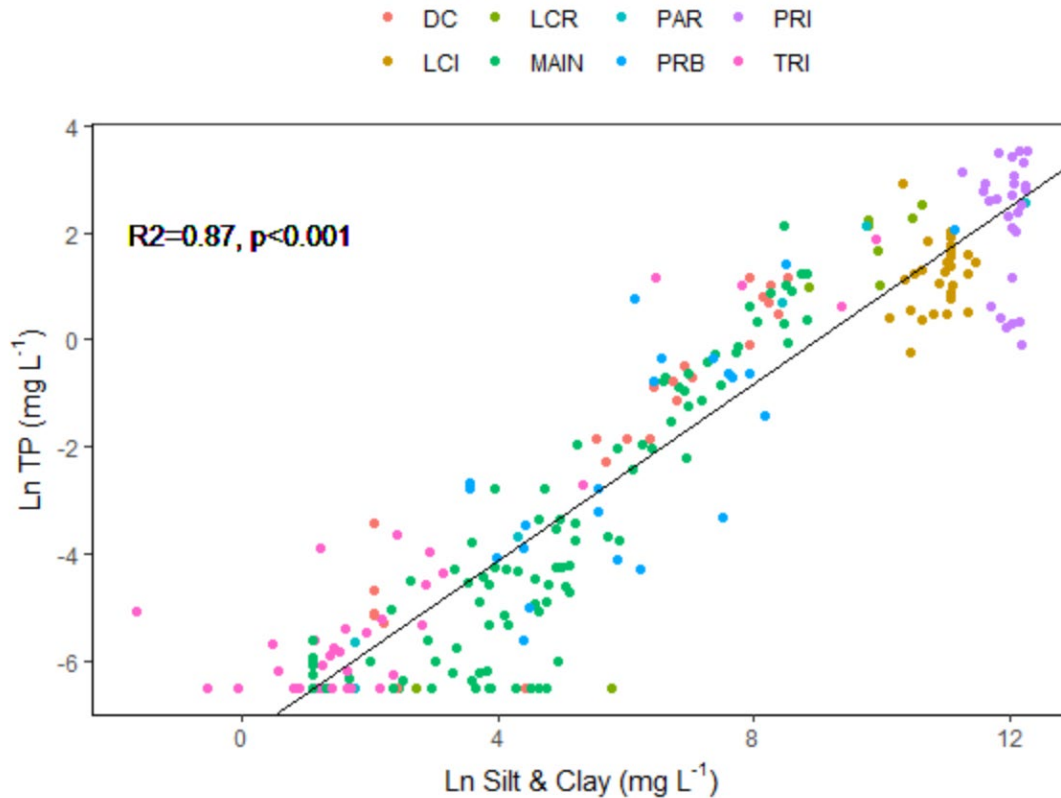


Figure 3. Positive natural log relationship between mainstem and tributary silt & clay concentration and total phosphorus (TP) concentration for samples collected in FY 2021 (n=155). Color indicates sample location and type (“DC”=Colorado River above Diamond Creek, “LCI”=Little Colorado River at Cameron Isco, “LCR”=Little Colorado River grab sample, “MAIN”=mainstem Colorado grab sample, “PAR”= Paria River grab sample, “PRB”=monthly USGS Paria River baseflow grab sample, “PRI”=Paria River Isco sample, and “TRI”= tributary grab sample (generally by community scientists). Preliminary data, subject to revision.

Factors affecting Spatio-temporal Variation in Phosphorus in the Colorado River Ecosystem

While there is temporally and spatially resolved information about riverine turbidity, solar inputs, discharge, and gross primary productivity (via continuous oxygen and temperature measurements) in the Colorado River Ecosystem (CRE), there is much less information about spatio-temporal variability of P and other nutrients in the river. Still, total P concentrations scale with silt and clay concentrations across mainstem Colorado River and tributary sites (Figure 3), which is promising for hindcasting ecosystem P.

Declining Lake Powell elevations are generally causing lower P water (that is higher in pH) to be released downstream (Deemer and others, 2023b). Storm-based inputs of P have also been identified as an important source of P to the river (Deemer and others, 2023b). Thus, we expect tributaries to be an increasingly dominant source of P the further you get from GCD, with the dam being the dominant source in the Glen Canyon reach (H1). Given large year to year variation in the summer storm season, we also expect the availability of P to be more variable in Marble and Grand Canyon than in Glen Canyon (H2).

Given the calcium carbonate-dominated geochemistry of the Little Colorado River (LCR; Stone and others, 2018), we expect storms to mobilize much higher concentrations of travertine-bound P than many of the other tributaries. While most of our measurements of P in the LCR are upstream of travertine formation (at the Little Colorado River at Cameron, AZ, USGS gaging station 09401200), we expect that during baseflow conditions a substantial fraction of the biologically available P passing Cameron is sorbed to sediments in the reach of the river between Blue Springs and the confluence. This binding of P can happen as gradually increasing pH drives calcium carbonate to precipitate, binding P in the process (Corman and others, 2016; Stone and others, 2018; Deemer and others, 2023a). Sampling in the lower 10 km of the LCR in the summer of 2022 showed that while total river P was high, biologically available P (as soluble reactive P) was consistently near or below detection. This is consistent with previous observations of higher biologically available P near Blue Spring and declining biologically available P moving downstream (Moody and Muehlbauer, unpublished data), suggesting uptake of P by travertine rich sediments. If true, large storms that mobilize LCR sediments are likely to flush this travertine-bound P into the CRe where temperature and pH change may drive biologically available P back into solution, representing a key source of P to the Grand Canyon reach of the Colorado River (H3). This would be consistent with GPP that is more limited by P upstream of the LCR than downstream (H4).

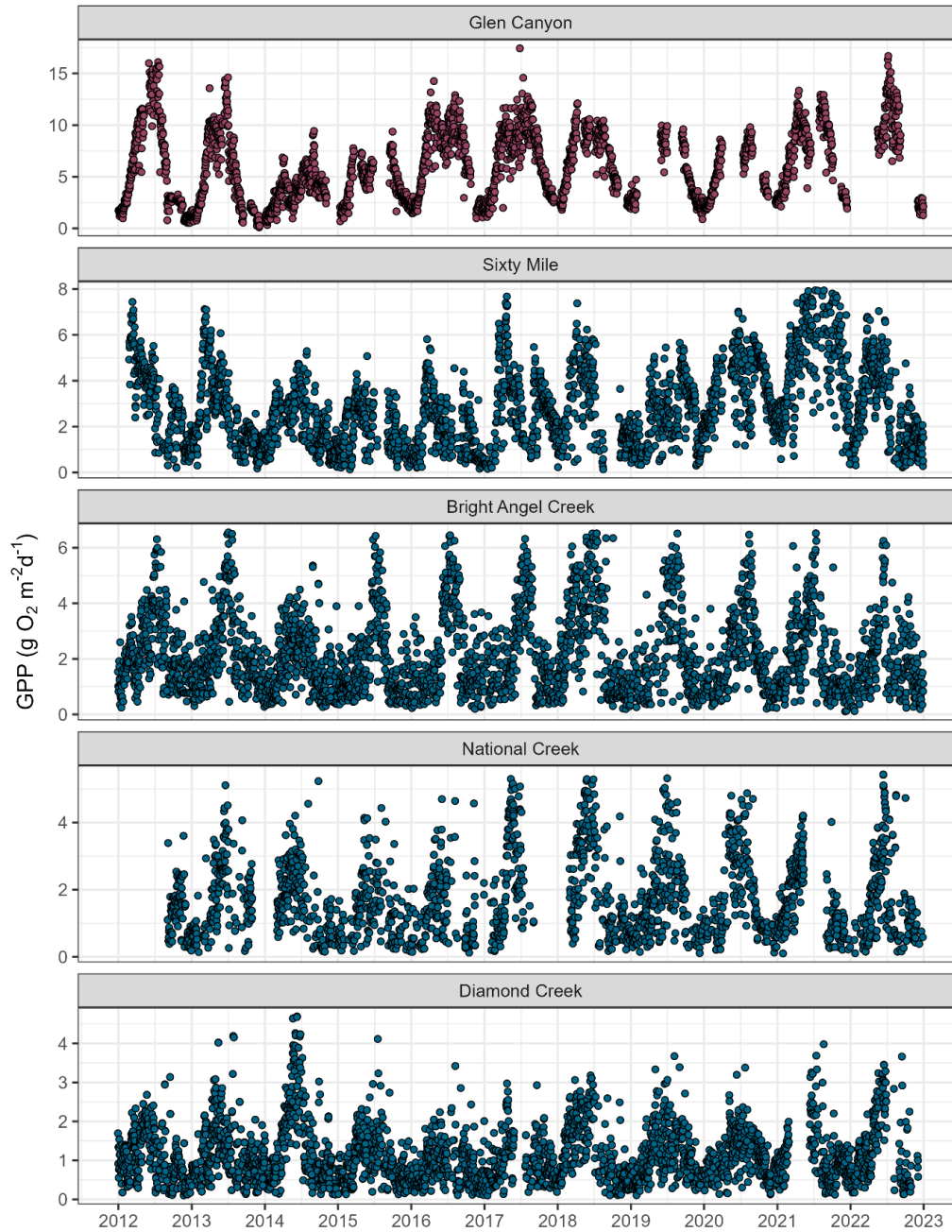


Figure 4. Daily gross primary productivity (GPP) estimates for five dissolved oxygen monitoring reaches of the CRe for the 12 years spanning 2012-2023. Only high-quality estimates are shown for Glen Canyon ($R^2 > 0.9$), and only estimates where $GPP > 0.1$ and $R_{hat} < 1.02$ are shown for the four downstream sites. From top to bottom, there are 2743, 3148, 3156, 2007, and 3391 daily estimates per site, or 241 days per year on average. GPP time series through 2020 for Sixty Mile, Bright Angel Creek, National Creek, and Diamond Creek are available in Deemer and others, 2022b. Glen Canyon and post-2020 Grand Canyon estimates are preliminary and subject to revision.

Controls on GPP in the Colorado River and Potential Food Web Implications

Our capacity to estimate daily (and even sub-daily) GPP from continuous dissolved oxygen measurements provides a unique opportunity to identify how dam management (Hall and others, 2015; Deemer and others, 2022a) and other ecological drivers affect production at the base of the food web. For example, the very fine timescale of our GPP estimates allow us to link sub-daily changes in flow to effects on GPP. During the last work plan, we developed a reproducible workflow for estimating GPP in Marble and Grand Canyon. We also demonstrated that higher rates of GPP support higher growth rates of the native flannelmouth sucker (Hansen and others, 2023a), and that these rates can be managed by dam operations (Deemer and others, 2022a). Still, we have only scratched the surface of understanding the spatial and temporal patterns in GPP in the CRe and the dam-related levers on these rates. Seasonal patterns are not consistent across river reaches (Figure 4), suggesting unique controls on production by reach. For example, GPP does not scale with temperature in Marble Canyon (“Sixty Mile” site), like it does farther downstream (Figure 5). A better understanding of how the seasonality of reach-scale GPP aligns with key life history stages of native fishes could provide insight into times when managing for high GPP will have greatest effect. An understanding of how GPP responds to flow disturbances (e.g., high flow experiments and spike flows), and the timing of flow disturbances, would also be instructive.

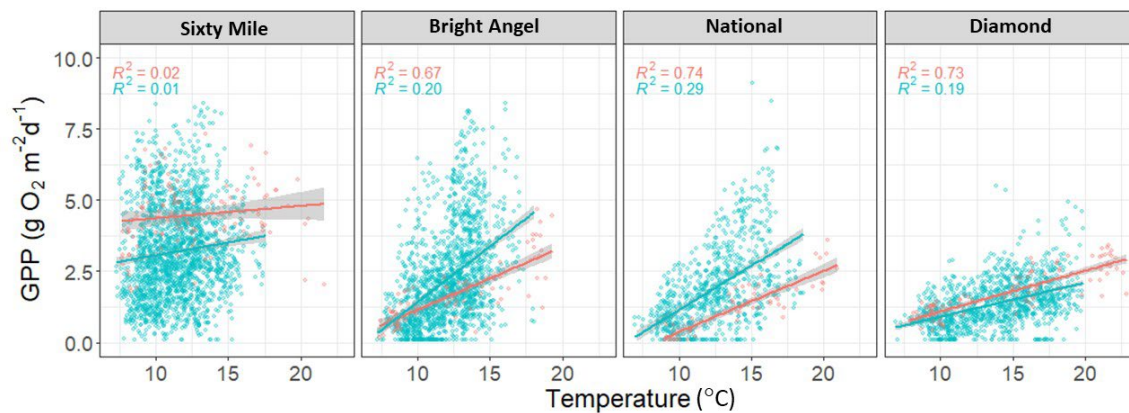


Figure 5. The relationship between gross primary productivity (GPP) and temperature for 2012-2021 (blue dots) and for 2022 (red dots) at four sites representing reaches ending at the USGS gaging stations: Colorado River above Little Colorado River near Desert View, AZ, “Sixty Mile” 09383100, Colorado River near Grand Canyon, AZ “Bright Angel” 09402500, Colorado River above National Canyon near Supai, AZ “National” 09404120, and Colorado River above Diamond Creek near Peach Springs, AZ “Diamond” 09404200. These data represent only those GPP estimates calculated in low turbidity conditions (turbidity <12 FNU). By removing GPP values where turbidity is high (>12 FNU), we can see a more significant relationship between habitat variables and GPP. GPP estimates from 2022 are colored in red to represent warm water releases from Glen Canyon Dam. Preliminary data, subject to revision.

Estimating GPP in the Glen Canyon reach is more complicated than further downstream (Payn and others, 2017). Recent improvements in modeling tailwater GPP provide a valuable opportunity to better understand ecological dynamics in the CRe (Bishop and others, 2024).

The Glen Canyon reach exhibits both a spring and a fall peak in GPP (Figure 6). Despite P concentrations being an important driver of secondary productivity in Glen Canyon (Korman and others, 2021; Yard and others, 2023), GPP does not always track P concentrations. This is likely due to a more complex mix of primary producers in this reach (macrophytes, diatoms, cyanobacteria, etc.; H6), each with varying physiological capacities (Lange and others, 2016) and environmental preferences (Spaulding and others, 2021). Some primary producers can access P in riverine sediments, while others are reliant on water column P. GPP derived from different producer communities varies in food web relevance as well, and we expect that the specifics of which drivers are most important to food web-relevant GPP (H4) will be improved by parsing patterns in distinct producer groups. Specifically, we expect that the spring GPP peak in Glen Canyon is driven by benthic diatoms (preferred food source of CRE consumers) growth responding to increased light as well as increased P from GCD outflow (Figure 6; H5).

In contrast, we hypothesize that the late summer GPP max is less relevant to consumers, being driven by extensive macrophyte growth responding to increased flow, increased temperature, and stable access to sediment P (H5). Under lower reservoir elevations, a greater fraction of Glen Canyon production may be ascribed to phytoplankton entrained from Lake Powell as well (H6).

The relationship between GPP and both aquatic insect drift and fish condition in Marble and Grand Canyon (Deemer 2020) suggests the primary producer community is generally a high-quality food source, but this may change with new water quality regimes under lower Lake Powell elevations. For example, significant proportions of diatom communities throughout the CRE are composed of species (e.g., *Diatoma vulgare* and *Rhoicosphenia abbreviata*) that have shown sensitivity to experimentally increased temperatures (Blinn and others, 1989), and we hypothesize that recent and predicted high water temperatures downstream would result in decreased biomass for such species (H7). This sensitivity is particularly salient because these species also belong to the diatom guild most preferred by local macroinvertebrate consumers (Colletti and others, 1987; Steinman and others, 1987).

Another example of important compositional change relates to potential growth in the abundance and spatial distribution of notable nuisance diatom species, including *Didymosphenia geminata* (“didymo”), and “*Cymbella* sp.”, two stalked benthic diatoms that form significant, macroscopic blooms (Bothwell and others, 2014; and see the Central Arizona Project website: <https://arcg.is/00fuGz>). Both taxa have been observed in the Colorado River downstream of GCD to varying degrees (Ian Bishop and Sarah Spaulding, personal observation, Figure 2). Both taxa in their own ways would likely benefit from decreased P (didymo) and increased temperature (*Cymbella* sp.), and we hypothesize that they may be more widespread than previously believed and that the system is primed for nuisance bloom behavior (H7).

Given recent water quality conditions and plausible future scenarios, we propose that CRE diatom composition be revisited to establish a modern baseline of abundance and distribution for the most food web-relevant producer community in the CRE.

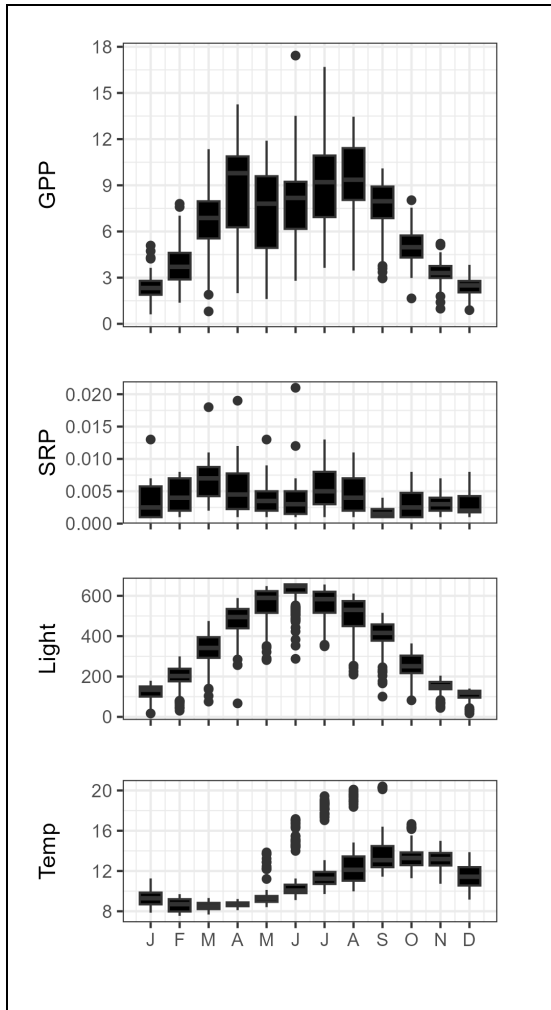


Figure 6. Monthly variability in Glen Canyon gross primary production (GPP) and important environmental drivers (2013-2023). GPP = each point is a daily estimate of GPP estimated for reach between -8 mile and Lees Ferry in units of $\text{g O}_2 \text{ m}^{-2} \text{ d}^{-1}$. SRP = each point represents a single monthly measurement of soluble reactive phosphorus (mg/L) measured in GCD draft tubes, where months with poor QA/QC results are omitted; Light = cloud-adjusted mean daily solar insolation in Glen Canyon (PPFD). Light estimates derived from Yard and others (2005) instantaneous solar flux model; cloud adjustment factor retrieved for same location from the National Solar Radiation Database, USA & Americas 30 minute, 4km data product; Temp = mean daily water temperature ($^{\circ}\text{C}$) at Lees Ferry. Preliminary data, subject to revision.

Quantitative Links among Flow, Nutrients, Primary Production, and Higher Trophic Levels

While fish are often food limited in the Colorado River downstream of Glen Canyon Dam (Cross and others, 2011, Hansen and others, 2023a), we lack a quantitative understanding of how food limitation varies over time and space and how it impacts different fish species. Food limitation can be defined formally as conditions when the energy produced by lower trophic levels and transferred to higher trophic levels does not meet demand from higher trophic levels.

Project Element E.2 focuses on identifying the causes of variation in the production of energy at the base of food webs; however, to understand impacts on fishes, we also require estimates of how demand varies. Quantifying demand requires estimates of abundance across different size classes and species of fish within a given reach, as well as the metabolic needs associated with growth in weight over a given time interval.

In Glen Canyon, recent work has begun to quantify this demand for rainbow trout to better understand (and eventually predict) trends in rainbow trout demography (M. Yard, unpublished data). These calculations rely not only on field measurement, but also on lab-based estimates of metabolic needs, which are well-studied for a species like rainbow trout. For other species, like flannelmouth sucker and humpback chub that dominate downstream biomass estimates, we lack direct estimates of metabolism, and bioenergetics assessments have often relied on lab estimates for related taxa.

This approach to estimating metabolism is problematic as there is evidence to suggest that species found in Grand Canyon may have much lower metabolic needs than related species (H8). For example, Dibble and others (2017) found that humpback chub held without food for 45 days did not significantly decrease in condition, while bonytail and roundtail chub (*Gila robusta*) declined significantly – an observation that has been noted in conversations with others doing experimental work with humpback chub. Quantifying this demand is important for understanding recent increases in populations of native fish species in Western Grand Canyon and determining potential limits to future population growth. If humpback chub and other native species in Grand Canyon have lower metabolic demands (H8) they will require substantially less food to maintain a given weight, and it may be possible to sustain much larger biomasses (abundances) of native fishes at a given food availability (H9). If these species have lower metabolic demands, it may also explain why declines in food availability have been associated with different demographic responses in native fishes as compared to rainbow trout.

Based on the life history of species, they often respond demographically in different ways when food availability increases or decreases. For rainbow trout, it appears that annual changes in food availability directly impact fish condition and growth, which in turn impacts reproduction (Korman and others, 2021). When food is abundant, survival of all sizes of rainbow trout is higher (Korman and others, 2021). When food supply declines, however, large fish face high metabolic deficits (Korman and others, 2021). Among native fishes, we have observed similar changes in fish condition and perhaps reproduction; however, impacts on growth and survival are either smaller or masked by effects of environmental factors. As a result, adult abundances of native fishes appear to be more stable in response to declines in food availability, but reproduction is quite variable (Yackulic, 2020). Nonetheless as native fish populations approach their carrying capacity, we hypothesize that long-term changes in environmental conditions (e.g., from changing quality of water released from Lake Powell) will play an important role in regulating overall abundances (H10).

Additions of higher trophic levels (i.e., piscivorous, warm-water nonnatives like smallmouth bass), are also expected to impact production and consumption at lower trophic levels and ecosystem models could provide another mechanism for predicting the potential impacts of nonnative expansions on native fish population sizes.

Proposed Work

In this workplan, we propose to revise three elements from the prior FY 2021-23 workplan (Project Elements E.1, E.2, and E.4). A new element (currently unfunded) focuses on ecosystem metabolism in the Western Grand Canyon (E.3). Several sub-elements propose to analyze data collected during the last work plan (mass balance P budgeting in E.1; vegetation mapping in Glen Canyon and GPP modeling exercises in E.2 (unfunded); bioenergetics modeling in E.4). One new sub-element proposes to survey benthic diatom populations in Grand Canyon and establish a modern molecular baseline for their abundance and distribution (E.2)

Project Element E.1. Phosphorus Budgeting in the Colorado River (Ongoing Study; Partially Funded)

Bridget Deemer¹, Ian Bishop¹, Robert Hall², David Topping¹, Tom Sabol¹, Theodore Kennedy¹, Charles Yackulic¹, Sasha Reed³

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Hypotheses and Science Questions

- H1: Glen Canyon Dam outflow is the biggest control on P concentrations in Glen Canyon and Marble Canyon, but this influence is dampened the further you move downstream (and with storm-based tributary inflows).
- H2: Interannual differences in tributary sourced P are much larger than interannual differences in P loading from Glen Canyon Dam.
- H3: Storm-driven P mobilization from the Little Colorado River can fertilize a substantial portion of the mainstem river due to the timescale of P desorption kinetics.

Methods

This project will consider major sources and potential sinks for P in the CRe using a P mass balance approach (in the spirit of Meyer and Likens, 1979). P loading from different potentially important sources will be estimated and compared to estimates of P export (or P that is transported downstream of the Colorado River above Diamond Creek near Peach Springs, AZ gaging station).

This will be done using empirical relationships between P and silt-and-clay, based on data collected during the last work plan (Figure 3) and will also utilize estimates of P equilibrium from sorption/desorption experiments that are being conducted this summer (FY 2024).

Ideally, such empirical relationships could be extended to link P concentrations with turbidity, which is an easily measured water quality parameter that is collected as part of Project A. Such a modeling approach could be combined with turbidity modeling work proposed in Project B4 to develop predictive models of P.

We will compare P loading from GCD with P loading from tributary storms to characterize the predominant source of P in different river reaches (H1). Monthly measurements at the outflow to Glen Canyon Dam are currently taken as part of the Interagency Lake Powell Water Quality Monitoring Program (see Appendix 1). These measurements will be used in combination with discharge to estimate P loading from the dam. Given the uncertainty associated with single P measurements, we will use these measurements together with QA/QC results (from equipment and reagent blanks) to train a seasonal model of P. Measurements of mainstem and tributary P concentrations were made in FY 2021 and FY2022 in coordination with citizen scientists. In total, 423 total P samples were collected, representing 23 tributaries as well as 8 sites within the mainstem Colorado River. More extensive sampling was focused in the Paria (27 baseflow and 109 storm samples) and LCR (26 baseflow and 76 storm samples), with 53 samples taken across 21 other tributaries. We expect that 129 mainstem samples will help constrain estimates of P outflow at Diamond Creek.

We will use data collected in FY 2024 to characterize the potential for LCR sediment-bound phosphorus to fertilize the mainstem Colorado River during storm events (H3). Specifically, we will use longitudinal sediment P concentrations collected at twelve sites together with sorption and desorption assays that characterize equilibrium P concentrations in both the LCR and the Colorado River near its confluence with the LCR.

To assess the capacity for overplus P uptake and the extent to which it represents a significant retention of available P during such storm events, an unfunded project proposed to measure P loading and P uptake at the Paria River confluence 1 day before, during and up to 7 days after a summer storm event. Because pre-storm P concentrations would likely affect uptake rates, we would attempt to conduct this experiment early in the monsoon season, or at the very least for a storm for which no substantial loading event has taken place over the past 10-14 days, so that benthic communities are acclimated to P limited conditions. Multiple algal substrates (e.g., cyanobacteria- and diatom-predominant biofilms) will be sampled to assess variable uptake rates of polyphosphate, the most common form in which cellular P is stored. Total P of the benthic samples would also be measured, as the extracellular biofilm matrix these algae construct is itself known to absorb significant quantities of P (Sundareshwar and others, 2011). Algal substrate sampling would be paired with frequent sampling of P in the Paria River outflow during the storm event, which allows us to relate uptake rates of the various algal communities back to P loading intensity of this specific storm event.

Finally, in addition to measuring uptake rates, we will monitor changing producer biomass during the post-event period by measuring benthic algal pigment fluorescence with a BenthosTorch (bbe-Moldaenke). Cobble substrate would be dark-acclimated for 15 minutes ahead of measurement to reduce the effects of variable light conditions on in situ fluorescence readings. These data, in combination with nearby DO-based reach-scale GPP measurements would allow us to assess how excess P is utilized in the short term after the storm event concludes.

Anticipated Use of Data

The capability to model river P concentrations based on suspended sediment concentrations will inform GPP modeling efforts described in E.2, allowing us to link P to rates of primary production (especially in further downstream sections where tributary inputs potentially override the influence of variation in the P in dam releases). The element will also support our broader ability to hindcast (and possibly even forecast) P dynamics, which will improve our understanding of linkages between climate, P and the food web.

An unfunded component of this project examines the hypothesis that benthic algae downstream of Lees Ferry sustain elevated GPP during the monsoon season in part through a nutrient acquisition strategy known as “overplus P uptake” (Harold, 1966; e.g., Lapointe and others, 2024), in which algal cells adapted to low-P environments opportunistically uptake extra P during storm events and store it for later use when baseline low-P conditions resume. This information could inform how we represent time lag effects after storms in canyon-wide GPP models.

Outcomes and Products

Outcomes

- We expect that this project element will constrain the potential importance of different P inputs to the canyon.

Products

- One to two journal articles describing P budget in Glen and Grand canyons.
- A model that hindcasts patterns in P inputs to the Colorado River.

Project Element E.2. Rates and Composition of Primary Producers in the Colorado River (Modified Study; Partially Funded)

Ian Bishop¹, Lindsay Hansen¹, Kimberly Dibble¹, Bridget Deemer¹, Dan Buscombe², Tom Sabol¹, Mike Yard³, Bob Tusso¹, Robert Hall⁴, Theodore Kennedy¹, Charles Yackulic¹

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Hypotheses and Science Questions

- H4: Phosphorus is more limiting to GPP in the spring than during other seasons. The degree of phosphorus limitation also varies spatially, with Glen and Marble Canyon being especially P limited.
- H5: GPP seasonality in Glen Canyon is driven by the distinct environmental preferences/tolerances of algal and macrophytic producers.
- H6: Shifts in Lake Powell plankton entrainment, rooted macrophytes, and macroalgal species within Glen Canyon correspond to changes in dam operations (e.g., high flow events) and changes to reservoir elevation.
- H7: Declining reservoir elevations and associated changing GCD outflow water quality negatively impacts the preferred food base (benthic diatom community) composition throughout the CRE.

Methods

We will continue developing estimates of GPP at the daily timestep to add to the continuous record that has been developed at the USGS gaging stations: Colorado River above Little Colorado River near Desert View, AZ, “Sixty Mile” 09383100, Colorado River near Grand Canyon, AZ “Bright Angel” 09402500, Colorado River above National Canyon near Supai, AZ “National” 09404120, and Colorado River above Diamond Creek near Peach Springs, AZ “Diamond” 09404200 and within Glen Canyon via dissolved oxygen measurements at two sites not served online (Figure 4). Bayesian methods have been developed for Marble and Grand Canyon using a light process error model and a reproducible workflow has been developed to update models based on continuous water quality data collected as part of Project A.

For Glen Canyon, modeling is complicated by the combination of sub-daily variation in flow and oxygen disequilibrium (due to the bottom water and mid-water releases from Lake Powell). We are currently developing a two-station method that builds upon earlier more time intensive modeling approaches (e.g., Payn and others, 2017) and that can estimate GPP fast enough to develop much longer time series (Bishop and others, 2024; Figure 4 top panel).

We are excited to explore controls on GPP time series throughout the CRe and propose to examine 1) response to disturbance, 2) evidence of nutrient limitation and lagged response to storm-based phosphorus inputs, and 3) reasons for asynchronous seasonal patterns throughout the river. We will implement a new approach to modeling primary producer biomass from GPP time series (Błaszczak and others, 2023) to quantify the resilience of the ecosystem to disturbance, and to examine the controls on this resilience. We will also use autoregressive integrated moving average (ARIMA) models for post-estimation analysis of controls on GPP.

The purpose of the proposed spring and fall benthic diatom surveys is to comprehensively document the distribution and character of the algal group most directly connected to local and valued aquatic food webs. It has been nearly 50 years since Czarnecki and others (1976) collected and analyzed benthic diatom communities from Lees Ferry down through the Western Grand Canyon, and much has changed since then. Dominant invertebrate grazer communities have shifted with the maturing presence of multiple introduced species (*Gammarus* and New Zealand mud snails (*Potamopyrgus antipodarum*); (Cross and others, 2010), hydrology has evolved through increasingly constrained load-following and regular high flow experiments (see Figure 2 in Sankey and others, 2018), and more recently elevational change in Lake Powell has driven strong changes to water quality, most notably temperature and phosphorus concentration (Dibble and others, 2021; Deemer and others, 2023b; U.S. Geological Survey, 2024).

For the FY 2025-27 work plan, these surveys will be opportunistically conducted by accompanying 2 seasonal monitoring river trips (April-May and September FY 2025), periods during which annual GPP peaks. At approximately 20 mainstem and confluence sites, a representative set of benthic substrates will be sampled, combined, and processed for microscopic and genetic metabarcoding analysis. Ian Bishop, a recently hired post-doc at GCMRC, has extensive expertise in collecting and analyzing such communities using both traditional microscopic and modern genomic methods, which will facilitate comparison to past and future literatures. Additional samples will be collected for targeted, low-effort microscopic scanning to improve spatial distribution resolution for important taxa (e.g., didymo).

The primary goal here is to produce species-level checklists, abundances, and distributional patterns, which will be related back to key environmental drivers. To further contextualize these data, we will assess the extent to which they correspond to two active independent data streams: 1) the monthly sampling of phytoplankton composition for lower Lake Powell and Lees Ferry (20+ years of data), and 2) longitudinal patterns of GPP measured via continuous DO monitoring throughout the study area. The former will help in assessing how current downstream GPP is subsidized by upstream communities, and the latter will provide a snapshot of the specific producer communities underlying high spring and fall GPP.

The unfunded vegetation mapping project aims to build on work completed during previous work plans. Already, we have worked towards the development of a deep convolutional neural network model by 1) utilizing image processing software to manually label vegetation classes on underwater images, 2) creating a library of vegetation labels to train model development, 3) compiling images to feed into the model framework, and 4) running segmentation models on >50,000 images. We proposed to focus work in the FY 2025-27 work plan on analyzing existing imagery, refining and publishing the deep learning model, and creating maps from which to compare future disturbance events or community change over time. This work would include an analysis of flow impacts to aquatic vegetation composition and cover in two sites that overlap long-term rainbow trout monitoring sites in Lees Ferry.

Anticipated Use of Data

This project element provides the underlying modeling to support the natural processes GPP metric and aims to disentangle the drivers of both rates and types of riverine primary production and their link back to fish production. This project will also help clarify how resilient the diatom base of the food web is to recent environmental change. It will establish a modern (molecular) baseline for the abundance and distribution of this important producer group throughout the canyons, including for salient, ecologically relevant taxa such as nitrogen-fixing (e.g., *Epithemia* spp.) and nuisance bloom-producing diatoms (Figure 2).

The unfunded vegetation mapping project would provide a machine learning model that could be used to classify images and produce maps of submerged aquatic vegetation in Lees Ferry. Such maps could serve as a baseline to answer questions related to the effects of dam operations and reservoir conditions on dominant primary producers that fuel the “green” food web consumed by aquatic invertebrates and fishes.

Outcomes and Products

Outcomes

- We expect that this project element will further our understanding of the controls on riverine GPP and how controls change spatially throughout the canyon. We also expect to improve our understanding of how GPP and P relate to the most biologically available primary producer pool in Glen Canyon. We expect to gain a better understanding how distinct algal and macrophyte communities each shape trends in whole-river GPP. Additionally, we will set the base upon which to evaluate future changes in aquatic vegetation communities in Glen Canyon to measure long-term trends and/or the short-term effects of experimental flows on primary producers. Lastly, we expect to establish a modern baseline for the most food web-relevant producer community in the CRe (diatoms) by surveying their abundance and distribution for the first time in nearly 50 years.

Products

- One to two journal articles identifying the most important controls on riverine GPP and describing how these controls vary spatially.
- One journal article describing aquatic vegetation communities in the CRe.

Project Element E.3. Understanding the Energetic Basis of the Food Web in Western Grand Canyon (New Study; Unfunded)

Eric Scholl¹, Ian Bishop¹, Bridget Deemer¹, Theodore Kennedy¹, Charles Yackulic¹

¹U.S. Geological Survey, Southwest Biological Science Center, Grand Canyon Monitoring and Research Center

Hypotheses and Science Questions

- Decomposition rates will be faster in Western Grand Canyon compared to Lees Ferry due to warmer water temperatures.
- The brown food web (decomposition) will be driven more by microbes in Western Grand Canyon compared to Lees Ferry due to a lower abundance of invertebrates and the absence of New Zealand mud snails.
- Cotton strip assays can serve as an effective monitoring tool to track the brown food web (decomposition and ecosystem respiration).

Methods

This project element aims to understand controls on the base of the food web in Western Grand Canyon. We propose to quantify rates and drivers of decomposition, substrate-scale respiration, and whole-ecosystem primary production and ecosystem respiration in the Western Grand Canyon by deploying a miniDOT (PME) logger for metabolism estimates and, to the extent that outside funding continues, leveraging a new gaging station near Columbine. While turbulent whitewater conditions preclude estimation of respiration rates in most of Grand Canyon, we expect the lower gradient reach approaching Pearce Ferry to support paired estimates of GPP and ecosystem respiration. We also propose targeted decomposition experiments on natural litter and cotton strip assays to understand controls on detritivore vs. microbe-driven decomposition and how changes in temperature, nutrient availability, and litter quality may influence food availability for higher trophic levels. By pairing integrative estimates of ecosystem metabolism with finer-scale decomposition and respiration measurements, our approach will offer a uniquely holistic perspective on the energetic pathways underpinning food webs in Western Grand Canyon.

Leaf decomposition at both Lees Ferry and Western Grand Canyon will be estimated during one season following protocols outlined by Scholl and others, 2024. We will collect leaves from the banks of the Colorado River in the fall of 2024 and allow the litter to dry in the laboratory prior to the start of the experiment. In the laboratory we will measure 4 g dry mass of leaf litter into 50 (5 replicates*5 collections [day 1,2,7,21,84]*2 locations) coarse mesh and 50 fine mesh packs.

Different mesh sizes will be used to isolate the potential effects of larger macroinvertebrate consumers on decomposition. Leaf packs will be deployed at locations with similar environmental conditions (e.g., low velocity) at both sites. At each site we will submerge five separate chains and randomly tie five coarse mesh and five fine mesh packs to each chain. On each collection date we will randomly collect one coarse and one fine mesh bag per chain, place the packs into labeled bags, and store on ice until laboratory processing. In the laboratory, samples will be rinsed, sorted from invertebrates, dried, weighed, and combusted to estimate ash-free dry mass (AFDM) of leaf packs throughout the experiment. Percent leaf mass remaining (AFDM) will be natural log-transformed and regressed over time and degree days to estimate decay rate (k) and temperature corrected decay rate (k_{dd}). We will then compare decay rate slope estimates among mesh sizes and locations to examine the effects of location, temperature, and invertebrates vs. microbes on decomposition.

At each site where natural litter bags are placed, we will also deploy standardized cotton strips following protocols outlined in Tiegs and others, 2013. We will construct 50 cotton strips (each ~80 mm x 25 mm strip from heavy-weight cotton fabric) so that 5 replicates will be collected on the same dates at natural sources (see above). Upon removal from the water column, we will measure biofilm respiration in stream-rinsed jars pre-fixed with non-invasive dissolved oxygen optodes (PreSens Precision Sensing GmbH, Regensburg, Germany; Deemer and others, 2023a). Initial dissolved oxygen (DO) measurements will be recorded for each jar, which will then be stored in stable temperature in the dark for 2-4 hours, after which final DO measurements will be taken. Three controls on each date will be used to measure ambient respiration in the water column and respiration rate for each strip will be estimated as

$$((DO_{stream} - DO_{strip\ post}) - (DO_{stream} - DO_{control\ post})) \times V_{H_2O\ jar} / m_{strip}/t$$

where DO_{stream} is the initial DO concentration of river water, $DO_{strip\ post}$ is the DO concentration after incubation, $DO_{control\ post}$ is the DO concentration of the control jars after incubation, $V_{H_2O\ jar}$ is the volume of river water used, m_{strip} is the dry mass of the cotton strip, and t is the incubation duration.

Following respiration measurements, each strip will be placed in a tray, gently cleaned using a paint brush and 95-100% ethanol and stored in individual aluminum envelopes. In the laboratory, each strip will be dried at 40°C, weighed, and measured for tensile strength using a Mark-10 tensiometer affixed to a motorized stand. The rate of tensile strength loss will be estimate as

$$k_{strip} = (-\ln(T_c/T_{sc}))/days$$

where T_s is the maximum tensile strength recorded for each incubated strip, T_{sc} is the mean tensile strength of the control strips, and days is incubation duration in days. Similar to natural leaves, we will also isolate the effects of temperature by standardizing for degree days.

At each site we will measure daily water temperature using either the USGS Colorado River at Lees Ferry gaging station (09380000) or a submerged HOBO[®] temperature logger in Western Grand Canyon.

Three replicates of total and dissolved P will be measured at three time points throughout the decomposition experiment ($n = 18$; 3 replicates*3 times * 2 locations) following protocols outlined by Deemer and others (2023b). Finally, estimates of whole-ecosystem metabolism (GPP and ER) will be made using a miniDOT and protocols outlined in Deemer and others, 2022a. We expect to successfully estimate ER in Western Grand Canyon given the relatively minimal whitewater in this reach (whereas microbubble entrainment precludes ER estimation at the long-term gaging stations in Figure 5; Hall and others, 2015).

Anticipated Use of Data

This project would elucidate energetic pathways underpinning food webs in Western Grand Canyon. The project would help test the hypothesis that warming water temperatures are supporting a more productive “brown” food web.

Outcomes and Products

Outcomes

- No outcomes (unfunded).

Products

- No products (unfunded).

Project Element E.4. Linking Ecosystem Metabolism to Higher Trophic Levels (Ongoing Study)

Charles B. Yackulic¹, Kimberly Dibble¹, Eric Scholl¹, Drew Eppheimer¹, David Ward², Eric Frye¹

¹U.S. Geological Survey, Southwest Biological Science Center, Grand Canyon Monitoring and Research Center

²U.S. Fish and Wildlife Service

Hypotheses and Science Questions

- H8: Humpback chub and flannelmouth sucker have lower basal metabolic demands than related taxa.

- H9: If humpback chub and flannelmouth sucker have lower basal metabolic demands than related taxa, the ecosystem can sustain large populations of these species despite relatively low primary production and these species can survive through relatively extended periods of low food availability.
- H10: Native fish species in the Western Grand Canyon may be approaching abundances at which food limitation becomes more important in regulating population dynamics.

Methods

The goal of this project is to develop ecosystem models that incorporate data collected at multiple trophic levels. This modeling will leverage ecosystem production rates estimated in Project Elements E.2 and E.3. together with estimates of consumption by fish populations to understand how bottom-up and top-down processes interact to drive ecosystem dynamics and determine carrying capacity for consumers (Figure 7). This type of approach was recently taken to understand energetic constraints on the rainbow trout population in Glen Canyon (Yard and others, 2023). The modeling will combine estimates of species and size class specific somatic growth and survival rates with estimates of basal metabolism to estimate energetic demand for individual fish, which will then be scaled up with species and size class specific abundance estimates to estimate the consumption required to support observed abundances, survival, and growth rates. These estimates of consumption will then be compared with estimates of energy production at lower trophic levels (i.e., ecosystem metabolism) or invertebrate availability (i.e., drift measurements) accounting for the transfer efficiency of energy across trophic levels to understand how changes in energy production impact fish population dynamics.

We will combine fish population abundances and somatic growth estimates from fixed sites (i.e., TRGD in Glen Canyon, Juvenile Chub Monitoring (JCM), and JCM-west – see Projects G and H), invertebrate drift data (Project F) with a better understanding of primary production (Project Element E.2), and lab measures of fish standard and active metabolic rates to develop ecosystem models. For the latter, acquiring a better understanding of respiratory physiology and the energy necessary for basic physiological functioning will provide better context for the effects of changes in dam operations and environmental conditions on the growth and production of native and nonnative fish species (Chabot and others, 2016a).

The metabolic rate (MR) of fishes is influenced by factors including body size, water temperature, activity level, consumption/food intake, physiological state, and anabolism (Chabot and others, 2016b). The basal, or standard metabolic rate (SMR) represents the minimum amount of energy needed for a fish to persist in its environment at a given temperature and does not include the added energy needed for growth, digestion, activity, and reproduction. Below SMR, physiological function in fishes is impaired and usually leads to mortality. Active metabolic rate (AMR) is at the other end of the spectrum and represents a fish's maximum aerobic metabolic rate at a specific temperature.

The difference between SMR (minimum) and AMR (maximum) provide an indication of the total amount of energy available to a fish (Norin and Malte, 2011; Chabot and others, 2016b). Laboratory experiments will be used to determine the standard and active metabolic rates of large-bodied native and nonnative fishes in the Grand Canyon that dominate the biomass of fish communities and for which there are no literature values. These data will be integrated into an aquatic ecosystem model that seeks to understand dynamics of fish communities throughout different reaches of the Grand Canyon from an energetic perspective.

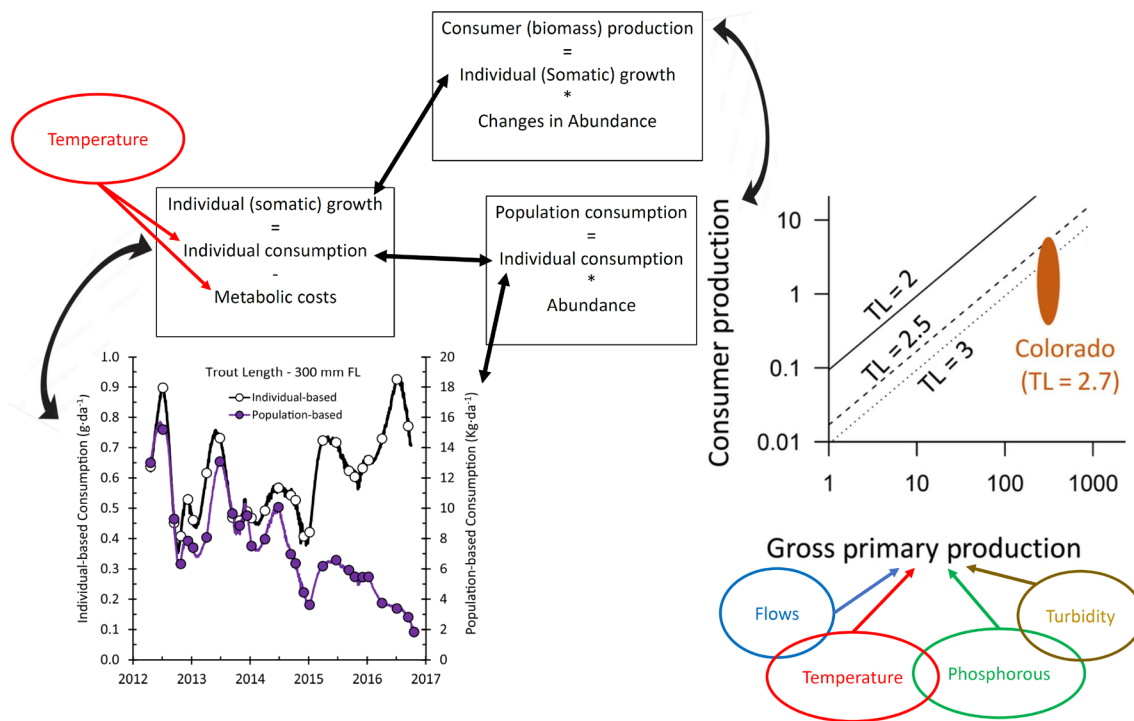


Figure 7. Conceptual model linking primary production to fish consumption, production and ultimately individual growth and population dynamics. Many of the processes linking primary production to fish population dynamics are affected by environmental drivers (e.g., flow, turbidity, thermal, and nutrient regimes), and we only identify a few linkages here. TL refers to trophic level and the value of 2.7 is based on prior studies of flannelmouth sucker (*Catostomus latipinnis*) (Cross and others, 2013). Figure is based on modification of figures presented in Rüegg and others (2020) and Yard and others (2023).

Measuring fish metabolism in controlled laboratory settings is relatively straightforward (Chabot and others, 2016b). Standard and active metabolic rates are quantified by measuring oxygen removal from the water column (i.e., measured oxygen uptake, MO_2 , or respiration) by a fish, and then converting oxygen uptake to units of energy used. Fish specimens representing juvenile, sub-adult, and adult size ranges (ranging from 50-500 mm TL) will be selected from laboratory-grown stock housed at the U.S. Forest Service Rocky Mountain Research Station Fish Laboratory in Flagstaff, Arizona.

Multiple size classes of fishes will be used in this experiment to allometrically scale metabolic rates. The size range selected will overlap with fish size as measured in the field, so results are applicable to field specimens. We will measure SMR and AMR in humpback chub and flannelmouth sucker, in addition to nonnative species such as smallmouth bass (and others), as long as we can obtain necessary size classes of specimens for use in laboratory research.

Fish will be kept in recirculating tanks at ambient temperature and fed a maintenance diet equivalent to approximately 0.5-1% body weight per day using species-specific or general aquaculture feed containing a mixture of essential amino and fatty acids, proteins, and fiber until trials commence.

Standard metabolic rates of individual fish will be measured using automated intermittent flow respirometry in a Loligo Systems Core Resting Respirometer in a chamber selected to match the size of each fish. Briefly, replicates of closed acrylic respirometer chambers will be submerged in temperature treatments representing a range of historic and potential future thermal regimes of the Colorado River in Grand Canyon (10°C, 15°C, 20°C, 25°C). An ultraviolet sterilizer of flow-through water will be used to minimize respiration from bacteria in the water column. Fully aerated water will be introduced into each chamber via a recirculating pump. Individual fish will be placed in each chamber to acclimate, and a series of oxygen consumption measurements will be taken. Active metabolic rates will be measured in a similar manner, but specimens will be actively chased in a tank prior to placement in each chamber (Killen and others, 2021).

Individual fish will undergo multiple trials to assess repeatability.

Since native fishes such as humpback chub are diurnally active as young-of-year but switch to nocturnal foraging as adults (Stone and Gorman 2005), we hypothesize that metabolic rates may differ between juveniles, sub-adults, and adults and during daytime and nighttime hours. Diel variation in metabolic rates will be determined using 24-h intermittent flow respirometry that accounts for photoperiod and temperature fluctuations. These measurements will account for physiological mechanisms in place that alter SMR in response to diel rhythms and light availability (e.g., Speers-Roesch and others, 2018, Wheeler and others, 2022), which could affect our interpretation of the energetic demands of fishes during different seasons in Grand Canyon.

Ecosystem models will be built by coupling seasonal estimates of GPP, invertebrate drift, and fish populations in a series of integrated difference equations that estimate quantities like transfer efficiencies using priors informed by intensive food web studies and ongoing diet studies. This work will expand on efforts already made to estimate the amount of invertebrate consumption required to support observed variation in rainbow trout biomass and growth in Glen Canyon by extending this approach to the more diverse fish communities found in the JCM and JCM-west reference reaches.

Anticipated Use of Data

This project element will integrate data on primary production, insect drift, fish growth, and population size to understand trophic linkages and better predict how the ecosystem will respond to changes in nutrients, temperature, and flow. The improved understanding of species-specific respiratory physiology will help predict how changes in dam operations will influence growth and production of native and nonnative fish species.

Outcomes and Products

Outcomes

- We expect that this project element will provide a better understanding of the amount of energy needed for native and nonnative fish growth. Ecosystem models will be developed that link primary production and changes in ecosystem drivers (temperature, flow, nutrients) to higher trophic levels across the CRe. This information could be used to develop an estimate of the carrying capacity of Western Grand Canyon to support fish communities.

Products

- One journal article that reports on the standard and active metabolic rates of native and nonnative fishes that lack parameter estimates (i.e., inputs into the bioenergetics and/or ecosystem model for species such as humpback chub and flannelmouth sucker).
- One to two journal articles describing correlations between GPP, invertebrate drift and fish condition/demography at various sites throughout the river and introducing a mathematical ecosystem model linking flow, temperature, nutrients, and energy needs to multiple trophic levels.

Budgets

Project Element Budget

| Fiscal Year 2025 | | | | | | | | | |
|---|------------------|----------------------|-----------------------|-----------------------|---------------------------|--------------------------|-----------------------------|------------------|------------------------------|
| Project E Controls on Ecosystem Productivity: Nutrients, Flow, and Temperature | Salaries | Travel & Training | Operating Expenses | Logistics Expenses | Cooperative Agreements | To other USGS Centers | AMP Special Burden Rate* | Total | USGS Contributing Funds** |
| | | | | | | | 21.86% | | Est Full Rate = 55% |
| E.1. Phosphorus budgeting in the Colorado River | \$19,449 | \$0 | \$3,500 | \$1,225 | \$0 | \$0 | \$5,285 | \$29,459 | |
| E.2. Rates and composition of primary producers in the Colorado River | \$217,868 | \$4,000 | \$24,637 | \$1,750 | \$0 | \$0 | \$54,276 | \$302,531 | |
| E.4. Productivity at higher trophic levels | \$96,625 | \$0 | \$3,000 | \$0 | \$0 | \$0 | \$21,781 | \$121,406 | |
| Total Project E | \$333,942 | \$4,000 | \$31,137 | \$2,975 | \$0 | \$0 | \$81,342 | \$453,396 | \$38,854 |

| Fiscal Year 2026 | | | | | | | | | |
|---|------------------|----------------------|-----------------------|-----------------------|---------------------------|--------------------------|-----------------------------|------------------|------------------------------|
| Project E Controls on Ecosystem Productivity: Nutrients, Flow, and Temperature | Salaries | Travel & Training | Operating Expenses | Logistics Expenses | Cooperative Agreements | To other USGS Centers | AMP Special Burden Rate* | Total | USGS Contributing Funds** |
| | | | | | | | 22.60% | | Est Full Rate = 55.5% |
| E.1. Phosphorus budgeting in the Colorado River | \$20,358 | \$0 | \$3,500 | \$700 | \$0 | \$0 | \$5,550 | \$30,109 | |
| E.2. Rates and composition of primary producers in the Colorado River | \$100,742 | \$4,000 | \$28,676 | \$1,750 | \$0 | \$0 | \$30,548 | \$165,717 | |
| E.4. Productivity at higher trophic levels | \$69,525 | \$0 | \$3,000 | \$0 | \$0 | \$0 | \$16,391 | \$88,915 | |
| Total Project E | \$190,625 | \$4,000 | \$35,176 | \$2,450 | \$0 | \$0 | \$52,489 | \$284,740 | \$24,611 |

| Fiscal Year 2027 | | | | | | | | | |
|---|------------------|----------------------|-----------------------|-----------------------|---------------------------|--------------------------|-----------------------------|------------------|------------------------------|
| Project E Controls on Ecosystem Productivity: Nutrients, Flow, and Temperature | Salaries | Travel & Training | Operating Expenses | Logistics Expenses | Cooperative Agreements | To other USGS Centers | AMP Special Burden Rate* | Total | USGS Contributing Funds** |
| | | | | | | | 23.40% | | Est Full Rate = 56% |
| E.1. Phosphorus budgeting in the Colorado River | \$44,632 | \$0 | \$3,500 | \$0 | \$0 | \$0 | \$11,263 | \$59,395 | |
| E.2. Rates and composition of primary producers in the Colorado River | \$63,290 | \$0 | \$18,044 | \$1,750 | \$0 | \$0 | \$19,442 | \$102,526 | |
| E.4. Productivity at higher trophic levels | \$50,485 | \$0 | \$3,000 | \$0 | \$0 | \$0 | \$12,516 | \$66,001 | |
| Total Project E | \$158,408 | \$0 | \$24,544 | \$1,750 | \$0 | \$0 | \$43,220 | \$227,922 | \$19,881 |

*GCDAMP Special Rate: Rate includes current DOI preferred rate (currently 15%; subject to change) and facilities rate (will vary annually). No USGS bureau overhead is charged (unique to the GCDAMP agreement).

**USGS Contributing Funds: The amount of funds required to cover the subsidy created by the reduced burden rate (i.e., GCDAMP Special Rate). As in previous years, SBSC/GCMRC will request these funds of the USGS cost-share program. These funds are not guaranteed.

Unfunded Project Element Budget

These tables show costs for the overplus P uptake experiments described in project element E.1, the aquatic vegetation mapping work described in project element E.2, and the overall cost of proposed work in element E.3.

| Fiscal Year 2025 | | | | | | | | |
|---|-----------------|----------------------|-----------------------|-----------------------|---------------------------|--------------------------|-----------------------------|-----------------|
| Project E Unfunded Project Elements | Salaries | Travel & Training | Operating Expenses | Logistics Expenses | Cooperative Agreements | To other USGS Centers | AMP Special Burden Rate* | Total |
| | | | | | | | 21.86% | |
| E.1. Phosphorus budgeting in the Colorado River (<i>P uptake experiments</i>) | \$0 | \$0 | \$9,345 | \$1,000 | \$0 | \$0 | \$2,255 | \$12,600 |
| E.2. Rates and composition of primary producers in the Colorado River (<i>vegetation mapping project</i>) | \$0 | \$0 | \$0 | \$0 | \$0 | \$0 | \$0 | \$0 |
| E.3. Understanding the energetic basis of the food web in Western Grand Canyon | \$20,811 | \$2,000 | \$10,950 | \$4,200 | \$0 | \$0 | \$8,276 | \$46,237 |
| Total Project E | \$20,811 | \$2,000 | \$20,295 | \$5,200 | \$0 | \$0 | \$10,531 | \$58,837 |

| Fiscal Year 2026 | | | | | | | | |
|---|------------------|----------------------|-----------------------|-----------------------|---------------------------|--------------------------|-----------------------------|------------------|
| Project E Unfunded Project Elements | Salaries | Travel & Training | Operating Expenses | Logistics Expenses | Cooperative Agreements | To other USGS Centers | AMP Special Burden Rate* | Total |
| | | | | | | | 22.60% | |
| E.1. Phosphorus budgeting in the Colorado River (<i>P uptake experiments</i>) | \$69,251 | \$2,000 | \$13,500 | \$1,000 | \$0 | \$0 | \$19,380 | \$105,131 |
| E.2. Rates and composition of primary producers in the Colorado River (<i>vegetation mapping project</i>) | \$10,710 | \$0 | \$0 | \$0 | \$0 | \$0 | \$2,420 | \$13,130 |
| E.3. Understanding the energetic basis of the food web in Western Grand Canyon | \$65,843 | \$1,000 | \$3,500 | \$0 | \$0 | \$0 | \$15,898 | \$86,241 |
| Total Project E | \$145,804 | \$3,000 | \$17,000 | \$1,000 | \$0 | \$0 | \$37,698 | \$204,502 |

| Fiscal Year 2027 | | | | | | | | |
|---|-----------------|----------------------|-----------------------|-----------------------|---------------------------|--------------------------|-----------------------------|-----------------|
| Project E Unfunded Project Elements | Salaries | Travel & Training | Operating Expenses | Logistics Expenses | Cooperative Agreements | To other USGS Centers | AMP Special Burden Rate* | Total |
| | | | | | | | 23.40% | |
| E.1. Phosphorus budgeting in the Colorado River (<i>P uptake experiments</i>) | \$0 | \$0 | \$0 | \$0 | \$0 | \$0 | \$0 | \$0 |
| E.2. Rates and composition of primary producers in the Colorado River (<i>vegetation mapping project</i>) | \$22,918 | \$0 | \$40,000 | \$0 | \$0 | \$0 | \$14,723 | \$77,641 |
| E.3. Understanding the energetic basis of the food web in Western Grand Canyon | \$0 | \$0 | \$0 | \$0 | \$0 | \$0 | \$0 | \$0 |
| Total Project E | \$22,918 | \$0 | \$40,000 | \$0 | \$0 | \$0 | \$14,723 | \$77,641 |

References Cited

- Andrews, C.M., and Deemer, B.R., 2022, Limnology data from Lake Powell, desert southwest USA: U.S. Geological Survey data release, <https://doi.org/10.5066/P9ZIKVYW>.
- Battin, T.J., Lauerwald, R., Bernhardt, E.S., Bertuzzo, E., Gener, L.G., Hall, R.O., Jr., Hotchkiss, E.R., Maavara, T., Pavelsky, T.M., Ran, L., Raymond, P., Rosentreter, J.A., and Regnier, P., 2023, River ecosystem metabolism and carbon biogeochemistry in a changing world: *Nature*, v. 613, no. 7944, p. 449–459, <https://doi.org/10.1038/s41586-022-05500-8>.
- Benenati, P.L., Shannon, J.P., and Blinn, D.W., 1998, Desiccation and recolonization of phytobenthos in a regulated desert river—Colorado River at Lees Ferry, Arizona, USA: *Regulated Rivers: Research and Management*, v. 14, no. 6, p. 519-532, [https://doi.org/10.1002/\(SICI\)1099-1646\(1998110\)14:6<519::AID-RRR518>3.0.CO;2-H](https://doi.org/10.1002/(SICI)1099-1646(1998110)14:6<519::AID-RRR518>3.0.CO;2-H).
- Bernhardt, E.S., Heffernan, J.B., Grimm, N.B., Stanley, E.H., Harvey, J.W., Arroita, M., Appling, A.P., Cohen, M.J., McDowell, W.H., Hall, R.O., Jr., Read, J.S., Roberts, B.J., Stets, E.G., and Yackulic, C.B., 2018, The metabolic regimes of flowing waters: *Limnology and Oceanography*, v. 63, no. S1, p. S99-S118, <https://doi.org/10.1002/lno.10726>.
- Bernhardt, E.S., Savoy, P., Vlah, M.J., Appling, A.P., Koenig, L.E., Hall, R.O., Jr., Arroita, M., Blaszcak, J.R., Carter, A.M., Cohen, M., Harvey, J.W., Heffernan, J.B., Helton, A.M., Hosen, J.D., Kirk, L., McDowell, W.H., Stanley, E.H., Yackulic, C.B., and Grimm, N.B., 2022, Light and flow regimes regulate the metabolism of rivers: *Proceedings of the National Academy of Sciences (PNAS)*, v. 119, no. 8, e2121976119, <https://doi.org/10.1073/pnas.2121976119>.

- Beville, S.T., Kerr, G.N., and Hughey, K.F.D., 2012, Valuing impacts of the invasive alga *Didymosphenia geminata* on recreational angling: *Ecological Economics*, v. 82, p. 1-10, <https://doi.org/10.1016/j.ecolecon.2012.08.004>.
- Bishop, I.W., Deemer, B.R., Hall, R.O., Sabol, T.A., and Yackulic, C.B., 2024, Primary production and disturbance dynamics in Glen Canyon tailwaters [poster]: Glen Canyon Dam Adaptive Management Program, Annual Reporting Meeting, Phoenix, Ariz., January 23-24, 2024: prepared by U.S. Geological Survey, Southwest Biological Science Center, Grand Canyon Monitoring and Research Center, Flagstaff, Ariz.
- Blaszczak, J.R., Yackulic, C.B., Shriver, R.K., and Hall, J., Robert O., 2023, Models of underlying autotrophic biomass dynamics fit to daily river ecosystem productivity estimates improve understanding of ecosystem disturbance and resilience: *Ecology Letters*, v. 26, no. 9, p. 1510–1522, <https://doi.org/10.1111/ele.14269>.
- Blinn, D.W., Truitt, R., and Pickart, A., 1989, Response of epiphytic diatom communities from the tailwaters of Glen Canyon Dam, Arizona, to elevated water temperature: *Regulated Rivers: Research and Management*, v. 4, no. 1, p. 91-96, <https://doi.org/10.1002/rrr.3450040108>.
- Bothwell, M.L., Taylor, B.W., and Kilroy, C., 2014, The Didymo story—The role of low dissolved phosphorus in the formation of *Didymosphenia geminata* blooms: *Diatom Research*, v. 29, no. 3, p. 229-236, <https://doi.org/10.1080/0269249X.2014.889041>.
- Chabot, D., Steffensen, J.F., and Farrell, A.P., 2016a, The determination of standard metabolic rate in fishes: *The Journal of Fish Biology*, v. 88, no. 1, p. 81-121, <https://doi.org/10.1111/jfb.12845>.
- Chabot, D., McKenzie, D.J., and Craig, J.F., 2016b, Metabolic rate in fishes—Definitions, methods and significance for conservation physiology: *Journal of Fish Biology*, v. 88, no. 1, p. 1-9, <https://doi.org/10.1111/jfb.12873>.
- Colletti, P.J., Blinn, W., Pickart, A., and Wagner, V., 1987, Influence of different densities of the mayfly grazer *Heptagenia criddlei* on lotic diatom communities: *Journal of the North American Benthological Society*, v. 6, no. 4, p. 270–280, <https://doi.org/10.2307/1467314>.
- Corman, J.R., Moody, E.K., and Elser, J.J., 2016, Calcium carbonate deposition drives nutrient cycling in a calcareous headwater stream: *Ecological Monographs*, v. 86, no. 4, p. 448-461, <https://doi.org/10.1002/ecm.1229>.
- Cross, W.F., Baxter, C.V., Donner, K.C., Rosi-Marshall, E.J., Kennedy, T.A., Hall, R.O., Jr., Wellard-Kelly, H.A., and Rogers, R.S., 2011, Ecosystem ecology meets adaptive management—Food web response to a controlled flood on the Colorado River, Glen Canyon: *Ecological Applications*, v. 21, no. 6, p. 2016-2033, <https://doi.org/10.1890/10-1719.1>.
- Cross, W.F., Baxter, C.V., Rosi-Marshall, E.J., Hall, R.O., Jr., Kennedy, T.A., Donner, K.C., Wellard Kelly, H.A., Seegert, S.E.Z., Behn, K., and Yard, M.D., 2013, Food-web dynamics in a large river discontinuum: *Ecological Monographs*, v. 83, no. 3, p. 311-337, <https://doi.org/10.1890/12-1727.1>.

- Cross, W.F., Rosi-Marshall, E.J., Behn, K.E., Kennedy, T.A., Hall, R.O., Fuller, A.E., and Baxter, C.V., 2010, Invasion and production of New Zealand mud snails in the Colorado River, Glen Canyon: Biological Invasions, v. 12, no. 9, p. 3033-3043, <https://doi.org/10.1007/s10530-010-9694-y>.
- Czarnecki, D.B., Blinn, D.W., and Tompkins, T., 1976, A periphytic microflora analysis of the Colorado River and major tributaries in Grand Canyon National Park and vicinity: Flagstaff, Northern Arizona University, submitted to National Park Service, Grand Canyon National Park, Technical report no. 6, 106 p., <http://www.riversimulator.org/Resources/NPS/GCresearch/1976no6microflora.pdf>.
- Deemer, B.R., 2020, Lake Powell as a regulator of downstream water quality and ecosystem productivity [presentation]: Phoenix, Ariz., U.S. Geological Survey, Glen Canyon Dam Adaptive Management Program Annual Reporting Meeting, January 13, 2020, <https://www.usbr.gov/uc/progact/amp/twg/2020-01-13-twg-meeting/20200113-AnnualReportingMeeting-LakePowellRegulatorDownstreamWaterQualityEcosystemProductivity-Presentation-508-UCRO.pdf>.
- Deemer, B.R., Yackulic, C.B., Hall, R.O., Jr., Dodrill, M.J., Kennedy, T.A., Muehlbauer, J.D., Topping, D.J., Voichick, N., and Yard, M.D., 2022a, Experimental reductions in subdaily flow fluctuations increased gross primary productivity for 425 river kilometers downstream: PNAS Nexus, v. 1, no. 3, pgac094, <https://doi.org/10.1093/pnasnexus/pgac094>.
- Deemer, B.R., Yard, M.D., Voichick, N., Goodenough, D.C., Bennett, G.E., Hall Jr., R.O., Dodrill, M.J., Topping, D.J., Gushue, T., Muehlbauer, J.D., Kennedy, T.A., and Yackulic, C.B., 2022b, Gross primary production estimates and associated light, sediment, and water quality data from the Colorado River below Glen Canyon Dam: U.S. Geological Survey data release, <https://doi.org/10.5066/P9ZS6YLV>.
- Deemer, B.R., Andrews, C.M., Strock, K.E., Voichick, N., Hensleigh, J., Beaver, J.R., and Radtke, R., 2023a, Over half a century record of limnology data from Lake Powell, desert southwest United States—From reservoir filling to present day (1964–2021): Limnology and Oceanography Letters, v. 8, no. 4, p. 580-594, <https://doi.org/10.1002/lol2.10310>.
- Deemer, B.R., Reibold, R.H., Fatta, A., Corman, J.R., Yackulic, C.B., and Reed, S.C., 2023b, Storms and pH of dam releases affect downstream phosphorus cycling in an arid regulated river: Biogeochemistry, v. 165, p. 57–74, <https://doi.org/10.1007/s10533-023-01064-5>.
- Dibble, K.L., Yackulic, C.B., Kennedy, T.A., Bestgen, K.R., and Schmidt, J.C., 2021, Water storage decisions will determine the distribution and persistence of imperiled river fishes: Ecological Applications, v. 31, no. 2, e02279, p. 1-9, <https://doi.org/10.1002/eap.2279>.
- Dibble, K.L., Yard, M.D., Ward, D.L., and Yackulic, C.B., 2017, Does bioelectrical impedance analysis accurately estimate the physiological condition of threatened and endangered desert fish species?: Transactions of the American Fisheries Society, v. 145, no. 5, p. 888-902, <https://doi.org/10.1080/00028487.2017.1302993>.

- Eppehimer, D.E., Yackulic, C.B., Bruckerhoff, L.A., Wang, J., Young, K.L., Bestgen, K.R., Mihalevich, B.A., and Schmidt, J.C., 2024, Declining reservoir elevations following a two-decade drought increase water temperatures and non-native fish passage facilitating a downstream invasion: bioRxiv, <https://doi.org/10.1101/2024.01.23.576966>.
- Furey, P.C., Kupferberg, S.J., and Lind, A.J., 2014, The perils of unpalatable periphyton—*Didymosphenia* and other mucilaginous stalked diatoms as food for tadpoles: *Diatom Research*, v. 29, no. 3, p. 267–280, <https://doi.org/10.1080/0269249X.2014.924436>.
- Gillis, C.-A., and Chalifour, M., 2010, Changes in the macrobenthic community structure following the introduction of the invasive algae *Didymosphenia geminata* in the Matapedia River (Québec, Canada): *Hydrobiologia*, v. 647, no. 1, p. 63-70, <https://doi.org/10.1007/s10750-009-9832-7>.
- Gloss, S.P., 1977, Application of the nutrient loading concept to Lake Powell, the effects of nutrient perturbations on phytoplankton productivity, and levels of nitrogen and phosphorus in the reservoir: Albuquerque, University of New Mexico, Ph.D. thesis, 225 p., https://digitalrepository.unm.edu/biol_etds/290/.
- Hall, R.O., Jr., Yackulic, C.B., Kennedy, T.A., Yard, M.D., Rosi-Marshall, E.J., Voichick, N., and Behn, K., 2015, Turbidity, light, temperature, and hydropeaking control primary productivity in the Colorado River, Grand Canyon: *Limnology and Oceanography*, v. 60, no. 2, p. 512-526, <https://doi.org/10.1002/lno.10031>.
- Hansen, L.E., Yackulic, C.B., Dickson, B.G., Deemer, B.R., and Best, R.J., 2023a, Linking ecosystem processes to consumer growth rates—Gross primary productivity as a driver of freshwater fish somatic growth in a resource-limited river: *Canadian Journal of Fisheries and Aquatic Sciences*, v. 80, no. 9, p. 1456-1469, <https://doi.org/10.1139/cjfas-2022-0229>.
- Hansen, L.E., Deemer, B.D., Yackulic, C.B., 2023b, A decade of GPP data in a changing river [poster]: Phoenix, Ariz., Glen Canyon Dam Adaptive Management Program Annual Reporting Meeting, January 24-25, 2023.
- Hansson, L.-A., Ekvall, M.K., He, L., Li, Z., Svensson, M., Urrutia-Cordero, P., and Zhang, H., 2020, Different climate scenarios alter dominance patterns among aquatic primary producers in temperate systems: *Limnology and Oceanography*, v. 65, no. 10, p. 2328-2336, <https://doi.org/10.1002/lno.11455>.
- Harold, F.M., 1966, Inorganic polyphosphates in biology—Structure, metabolism, and function: *Bacteriological Reviews*, v. 30, no. 4, p. 772-794, <https://doi.org/10.1128/br.30.4.772-794.1966>.
- Hood, J.M., Benstead, J.P., Cross, W.F., Huryn, A.D., Johnson, P.W., Gíslason, G.M., Junker, J.R., Nelson, D., Ólafsson, J.S., and Tran, C., 2018, Increased resource use efficiency amplifies positive response of aquatic primary production to experimental warming: *Global Change Biology*, v. 24, no. 3, p. 1069-1084, <https://doi.org/10.1111/gcb.13912>.
- Kazanjian, G., Flury, S., Attermeyer, K., Kalettka, T., Kleeberg, A., Premke, K., Köhler, J., and Hilt, S., 2018, Primary production in nutrient-rich kettle holes and consequences for nutrient

- and carbon cycling: *Hydrobiologia*, v. 806, no. 1, p. 77–93, <https://doi.org/10.1007/s10750-017-3337-6>.
- Kelly, V.J., 2001, Influence of reservoirs on solute transport—A regional-scale approach: Hydrological Processes—Special Issue: Water Quality of Large US Rivers: Results from the U.S. Geological Survey's National Stream Quality Accounting Network, v. 15, no. 7, p. 1227-1249, <https://doi.org/10.1002/hyp.211>.
- Killen, S.S., Christensen, E.A.F., Cortese, D., Závorka, L., Norin, T., Cotgrove, L., Crespel, A., Munson, A., Nati, J.J.H., Papatheodoulou, M., and McKenzie, D.J., 2021, Guidelines for reporting methods to estimate metabolic rates by aquatic intermittent-flow respirometry: *Journal of Experimental Biology*, v. 224, no. 18, jeb242522, <https://doi.org/10.1242/jeb.242522>.
- Korman, J., Yard, M.D., Dzul, M.C., Yackulic, C.B., Dodrill, M.J., Deemer, B.R., and Kennedy, T.A., 2021, Changes in prey, turbidity, and competition reduce somatic growth and cause the collapse of a fish population: *Ecological Monographs*, v. 91, no. 1, e01427, p. 1-20, <https://doi.org/10.1002/ecm.1427>.
- Krist, A.C., and Charles, C.C., 2012, The invasive New Zealand mudsnail, *Potamopyrgus antipodarum*, is an effective grazer of algae and altered the assemblage of diatoms more than native grazers: *Hydrobiologia*, v. 694, p. 143-151, <https://doi.org/10.1007/s10750-012-1138-5>.
- Lange, K., Townsend, C.R., and Matthaei, C.D., 2016, A trait-based framework for stream algal communities: *Ecology and Evolution*, v. 6, no. 1, p. 23-36, <https://doi.org/10.1002/ece3.1822>.
- Lapointe, A., Kocademir, M., Bergman, P., Ragupathy, I.C., Laumann, M., Underwood, G.J.C., Zumbusch, A., Spiteller, D., and Kroth, P.G., 2024, Characterization of polyphosphate dynamics in the widespread freshwater diatom *Achnanthes minutissimum* under varying phosphorus supplies: *Journal of Phycology*, online, <https://doi.org/10.1111/jpy.13423>.
- Larson, A., and Carreiro, J., 2008, Relationships between nuisance blooms of *Didymosphenia geminata* and measures of aquatic community composition in Rapid Creek, South Dakota, in Bothwell, M.L., and Spaulding, S.A., eds., *Proceedings of the 2007 International Workshop on Didymosphenia geminata*, Canadian Technical Report of Fisheries and Aquatic Science 2795: Government of Canada, Department of Fisheries and Oceans, 45-49 p., https://publications.gc.ca/collections/collection_2012/mpo-dfo/Fs97-6-2795-eng.pdf.
- Lowman, H.E., Shriver, R.K., Hall Jr., R.O., Harvey, J., Savoy, P., Yackulic, C.B., and Blaszcak, J.R., 2024, Macroscale controls determine the recovery of river ecosystem productivity following flood disturbances: *Proceedings of the National Academy of Sciences*, v. 121, no. 5, e2307065121, p. 1-9, <https://doi.org/10.1073/pnas.2307065121>.
- Melis, T.S., Korman, J., and Kennedy, T.A., 2012, Abiotic and biotic responses of the Colorado River to controlled floods at Glen Canyon Dam, Arizona, USA: *River Research and Applications*, v. 28, no. 6, p. 764-776, <https://doi.org/10.1002/rra.1503>.

- Meyer, J.L., and Likens, G.E., 1979, Transport and transformation of phosphorus in a forest stream ecosystem: *Ecology*, v. 60, no. 6, p. 1255-1269, <http://www.jstor.org/stable/1936971>.
- Norin, T., and Malte, H., 2011, Repeatability of standard metabolic rate, active metabolic rate and aerobic scope in young brown trout during a period of moderate food availability: *Journal of Experimental Biology*, v. 214, no. 10, p. 1668-1675, <https://doi.org/10.1242/jeb.054205>.
- Payn, R.A., Hall, R.O., Jr., Kennedy, T.A., Poole, G.C., and Marshall, L.A., 2017, A coupled metabolic-hydraulic model and calibration scheme for estimating whole-river metabolism during dynamic flow conditions: *Limnology and Oceanography Methods*, v. 15, no. 10, p. 847-866, <https://doi.org/10.1002/lom3.10204>.
- Rüegg, J., Conn, C.C., Anderson, E.P., Battin, T.J., Bernhardt, E.S., Canadell, M.B., Bonjour, S.M., Hosen, J.D., Marzolf, N.S., and Yackulic, C.B., 2020, Thinking like a consumer— Linking aquatic basal metabolism and consumer dynamics: *Limnology and Oceanography Letters*, v. 6, no. 1, p. 1-17, <https://doi.org/10.1002/lo12.10172>.
- Runge, M.C., LaGory, K.E., Russell, K., Balsom, J.R., Butler, R.A., Coggins, L.G., Jr., Grantz, K., Hayse, J., Hlohowskyj, I., Korman, J., May, J.E., O'Rourke, D.J., Poch, L.A., Prairie, J.R., VanKuiken, J.C., Van Lonkhuyzen, R.A., Varyu, D.R., Verhaaren, B.T., Veselka, T.D., Williams, N.T., Wuthrich, K.K., Yackulic, C.B., Billerbeck, R.P., and Knowles, G.W., 2015, Decision analysis to support development of the Glen Canyon Dam long-term experimental and management plan: U.S. Geological Survey Scientific Investigations Report 2015-5176, 64 p., <https://doi.org/10.3133/sir20155176>.
- Sabo, J.L., Caron, M., Doucett, R.R., Dibble, K.L., Ruhi, A., Marks, J.C., Hungate, B.A., and Kennedy, T.A., 2018, Pulsed flows, tributary inputs, and food web structure in a highly regulated river: *Journal of Applied Ecology*, v. 55, no. 4, p. 1884-1895, <https://doi.org/10.1111/1365-2664.13109>.
- Sankey, J.B., Caster, J.J., Kasprak, A., and East, A.E., 2018, The response of source-bordering aeolian dunefields to sediment-supply changes 2—Controlled floods of the Colorado River in Grand Canyon, Arizona, USA: *Aeolian Research*, v. 32, p. 154-169, <https://doi.org/10.1016/j.aeolia.2018.02.004>.
- Scholl, E.A., Hanus, K.R., Gardner, T.W., and Kennedy, T.A., 2024, Multiple stressors mediate the effects of warming on leaf decomposition in a large regulated river: *Ecosphere*, v. 15, no. 3, e4804, <https://doi.org/10.1002/ecs2.4804>.
- Spaulding, S.A., Potapova, M.G., Bishop, I.W., Lee, S.S., Gasperak, T.S., Jovanoska, E., Furey, P.C., and Edlund, M.B., 2021, *Diatoms.org*—Supporting taxonomists, connecting communities: *Diatom Research*, v. 36, no. 4, p. 291–304, <https://doi.org/10.1080/0269249X.2021.2006790>.
- Speers-Roesch, B., Norin, T., and Driedzic, W.R., 2018, The benefit of being still—Energy savings during winter dormancy in fish come from inactivity and the cold, not from metabolic rate depression: *Proceedings of the Royal Society B—Biological Sciences*, v. 285, no. 1886, article 20181593, p. 1-10, <https://doi.org/10.1098/rspb.2018.1593>.

- Steinman, A.D., McIntire, C.D., Gregory, S.V., Lamberti, G.A., and Ashkenas, L.R., 1987, Effects of herbivore type and density on taxonomic structure and physiognomy of algal assemblages in laboratory streams: *Journal of the North American Benthological Society*, v. 6, no. 3, p. 175-188, <https://doi.org/10.2307/1467509>.
- Stone, D.M., and Gorman, O.T., 2006, Ontogenesis of endangered humpback chub (*Gila cypha*) in the Little Colorado River, Arizona: *The American Midland Naturalist*, v. 155, no. 1, p. 123-135, [https://doi.org/10.1674/0003-0031\(2006\)155\[0123:OOEHCG\]2.0.CO;2](https://doi.org/10.1674/0003-0031(2006)155[0123:OOEHCG]2.0.CO;2).
- Stone, D.M., Young, K.L., Mattes, W.P., and Cantrell, M.A., 2018, Abiotic controls of invasive nonnative fishes in the Little Colorado River, Arizona: *The American Midland Naturalist*, v. 180, no. 1, p. 119-142, <https://doi.org/10.1674/0003-0031-180.1.119>.
- Sundareshwar, P.V., Upadhayay, S., Abessa, M., Honomichl, S., Berdanier, B., Spaulding, S.A., Sandvik, C., and Trennepohl, A., 2011, *Didymosphenia geminata*—Algal blooms in oligotrophic streams and rivers: *Geophysical Research Letters*, v. 38, no. 10, <https://doi.org/10.1029/2010GL046599>.
- Tiegs, S.D., Clapcott, J.E., Griffiths, N.A., and Boulton, A.J., 2013, A standardized cotton-strip assay for measuring organic-matter decomposition in streams: *Ecological Indicators*, v. 32, p. 131-139, <https://doi.org/10.1016/j.ecolind.2013.03.013>.
- Topping, D.J., Rubin, D.M., and Vierra, L.E., Jr., 2000, Colorado River sediment transport—1. Natural sediment supply limitation and the influence of the Glen Canyon Dam: *Water Resources Research*, v. 36, no. 2, p. 515-542, <https://doi.org/10.1029/1999WR900285>.
- U.S. Department of Interior, 2016, Glen Canyon Dam Long-Term Experimental and Management Plan final Environmental Impact Statement (LTEMP EIS): U.S. Department of the Interior, Bureau of Reclamation, Upper Colorado Region, National Park Service, Intermountain Region, online, <http://ltempeis.anl.gov/documents/final-eis/>.
- U.S. Geological Survey, Grand Canyon Monitoring and Research Center, 2024, U.S. Geological Survey Grand Canyon Monitoring and Research Center Proceedings of the Fiscal Year 2023 Annual Reporting Meeting to the Glen Canyon Dam Adaptive Management Program: Flagstaff, Ariz., prepared by U.S. Geological Survey, Southwest Biological Science Center, submitted to Bureau of Reclamation, Phoenix, Ariz., January 23-25, 2024, 189 p., <https://pubs.usgs.gov/publication/70251210>.
- Wellard Kelly, H.A., Rosi-Marshall, E.J., Kennedy, T.A., Hall, R.O., Cross, W.F., and Baxter, C.V., 2013, Macroinvertebrate diets reflect tributary inputs and turbidity-driven changes in food availability in the Colorado River downstream of Glen Canyon Dam: *Freshwater Science*, v. 32, no. 2, p. 397-410, <https://doi.org/10.1899/12-088.1>.
- Wheeler, C.R., Kneebone, J., Heinrich, D., Strugnell, J.M., Mandelman, J.W., and Rummer, J.L., 2022, Diel rhythm and thermal independence of metabolic rate in a benthic shark: *Journal of Biological Rhythms*, v. 37, no. 5, p. 484-497, <https://doi.org/10.1177/07487304221107843>.
- Wildman, R.A., and Vernieu, W.S., 2017, Turbid releases from Glen Canyon Dam, Arizona, following rainfall-runoff events of September 2013: *Lake and Reservoir Management*, v. 33, no. 3, p. 211-216, <https://doi.org/10.1080/10402381.2017.1293756>.

- Yackulic, C.B., 2020, Predictions, populations, and energetic constraints [presentation]: Phoenix, Ariz., Glen Canyon Dam Adaptive Management Program, Annual Reporting Meeting, January 13, 2020, U.S. Geological Survey, Southwest Biological Science Center, Grand Canyon Monitoring and Research Center, <https://www.usbr.gov/uc/progact/amp/twg/2020-01-13-twg-meeting/20200113-AnnualReportingMeeting-PredictionsPopulationsEnergeticConstraints-Presentation-508-UCRO.pdf>.
- Yard, M.D., Bennett, G.E., Mietz, S.N., Coggins, L.G., Jr., Stevens, L.E., Hueftle, S.J., and Blinn, D.W., 2005, Influence of topographic complexity on solar insolation estimates for the Colorado River, Grand Canyon, AZ: Ecological Modelling, v. 183, no. 2-3, p. 157-172, <https://doi.org/10.1016/j.ecolmodel.2004.07.027>.
- Yard, M.D., Yackulic, C.B., Korman, J., Dodrill, M.J., and Deemer, B.R., 2023, Declines in prey production during the collapse of a tailwater rainbow trout population are associated with changing reservoir conditions: Transactions of the American Fisheries Society, v. 152, no. 1, p. 35-50, <https://doi.org/10.1002/tafs.10381>.

Project F: Aquatic Invertebrate Ecology

Investigators

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Project Summary and Purpose

The primary focus of Project F is continuation of long-term monitoring needed to track ecosystem response to Macroinvertebrate Production Flows (Bug Flows) and other Long-Term Experimental and Management Plan (LTEMP, Department of Interior 2016) experiments. Research by our group has demonstrated that the scarcity of mayflies, stoneflies, and caddisflies from the Colorado River is partly due to acute mortality of insect eggs arising from hourly changes in discharge associated with hydropower generation (Kennedy and others, 2016; Miller and others, 2020; Figure 1). In May–August of 2018–2020 and 2022, Glen Canyon Dam operations were experimentally modified to evaluate whether low steady flows on weekends would increase the production and diversity of aquatic insects in the Colorado River ecosystem (CRe). These experimental Bug Flows involved hourly flow fluctuations for hydropower generation during weekdays, coupled with steady, low flows on weekends to reduce aquatic insect egg desiccation and mortality. Project F is tracking ecosystem response to the Bug Flows experiment and other ongoing or potential management actions using community science monitoring of aquatic insects (F.1), monitoring of invertebrate drift (F.1 and F.2), monitoring of invertebrate communities in tributaries and the mainstem Colorado River using environmental-DNA (eDNA; F.3), and quantifying feeding habits, parasite loads, and overall health of native and nonnative fishes using DNA analysis of feces and stable isotope analysis of fin clips (F.4). To reduce costs and facilitate processing of existing samples and analysis and synthesis of existing data, we are proposing to reduce all monitoring and sample collections, and some sample collections will be discontinued entirely (e.g., all bat monitoring will be discontinued, sticky trap and light trap sampling in Glen Canyon will be discontinued, and annual collections of invertebrate drift samples throughout Grand Canyon will be discontinued).

Research and monitoring of invertebrates and fish feeding habits described in Project F also provides essential context and data that are used by other projects. For example, invertebrate monitoring data are used by Project E (controls on ecosystem productivity) to identify the extent to which changing nutrient levels are propagating up through the food web.

Data on invertebrate populations and fish feeding habits also aid interpretation of seasonal and annual trends in humpback chub (*Gila cypha*; Project G), rainbow trout (*Oncorhynchus mykiss*; Project H), and smallmouth bass (*Micropterus dolomieu*; Project I), because aquatic invertebrates represent the food base for these species of fish during some, or all, of their life-stages. Project F also integrates and uses data from other projects, particularly Project A (streamflow, water quality, and sediment transport), to identify how changing environmental conditions affect invertebrate populations.

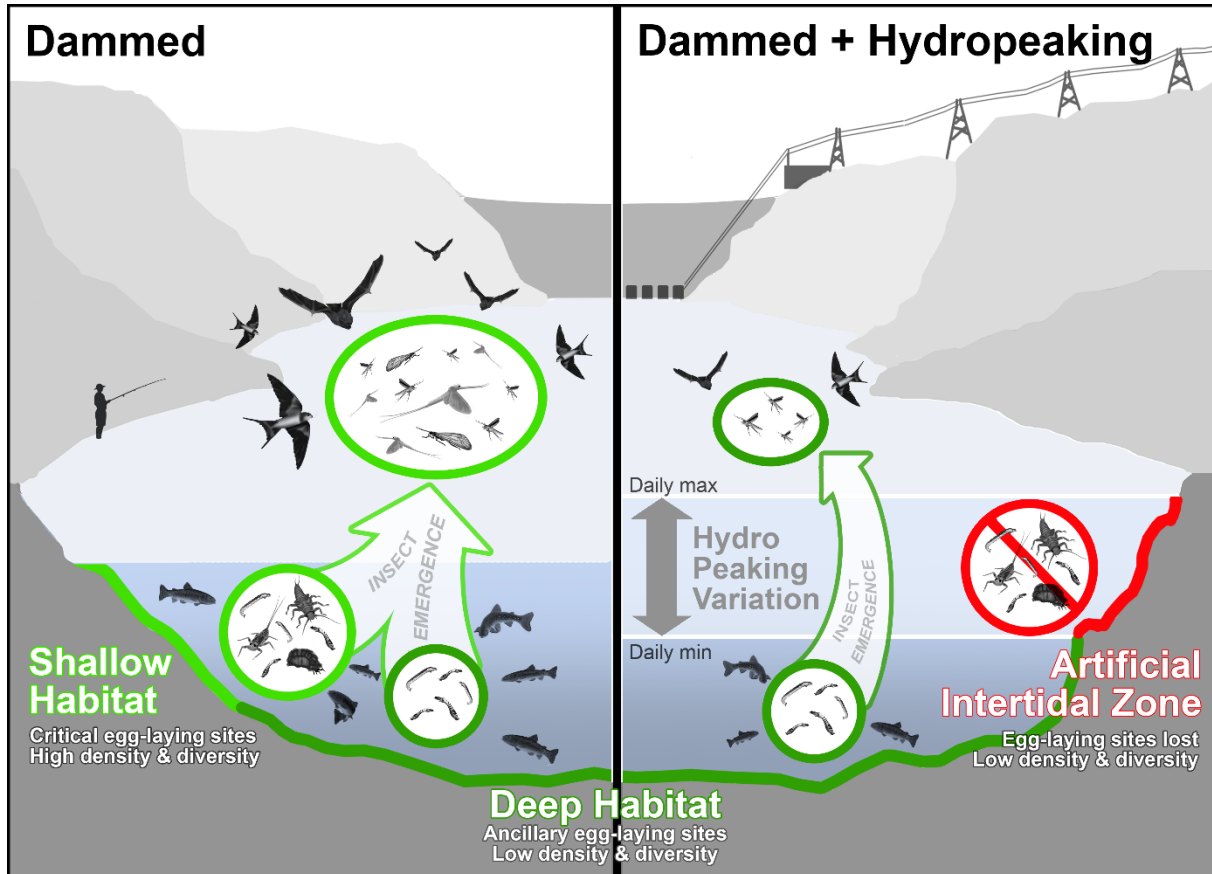


Figure 1. Conceptual model highlighting the essential role that aquatic insects play in river food webs. Ecologically important insect groups such as mayflies, stoneflies, and caddisflies cement their eggs along river-edge habitats, making them especially sensitive to hydropower production practices that affect these edge habitats. The Bug Flows experiment seeks to mitigate these negative impacts by periodically providing favorable egg laying conditions for aquatic insects, which is expected to shift food webs to look more like the left side of this diagram. From Kennedy and others, 2016.

Long-Term Experimental and Management Plan (LTEMP) Resource Goals

Project F supports adaptive ecosystem management by addressing important LTEMP goals and other information needs. A productive and diverse aquatic food base is an indicator of healthy *Natural Processes*. Thus, Project F aligns with the associated LTEMP goal:

Restore, to the extent practicable, ecological patterns and processes within their range of natural variability, including the natural abundance, diversity, and genetic and ecological integrity of the plant and animal species native to those ecosystems.

Project F also supports LTEMP goals for

- Humpback Chub
- Other Native Fishes
- Rainbow Trout Fishery
- Nonnative Invasive Species

by quantifying the feeding habits of fishes (see F.4) and by monitoring invertebrate prey resources in the mainstem and tributaries (see F.1, F.2, F.3), which informs ongoing fish monitoring (e.g., Korman and others, 2021).

Background

Aquatic invertebrates exhibit movements and behaviors that are ecologically important (Figure 2), not only because these processes are critical to invertebrate life cycles and population dynamics, but because these movements make invertebrates vulnerable to predation by wildlife populations. For instance, although most stream invertebrates are benthic (bottom dwelling), invertebrates are also regularly found drifting with the river current. This process of drift is essential to invertebrate dispersal and colonization and therefore critical to population maintenance of these animals (Brittain and Eikeland, 1988).

Similarly, of the many insect stream invertebrates, nearly all transition to becoming winged, air-breathing adults via emergence (Figure 2). Each of these are critical stages in invertebrate life cycles (Huryn and Wallace, 2000), but also processes by which they become vulnerable to fishes and terrestrial food webs, often serving as prey for animals like birds, bats, spiders, and lizards (Baxter and others, 2005). For example, invertebrate drift biomass ($\text{g}\cdot\text{m}^{-3}$) was the single-best predictor of rainbow trout growth over a 5-year study spanning 5 sampling sites and 80 river miles (Korman and others, 2021). Notably, invertebrate drift biomass was a better predictor of rainbow trout growth and survival than water temperature, turbidity, intra-specific competition, and other variables.

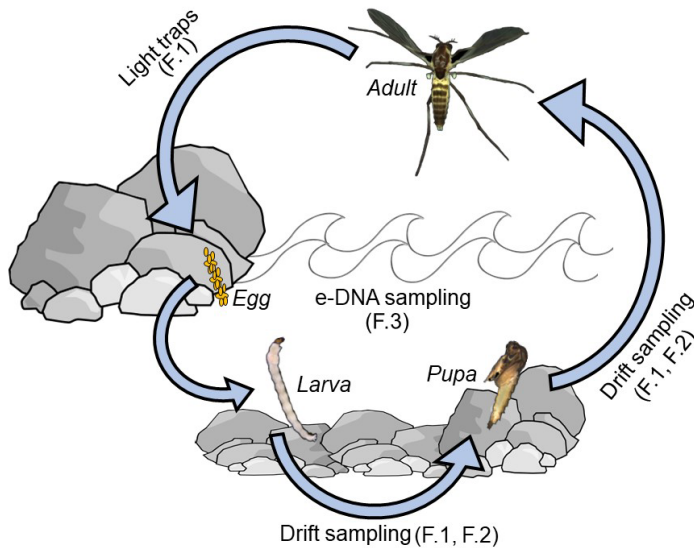


Figure 2. Typical life cycle of an aquatic insect. Non-biting midges (Chironomidae), which represent the dominant aquatic insect in the Colorado River food web, are shown. The diversity of sampling methods utilized in Project F target different life stages of insects. Aquatic insects are susceptible to fish predation at all life stages, but especially during the drifting life stage just prior to adult emergence. Research by Project F has shown that fluctuating flows associated with hydropower generation cause mortality at the egg stage, disrupting insect life cycles. The Bug Flows experiment seeks to restore the Natural Processes that sustain aquatic insects and food webs by periodically providing stable flows that enhance aquatic insect egg survival.

Research by our group has demonstrated that the scarcity of mayflies, stoneflies, and caddisflies from the Colorado River is partly due to acute mortality of insect eggs arising from hourly changes in discharge associated with hydropower generation (Kennedy and others, 2016; Miller and others, 2020). In May–August of 2018–2020 and 2022, Glen Canyon Dam operations were experimentally modified to evaluate whether low steady flows on weekends would increase the production and diversity of aquatic insects in the Colorado River ecosystem (CRe). These experimental Bug Flows involved hourly flow fluctuations for hydropower generation during weekdays, coupled with steady, low flows on weekends to reduce aquatic insect egg desiccation and mortality.

Experimental Bug Flows are associated with significant increases in the abundance of key invertebrate prey items (Figure 3) that in turn fuel growth of fishes, bats, and other wildlife throughout the Colorado River ecosystem (Figure 4; Cross and others, 2013; Metcalfe and others, 2023). Bug Flows also leads to ecologically meaningful increases in gross primary production (GPP) throughout the CRe (i.e., 40% increase) resulting in an additional 350 metric tons of high-quality algae carbon available to fuel river and riparian food webs each year (Deemer and others, 2022). These increases in GPP have been linked to higher growth rates of native flannelmouth sucker (*Catostomus latipinnis*) and humpback chub (Hansen and others, 2023; Hansen and others, 2024). Bug Flows also increases growth rates and angler catch rates of rainbow trout (Metcalfe and others, 2020a; Korman and others, 2023).

Additional years of Bug Flows testing in FY 2025-27 would improve our understanding of ecosystem response to these flows and relation of this flow experiment to other LTEMP goals.

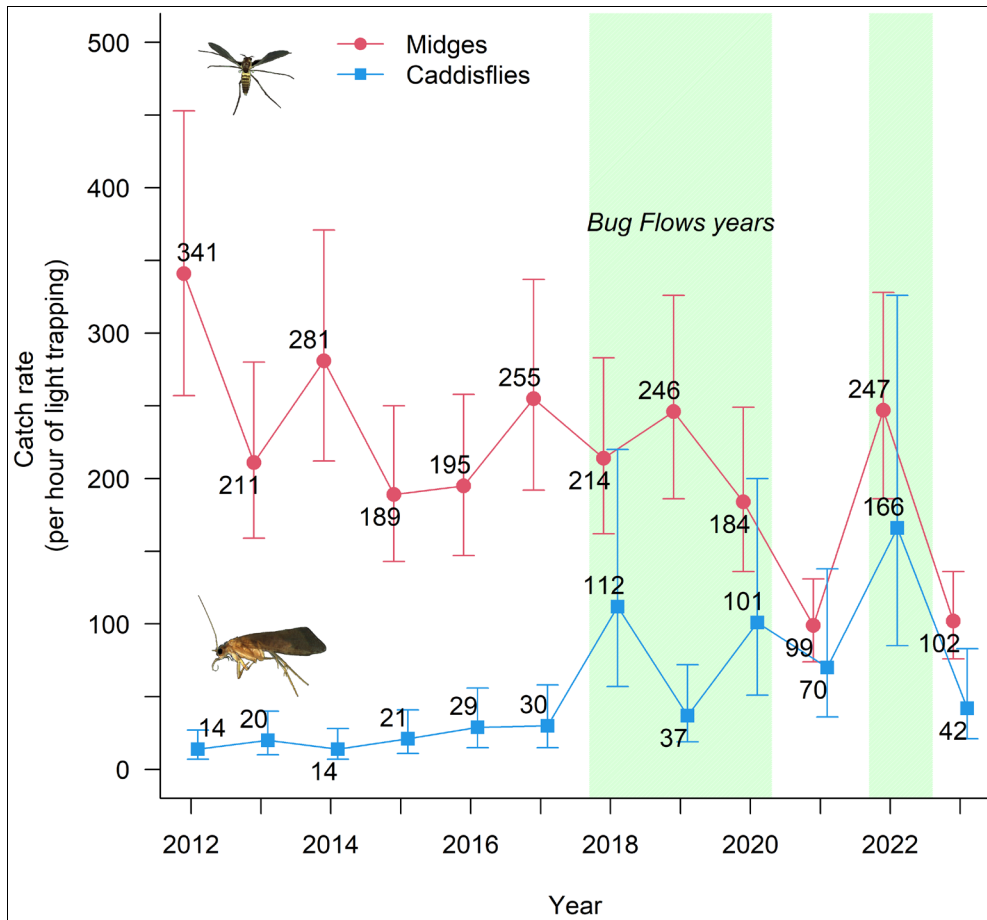


Figure 3. Community science light trap catches of midges and caddisflies by year (solid lines and circles). Years of Bug Flows testing are highlighted in green. Annual average values appear above each point and are estimated from a mixed-effects model that accounts for variation in sampling effort across reaches and across years. Error bars represent one standard error. There is strong model support for a positive Bug Flows effect on catch rates of both caddisflies and midges i.e., inclusion of Bug Flows as a fixed effect lowers the model AIC (Akaike Information Criterion, a metric of how well a model fits the data) by 28 and 36 for midge and caddisfly models, respectively; note that AIC decreases >8 are considered strong model support for a given predictor. Inclusion of High Flow Experiments, which have occurred sporadically over the period of record, does not improve model performance based on AIC. We are unable to evaluate whether environmental drivers such as temperature or suspended sediment conditions improve model performance over the period of record owing to staffing vacancies in Project A, which were only recently filled. Unpublished data, subject to change, do not cite or distribute.

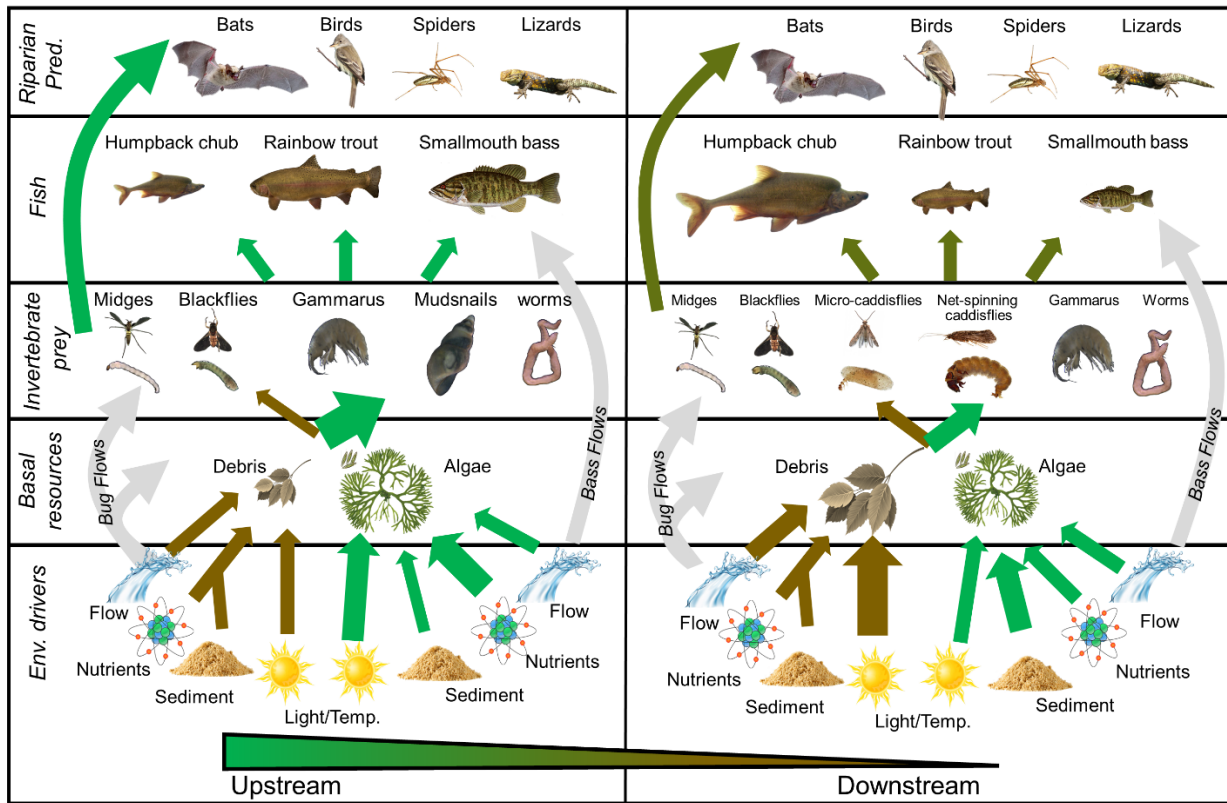


Figure 4. Simplified conceptual model of the Colorado River ecosystem. Arrows indicate linkages between food web components, with the strength of the arrow roughly indicating the strength of the linkage. Note the arrow between flow and aquatic insects, which is the linkage being tested by the Bug Flows experiment. The height of the bottom triangle indicates the amount of whole-ecosystem basal productivity (decreasing from Lees ferry to Western Grand Canyon) and the color indicates the primary source of energy (green = gross primary production; brown = detritus) fueling upper trophic levels.

Proposed Work

Project Element F.1. Aquatic Invertebrate Monitoring in Marble and Grand Canyons (Modified Study; Partially Funded)

Hypotheses and Science Questions

Our ability to test hypotheses and answer science questions concerning the food base will depend in part on which flow experiments occur during FY 2025-27. Testing of experimental Bug Flows in 2018-2020 and 2022 has provided a wealth of data and strong support for the following hypotheses, which were first included in the FY 2021-23 TWP:

- H1: Bug Flows increase the abundance of midges (Chironomidae) in the CRe by improving survival of sensitive insect eggs (Figures 1, 2).

- H2: Bug Flows increase the abundance of EPT (Ephemeroptera [mayflies], Plecoptera [stoneflies], Trichoptera [caddisflies]) in the CRe by improving survival of sensitive insect eggs (Figures 1, 2).

Cool-Mix flows that are intended to suppress reproduction and recruitment of nonnative warm water fishes are being tested in FY 2024 and may also be tested during FY 2025-27. These flows cool water temperatures, which will slow invertebrate growth rates, but the effects of water temperature on the food base may be counteracted by increases in nutrient concentrations arising from bypass releases that pull water from lower in Lake Powell reservoir where nutrient concentrations are higher. Thus, we hypothesize that:

- H3: Cool-Mix flows have neutral effects on the invertebrate prey base through the counteracting effects of cooler water temperatures and higher nutrient concentrations.

Methods

This project element focuses on identifying links between Glen Canyon Dam operations, environmental conditions (e.g., tributary flooding, water temperature), and the downstream aquatic food base. We focus our efforts on monitoring invertebrate populations during periods of movement (i.e., emergence and drift), because these drift and emergence data can be used to make inferences about the health and status of invertebrate populations (Kennedy and others, 2014; Kennedy and others, 2016) and also provide a direct measure of the food base available to humpback chub, rainbow trout, bats, and other wildlife populations (see Projects G and H, especially).

The main thrust of F.1 is the community science monitoring of emergent aquatic insects, where river guides, education groups, private boaters, and other members of the public deploy a simple light trap each night in camp in a standardized fashion to collect samples of adult aquatic insects that have emerged from the Colorado River (Kennedy and others, 2016; Metcalfe and others, 2020b). At the conclusion of their river trips, community scientists return samples to our USGS laboratory for processing. Laboratory processing of community science light trap samples includes counting and identifying aquatic insects to family or genus, whereas terrestrial insects are identified to order or family. Community science participants are provided a modest stipend (\$20) for each light trap sample they collect to ensure high quality data.

To reduce costs, we propose to discontinue bat monitoring via both community science and at fixed stations throughout Glen and Grand Canyon (Kennedy and others 2024). Community science bat monitoring started in 2017 and is done in conjunction with community science light trapping of aquatic insects. In short, acoustic recording devices (Echo-Meter Touch running from a tablet computer) are deployed concurrent with light trap samples to record bat activity levels by species (i.e., number of feeding calls per species per hour, which is positively related to bat abundance/density). Modeling and analysis of these paired samples demonstrate that bat activity is strongly and positively related to the abundance of non-biting midges (i.e., Chironomidae).

The abundance of non-biting midges showed statistically significant increases during Bug Flows testing (Figure 3), thus establishing a positive link between this LTEMP flow experiment and the activity/abundance of bats in the Colorado River ecosystem. In 2023, eight higher-quality acoustic monitors (Song Meter 4) were purchased and deployed at fixed sites in Glen and Grand Canyon. The locations for these fixed site monitors were chosen in collaboration with Grand Canyon National Park staff with the goal of building upon the Parks existing network of identical bat activity sensors. Both these bat monitoring efforts will be discontinued in FY 2025 to reduce costs. Additionally, we propose to discontinue monitoring of invertebrate drift on annual spring river trips. However, we propose to continue collecting invertebrate drift samples at the Juvenile Chub Monitoring site during those trips to provide data on invertebrate prey availability to inform humpback chub models.

Anticipated Use of Data

The community science light trapping data of adult aquatic insects and the invertebrate drift samples collected at Juvenile Chub Monitoring sites will be used to track changes in the abundance and composition of the invertebrate prey base and to inform fish modeling efforts.

Outcomes and Products

The community science light trapping data will be presented annually at the Annual Reporting Meeting and will be used to compute the EPT percent metric that we describe in the forthcoming Metrics document. The light trapping data on adult aquatic insects will also be analyzed and published in peer-reviewed journal articles. Data on invertebrate drift at Juvenile Chub Monitoring sites will be incorporated into various fishery publications and analyses.

Project Element F.2. Aquatic Invertebrate Monitoring in Glen Canyon (Modified Study; Partially Funded)

Hypotheses and Science Questions

Our ability to test hypotheses and answer science questions concerning the food base will depend in part on which flow experiments occur during FY 2025-27. Testing of experimental Bug Flows in 2018-2020 and 2022 has provided a wealth of data and strong support for the following hypotheses, which were first included in the FY 2021-23 TWP:

- H1: Bug Flows increase the abundance of midges (Chironomidae) in the CRe by improving survival of sensitive insect eggs (Figures 1, 2).
- H2: Bug Flows increase the abundance of EPT (Ephemeroptera [mayflies], Plecoptera [stoneflies], Trichoptera [caddisflies]) in the CRe by improving survival of sensitive insect eggs (Figures 1, 2).

Cool-Mix flows that are intended to suppress reproduction and recruitment of nonnative warm water fishes are being tested in FY 2024 and may also be tested during FY 2025-27. These flows cool water temperatures, which will slow invertebrate growth rates, but the effects of water temperature on the food base may be counteracted by increases in nutrient concentrations arising from bypass releases that pull water from lower in Lake Powell reservoir where nutrient concentrations are higher. Thus, we hypothesize that:

- H3: Cool-Mix flows have neutral effects on the invertebrate prey base through the counteracting effects of cooler water temperatures and higher nutrient concentrations.

Methods

This element is a continuation of a Glen Canyon monitoring program that has been ongoing since 2007. It represents a valuable long-term dataset for identifying status and trends in the aquatic food base supporting rainbow trout populations in Glen Canyon (see Figure 5; Korman and others, 2017; Korman and others, 2021). This monitoring is carried out using published methods developed by the food base group (Copp and others, 2014; Kennedy and others, 2014; Baxter and others, 2017; Muehlbauer and others, 2017).

Invertebrate drift will be sampled 4 times per year concurrent with Trout Reproduction and Growth Dynamics monitoring trips (see Project H). Sampling will occur at 5 sites distributed from Glen Canyon Dam (RM -16) to the Lees Ferry boat ramp (RM 0) using methods described in Kennedy and others (2014). In brief, drift nets with a 0.25 mm mesh size are deployed from a boat in the center of the channel at a fixed depth (7 feet) using a hand-powered winch and a 75lb sounding weight keeps the nets oriented into the current. Flow meters on the net-mouth and on the sounding weight are compared and used to verify that nets have not become clogged during the 5 minute deployment (Muehlbauer and others, 2017). Five-minute deployments typically filter ~50 m³ of water.

The proposed study design (4 sampling bouts per year, 5 locations per bout) represents a reduction in monitoring effort to reduce costs compared to the ~monthly monitoring that has been conducted since 2007. Additionally, we propose to discontinue monthly sticky trap (Smith and others 2014) and light trap sampling of adult aquatic insects in Glen Canyon. These sample collections were started in ~2014, but both require an overnight stay in Lees Ferry thereby increasing cost.

Anticipated Use of Data

Data derived from quarterly sampling of invertebrate drift in Glen Canyon during FY 2025-27 will be provided to fishery colleagues to inform modeling of rainbow trout and other fishes in Glen Canyon.

Outcomes and Products

Data derived from quarterly monitoring of invertebrate drift in Glen Canyon during FY 2025-27 will be incorporated into various rainbow trout publications. Funding from this element during FY 2025-27 will also support staff time to continue modeling, analyzing, and publishing the long-term Glen Canyon monitoring data collected from 2008-present when monitoring efforts were more comprehensive.

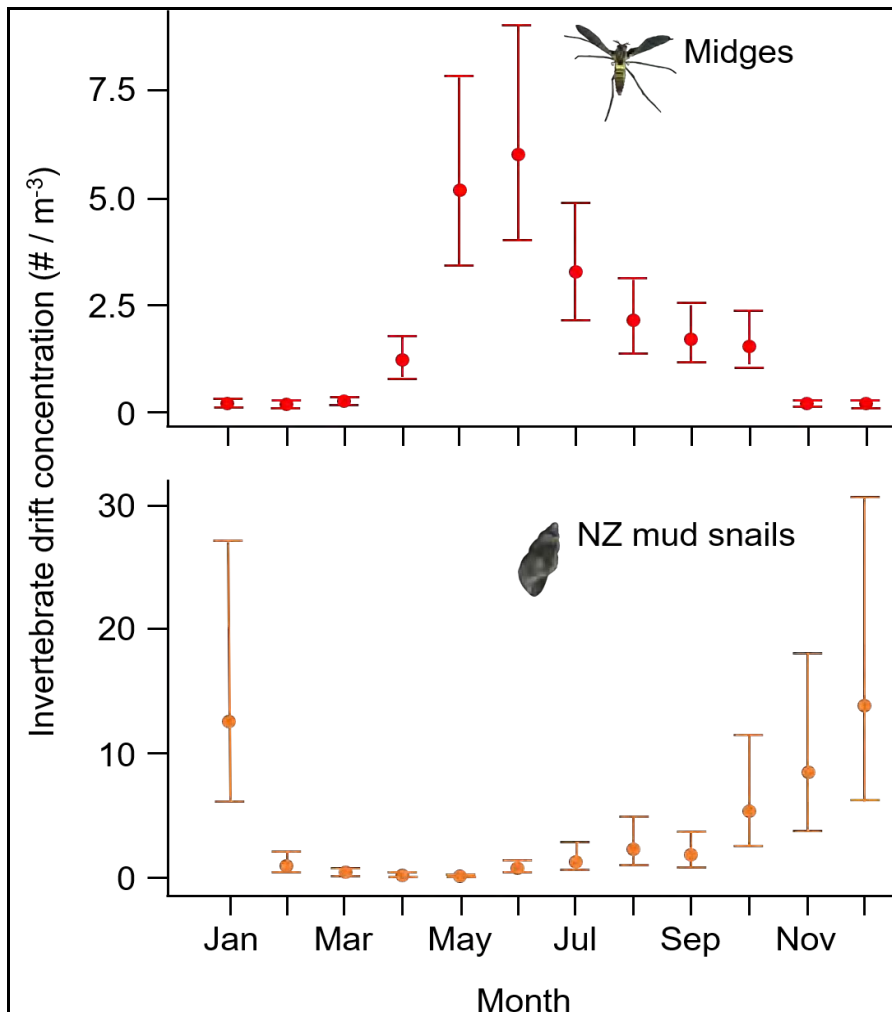


Figure 5. Drift concentrations ($\#/m^3$) of midges and New Zealand mud snails from Glen Canyon monthly monitoring. Points represent the mean \pm one standard error estimated from a mixed-effects model. Period of record is 2007-2022. Note the seasonal differences, with more midge activity in summer months and more mud snail activity in the winter. Unpublished data, subject to change, do not cite or distribute.

Project Element F.3. Aquatic Invertebrate Monitoring in the Colorado River and Grand Canyon Tributaries (Ongoing Study; Funded)

eDNA monitoring of tributary streams is relevant to management, because these streams are important spawning and rearing habitat for native fishes, and some are also sites of humpback chub translocations. For this reason, understanding the diversity of aquatic food base resources available to these fishes can influence decisions about whether to translocate more fishes into these streams, and in identifying candidate streams for future translocations. Further, tributaries represent sources of aquatic insects that could recolonize the mainstem Colorado River. Understanding the locations of these aquatic insect populations therefore provides insight into where we might first expect to see sensitive Ephemeroptera (mayfly), Plecoptera (stonefly) and Trichoptera (caddisfly) taxa colonizing the mainstem.

eDNA monitoring demonstrates that tributaries in Grand Canyon support more diverse invertebrate communities compared to the mainstem Colorado River, with considerable variation in invertebrate communities across tributaries and very little variation in communities across 280 miles of the mainstem (Figure 6; Lytle and others, 2023). Notably, eDNA monitoring of invertebrate diversity in tributaries has documented twice the number of invertebrate genera compared to traditional benthic surveys of invertebrates (e.g., Oberlin and others, 1999). Specifically, Oberlin and others (1999) documented 42 genera of invertebrates across 10 tributaries while eDNA monitoring has documented 84 invertebrate genera across 18 tributaries. Our eDNA results are providing much finer taxonomic resolution than traditional methods, with over 700 OTUs identified thus far across the Grand Canyon ecosystem and evidence for population-level differences within some taxa, and clear differences between mainstem and tributary communities (Figure 6). Using funding from a 3-year USGS/NPS-Water Quality Partnership grant awarded to Theodore Kennedy (USGS), Emily Omana-Smith (NPS) and David Lytle (Oregon State University), DNA that has been extracted from water samples is also being amplified and sequenced using other primer sets that are specific to metazoans to detect parasite and pathogen communities in tributaries and the mainstem.

Hypotheses and Science Questions

We hypothesize that:

- H4: Tributaries support more diverse invertebrate and parasite communities compared to the mainstem Colorado River.

We will also evaluate what factors explain variation in invertebrate and parasite diversity across tributaries with a focus on the potential role of water quality conditions (e.g., pH, alkalinity, heavy metals, nutrients). Better understanding the factors that control the diverse invertebrate assemblages that are present across different tributaries will provide insight into the role that water quality may play in promoting or limiting colonization of the mainstem Colorado River by tributary invertebrates.

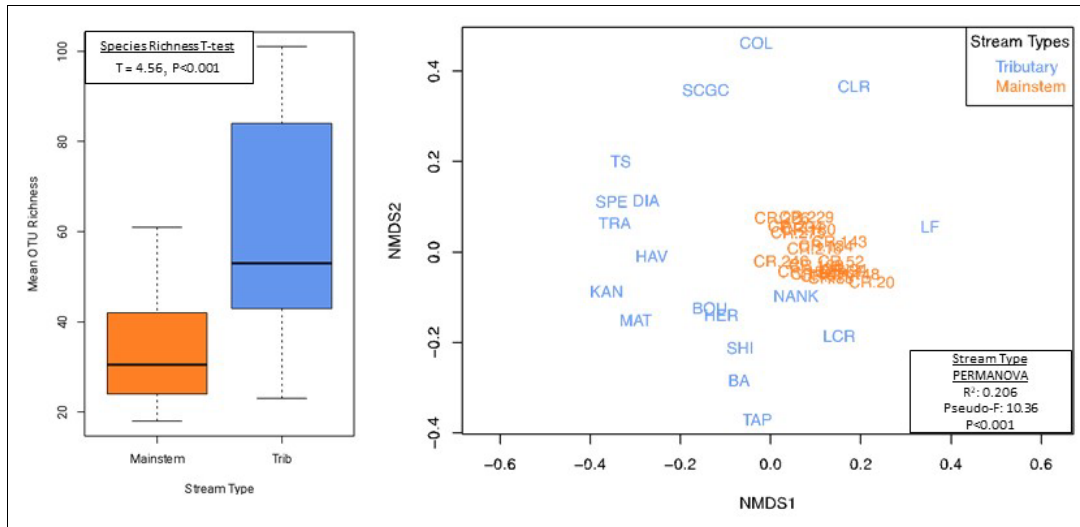


Figure 6. Box-plot (left panel) and non-metric multidimensional scaling plot (NMDS; right panel) of eDNA monitoring results from the mainstem Colorado River and tributaries in Grand Canyon. The box plot shows the mean number of operational taxonomic units (OTUs, which are equivalent to species) detected at each of 18 mainstem sampling locations and 18 tributaries. The NMDS plot depicts the variation in species composition of invertebrates detected in each tributary (BLUE) and the mainstem (ORANGE). Note the large amount of variation in invertebrate community composition across tributaries and the minimal variation in invertebrate community composition across 18 mainstem sites. Unpublished data, subject to change, do not cite or distribute. Figure courtesy of David Lytle and modified from Lytle and others (2023), Oregon State University.

Methods

This element involves monitoring the aquatic invertebrate community within tributaries and the mainstem in Grand Canyon using eDNA. All organisms that are living in a stream or river shed DNA. Water samples are collected from these habitats and the extracted genomic DNA can be stored indefinitely, providing a snapshot of the ecosystem in time and space. Sample DNA can then be sequenced to identify invertebrates, algae, parasites, pathogens, fishes, or any other organisms that were present. Our current effort focuses on aquatic insect identification, but our sample library could be used later to determine the occurrence and distribution of any species of interest.

Water samples for eDNA analysis of invertebrate diversity have been collected annually on April food base monitoring river trips starting in 2021 and processed and analyzed by cooperators at Oregon State University. Sample collection entails filtering four, 1-liter replicates of water per sampling location onto 0.4 μ m filters and preserving the filter in a buffer. Sampling locations include major tributaries in Grand Canyon (n= 18) and the mainstem Colorado River just upstream of these tributary confluences. In the laboratory, analysis of eDNA water samples entails amplifying a 142 base pair fragment of the mitochondrially encoded cytochrome c oxidase I gene (COI) with a degenerate primer set that targets aquatic invertebrates (Leese and others, 2021). These are then sequenced using an Illumina NextSeq platform, producing over 1 million reads per sample. Sequences are then filtered and identified to species (technically, operational taxonomic units or OTUs) using the R package JAMP.

We have been collaborating with USGS Arizona Water Science Center Staff on April river trips since 2022 to collect detailed water quality information (e.g., anions, cations, nutrients, trace metals, pH, discharge, etc.) on tributaries and mainstem locations to inform invertebrate community analyses derived from eDNA sampling.

To reduce costs, we will pause collection of new eDNA samples in FY 2025. Archived eDNA water samples dating back to 2021 and new samples that will be collected during FY 2026-27 can also be used to detect the presence/absence of nonnative invertebrates, parasites, pathogens, and fishes whose ranges and distribution are expanding or contracting.

The cooperative agreement associated with F.3 to Oregon State University will fund a new PhD student to perform DNA extractions on new samples and model the distribution and potential expansion of parasites, invertebrates, and fish using these eDNA water sampling data.

Anticipated Use of Data

The eDNA samples collected as part of this element will be used to document the distribution and species richness of invertebrates, parasites and pathogens, and fish throughout the mainstem Colorado River and all major tributaries. Developing a comprehensive baseline for invertebrate, parasite, and fish species composition and distributions throughout the Grand Canyon ecosystem using eDNA will be valuable for documenting future changes that may arise owing to changing environmental conditions, adaptive management experimentation, or other factors.

Outcomes and Products

Data on the distribution and species richness of invertebrates, parasites and pathogens, and fish derived from eDNA sampling will be presented at Annual Reporting Meetings, it will be published in a forthcoming PhD dissertation (anticipated in May 2025), and it will be published in multiple peer-reviewed journal articles.

Project Element F.4. Fish Diet and Health Studies (Ongoing Study; Funded)

Current food base monitoring approaches were informed by detailed food web studies and invertebrate and fish diet analysis of samples collected from 2006–2009 (Cross and others, 2013; Kennedy and others, 2013; Sabo and others, 2018). These studies identified that algae fuels growth of invertebrate populations everywhere, even at turbid downstream sites where algae are scarce (Wellard Kelly and others, 2013). Based on these insights, GCMRC scientists and collaborators developed techniques for continuously monitoring the algae-portion of the food base using dissolved oxygen budgeting (see Project E), which has shed light on the role of dam operations and environmental factors in regulating algae growth (Hall and others, 2015; Deemer and others, 2022).

Early diet studies also identified that aquatic insects were key prey for native and desired nonnative fishes (Cross and others, 2013; Zahn Seegert and others, 2014), but the overall low production and diversity of aquatic insects in the CRe appeared to be a major constraint on fish populations overall (Kennedy and others, 2013). Thus, GCMRC scientists developed new techniques for studying the invertebrate-portion of the food base, including community science light trapping of the understudied adult life stage of aquatic insects (Kennedy and others, 2016).

Owing to numerous changes in the food base and fish communities since the last detailed food web studies ~15 years ago, we will assess feeding habits of flannelmouth sucker, humpback chub, rainbow trout, and smallmouth bass. We will focus on non-lethal methods in line with tribal concerns regarding the taking of life. Specifically, we will collect fin clips from fishes to analyze for stable isotopes, and we will collect fecal samples for DNA analysis. Carbon and nitrogen stable isotope analysis of fin clips provide a non-lethal indication of long-term feeding habits (i.e., recent months) including the trophic position of fishes and the relative importance of algae vs. terrestrial detritus to fish production overall (Layman and others 2012). This long-term information on general feeding habits obtained from stable isotopes complements the more detailed, species-level, snapshot of feeding habits that we will obtain from DNA analysis of fish feces (i.e., what the fish ate that day). This detailed information concerning species presence/absence in diets, when combined with direct eDNA measurements of invertebrate distributions across the Colorado River ecosystem (F.3, above), can be used to estimate source and contribution (tributary vs. mainstem, with tributary identity and mainstem river mile location as cofactors) of the invertebrate food base to fish feeding habits. Because our eDNA methods also detect DNA from terrestrial insects when present, we will be able to quantify the proportion of terrestrial vs. aquatic-origin insect species in fish diets. Understanding direct links between prey distribution and utilization of prey by fish is an important element in management and conservation of native and threatened fish such as the humpback chub.

Hypotheses and Science Questions

Owing to changes in environmental conditions, the aquatic food base, and fish communities since 2006-2009, we will test the following hypotheses:

- H1: The trophic niche breadth (i.e., a metric describing the diversity of prey resources utilized by a population) for flannelmouth sucker, humpback chub, and rainbow trout has increased compared to 2006-2009 owing to more abundant and diverse invertebrate prey resources.
- H2: The trophic niche breadth (i.e., a metric describing the diversity of prey resources utilized by a population) for flannelmouth sucker, humpback chub, and rainbow trout that are proximate to tributaries is larger than the niche breadth for fish that are distant from tributaries, because tributaries harbor diverse invertebrate assemblages and flooding from tributaries contributes terrestrial prey resources.

- H3: The trophic position for humpback chub has increased compared to 2006-2009 owing to more abundant fish prey resources (e.g., juvenile flannelmouth suckers).
- H4: The contribution of algae-derived carbon to flannelmouth sucker, humpback chub, and rainbow trout production declines in the downstream direction owing to downstream decreases in algae production.
- H5: The overall contribution of terrestrially derived carbon to flannelmouth sucker and humpback chub production has increased compared to 2006-2009, particularly at sites in western Grand Canyon, owing to increases in water temperature that drive increases in microbial growth on leaf litter and other types of terrestrial carbon thereby rendering it more nutritious.

Methods

DNA analysis of fish fecal samples will be done at Oregon State University using the same laboratory processing as described above for eDNA water sampling in F.3; interpretation and analysis of these fecal DNA data will be done by USGS staff. Using funding from a USGS/NPS-Water Quality Partnership grant, we are supporting a PhD student to analyze fish fecal samples for parasites and pathogens to provide baseline information on these communities. Changes in water quality associated with aridification and declining elevations in Lake Powell reservoir are changing the environment for, and risks associated with, parasites and pathogens. Additionally, increases in fish passage through Glen Canyon Dam associated with declining reservoir elevations are introducing new species of fish and their parasites to the CRE. Feeding habits of smallmouth bass that are euthanized as part of ongoing removal efforts will be accomplished using traditional gut content analysis. To reduce costs, we propose to pause sample collections associated with this element in FY 2025.

Anticipated Use of Data

DNA analysis of fish feces and carbon and nitrogen stable isotope analysis of fish fin clips will be used to document fish feeding habits, trophic position, trophic niche breadth, and other measures of food web structure. By comparing present-day food web structural metrics to similar metrics computed from 2006-2009, these data will provide insight into the causes of native fish population increases that occurred during the 2010s. These data will also provide a baseline for comparison if fish communities continue to change in the future owing to changing environmental conditions or other factors (e.g., population expansion of smallmouth bass or razorback sucker, reintroduction of Colorado pikeminnow).

Outcomes and Products

Data on feeding habits and parasite communities of fish derived from DNA analysis of feces and stable isotope analysis of fin clips is available to be presented at Annual Reporting Meetings, it will be published in a forthcoming PhD dissertation (anticipated in 2027), and multiple peer-reviewed journal articles.

Budgets

Project Element Budget

| Fiscal Year 2025 | | | | | | | | | |
|--|------------------|----------------------|-----------------------|-----------------------|---------------------------|--------------------------|-----------------------------|------------------|------------------------------|
| Project F Aquatic Invertebrate Ecology | Salaries | Travel & Training | Operating Expenses | Logistics Expenses | Cooperative Agreements | To other USGS Centers | AMP Special Burden Rate* | Total | USGS Contributing Funds** |
| | | | | | | | 21.86% | | Est Full Rate = 55% |
| F.1. Invertebrate and bat monitoring in Marble and Grand Canyons | \$228,969 | \$1,000 | \$17,000 | \$0 | \$0 | \$0 | \$53,995 | \$300,964 | |
| F.2. Aquatic invertebrate monitoring in Glen Canyon | \$170,121 | \$500 | \$1,500 | \$875 | \$0 | \$0 | \$37,822 | \$210,818 | |
| F.3. Aquatic invertebrate monitoring of Grand Canyon tributaries | \$22,756 | \$500 | \$0 | \$0 | \$45,825 | \$0 | \$6,459 | \$75,540 | |
| F.4. Invertebrate and fish diet studies | \$108,428 | \$500 | \$1,500 | \$0 | \$0 | \$0 | \$24,143 | \$134,571 | |
| Total Project F | \$530,274 | \$2,500 | \$20,000 | \$875 | \$45,825 | \$0 | \$122,419 | \$721,893 | \$57,817 |

| Fiscal Year 2026 | | | | | | | | | |
|--|------------------|----------------------|-----------------------|-----------------------|---------------------------|--------------------------|-----------------------------|------------------|------------------------------|
| Project F Aquatic Invertebrate Ecology | Salaries | Travel & Training | Operating Expenses | Logistics Expenses | Cooperative Agreements | To other USGS Centers | AMP Special Burden Rate* | Total | USGS Contributing Funds** |
| | | | | | | | 22.60% | | Est Full Rate = 55.5% |
| F.1. Invertebrate and bat monitoring in Marble and Grand Canyons | \$234,793 | \$1,000 | \$17,000 | \$0 | \$0 | \$0 | \$57,131 | \$309,924 | |
| F.2. Aquatic invertebrate monitoring in Glen Canyon | \$122,673 | \$1,000 | \$1,500 | \$875 | \$0 | \$0 | \$28,487 | \$154,535 | |
| F.3. Aquatic invertebrate monitoring of Grand Canyon tributaries | \$36,189 | \$1,000 | \$0 | \$8,669 | \$48,175 | \$0 | \$11,809 | \$105,843 | |
| F.4. Invertebrate and fish diet studies | \$89,472 | \$1,000 | \$1,500 | \$8,669 | \$0 | \$0 | \$22,745 | \$123,386 | |
| Total Project F | \$483,127 | \$4,000 | \$20,000 | \$18,213 | \$48,175 | \$0 | \$120,172 | \$693,687 | \$55,670 |

| Fiscal Year 2027 | | | | | | | | | |
|--|------------------|----------------------|-----------------------|-----------------------|---------------------------|--------------------------|-----------------------------|------------------|------------------------------|
| Project F Aquatic Invertebrate Ecology | Salaries | Travel & Training | Operating Expenses | Logistics Expenses | Cooperative Agreements | To other USGS Centers | AMP Special Burden Rate* | Total | USGS Contributing Funds** |
| | | | | | | | 23.40% | | Est Full Rate = 56% |
| F.1. Invertebrate and bat monitoring in Marble and Grand Canyons | \$171,354 | \$1,000 | \$17,000 | \$0 | \$0 | \$0 | \$44,309 | \$233,663 | |
| F.2. Aquatic invertebrate monitoring in Glen Canyon | \$139,043 | \$1,000 | \$1,500 | \$875 | \$0 | \$0 | \$33,326 | \$175,744 | |
| F.3. Aquatic invertebrate monitoring of Grand Canyon tributaries | \$42,302 | \$1,000 | \$0 | \$8,961 | \$49,350 | \$0 | \$13,710 | \$115,324 | |
| F.4. Invertebrate and fish diet studies | \$76,487 | \$1,000 | \$1,500 | \$8,961 | \$0 | \$0 | \$20,580 | \$108,527 | |
| Total Project F | \$429,186 | \$4,000 | \$20,000 | \$18,797 | \$49,350 | \$0 | \$111,924 | \$633,257 | \$50,804 |

*GCDAMP Special Rate: Rate includes current DOI preferred rate (currently 15%; subject to change) and facilities rate (will vary annually). No USGS bureau overhead is charged (unique to the GCDAMP agreement).

**USGS Contributing Funds: The amount of funds required to cover the subsidy created by the reduced burden rate (i.e., GCDAMP Special Rate). As in previous years, SBSC/GCMRC will request these funds of the USGS cost-share program. These funds are not guaranteed.

References Cited

- Baxter, C.V., Fausch, K.D., and Saunders, W.C., 2005, Tangled webs—Reciprocal flows of invertebrate prey link streams and riparian zones: *Freshwater Biology*, v. 50, no. 2, p. 201–220, <https://doi.org/10.1111/j.1365-2427.2004.01328.x>.
- Baxter, C.V., Kennedy, T.A., Miller, S.W., Muehlbauer, J.D., and Smock, L.A., 2017, Macroinvertebrate drift, adult insect emergence and oviposition, *in* Hauer, F.R., and Lamberti, G.A., eds., *Methods in Stream Ecology*, vol. 1, 3rd ed., *Ecosystem Structure*: Academic Press, p. 435-456, <https://doi.org/10.1016/B978-0-12-416558-8.00021-4>.
- Brittain, J.E., and Eikeland, T.J., 1988, Invertebrate drift—A review: *Hydrobiologia*, v. 166, no. 1, p. 77-93, <https://doi.org/10.1007/BF00017485>.
- Copp, A.J., Kennedy, T.A., and Muehlbauer, J.D., 2014, Barcodes are a useful tool for labeling and tracking ecological samples: *Bulletin of the Ecological Society of America*, v. 95, no. 3, p. 293-300, <https://doi.org/10.1890/0012-9623-95.3.293>.
- Cross, W.F., Baxter, C.V., Rosi-Marshall, E.J., Hall, R.O., Jr., Kennedy, T.A., Donner, K.C., Wellard Kelly, H.A., Seegert, S.E.Z., Behn, K., and Yard, M.D., 2013, Food-web dynamics in a large river discontinuum: *Ecological Monographs*, v. 83, no. 3, p. 311-337, <https://doi.org/10.1890/12-1727.1>.
- Deemer, B.R., Yackulic, C.B., Hall, R.O., Jr., Dodrill, M.J., Kennedy, T.A., Muehlbauer, J.D., Topping, D.J., Voichick, N., and Yard, M.D., 2022, Experimental reductions in subdaily flow fluctuations increased gross primary productivity for 425 river kilometers downstream: *PNAS Nexus*, v. 1, no. 3, pgac094, <https://doi.org/10.1093/pnasnexus/pgac094>.
- Hall, R.O., Jr., Yackulic, C.B., Kennedy, T.A., Yard, M.D., Rosi-Marshall, E.J., Voichick, N., and Behn, K., 2015, Turbidity, light, temperature, and hydropeaking control primary productivity in the Colorado River, Grand Canyon: *Limnology and Oceanography*, v. 60, no. 2, p. 512-526, <https://doi.org/10.1002/lno.10031>.
- Hansen, L.E., Yackulic, C.B., Dickson, B.G., Deemer, B.R., and Best, R.J., 2023, Linking ecosystem processes to consumer growth rates—Gross primary productivity as a driver of freshwater fish somatic growth in a resource-limited river: *Canadian Journal of Fisheries and Aquatic Sciences*, v. 80, no. 9, p. 1456-1469, <https://doi.org/10.1139/cjfas-2022-0229>.
- Hansen, L.E., Eppheimer, D.E., Dzul, M.C., Deemer, B.R., and Yackulic, C.B., 2024, Humpback chub growth becomes resource dependent when released from temperature limitation [poster presentation]: Phoenix, Ariz., Glen Canyon Dam Adaptive Management Program, Annual Reporting Meeting, January 24-26, 2024, https://gcdamp.com/images_gcdamp_com/1/19/ARM_HBC_Growth_Poster_2024_V2_final.pdf.
- Huryn, A.D., and Wallace, J.B., 2000, Life history and production of stream insects: *Annual Review of Entomology*, v. 45, p. 83-110, <https://doi.org/10.1146/annurev.ento.45.1.83>.
- Kennedy, T.A., Cross, W.F., Hall, R.O., Jr., Baxter, C.V., and Rosi-Marshall, E.J., 2013, Native and non-native fish populations of the Colorado River are food limited—Evidence from new

- food web analyses: U.S. Geological Survey Fact Sheet 2013-3039, 4 p., <https://pubs.usgs.gov/fs/2013/3039/>.
- Kennedy, T.A., Muehlbauer, J.D., Yackulic, C.B., Lytle, D.A., Miller, S.W., Dibble, K.L., Kortenhoeven, E.W., Metcalfe, A.N., and Baxter, C.V., 2016, Flow management for hydropower extirpates aquatic insects, undermining river food webs: *BioScience*, v. 66, no. 7, p. 561-575, <https://doi.org/10.1093/biosci/biw059>.
- Kennedy, T.A., Scholl, E., Metcalfe, A., Muehlbauer, J., Yackulic, C., Ford, M., Szydlo, C., 2024, Project F Update: Leaf decomposition, bat monitoring, and aquatic insects [presentation]: Phoenix, Ariz., Glen Canyon Dam Adaptive Management Program, Annual Reporting Meeting, January 25, 2024. <https://www.usbr.gov/uc/progact/amp/twg/2024-01-25-twg-meeting/20240125-AnnualReportingMeeting-ProjectFUpdateLeadDecompositionBatMonitoringAquaticInsects-508-UCRO.pdf>
- Kennedy, T.A., Yackulic, C.B., Cross, W.F., Grams, P.E., Yard, M.D., and Copp, A.J., 2014, The relation between invertebrate drift and two primary controls, discharge and benthic densities, in a large regulated river: *Freshwater Biology*, v. 59, no. 3, p. 557-572, <https://doi.org/10.1111/fwb.12285>.
- Korman, J., Yard, M.D., and Kennedy, T.A., 2017, Trends in rainbow trout recruitment, abundance, survival, and growth during a boom-and-bust cycle in a tailwater fishery: *Transactions of the American Fisheries Society*, v. 146, no. 5, p. 1043-1057, <https://doi.org/10.1080/00028487.2017.1317663>.
- Korman, J., Yard, M.D., Dzul, M.C., Yackulic, C.B., Dodrill, M.J., Deemer, B.R., and Kennedy, T.A., 2021, Changes in prey, turbidity, and competition reduce somatic growth and cause the collapse of a fish population: *Ecological Monographs*, v. 91, no. 1, e01427, p. 1-20, <https://doi.org/10.1002/ecm.1427>.
- Korman, J., Deemer, B., Yackulic, C.B., Kennedy, T.A., and Giardina, M., 2023, Drought related changes in water quality surpass effects of experimental flows on trout growth downstream of Lake Powell reservoir: *Canadian Journal of Fisheries and Aquatic Sciences*, v. 80, no. 3, p. 424-438, <https://doi.org/10.1139/cjfas-2022-0142>.
- Layman, C.A., Araujo, M.S., Boucek, R., Hammerschlag-Peyer, C.M., Harrison, E., Jud, Z.R., Matich, P., Rosenblatt, A.E., Vaudo, J.J., Yeager, L.A., Post, D.M., and Bearhop, S., 2012, Applying stable isotopes to examine food-web structure—An overview of analytical tools: *Biological Reviews*, v. 87, no. 3, p. 545-562, <https://doi.org/10.1111/j.1469-185X.2011.00208.x>.
- Leese, F., Sander, M., Buchner, D., Elbrecht, V., Haase, P., and Zizka, V.M.A., 2021, Improved freshwater macroinvertebrate detection from environmental DNA through minimized nontarget amplification: *Environmental DNA*, v. 3, no. 1, p. 261-276, <https://doi.org/10.1002/edn3.177>.
- Lytle, D.A., Kurthen, A., and Freedman, J., 2023, Molecular and modeling tools for tracking food base dynamics in changing environments [presentation]: Phoenix, Ariz., Glen Canyon Dam Adaptive Management Program, Annual Reporting Meeting, January 26, 2023,

<https://www.usbr.gov/uc/progact/amp/twg/2023-01-26-twg-meeting/20230126-AnnualReportingMeeting-MolecularModelingToolsTrackingFoodBaseDynamicsChangingEnvironments-508-UCRO.pdf>.

- Metcalf, A.N., Fritzinger, C.A., Weller, T.J., Dodrill, M.J., Muehlbauer, J.D., Yackulic, C.B., Holton, P.B., Szydlo, C.M., Durning, L.E., Sankey, J.B., and Kennedy, T.A., 2023, Insectivorous bat foraging tracks the availability of aquatic flies (Diptera): *The Journal of Wildlife Management*, v. 87, no. 5, e22414, <https://doi.org/10.1002/jwmg.22414>.
- Metcalf, A.N., Muehlbauer, J.D., Kennedy, T.A., and Ford, M.A., 2020a, Bug flows—Don't count your midges until they hatch: *Boatman's Quarterly Review*, v. 32, no. 4, winter 2019-2020, p. 8-11, https://www.researchgate.net/publication/339569673_Bug_flows_Don't_count_your_midges_until_they_hatch.
- Metcalf, A.N., Muehlbauer, J.D., Kennedy, T.A., Yackulic, C.B., Dibble, K.L., and Marks, J.C., 2020b, Net-spinning caddisfly distribution in large regulated rivers: *Freshwater Biology*, v. 66, no. 1, p. 89-101, <https://doi.org/10.1111/fwb.13617>.
- Miller, S.W., Schroer, M., Fleri, J.R., and Kennedy, T.A., 2020, Macroinvertebrate oviposition habitat selectivity and egg-mass desiccation tolerance—Implications for population dynamics in large regulated rivers: *Freshwater Science*, v. 39, no. 3, p. 584–599, <https://doi.org/10.1086/710237>.
- Muehlbauer, J.D., Kennedy, T.A., Copp, A.J., and Sabol, T.A., 2017, Deleterious effects of net clogging on the quantification of stream drift: *Canadian Journal of Fisheries and Aquatic Sciences*, v. 74, no. 4, p. 1041-1048, <https://doi.org/10.1139/cjfas-2016-0365>.
- Oberlin, G.E., Shannon, J.P., and Blinn, D.W., 1999, Watershed influence on the macroinvertebrate fauna of ten major tributaries of the Colorado River through Grand Canyon, Arizona: *The Southwestern Naturalist*, v. 44, no. 1, p. 17-30, <http://www.jstor.org/stable/30055398>.
- Sabo, J.L., Caron, M., Doucett, R.R., Dibble, K.L., Ruhi, A., Marks, J.C., Hungate, B.A., and Kennedy, T.A., 2018, Pulsed flows, tributary inputs, and food web structure in a highly regulated river: *Journal of Applied Ecology*, v. 55, no. 4, p. 1884-1895, <https://doi.org/10.1111/1365-2664.13109>.
- Smith, J.T., Kennedy, T.A., and Muehlbauer, J.D., 2014, Building a better sticky trap—Description of an easy-to-use trap and pole mount for quantifying the abundance of adult aquatic insects: *Freshwater Science*, v. 33, no. 3, p. 972-977, <https://doi.org/10.1086/676998>.
- U.S. Department of the Interior, 2016, Glen Canyon Dam Long-Term Experimental and Management Plan final Environmental Impact Statement (LTEMP EIS): U.S. Department of the Interior, Bureau of Reclamation, Upper Colorado Region, National Park Service, Intermountain Region, online, <http://ltempeis.anl.gov/documents/final-eis/>.
- Wellard Kelly, H.A., Rosi-Marshall, E.J., Kennedy, T.A., Hall, R.O., Cross, W.F., and Baxter, C.V., 2013, Macroinvertebrate diets reflect longitudinal and seasonal changes in food

availability downstream of a large dam: *Limnology and Oceanography*, v. 32, no. 2, p. 397-410, <https://doi.org/10.1899/12-088.1>.

Zahn Seegert, S.E., Rosi-Marshall, E.J., Baxter, C.V., Kennedy, T.A., Hall, R.O., Jr., and Cross, W.F., 2014, High diet overlap between native small-bodied fishes and non-native fathead minnow in the Colorado River, Grand Canyon, Arizona: *Transactions of the American Fisheries Society*, v. 143, no. 4, p. 1072-1083, <https://doi.org/10.1080/00028487.2014.901250>.

Project G: Humpback Chub Population Dynamics Throughout the Colorado River Ecosystem

Investigators

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Project Summary and Purpose

The FY 2025-27 Triennial Work Plan coincides with a period that will likely present new challenges to managing fishery resources within the Colorado River ecosystem (CRe) in Grand Canyon. Most of our knowledge of humpback chub (*Gila cypha*) in Grand Canyon comes from studies that occurred under colder water conditions in the mainstem Colorado River following construction of Glen Canyon Dam. Warming of the mainstem river by a few degrees in recent years coincides with large changes in fish communities in Grand Canyon (e.g., increase/expansion of humpback chub in western Grand Canyon, first observations of juvenile young-of-year [yoy] from several warm-water nonnatives) providing further evidence that water temperature is a key driver of fish communities (Lessard and Hayes, 2003; Dibble and others, 2021). While temperature increases and drought are likely, the rules that will determine Lake Powell elevations and annual release volumes are unknown, and future hydrologic conditions are uncertain, making it difficult to predict how much time will be spent at lower reservoir elevations at which releases through penstocks will be as warm or warmer than releases in 2022 and 2023. Additionally, modifications to LTEMP flows are being proposed that may lead to increased frequency of spring-timed high flow experiments (HFEs), as well as flows to disadvantage warm-water nonnatives including smallmouth bass (*Micropterus dolomieu*) that may have direct and indirect impacts on humpback chub and other native fishes. Project G is designed to estimate the effects of managed, and unmanaged, drivers on the growth, survival, and juvenile production of humpback chub to quantify impacts of management (in the context of unmanaged variation) and improve our ability to forecast future population dynamics. Project G is also designed to estimate state variables (abundance, distribution) for the 2016 Biological Opinion (U.S. Fish and Wildlife Service, 2016), and associated fieldwork and modeling efforts are supported by the 2024 Interim Guidelines Biological Opinion (U.S. Fish and Wildlife Service, 2024). Additionally, Project G includes projects to address proposed LTEMP Performance Metrics. Many project elements in Project G are long-term monitoring projects to estimate humpback chub abundances at various life stages throughout different parts of Grand Canyon.

In addition, monitoring from fixed sites (e.g., juvenile chub monitoring [JCM]-east, JCM-west, lower LCR) provides estimates of demographics (e.g., life-stage specific abundances) and population processes (e.g., survival, growth) which can allow us to forecast population dynamics under different management scenarios. Past data collected by these monitoring projects provide a baseline for comparison under past and future climate scenarios and can also provide information about if and how potential increases in warm-water nonnative fishes (e.g., smallmouth bass, green sunfish [*Lepomis cyanellus*], walleye [*Sander vitreus*]) affect humpback chub population dynamics. For this reason, continuing long-term monitoring is a priority in Project G for FY 2025-27. Importantly, while the focus of these trips is humpback chub sampling, these monitoring trips sample other native fishes (e.g., razorback sucker [*Xyrauchen texanus*], flannelmouth sucker [*Catostomus latipinnis*], bluehead sucker [*Catostomus discobolus*], and speckled dace [*Rhinichthys osculus*]) and nonnative fishes and can help assess changes in Grand Canyon's fish assemblage. But, since project N is largely unfunded, data analysis of these other native fishes will be limited.

Additionally, monitoring outlined in Project G will help document whether high age-0 production of humpback chub in 2022 and 2023 in the Colorado River ecosystem experience resource limitation, density-dependence, and natural population fluctuations in the near future. In both JCM-east and JCM-west, monitoring efforts have documented strong cohorts (i.e., lots of age-0 fish) in 2022 and 2023, years when adult humpback chub abundances were also high. It is unknown whether juvenile humpback chub born in 2022 and 2023 will successfully recruit to the adult population, or whether resource limitation and density-dependence will cause high mortality of these fish, similar to the population decline documented for rainbow trout (*Oncorhynchus mykiss*) in 2014 and 2015 that occurred after a large production pulse in 2011 (Korman and others, 2017). Similarly, the LCR had high production of age-0 humpback chub in spring 2023, but these fish were all of small size, and it is unknown how many from this cohort will survive winter in the LCR.

In summary, Project G includes project elements to estimate abundances required by the 2016 Biological Opinion (G.1, G.2, G.3, G.4) by monitoring humpback chub in the LCR-spawning population by sampling the LCR and JCM-east reach in the Colorado River (Project Elements G.2, G.3). Additionally, this project includes sampling in western Grand Canyon via continuation of mark-recapture in the JCM-west reach (210.5-214 river miles downstream of Lee Ferry) and extensive spatial sampling via the aggregation trips (Project elements G.5, G.6). Mark-recapture data from these trips will be supplemented with data from autonomous PIT tag antennas (Project Elements G.4, G.9), such as the LCR multiplexer array and submersible antennas, as these technologies have proven effective at detecting larger adults. Data collected from the above-mentioned field efforts will be analyzed to help learn more about humpback chub life history and to guide management efforts (Project Element G.1). In an effort to help address To promote animal welfare, Grand Canyon Monitoring and Research Center is undergoing a process to develop a formal protocol for the Institutional Animal Care and Use Committee (IACUC) to ensure best practices when handling animals to minimize their pain and distress.

Long-Term Experimental and Management Plan (LTEMP) Resource Goals

- Humpback Chub
- Other Native Fishes
- Nonnative Species

Background

Humpback Chub in Grand Canyon

We distinguish three different groups of humpback chub within Grand Canyon: 1) LCR spawners, 2) mainstem spawners in western Grand Canyon (i.e., between the confluence of Havasu Creek and Pearce Ferry rapid), and 3) humpback chub in other tributaries (e.g., Havasu Creek, Shinumo Creek, Bright Angel Creek). Of these three groups, LCR-spawning humpback chub represent the most established component of the population, as this group was able to persist in relatively high abundance in Grand Canyon despite construction of Glen Canyon Dam and resulting reductions in water temperature (Wright and others, 2009).

In contrast, mainstem spawners in western Grand Canyon were rare up until 10-15 years ago, but since 2016 humpback chub have become numerous and widespread throughout western Grand Canyon and represents a relatively ‘new’ component of the humpback chub population (Van Haverbeke and others, 2017). Humpback chub in other Grand Canyon tributaries are less abundant than the LCR and western Grand Canyon groups and are largely the result of translocations by the National Park Service. Humpback chub in the first two groups (LCR-spawning, and western Grand Canyon mainstem-spawning) are monitored by U.S. Fish and Wildlife Service (USFWS) and USGS, whereas humpback chub in other tributaries are monitored by National Park Service (NPS) outside of the Glen Canyon Dam Adaptive Management Program (GCDAMP).

LCR-Spawning Humpback Chub (Project Elements G.1, G.2, G.3, G.4, G.7, G.8)

After closure of Glen Canyon Dam in the mid-1960s, water temperatures in the mainstem Colorado River plummeted due to hypolimnetic releases from Lake Powell (Wright and others, 2009). Lab studies confirmed that generally water temperatures in this segment of the Colorado River were too cold for successful humpback chub reproduction (Hamman, 1982; Gorman and VanHoosen, 2000; Robinson and Childs, 2001). Humpback chub were able to successfully spawn and recruit in the LCR, which had warmer water temperatures and also a more natural flood regime compared to the Colorado River mainstem (Kaeding and Zimmerman, 1983; Robinson and Childs, 2001).

Humpback chub abundance in the LCR decreased in the early 2000s (Coggins, Jr. and others, 2006), prompting trout removal efforts in the mainstem near the LCR confluence from 2003-2006.

Humpback chub abundances started increasing in the mid-2000s, shortly after trout removals and also during a period of warming mainstem water temperatures, leading to uncertainty about what mechanism(s) ultimately were responsible for the population rebound (Coggins, Jr. and others, 2011). Additionally, un-modeled temporary emigration led to retrospective negative bias in humpback chub abundances (Coggins, 2008), suggesting the need for a different modeling approach that included monitoring both in the LCR and mainstem Colorado River. Since the mid-2000s, adult population abundances have increased and remained above 9000 for LCR-spawning humpback chub (Dzul and others, 2021b).

Ecological studies of LCR-spawning humpback chub indicated a dual life history strategy – some individuals (hereafter LCR residents) in this population never leave the LCR and others (hereafter LCR migrants) that are originally born into the LCR out-migrate to the Colorado River as juveniles or subadults and then migrate back to the LCR to spawn as adults (Yackulic and others, 2014; Dzul and others, 2021a). LCR residents typically exhibit a ‘fast’ life history strategy, where they grow to adulthood within 3 years and live 1-2 years upon reaching adulthood.

In contrast, LCR migrants have slower growth due to inhabiting colder water temperatures¹, so they typically take 3-6 years to reach adulthood but have longer lifespans (typically live on ~5-8 years after reaching adulthood but can live over 30 years). Skipped spawning occurs in this population, however, incorporation of antennas into population models suggests that skipped spawning is less common than previously reported by other models (Yackulic and others, 2014; Dzul and others, 2021a).

Abundances of LCR-spawning humpback chub are used to evaluate the need for management actions as part of the Biological Opinion in the LTEMP ROD (U.S. Department of the Interior, 2016; U.S. Fish and Wildlife Service, 2016). Specifically, the Biological Opinion identifies three tiers for management actions based on adult abundances of > 9000 (tier-0), 7000-9000 (tier-1), and < 7000 (tier-2). Subadult abundances are also included to ensure that adult abundances are on-track to meet the threshold abundances associated with each tier. These tiers can ‘trigger’ management actions – for example, if adult humpback chub abundance is between 7000-8999, then additional management actions (namely, humpback chub translocations) are warranted and if adult abundance falls below 7000 and rainbow trout densities are moderate or high (>60 individuals per km), trout removal efforts are required.

¹However, water temperatures have been warming in recent years (Dibble and others, 2021). Lake Powell reservoir levels in particular were very low in 2022, resulting in warm water temperatures and fast growth of humpback chub (Figure 3).

Abundance values for tier determination are estimated from Project Elements G.1, G.2, and G.3. Chute Falls translocations (G.7) represent an action that can be used to improve humpback chub abundances in tier-1.

The future health of the LCR-spawning population is dependent on future conditions, which are uncertain (Figure 1).

Predictions of future climate suggest increased warming and more severe water shortages for the U.S. Southwest (Udall and Overpeck, 2017), leading to increased likelihood of the following stressors: 1) reduced flooding in the LCR due to heat-stressed drought (Udall and Overpeck, 2017), 2) warmer mainstem water temperatures due to lower Lake Powell water levels² (Dibble and others, 2021), 3) population expansion/growth of nonnative piscivorous fishes (e.g., smallmouth bass, green sunfish, walleye; Dibble and others, 2021).

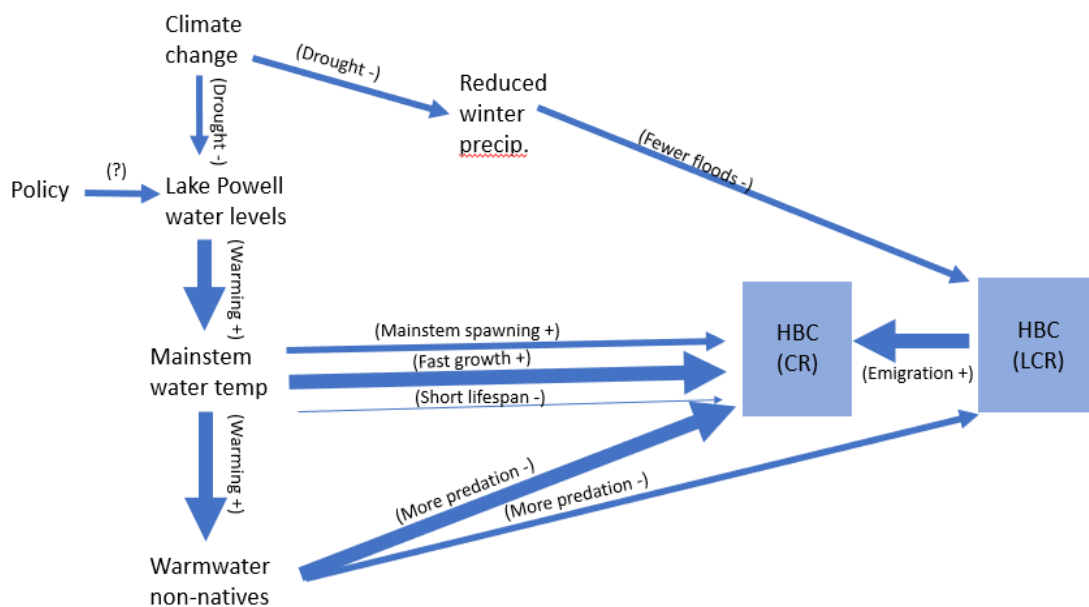


Figure 1. Conceptual model relating stressors to humpback chub (*Gila cypha*) (HBC) abundances in the Little Colorado River (LCR) and Colorado River (CR) for LCR-spawning humpback chub. Arrows represent processes and the sign in the descriptor is the hypothesized direction for the effect. The width of the arrow represents the certainty and strength of the relationship, with thick arrows representing stronger, more certain relationships than thin arrows.

²However, note Lake Powell water levels are impacted by policy, which is uncertain after 2026.

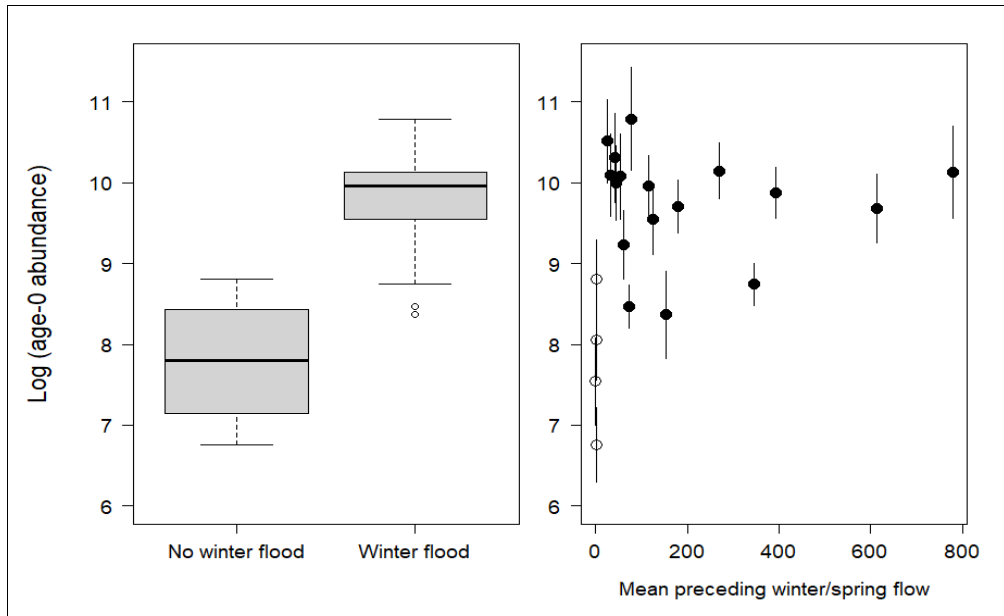


Figure 2. Log of age-0 humpback chub (*Gila cypha*) abundance in the Little Colorado River (LCR) in July plotted by years (2001-2020) with and without preceding winter floods (i.e., from 1-November to 1-April; left panel). Log of age-0 abundance in the LCR in July as a function of mean preceding winter flow (cubic feet per second - cfs) measured at the Cameron gage (USGS gage 09402000). Filled circles represent years with preceding winter floods and open points represent years without preceding winter floods. Figure from Dzul (2021).

The high likelihood of heat-induced drought (Udall and Overpeck, 2017) suggests a future with reduced river flooding, which could have negative impacts to age-0 humpback chub in the LCR. Monitoring efforts show substantial variability in age-0 humpback chub abundances from year to year (Dzul, 2021). Age-0 abundance in the LCR is low in years without preceding winter floods (Van Haverbeke and others, 2013), but even small winter floods can lead to production of large age-0 cohorts (Dzul, 2021; Figure 2). How floods affect age-0 production is unknown but could be due to factors such as siltation of gravels leading to poor egg survival, increases in nonnative fishes associated with baseflow years (Stone and others, 2018) leading to increased predation and lower survival of eggs/larvae, lack of spawning cue leading to low reproduction, or low food availability leading to poor larval survival. Thus, if the U.S. Southwest continues to incur drought, we predict that the LCR will support fewer age-0 humpback chub due to reduction in winter floods. Note, however, this prediction is uncertain because even small floods have been associated with strong age-0 production.

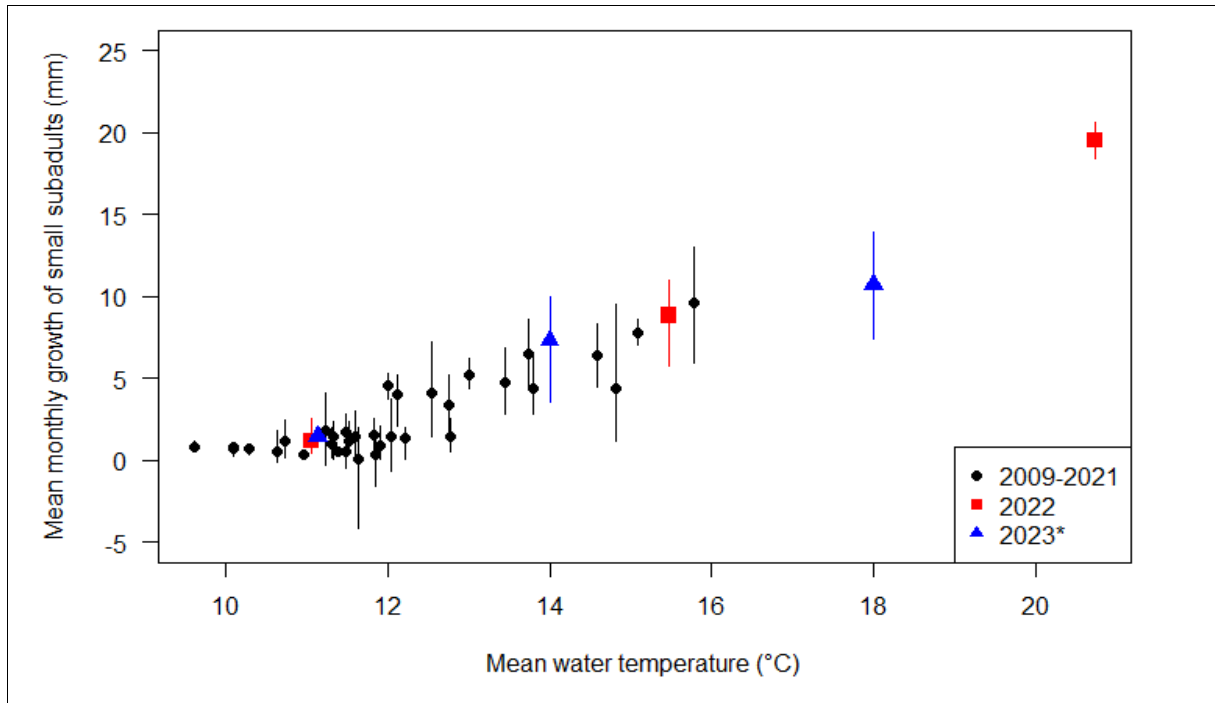


Figure 3. Observed growth of subadult humpback chub (*Gila cypha*) (starting length 100-149mm total length) captured on consecutive sampling trips as a function of mean water temperatures over that interval. Colors and shapes of points correspond to the year of the observation. Water temperature data were obtained from USGS gage 0938310. *Water temperatures from 2023 were not available and were approximated.

Warmer water temperatures in the mainstem could have multiple effects on humpback chub. Growth models of humpback chub in Colorado River in Grand Canyon establish that growth is strongly influenced by temperature (Gorman and VanHoosen, 2001; Dzul and others, 2016). In fact, in 2022, large increases in water temperature coincided with a period of much faster growth than had been observed since closure of Glen Canyon Dam (Figure 3). Increased growth in the mainstem may be beneficial to humpback chub, because fish spend less time at small size where they are more vulnerable to predation.

However, warmer water temperatures can also increase metabolic demands and oxidative stress, sometimes leading to reduced lifespan. If mainstem water temperatures are warmer in eastern Grand Canyon, we predict faster growth for LCR migrants and (possibly) reduced adult lifespan (Lee and others, 2013; Taylor and others, 2019).

Another potential effect of warming water temperatures in eastern Grand Canyon is mainstem spawning near the LCR confluence. A study of early life history humpback chub observed that survival from the egg to swim-up fry phase was ~30x greater at 16-17°C (56%) compared to 12-13°C (2%; Hamman 1982). In 2022, a year with low Lake Powell water levels, the Colorado River water temperatures near the LCR confluence (USGS gage 09383100) measured 150 days over 16°C and 92 days over 20°C, suggesting thermal conditions may be suitable for humpback chub spawning and larval development.

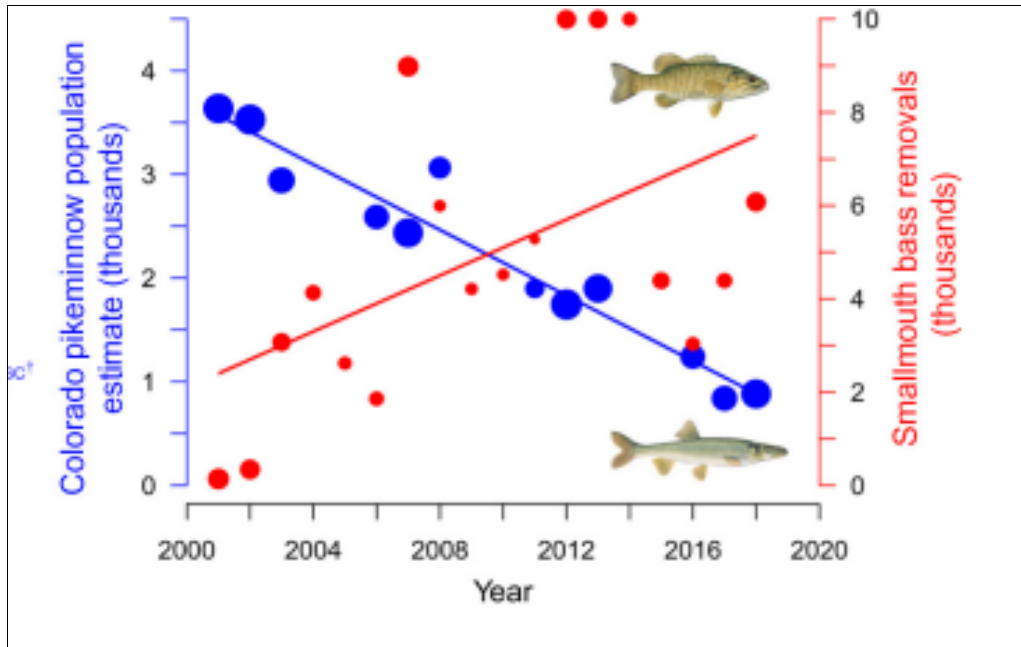


Figure 4. Relationship between estimates of Colorado pikeminnow (*Ptychocheilus lucius*) abundance and smallmouth bass (*Micropterus dolomieu*) catch (through removals) in the upper Colorado River basin. Figure from Dibble and others, 2021.

Warmer water temperatures could also lead to population growth of piscivorous, warm-water nonnative fishes, which could have negative impacts to humpback chub, as has been observed for native fish populations in the upper Colorado River basin (Dibble and others, 2021; Figure 4). Certain nonnative fishes may not be able to survive or reproduce in the LCR due to the LCR's harsh abiotic conditions (e.g., large floods, high levels of dissolved carbon dioxide; Stone and others, 2018). Thus, it is possible that humpback chub in the LCR may be less vulnerable to predation than mainstem humpback chub. However, this is highly uncertain.

In addition to emerging threats and ecosystem changes, management actions (e.g., spring HFEs to build sand bars, temperature changes/flow disturbances to disadvantage problematic nonnative fishes) may occur to help improve LTEMP resource goals in the Colorado River ecosystem.

The long-term record of LCR-spawning humpback chub population abundances can serve a baseline for comparing impacts of natural variability in conditions, emerging stressors, and management actions. Applying new modeling approaches to continued monitoring data will be necessary to evaluate drivers and impacts to humpback chub population dynamics. Modeling approaches will focus on linking environmental covariates (i.e., water temperature, flow, turbidity) to population processes (e.g., growth, survival).

Humpback Chub in Western Grand Canyon (Project Elements G.5, G.6, G.9)

Up until the last decade, humpback chub in western Grand Canyon were associated with two aggregation³ sites (Colorado River near Havasu Creek and Pumpkin Spring), which were estimated to have abundances of 13 and 5 adult humpback chub, respectively (Valdez and Ryel, 1997). Starting as early as 2006, increases in catch of humpback chub were observed at aggregation sites (Persons and others, 2017), and, by 2017, humpback chub became widespread throughout the western portion of Grand Canyon downstream of the confluence with Havasu Creek (Van Haverbeke and others, 2017; Rogowski and others, 2018).

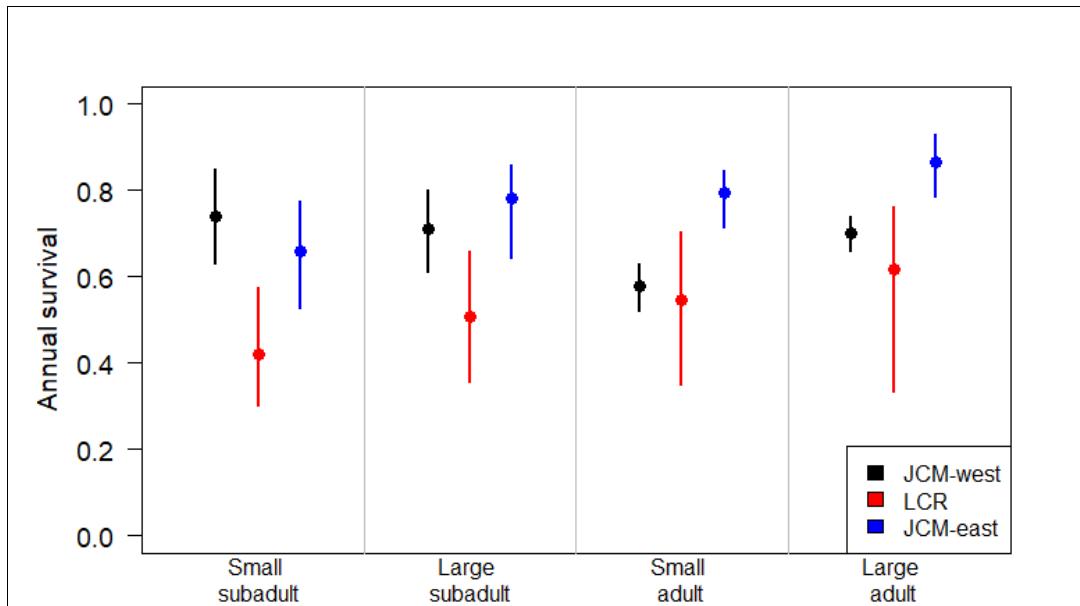


Figure 5. Graph of survival probabilities for humpback chub for different life stages (small subadult- 100-149mm TL; large subadult- 150-199mm TL, small adult- 200-249mm TL, large adult ≥ 250 mm TL) in three different locations (LCR, JCM-east, JCM-west). Survival for JCM-east and LCR humpback chub corresponds to true survival. In contrast, survival in JCM-west represents apparent survival (i.e., probability fish survives and remains in the study reach).

Ecological studies of western Grand Canyon humpback chub show this group is unique from LCR-spawners and that their life history is intermediate to that of LCR-residents and LCR-migrants on the fast-slow spectrum (Dzul and others, 2023). Specifically, western Grand Canyon chub exhibited faster growth and lower asymptotic body size compared to LCR migrants and slower growth/higher asymptotic body size compared to LCR residents.

³Aggregations were a term first used by Valdez and Ryel (1997) to describe the patchy spatial distribution of humpback chub in the Colorado River in Grand Canyon during the late 20th century. Most aggregations sites occurred in areas near sources of warm water (e.g., near the mouths or tributaries or in warm-water springs) that were interspersed throughout the river. Valdez and Ryel (1997) identify nine aggregations, but only two of these are in what is considered to be western Grand Canyon (i.e., downstream of Havasu Creek).

Apparent survival of subadult humpback chub in western Grand Canyon was relatively high and more comparable to that of LCR migrants (Figure 5). Taken together, the faster growth and higher survival of subadults partly explains the rapid population growth observed in western Grand Canyon.

Adult apparent survival (i.e., the probability of surviving and remaining in the study reach) displayed a more tenuous pattern because, unlike both LCR migrants and LCR residents, apparent survival in western Grand Canyon did not increase with fish size (Dzul and others, 2023; Figure 5). Adult apparent survival in western Grand Canyon tended to be relatively low (more comparable to LCR residents), and it is unknown whether the lower observed apparent survival probability is due to high mortality or high emigration from the study area. Apparent survival probabilities were lowest in spring, a time when humpback chub are spawning. Previous movement studies of LCR-spawning and upper basin humpback chub suggest humpback chub tend to exhibit only localized movements around a small spatial area (Valdez and Ryel, 1995, Kaeding and others, 1990), though LCR-migrants did exhibit longer-distance spawning migrations into the LCR in spring (Valdez and Ryel, 1995). It is unknown whether western Grand Canyon humpback chub are more mobile compared to LCR-spawning humpback chub and(or) whether this group migrates during the spawning season. To help address this question, we have introduced an antenna monitoring project that uses citizen science to spatially expand and randomize detection probability within Grand Canyon (Project Element G.9).

Proposed Work

Project Element G.1. Humpback Chub Population Monitoring (Ongoing Study; Partially Funded)

Maria Dzul, Lindsay Hansen, Ben Miller, Brian Healy, Kim Dibble, Charles B. Yackulic

U.S. Geological Survey, Southwest Biological Science Center, Grand Canyon Monitoring and Research Center

Evaluate Population Dynamics Using Mark-Recapture Data

We will develop and refine models using data from existing (and ongoing) field sampling to help inform management efforts and conservation actions for humpback chub. Proposed foci for modeling efforts may include 1) estimating abundance of various size classes in the LCR-spawning population as defined in the 2016 Biological Opinion (e.g., Figure 6; U.S. Fish and Wildlife Service, 2016), 2) including antenna detections from a wide geographic area in Grand Canyon to obtain a better estimate of adult survival in JCM-west (i.e., to differentiate emigration from mortality), 3) estimate abundance of humpback chub adults in western Grand Canyon based on data collected by humpback chub aggregations sampling trips, and 4) estimate the effects of environmental covariates on demographic parameters (e.g., survival, growth, movement), which may help predict humpback chub responses to management actions and future scenarios, and 5)

developing an occupancy model that accounts for detection probability to look at long-term, large-scale changes to humpback chub distribution with the Colorado River ecosystem.

Additionally, these models will help support LTEMP Performance Metrics proposed for humpback chub, including 1) the current tier for LCR-spawning humpback chub in the 2016 Biological Opinion (US Fish and Wildlife Service, 2016), 2) an estimate of humpback chub abundances across all Grand Canyon, and 3) detection/non-detection (or presence/absence) of different life stages of humpback chub across Grand Canyon.

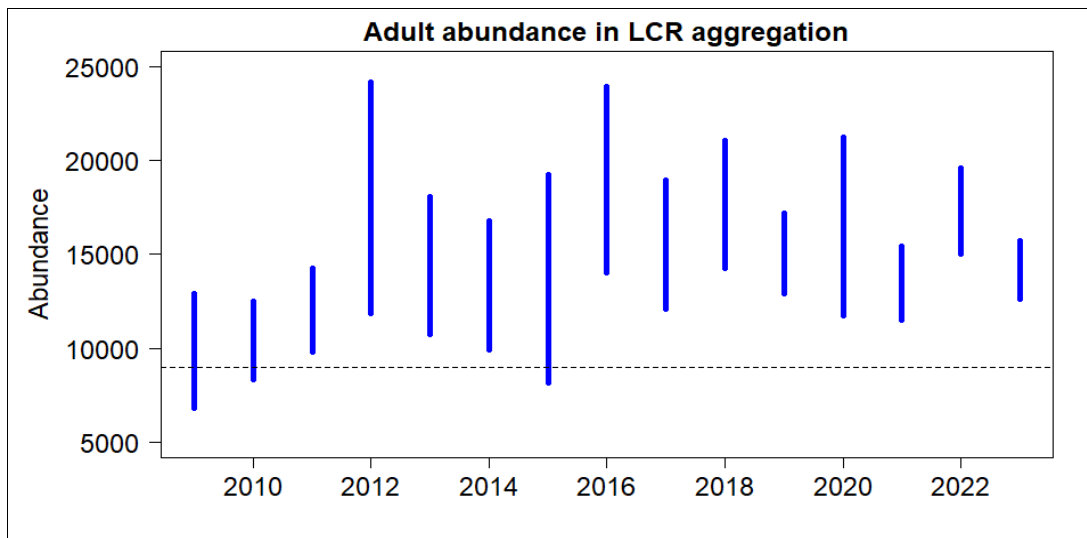


Figure 6. Estimates of adult humpback chub (*Gila cypha*) (≥ 200 mm total length) abundance for LCR-spawning humpback chub used for reporting for the LTEMP Biological Opinion (U.S. Department of the Interior, 2016; U.S. Fish and Wildlife Service, 2016). Bars represent 95% credible intervals.

Hypotheses and Science Questions

- Are LCR-spawning humpback chub increasing, decreasing, or stable?
 - Which life history stages and demographic processes are influencing trend?
 - Are management actions needed to increase abundances to meet compliance (i.e., triggers)?
- Are western Grand Canyon humpback chub increasing, decreasing, or stable?
 - Which life history stages and demographic processes are influencing trend?
- How are LCR-spawning humpback chub population dynamics (i.e., survival, growth, movement, recruitment) affected by environmental factors (e.g., water temperature, turbidity, flow)?
- How are humpback chub survival and growth impacted by nonnative fishes in the LCR and in eastern Grand Canyon? (contingent on increases in nonnative fishes).

- How are humpback chub population dynamics and abundances in the LCR and in eastern Grand Canyon affected by management actions (e.g., spring HFEs, proposed flows for nonnative suppression)?

Methods

Descriptions of the field methods and multistate model used for reporting to GCDAMP stakeholders can be obtained in Yackulic and others (2014), Yackulic and others (2020), and Dzul and others (2021b).

Anticipated Use of Data

We will continue to provide annual summaries of life-stage specific abundances and demographic processes that serve as triggers for management actions in the 2016 Biological Opinion (e.g., abundance, survival, recruitment, and growth rates; U.S. Fish and Wildlife Service, 2016) and continue to develop predictive models that have served as decision support tools for the LTEMP EIS and post-2016 guidelines. Also, we will provide summaries of LTEMP Performance Metrics to the GCDAMP and evaluate the effects of LTEMP flow experiments on humpback chub population dynamics. Sustained, multiyear increases or decreases in humpback chub will be analyzed to evaluate potential drivers (e.g., nonnative species, discharge, water temperature) and inform management of Glen Canyon Dam.

Outcomes and Products

Scientists will present at the annual reporting meeting to GCDAMP stakeholders on updates to trends in humpback chub and (if there's interest) other native fishes. We anticipate one manuscript describing abundance estimation for humpback chub in western Grand Canyon and one manuscript describing environmental influences on humpback chub growth. Other manuscripts may include analyses of environmental influences on age-0 abundances in the LCR, drivers of range expansion/retraction of humpback chub and other fishes in Grand Canyon, and evaluation of environmental drivers on humpback chub survival.

Project Element G.2. Annual Spring/Fall Abundance Estimates of Humpback Chub in the Lower 13.6 km of the LCR (Ongoing Study)

Michael Pillow, David Ward, Pilar Rinker

U.S. Fish and Wildlife Service

USFWS Sampling Trips

The USFWS has been conducting four sampling trips into the LCR each year since 2001, and this data set is used to estimate humpback chub abundance at different life stages.

These abundance estimates document substantial temporal changes to adult population size since 2000, most notable of which is the increase in adult abundance that has occurred since 2007. While all four trips sample all life stages, spring trips generally inform adult (Figure 7), large subadult, and age-1 abundance estimates in the LCR, as well as provide information about the migratory component of LCR-spawners, whereas the fall trips provide estimates of age-0 abundances in the LCR and abundances of LCR residents (Van Haverbeke and others, 2013). Spring LCR trips generally are the best opportunity for capturing (and issuing tags to) adult humpback chub, because capture probabilities for hoop nets are higher for adults in the LCR compared to the mainstem.

Hypotheses and Science Questions

- What are the lifestage-specific trends in humpback chub (e.g., >100mm, ≥150mm, ≥200mm TL; Figure 7) in the lower 13.6 km of the LCR during the spring and fall?
- How many age-0 humpback chub (40-99 mm TL) remain in the LCR during fall (i.e., after some variable proportion of age-0 humpback chub have emigrated to the mainstem)?
- What are trends in other LCR fishes and external parasites (i.e., *Lernaea cyprinacea*)?
- What are trends in sexual condition and characteristics?

Methods

USFWS conducts four 9-day sampling trips (typically in April, May, September, and October) to the lower 13.56 kilometers of the LCR to capture fishes using three passes of hoop nets. Upon capture, fishes are measured for total length and fork length and scanned with a PIT tag reader. Untagged fishes may be marked with either VIE or PIT tag, depending on their size and species. Fishes are released. More detailed methods are described in Van Haverbeke and others (2013) and Persons and others (2015).

Anticipated Use of Data

Model results will be reported to GCDAMP at the annual reporting meeting and as part of compliance for the Biological Opinion (US Fish and Wildlife Service, 2016), which can be used to determine the need for management actions such as translocations or trout removals. Specifically, the large subadult triggers in the spring in the LCR and the estimates of total adult abundance in the fall are informed by these efforts.

Outcomes and Products

USFWS scientists will present at the annual reporting meeting to GCDAMP stakeholders on updates to trends in LCR fishes, including humpback chub and (if there's interest) other native fishes and(or) nonnative fishes. Additionally, USFWS provides trip reports to GCMRC on a biannual basis.

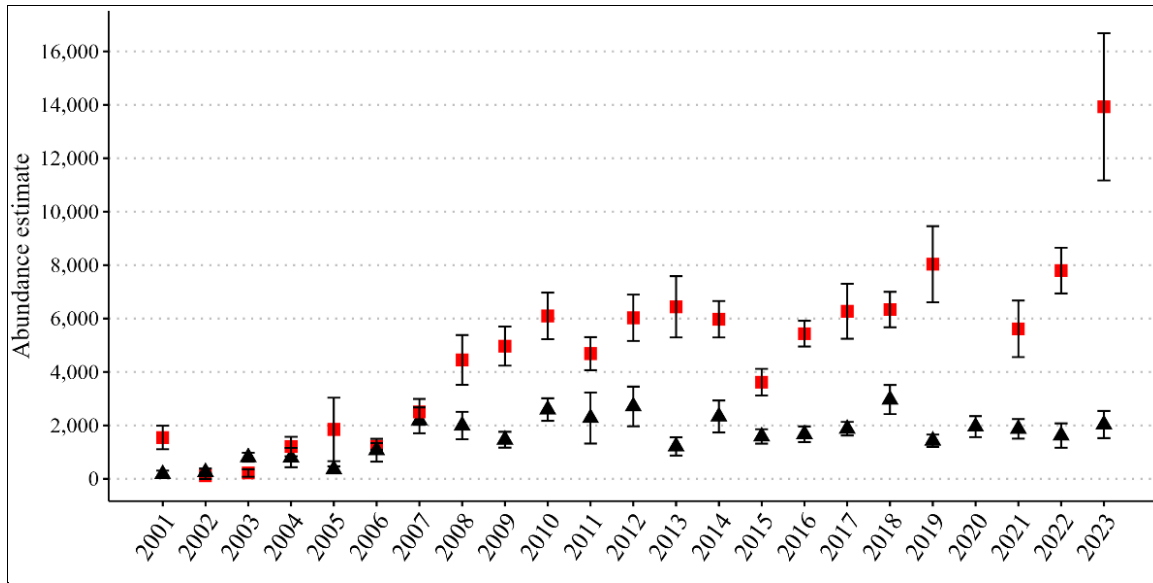


Figure 7. Spring (red squares) and fall (blue triangles) abundance estimates of adult humpback chub (*Gila cypha*) (≥ 200 mm total length) in the Little Colorado River (LCR) from U.S. Fish and Wildlife Service lower LCR monitoring trips.

Project Element G.3. Juvenile Chub Monitoring (JCM) near the LCR Confluence (Ongoing Study)

Lindsay Hansen, Ben Miller, Maria Dzul

U.S. Geological Survey, Southwest Biological Science Center, Grand Canyon Monitoring and Research Center

Sampling Trips to the JCM-east Reach and Lower LCR

This project element is a continuation of previous monitoring work that commenced in 2012 and includes three annual sampling trips to the JCM-east reach (located 62.7-66.0 river miles downstream of Lees Ferry) as well as one annual sampling trip to the lower LCR (July) to estimate abundance and outmigration of age-0 humpback chub born the previous spring. Because these trips visit fixed sites and use mark-recapture methods, this project element provides information about demographic processes, such as survival and growth (Dzul and others, 2016, Dzul and others, 2023) as well as abundances (Dzul and others, 2021b).

The relatively long-term data set from Project Elements G.2 (25 years) and G.3 (16 years) provides baseline data about population dynamics, which can serve to help compare how future changes to the CRe (e.g., increased water temperatures, more warm-water nonnative fishes) may affect humpback chub population dynamics.

Previous work has suggested substantial variability in age-0 humpback chub production and outmigration from the LCR to the CR (Yackulic and others, 2014; Dzul 2021; Figure 8). The July LCR trip focuses on estimating abundances of age-0 fish that were born the preceding spring, to learn more about variability in age-0 production across years in the LCR. The mainstem JCM trips can recapture fishes marked in the LCR and provide information about outmigration probability and abundance of juveniles/subadults in the Colorado River (i.e., that will become LCR migrants as adults). Because most (~80%) LCR-spawning adults are migratory (i.e., move between the Colorado River and LCR), understanding the key drivers and natural variability in recruitment and partial migration can help generate predictive models for humpback chub that spawn in the LCR. If the Colorado River experiences more warming in the future, it is possible that some humpback chub in the Colorado River near the LCR confluence may start reproducing in the mainstem, and Project Element G.3 may help inform whether or not this occurs.

Lastly, sampling in the Colorado River addresses the shortcomings of previous models (Coggins, 2008) that exhibited retrospective biases in humpback chub abundances, likely due to relying solely on LCR monitoring and not accounting for the possibility that some adults may not visit the LCR every year to spawn. Additionally, if the Colorado River warms, it is possible that more mainstem spawning may occur, so that some adults may never visit the LCR and would be missed by LCR sampling efforts.

Hypotheses and Science Questions

- Are humpback chub in JCM-east increasing, decreasing, or stable?
 - Which life history stages and demographic processes are influencing trend?
 - Are management actions needed to increase abundances to meet compliance (i.e., triggers)?
- Are there trends in rainbow trout and brown trout (*Salmo trutta*) abundance near the LCR confluence (Yackulic and others, 2018)? Is there a correlation between abundance of trout and(or) other nonnative fishes and humpback chub in JCM-east?
- How many age-0 humpback chub are produced each year? How are humpback chub abundances and outmigration to the Colorado River affected by floods and other environmental factors?

- Will mainstem spawning become more prevalent for native fishes in eastern Grand Canyon if water temperatures warm?
- How are humpback chub population dynamics and abundances in the LCR and in eastern Grand Canyon affected by management actions (e.g., spring HFEs, proposed flows for nonnative suppression)?

Methods

This project element includes three trips each year to the JCM-east reach (May, June/July and October) and one trip each year to the LCR (late June – before the onset of monsoon season). JCM trips will use a variety of gear types to sample fishes (i.e., hoop nets, portable remote PIT-tag antennas, nighttime electrofishing), and LCR trips use hoop nets as well as seines to target small, age-0 humpback chub. Detailed methods can be found in Dzul (2021).

Anticipated Use of Data

Data collected by this effort are modeled in Project Element G.1 to obtain abundances, then reported to GCDAMP at the annual reporting meeting as part of compliance for the Biological Opinion (US Fish and Wildlife Service, 2016). Specifically, the large subadult triggers in JCM-east and the estimates of total adult abundance are informed by these efforts. Estimates related to triggers in the Biological Opinion can be used to determine the need for management actions such as translocations or trout removals.

Outcomes and Products

We anticipate one manuscript describing the effects of environmental factors on humpback chub growth, which is informed by mark-recapture data from JCM-east and JCM-west. Additionally, we hope to publish one manuscript describing age-0 humpback chub population dynamics based on sampling of JCM-east (variability in abundances and outmigration to the Colorado River).

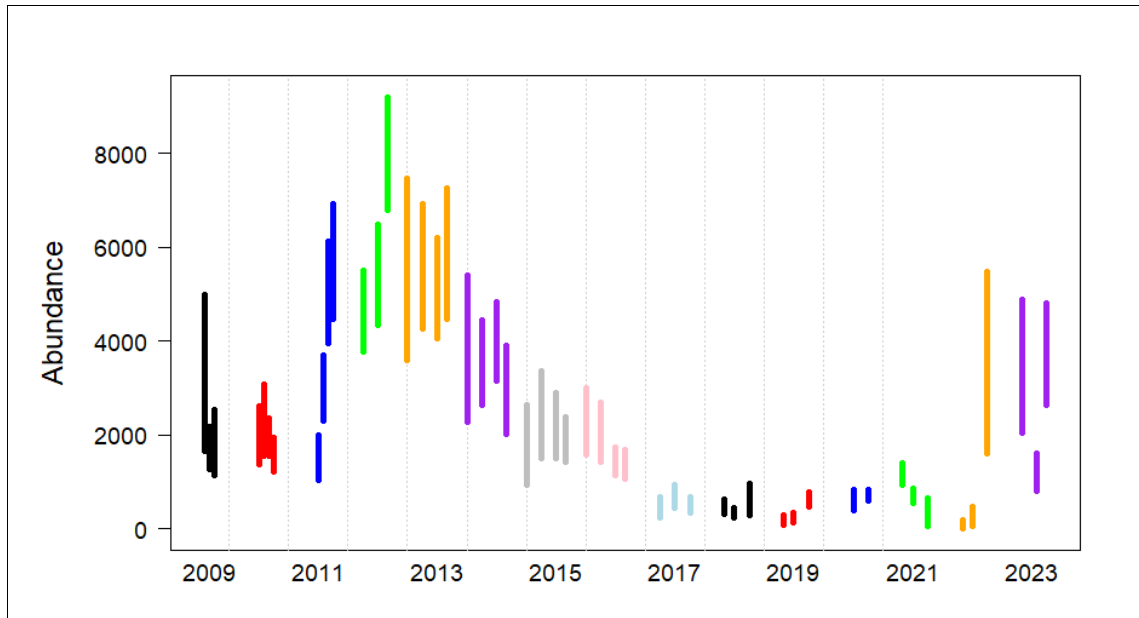


Figure 8. Abundance estimates of juvenile (<100mm TL) humpback chub (*Gila cypha*) in JCM-east sampling reach in the Colorado River. Bars correspond to 95% credible intervals.

Project Element G.4. Remote PIT Tag Array Monitoring in the LCR (Modified Study; Partially Funded)

Maria Dzul

U.S. Geological Survey, Southwest Biological Science Center, Grand Canyon Monitoring and Research Center

Assessing Humpback Chub Movement and Improving Detection of Large Fish

Unlike other monitoring types (e.g., hoop nets, electrofishing, seines) which require physical captures, PIT-tag arrays read and record codes from tagged fish that swim over antennas anchored to the river bottom or riverbank. Accordingly, these arrays provide a method for boosting recapture events without requiring additional fish handling (Figure 9). Importantly, these systems are particularly useful for detecting large (>250mm TL) humpback chub which are difficult to capture using hoop nets (Dzul and others, 2021a, 2024). In the LCR, models without antenna data underestimated the proportion and number of migratory adult humpback chub that moved into the LCR in spring months and also underestimated adult survival compared to models with these data (Dzul and others, 2021a). Taken together, these results illustrate that the benefit of including PIT-tag arrays is not solely based on their ability to increase detection probabilities for all humpback chub (Figure 9), but rather their ability to increase detection of a subset of the humpback chub population that is relatively invulnerable to capture.

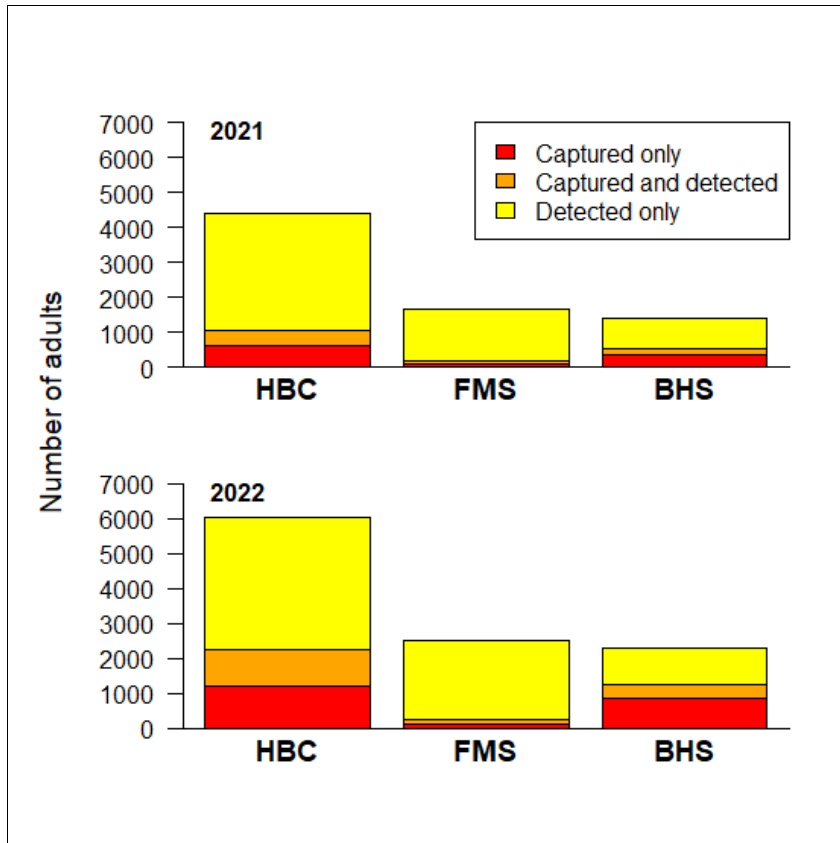


Figure 9. Comparison of adult humpback chub (*Gila cypha* - HBC), flannelmouth sucker (*Catostomus latipinnis* - FMS) and bluehead sucker (*Catostomus discobolus* - BHS) that were only physically captured (red), physically captured and detected on the MUX (orange), and only detected on the MUX (yellow) during spring of 2021 and 2022 in the Little Colorado River (LCR). Figure from Dzul and others (2024).

PIT tag array detections in the LCR includes continuous detections from two sources: a multiplexer system (MUX) and a network of shore-based single antennas (NET). The NET system (described in FY 2021-23 TWP) is being discontinued due to redundancy with the MUX and funding limitations. The MUX provides detection coverage across the entire channel across two arrays (downstream and upstream arrays) to inform movement directionality. The MUX is located in the LCR, ~1.8 river km upstream of the confluence with the Colorado River and was initially installed in 2009 and 2011 but replaced with an improved system in 2020. Maintenance and installation of the MUX has been funded by Reclamation outside of GCDAMP, however, GCDAMP funding is requested because USGS staff are asked to help with minor repairs, setting up trip logistics for maintenance trips, and assisting on maintenance/repair visits. Additionally, there are some USGS salary costs associated with data management and modeling antenna detection to improve population models for humpback chub and other native fishes. The MUX can also be used to detect movement of nonnative fishes that are tagged (e.g., rainbow trout, channel catfish [*Ictalurus punctatus*]) to determine whether or not nonnative species can persist in the LCR (Dzul and others, 2018).

Hypotheses and Science Questions

- Are there environmental cues for movement of humpback chub into and out of the LCR each spring?
- What is the abundance of humpback chub that visit the LCR each year in spring?
- Are there trends in rainbow trout and brown trout (*Salmo trutta*) abundance near the LCR confluence (Yackulic and others, 2018)? Is there a correlation between abundance of trout and(or) other nonnative fishes and humpback chub in JCM-east?
- How many age-0 humpback chub are produced each year? How are humpback chub abundances and outmigration to the Colorado River affected by floods and other environmental factors?
- Will mainstem spawning become more prevalent for native fishes in eastern Grand Canyon if water temperatures warm?

Methods

This project element includes three trips each year to the JCM-east reach (May, June/July and October) and one trip each year to the LCR (late June – before the onset of monsoon season). JCM trips will use a variety of gear types to sample fishes (i.e., hoop nets, portable remote PIT-tag antennas, nighttime electrofishing), and LCR trips use hoop nets as well as seines to target small, age-0 humpback chub. Detailed methods can be found in Dzul (2021) and Dzul (2021b).

Anticipated Use of Data

Data collected by this effort are modeled in Project Element G.1 to obtain abundances, then reported to GCDAMP at the annual reporting meeting as part of compliance for the Biological Opinion (US Fish and Wildlife Service, 2016). Specifically, the large subadult triggers in JCM-east and the estimates of total adult abundance are informed by these efforts. Estimates related to triggers in the Biological Opinion can be used to determine the need for management actions such as translocations or trout removals.

Outcomes and Products

We anticipate one manuscript describing the effects of environmental factors on humpback chub growth, which is informed by mark-recapture data from JCM-east and JCM-west. Additionally, we hope to publish one manuscript describing age-0 humpback chub population dynamics based on sampling of JCM-east (variability in abundances and outmigration to the Colorado River).

Project Element G.5. Monitoring Humpback Chub Aggregation Relative Abundance and Distribution (Ongoing Study; Partially Funded)

Pilar Rinker, David Ward, Michael Pillow

U.S. Fish and Wildlife Service

Canyon-wide Sampling of Fish throughout Grand Canyon with an Emphasis on Humpback Chub

Aggregations of humpback chub in Grand Canyon are biologically important because they provide redundancy and resiliency for the species. Annually monitoring the status and trends of humpback chub aggregations and conducting periodic surveys in between aggregations to identify additional aggregations are Conservation Measures listed in the Biological Opinion (U.S. Fish and Wildlife Service, 2016). Notably, this project element was key to first detecting the increase in humpback chub in western Grand Canyon in 2014 (Van Haverbeke and others, 2017) and, in turn, informing species-level conservation and spurring numerous research efforts to better understand drivers in this new population.

Project Element G.5 funds an additional boat that can be used to seine backwaters to monitor juvenile humpback chub and warm-water nonnatives (funded in FY 2025 and FY 2026, not funded in FY 2027). One partial trip (launching at Diamond Creek and taking out at Pearce Ferry) was requested in FY 2026 but remains unfunded. This will limit the ability for the aggregations trip to learn more about capture probabilities for humpback chub, especially if water temperature warm in future years, as many of the capture probabilities from the aggregations trip comes from years with cooler water temperatures.

Hypotheses and Science Questions

- How many adult humpback chub are in western Grand Canyon? Are abundances increasing, stable, or decreasing?
- What is the spatial distribution of humpback chub in western Grand Canyon?
- Is there evidence of successful reproduction for native and nonnative species in backwaters in Grand Canyon?

Methods

This project will conduct one mainstem sampling trip per year focused on aggregations sites. The annual aggregations trip will focus on hoop net monitoring of the known aggregations (e.g., RM 30-36, LCR, Bright Angel, Shinumo, Stephens Aisle/Middle Granite Gorge, Havasu, Pumpkin Spring) as well as other sites that are targeted for biological interest.

Aggregations monitoring obtains catch information for numerous reaches throughout Grand Canyon using baited hoop nets (Van Haverbeke and others, 2017), then periodically re-visits some sites a few weeks later to re-sample and obtain capture probability estimates. Once a range of capture probabilities are observed, models can generate abundances (Figure 10). Additionally, habitat and spatial information are used in models to help account for non-random sampling and spatial effects.

Anticipated Use of Data

Data from humpback chub aggregations trips will be used to generate abundance estimates for western Grand Canyon, which will be reported to GCDAMP stakeholders at the annual reporting meeting. Seining data will be used to detect nonnatives inhabiting backwaters and qualitatively assess mainstem production of age-0 humpback chub.

Outcomes and Products

We anticipate one manuscript describing abundance estimation in western Grand Canyon.

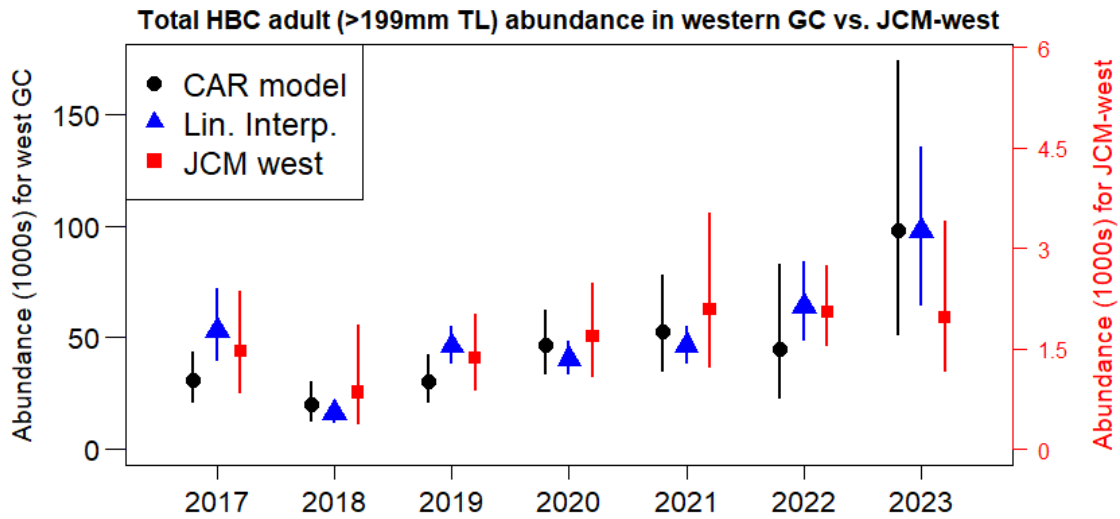


Figure 10. Unpublished data of abundances of humpback chub in western Grand Canyon (157-280 river miles downstream of Lees Ferry) as estimated by conditional auto-regressive models (CAR) and models with linear interpolation (Lin. Interp.). Additionally, abundances of humpback chub (*Gila cypha*) in JCM-west are modeled (left axis) to evaluate differences in trend across sites and sampling efforts.

Project Element G.6. Juvenile Humpback Chub Monitoring – West (Ongoing Study; Unfunded)

Lindsay Hansen, Ben Miller, Maria Dzul

U.S. Geological Survey, Southwest Biological Science Center, Grand Canyon Monitoring and Research Center

Sampling Fish in Fall Canyon Reach to Learn More About Survival, Abundance, and Growth

Humpback chub in western Grand Canyon have increased dramatically in the last 5-10 years (Van Haverbeke and others, 2017) and JCM-west monitoring was established to determine how survival, growth, abundance, and recruitment differs for this relatively ‘new’ group compared to the more established group of LCR-spawning humpback chub. Understanding the drivers of this population was a goal identified in the Conservation Measures of the Biological Opinion. Furthermore, the 2016 fisheries PEP specifically recommended additional study in the lower part of the CRE. To address these goals, monitoring of the JCM-west reach (i.e., in Fall Canyon, located 210.5-214.0 river miles downstream of Lees Ferry) commenced in fall 2017 and continued through 2024. Based on mark-recapture data from JCM-west, Dzul and others (2023) found different patterns in life-stage specific survival and growth rates compared to eastern Grand Canyon humpback chub (Figure 5, Figure 11) and may experience less temperature and food limitation than LCR-migrants and LCR-residents, respectively. Because humpback chub in JCM-west experience warmer water temperatures than humpback chub in JCM-east, and because water temperatures may warm in the future due to drought and low water levels in Lake Powell, humpback chub in JCM-west may serve as a predictor for changes that may occur to mainstem humpback chub near the LCR confluence.

Hypotheses and Science Questions

- What are life-stage specific survival and growth rates of humpback chub in western Grand Canyon, and how to these compare to those in eastern Grand Canyon and the LCR?
- What are the environmental drivers of humpback chub survival and growth in western Grand Canyon?
- How are humpback chub population dynamics and abundances in western Grand Canyon affected by management actions (e.g., spring HFES, proposed flows for nonnative suppression)?

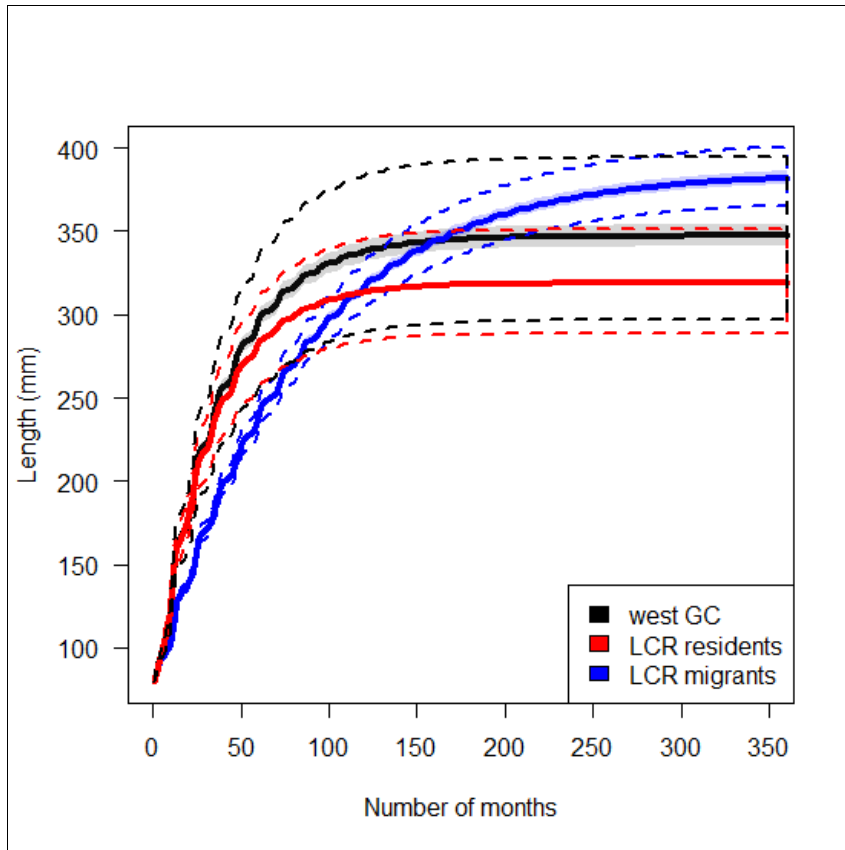


Figure 11. Growth curves for three groups of humpback chub (*Gila cypha*): Little Colorado River (LCR) residents (which spend their entire life in the LCR), LCR migrants (which migrate between the LCR and the main-stem Colorado River), and western Grand Canyon. The solid lines represent growth based on mean asymptotic length (L_{∞}), the shaded polygon represents 95% credible intervals for uncertainty in mean asymptotic length, and the dotted lines represent 2.5% and 97.5% quantiles for estimated asymptotic lengths across individuals. Graph is from Dzul and others (2023).

Methods

Detailed methods can be obtained in Dzul and others, 2023. In FY 2025-27, we proposed sampling of JCM-west reach occur during the same trips (i.e., May, June/July, October) and using the same sampling methods as in JCM-east (Project Element G.3). Due to cost increases and funding limitations, monitoring of JCM-west is currently unfunded in this work plan, so monitoring trips will only visit JCM-east and most personnel will hike out using Bright Angel Trail (except for the July trip, when everyone will continue on the river to Diamond Creek). Due to JCM-west currently being proposed as unfunded, inferences about how management actions, nonnative species, and other stressors affect survival and growth of humpback chub will mainly be limited to studies of LCR-spawners.

Anticipated Use of Data

Data from previous years may be used in analyses to help evaluate environmental drivers of growth and survival. Data collection in FY 2025-27 could be used in conjunction with antenna data from Project Element G.9 to obtain more accurate estimates of adult survival.

Outcomes and Products

One manuscript or report describing updated humpback chub movement and survival estimates in western Grand Canyon.

Project Element G.7. Chute Falls Translocations (Ongoing Study; Unfunded)

Michael Pillow, David Ward, Pilar Rinker

U.S. Fish and Wildlife Service

Management (and Monitoring) to Increase Humpback Chub Adult Abundance in the LCR-Spawning Population

Translocation and monitoring of humpback chub upstream of Chute Falls has been in place as a conservation action in Biological Opinions since 2002 (U.S. Fish and Wildlife Service, 2016, 2002). To date, over 4500 juvenile humpback chub have been translocated upstream of Chute Falls. Models suggest that humpback chub translocated above Chute Falls experience fast growth and high survival and that Chute Falls translocations are a beneficial (but limited) management tool for increasing abundance of humpback chub that can be used instead of nonnative removals in certain situations (Yackulic and others, 2021; Figure 12). In recent years (i.e., 2019, 2022) biologists have documented numerous unmarked subadult humpback chub, indicating that humpback chub may be spawning and recruiting above Chute Falls, potentially increasing the benefit of this management action.

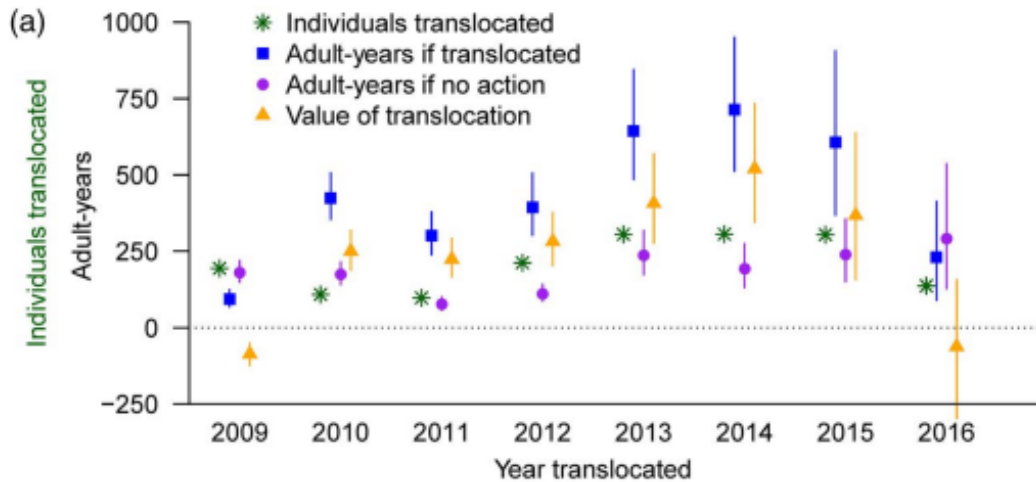


Figure 12. Population consequences of humpback chub (*Gila cypha*) translocations from individual years. The value of translocations (orange triangles) was calculated as the difference between the expected number of adult-years after translocation (blue squares) and the expected number of adult-years if translocation had not occurred (purple circles) and varies in part based on the number of chub translocated (green asterisks). Figure from Yackulic and others (2021).

In conjunction with translocation activities of humpback chub upstream of Chute Falls, this project element supports work with the USFWS Southwest Native Aquatic Resources and Recovery Center at Dexter, NM to maintain a long-term genetic refuge of humpback chub.

Additionally, humpback chub collected from the LCR by this effort have been used by NPS to provide juvenile humpback chub for translocation activities into Shinumo Creek and Havasu Creek.

Funding is used to support one additional camp (3-4 people) during the May USFWS sampling trip for monitoring work, and additional people and helicopter time during the October trip to catch and translocate fish from the lower LCR to above Chute Falls.

Hypotheses and Science Questions

- Do translocations of humpback chub above Chute Falls benefit the lower LCR population?
- How many humpback chub are above Chute Falls?
- Is there evidence of reproduction above Chute Falls?

Methods

Detailed methods can be obtained in Yackulic and others, 2021.

Anticipated Use of Data

Data would be used to assess the effectiveness of translocations above Chute Falls as a conservation tool for humpback chub, both in terms of adult abundances and potential for reproduction.

Outcomes and Products

Results would be presented at the GCDAMP annual reporting meeting.

Project Element G.8. Sampling of Springs in the Upper LCR (New Study; Unfunded)

David Ward

U.S. Fish and Wildlife Service

Sampling of the Upper LCR (~32 River Kilometers from Confluence) for Humpback Chub

In the past, fish monitoring of the LCR has focused primarily on areas where humpback chub are known to occur (i.e., downstream of Blue Springs, located 21 river kilometers (rkm) above the confluence with the Colorado River), with upstream sections less frequently monitored. Above Blue Springs, flow in the LCR is intermittent until reaching the White Mountains, near the headwaters of the LCR. When the LCR is not flooding, this intermittent section of the LCR is comprised of numerous springs interspersed through patches of dry riverbed, and when floods occur, the increased flow coming down the LCR corridor connects this intermittent segment to the rest of the LCR. This intermittent habitat has been recognized as a potential source of nonnative species getting washed downstream from the upper LCR, but it has not been considered to be suitable habitat for native fishes moving upstream from Blue Springs. Accordingly, biologists from USFWS were surprised to capture both juvenile and adult humpback chub in this stretch of river (~rkm 32.5) as part of their nonnative fish surveys in June 2023 (funded in Project I in the FY 2021-23 TWP).

The LCR experienced large floods during the winter of 2022-2023, and it is unknown whether humpback chub found in the upper LCR in June 2023 were recent migrants from the lower reaches (i.e., swam up from the lower LCR during the 2023 flood), or whether humpback chub have persisted in these spring habitats for numerous years. Furthermore, if humpback chub are recent transplants, it is unknown how long can they persist/reproduce in these patchy spring habitats or whether these fish are likely to die shortly after upstream movement and subsequent isolation. This project aims to understand habitat use and movement dynamics of humpback chub in the upper LCR, and will help evaluate 1) to what extent (if any) LCR population estimates of humpback chub are biased due to restricted spatial sampling (i.e., lower LCR only), 2) whether movement into the upper LCR acts as a source of mortality for adult HBC in flood years (i.e., adults swim up, get stranded in springs, and eventually die), and 3) whether upstream

springs can act as refugia for humpback chub if the mainstem Colorado River is invaded by nonnative warm-water piscivores (e.g., smallmouth bass, walleye, green sunfish) that may threaten recovery. Additionally, fish surveys of the upper LCR would continue to help biologists evaluate to what extent these springs act as sources of nonnative fishes that can seed the lower LCR during floods.

Hypotheses and Science Questions

- Are humpback chub rare or common in the upper LCR?
- Is the upper LCR suitable habitat for humpback chub translocations?
- Is the upper LCR a source or sink of humpback chub?

Methods

Conduct two 5-day trips to the upper LCR in FY 2025 (1 in spring, 1 in fall) to evaluate the spatial distribution of humpback chub and other fishes in the intermittent flow habitats of the upper LCR. Collect eDNA samples to supplement physical capture data.

Anticipated Use of Data

Data would be used to evaluate the spatial extent and relative abundances of humpback chub above Blue Springs, and possibly to determine the source of humpback chub that inhabit this area (i.e., movement from below Blue Springs or local reproduction). Data could be used to assess habitat quality to determine whether the site is suitable for translocations.

Outcomes and Products

Results would be presented at the GCDAMP annual reporting meeting.

Project Element G.9. Movement in Western Grand Canyon from System-Wide Antenna Monitoring (New Study; Partially Funded)

Eric Frye, Maria Dzul

U.S. Geological Survey, Southwest Biological Science Center, Grand Canyon Monitoring and Research Center

Increase Use of Submersible Antennas throughout Grand Canyon to Detect Fishes in Less-Frequently Sampled River Reaches

Because the Colorado River in Grand Canyon cannot be sampled in its entirety, biologists must frequently make assumptions about fish movement in order to fit population models to estimate abundance and survival.

Often, movement out of a study site (i.e., permanent emigration) is confounded with mortality so that biologists can only obtain estimates of ‘apparent’ survival (i.e., probability a fish survives and remains in the study site).

Apparent survival estimates of fishes in JCM-west are relatively low for adult humpback chub, suggesting either increased emigration or increased mortality of adult humpback chub in western Grand Canyon compared to near the LCR. Estimates of apparent survival can impair population inference, because movement and mortality have different implications for population dynamics. By searching for marked fishes in river reaches that are less frequently visited by fish monitoring trips, biologists would learn more about fish movement patterns and this updated information could be used to improve population models.

Due to cost increases and funding limitations, this project element is currently proposed to be funded only in FY 2025.

Hypotheses and Science Questions

- Can citizen science be used as a tool for improving detection of fishes marked with PIT tags?
- What are the movement patterns of humpback chub in western Grand Canyon, and how does humpback chub movement affect survival estimation?

Methods

Submersible antennas that could be deployed on numerous scientific trips (e.g., Arizona Game and Fish system-wide fish monitoring trips that randomly sample reaches) and also deployed by river boatmen as part of a citizen science project to help collect data about movement of fishes. These antennas would be baited, deployed overnight throughout a section of river, and collected the following morning. Citizen scientists would be paid for data collection (fee TBD, ~\$40).

Anticipated Use of Data

Detection data will be used to evaluate patterns in movement of fishes in Grand Canyon, and ideally to help improve survival estimation for humpback chub.

Outcomes and Products

Results from this pilot study will be presented at the annual reporting meeting in January 2026 or 2027.

Budgets

Project Element Budget

| Fiscal Year 2025 | | | | | | | | | |
|--|------------------|----------------------|-----------------------|-----------------------|---------------------------|--------------------------|-----------------------------|--------------------|------------------------------|
| Project G Humpback Chub Population Dynamics throughout the Colorado River Ecosystem | Salaries | Travel & Training | Operating Expenses | Logistics Expenses | Cooperative Agreements | To other USGS Centers | AMP Special Burden Rate* | Total | USGS Contributing Funds** |
| | | | | | | | 21.86% | | Est Full Rate = 55% |
| G.1. Humpback chub population modeling | \$145,751 | \$8,000 | \$12,000 | \$0 | \$0 | \$0 | \$36,238 | \$201,989 | |
| G.2. Annual spring/fall HBC abundance estimates in the lower 13.6 km of the LCR | \$4,719 | \$0 | \$20,222 | \$102,192 | \$406,258 | \$0 | \$39,983 | \$573,373 | |
| G.3. Juvenile chub monitoring near the LCR confluence (JCM-East) | \$150,152 | \$2,000 | \$31,156 | \$301,244 | \$0 | \$0 | \$105,938 | \$590,489 | |
| G.4. Remote PIT-tag array monitoring in the LCR | \$22,136 | \$0 | \$5,000 | \$3,000 | \$0 | \$0 | \$6,589 | \$36,724 | |
| G.5. Monitoring humpback chub aggregation relative abundance and distribution | \$3,938 | \$0 | \$12,436 | \$79,366 | \$142,984 | \$0 | \$25,221 | \$263,945 | |
| G.9. Movement in western Grand Canyon from system-wide antenna monitoring | \$10,742 | \$0 | \$12,000 | \$0 | \$0 | \$0 | \$4,972 | \$27,713 | |
| Total Project G | \$337,437 | \$10,000 | \$92,814 | \$485,802 | \$549,242 | \$0 | \$218,940 | \$1,694,235 | \$96,708 |

| Fiscal Year 2026 | | | | | | | | | |
|--|------------------|----------------------|-----------------------|-----------------------|---------------------------|--------------------------|-----------------------------|--------------------|------------------------------|
| Project G Humpback Chub Population Dynamics throughout the Colorado River Ecosystem | Salaries | Travel & Training | Operating Expenses | Logistics Expenses | Cooperative Agreements | To other USGS Centers | AMP Special Burden Rate* | Total | USGS Contributing Funds** |
| | | | | | | | 22.60% | | Est Full Rate = 55.5% |
| G.1. Humpback chub population modeling | \$185,221 | \$8,000 | \$5,000 | \$0 | \$0 | \$0 | \$44,798 | \$243,019 | |
| G.2. Annual spring/fall HBC abundance estimates in the lower 13.6 km of the LCR | \$2,905 | \$0 | \$20,222 | \$105,220 | \$412,700 | \$0 | \$41,388 | \$582,435 | |
| G.3. Juvenile chub monitoring near the LCR confluence (JCM-East) | \$152,119 | \$2,000 | \$29,656 | \$314,734 | \$0 | \$0 | \$112,663 | \$611,172 | |
| G.4. Remote PIT-tag array monitoring in the LCR | \$23,242 | \$0 | \$3,000 | \$3,000 | \$0 | \$0 | \$6,609 | \$35,851 | |
| G.5. Monitoring humpback chub aggregation relative abundance and distribution | \$2,905 | \$0 | \$12,436 | \$84,639 | \$145,034 | \$0 | \$26,947 | \$271,960 | |
| G.9. Movement in western Grand Canyon from system-wide antenna monitoring | \$0 | \$0 | \$0 | \$0 | \$0 | \$0 | \$0 | \$0 | |
| Total Project G | \$366,393 | \$10,000 | \$70,314 | \$507,593 | \$557,734 | \$0 | \$232,404 | \$1,744,438 | \$101,126 |

| Fiscal Year 2027 | | | | | | | | | |
|--|------------------|----------------------|-----------------------|-----------------------|---------------------------|--------------------------|-----------------------------|--------------------|------------------------------|
| Project G Humpback Chub Population Dynamics throughout the Colorado River Ecosystem | Salaries | Travel & Training | Operating Expenses | Logistics Expenses | Cooperative Agreements | To other USGS Centers | AMP Special Burden Rate* | Total | USGS Contributing Funds** |
| | | | | | | | 23.40% | | Est Full Rate = 56% |
| G.1. Humpback chub population modeling | \$194,482 | \$8,000 | \$5,000 | \$0 | \$0 | \$0 | \$48,551 | \$256,033 | |
| G.2. Annual spring/fall HBC abundance estimates in the lower 13.6 km of the LCR | \$3,051 | \$0 | \$20,222 | \$108,272 | \$422,944 | \$0 | \$43,470 | \$597,958 | |
| G.3. Juvenile chub monitoring near the LCR confluence (JCM-East) | \$153,851 | \$2,000 | \$30,156 | \$324,740 | \$0 | \$0 | \$119,515 | \$630,262 | |
| G.4. Remote PIT-tag array monitoring in the LCR | \$24,405 | \$0 | \$3,000 | \$3,000 | \$0 | \$0 | \$7,115 | \$37,519 | |
| G.5. Monitoring humpback chub aggregation relative abundance and distribution | \$3,051 | \$0 | \$12,436 | \$66,912 | \$147,083 | \$0 | \$23,694 | \$253,176 | |
| G.9. Movement in western Grand Canyon from system-wide antenna monitoring | \$0 | \$0 | \$0 | \$0 | \$0 | \$0 | \$0 | \$0 | |
| Total Project G | \$378,839 | \$10,000 | \$70,814 | \$502,924 | \$570,027 | \$0 | \$242,344 | \$1,774,948 | \$103,611 |

*GCDAMP Special Rate: Rate includes current DOI preferred rate (currently 15%; subject to change) and facilities rate (will vary annually). No USGS bureau overhead is charged (unique to the GCDAMP agreement).

**USGS Contributing Funds: The amount of funds required to cover the subsidy created by the reduced burden rate (i.e., GCDAMP Special Rate). As in previous years, SBSC/GCMRC will request these funds of the USGS cost-share program. These funds are not guaranteed.

Unfunded Project Element Budget

| Fiscal Year 2025 | | | | | | | | |
|---|-----------------|-------------------|--------------------|--------------------|------------------------|-----------------------|--------------------------|------------------|
| Project G Unfunded Project Elements | Salaries | Travel & Training | Operating Expenses | Logistics Expenses | Cooperative Agreements | To other USGS Centers | AMP Special Burden Rate* | Total |
| | | | | | | | 21.86% | |
| G.6. Juvenile chub monitoring - Western Grand Canyon (JCM-West) | \$47,674 | \$1,000 | \$29,836 | \$143,848 | \$0 | \$0 | \$48,474 | \$270,832 |
| G.7. Chute Falls translocations | \$0 | \$0 | \$930 | \$15,877 | \$82,649 | \$0 | \$6,143 | \$105,599 |
| G.8. Sampling of springs in the upper LCR | \$0 | \$0 | \$620 | \$20,936 | \$54,240 | \$0 | \$6,326 | \$82,122 |
| Total Project G | \$47,674 | \$1,000 | \$31,386 | \$180,661 | \$136,889 | \$0 | \$60,943 | \$458,553 |

| Fiscal Year 2026 | | | | | | | | |
|---|-----------------|-------------------|--------------------|--------------------|------------------------|-----------------------|--------------------------|------------------|
| Project G Unfunded Project Elements | Salaries | Travel & Training | Operating Expenses | Logistics Expenses | Cooperative Agreements | To other USGS Centers | AMP Special Burden Rate* | Total |
| | | | | | | | 22.60% | |
| G.6. Juvenile chub monitoring - Western Grand Canyon (JCM-West) | \$51,224 | \$1,000 | \$29,836 | \$157,187 | \$0 | \$0 | \$54,070 | \$293,317 |
| G.7. Chute Falls translocations | \$0 | \$0 | \$930 | \$16,347 | \$83,697 | \$0 | \$6,416 | \$107,390 |
| G.8. Sampling of springs in the upper LCR | \$11,712 | \$0 | \$15,000 | \$0 | \$0 | \$0 | \$6,037 | \$32,749 |
| Total Project G | \$62,936 | \$1,000 | \$45,766 | \$173,534 | \$83,697 | \$0 | \$66,523 | \$433,456 |

| Fiscal Year 2027 | | | | | | | | |
|---|-----------------|-------------------|--------------------|--------------------|------------------------|-----------------------|--------------------------|------------------|
| Project G Unfunded Project Elements | Salaries | Travel & Training | Operating Expenses | Logistics Expenses | Cooperative Agreements | To other USGS Centers | AMP Special Burden Rate* | Total |
| | | | | | | | 23.40% | |
| G.6. Juvenile chub monitoring - Western Grand Canyon (JCM-West) | \$54,981 | \$1,000 | \$29,836 | \$162,480 | \$0 | \$0 | \$58,101 | \$306,398 |
| G.7. Chute Falls translocations | \$0 | \$0 | \$930 | \$16,820 | \$84,764 | \$0 | \$6,696 | \$109,210 |
| G.8. Sampling of springs in the upper LCR | \$12,532 | \$0 | \$15,000 | \$0 | \$0 | \$0 | \$6,443 | \$33,975 |
| Total Project G | \$67,513 | \$1,000 | \$45,766 | \$179,300 | \$84,764 | \$0 | \$71,240 | \$449,583 |

References Cited

- Coggins, L.G., Jr., 2008, Active adaptive management for native fish conservation in the Grand Canyon—Implementation and evaluation: Gainesville, University of Florida, Ph.D. dissertation, 173 p., <https://ufdc.ufl.edu/UFE0021996/00001>.
- Coggins, L.G., Jr., Pine, W.E., III, Walters, C.J., Van Haverbeke, D.R., Ward, D., and Johnstone, H.C., 2006, Abundance trends and status of the Little Colorado River population of humpback chub: North American Journal of Fisheries Management, v. 26, no. 1, p. 233-245, <https://doi.org/10.1577/M05-075.1>.

- Coggins, L.G., Jr., Yard, M.D., and Pine, W.E., 2011, Nonnative fish control in the Colorado River in Grand Canyon, Arizona—An effective program or serendipitous timing?: Transactions of the American Fisheries Society, v. 140, no. 2, p. 456-470, <https://doi.org/10.1080/00028487.2011.572009>.
- Dibble, K.L., Yackulic, C.B., Kennedy, T.A., Bestgen, K.R., and Schmidt, J.C., 2021, Water storage decisions will determine the distribution and persistence of imperiled river fishes: Ecological Applications, v. 31, no. 2, e02279, p. 1-9, <https://doi.org/10.1002/eap.2279>.
- Dzul, M.C., 2021, Life history and population dynamics of humpback chub in the Grand Canyon: Fort Collins, Colorado State University, Ph.D. dissertation, 253 p., https://mountainscholar.org/bitstream/handle/10217/233838/Dzul_colostate_0053A_16731.pdf?sequence=1&isAllowed=y.
- Dzul, M.C., Kendall, W.L., Yackulic, C.B., Van Haverbeke, D.R., Mackinnon, P., Young, K., Pillow, M.J., and Thomas, J.R., 2024, Estimating migration timing and abundance in partial migratory systems by integrating continuous antenna detections with physical captures: Journal of Animal Ecology, v. 93, no. 7, p. 796-811, <https://doi.org/10.1111/1365-2656.14076>.
- Dzul, M.C., Kendall, W.L., Yackulic, C.B., Winkelman, D.L., Van Haverbeke, D.R., and Yard, M.D., 2021a, Partial migration and spawning movements of humpback chub in the Little Colorado River are better understood using data from autonomous PIT tag antennas: Canadian Journal of Fisheries and Aquatic Sciences, v. 78, no. 8, p. 1057-1072, <https://doi.org/10.1139/cjfas-2020-0291>.
- Dzul, M.C., Yackulic, C.B., Giardina, M., Van Haverbeke, D.R., and Yard, M., 2023, Vital rates of a burgeoning population of humpback chub in western Grand Canyon: Transactions of the American Fisheries Society, v. 153, no. 4, p. 443-459, <https://doi.org/10.1002/tafs.10415>.
- Dzul, M.C., Yackulic, C.B., Kendall, W.L., Winkelman, D.L., Conner, M.M., and Yard, M.D., 2021b, Incorporating antenna detections into abundance estimates of fish: Canadian Journal of Fisheries and Aquatic Sciences, v. 79, no. 3, p. 436–447, <https://doi.org/10.1139/cjfas-2021-0003>.
- Dzul, M.C., Yackulic, C.B., and Korman, J., 2018, Estimating disperser abundance using open population models that incorporate data from continuous detection PIT arrays: Canadian Journal of Fisheries and Aquatic Sciences, v. 75, no. 9, p. 1393-1404, <https://doi.org/10.1139/cjfas-2017-0304>.
- Dzul, M.C., Yackulic, C.B., Korman, J., Yard, M.D., and Muehlbauer, J.D., 2016, Incorporating temporal heterogeneity in environmental conditions into a somatic growth model: Canadian Journal of Fisheries and Aquatic Sciences, v. 74, no. 3, p. 316-326, <https://doi.org/10.1139/cjfas-2016-0056>.
- Gorman, O.T., and VanHoosen, R.R., 2000, Experimental growth of four native Colorado River fishes at temperatures of 12, 18, and 24°C—draft final report: Ashland, Wisc., and Willow Beach, Ariz., U.S. Geological Survey, Great Lakes Science Center and U.S. Fish and Wildlife Service, Willow Beach National Fish Hatchery, submitted to U.S. Geological

- Survey, Grand Canyon Monitoring and Research Center, 35 p., <https://www.researchgate.net/publication/260037968> Experimental growth of four native Colorado River fishes at temperatures of 12 18 and 24oC.
- Hamman, R.L., 1982, Spawning and culture of humpback chub: The Progressive Fish-Culturist, v. 44, no. 4, p. 213-216, [https://doi.org/10.1577/1548-8659\(1982\)44\[213:SACOHC\]2.0.CO;2](https://doi.org/10.1577/1548-8659(1982)44[213:SACOHC]2.0.CO;2).
- Kaeding, L.R., Burdick, B.D., Schrader, P.A., and McAda, C.W., 1990, Temporal and spatial relations between the spawning of humpback chub and roundtail chub in the upper Colorado River: Transactions of the American Fisheries Society, v. 119, no. 1, p. 135-144, [https://doi.org/10.1577/1548-8659\(1990\)119<0135:TASRBT>2.3.CO;2](https://doi.org/10.1577/1548-8659(1990)119<0135:TASRBT>2.3.CO;2).
- Kaeding, L.R., and Zimmerman, M.A., 1983, Life history and ecology of the humpback chub in the Little Colorado and Colorado Rivers of the Grand Canyon: Transactions of the American Fisheries Society, v. 112, no. 5, p. 577-594, [https://doi.org/10.1577/1548-8659\(1983\)112<577:LHAEOT>2.0.CO;2](https://doi.org/10.1577/1548-8659(1983)112<577:LHAEOT>2.0.CO;2).
- Korman, J., Yard, M.D., and Kennedy, T.A., 2017, Trends in rainbow trout recruitment, abundance, survival, and growth during a boom-and-bust cycle in a tailwater fishery: Transactions of the American Fisheries Society, v. 146, no. 5, p. 1043-1057, <https://doi.org/10.1080/00028487.2017.1317663>.
- Lee, W.-S., Monaghan, P., and Metcalfe, N.B., 2013, Experimental demonstration of the growth rate–lifespan trade-off: Proceedings of the Royal Society B—Biological Sciences, v. 280, no. 1752, 20122370, <https://doi.org/10.1098/rspb.2012.2370>.
- Lessard, J.L., and Hayes, D.B., 2003, Effects of elevated water temperature on fish and macroinvertebrate communities below small dams: River Research and Applications, v. 19, no. 7, p. 721-732, <https://doi.org/10.1002/rra.713>.
- Persons, W.R., Ward, D.L., and Avery, L.A., 2015, Standardized methods for Grand Canyon fisheries research 2015 (ver. 1.1, January 2015): U.S. Geological Survey, Techniques and Methods, book 2, chapter A12, 19 p., <https://www.usgs.gov/publications/standardized-methods-grand-canyon-fisheries-research-2015>.
- Persons, W.R., Van Haverbeke, D.R., and Dodrill, M.J., 2017, Colorado River fish monitoring in Grand Canyon, Arizona—2002–14 humpback chub aggregations: U.S. Geological Survey Open-File Report 2016-1177, 43 p., <https://doi.org/10.3133/ofr20161177>.
- Robinson, A.T., and Childs, M.R., 2001, Juvenile growth of native fishes in the Little Colorado River and in a thermally modified portion of the Colorado River: North American Journal of Fisheries Management, v. 21, no. 4, p. 809-815, [https://doi.org/10.1577/1548-8675\(2001\)021<0809:JGONFI>2.0.CO;2](https://doi.org/10.1577/1548-8675(2001)021<0809:JGONFI>2.0.CO;2).
- Rogowski, D.L., Osterhoudt, R.J., Mohn, H.E., and Boyer, J.K., 2018, Humpback chub (*Gila cypha*) range expansion in the western Grand Canyon: Western North American Naturalist, v. 78, no. 1, article 4, <https://scholarsarchive.byu.edu/wnan/vol78/iss1/4>.

- Stone, D.M., Young, K.L., Mattes, W.P., and Cantrell, M.A., 2018, Abiotic controls of invasive nonnative fishes in the Little Colorado River, Arizona: *The American Midland Naturalist*, v. 180, no. 1, p. 119-142, <https://doi.org/10.1674/0003-0031-180.1.119>.
- Taylor, B.M., Choat, J.H., DeMartini, E.E., Hoey, A.S., Marshall, A., Priest, M.A., Rhodes, K.L., and Meekan, M.G., 2019, Demographic plasticity facilitates ecological and economic resilience in a commercially important reef fish: *Journal of Animal Ecology*, v. 88, no. 12, p. 1888-1900, <https://doi.org/10.1111/1365-2656.13095>.
- Udall, B., and Overpeck, J., 2017, The twenty-first century Colorado River hot drought and implications for the future: *Water Resources Research*, v. 53, no. 3, p. 2404-2418, <https://doi.org/10.1002/2016WR019638>.
- U.S. Department of Interior, 2016, Record of Decision for the Glen Canyon Dam Long-Term Experimental and Management Plan final Environmental Impact Statement (LTEMP ROD): Salt Lake City, Utah, U.S. Department of the Interior, Bureau of Reclamation, Upper Colorado Region, National Park Service, Intermountain Region, 196 p., http://ltempeis.anl.gov/documents/docs/LTEMP_ROD.pdf.
- U.S. Fish and Wildlife Service, 2002, Humpback chub (*Gila cypha*) recovery goals—Amendment and supplement to the humpback chub recovery plan: Denver, Colo., U.S. Fish and Wildlife Service, Mountain Prairie Region, 71 p., plus appendices.
- U.S. Fish and Wildlife Service, 2016, Biological Opinion for the Glen Canyon Dam Long-term Experimental Plan, Coconino County, Arizona—Attachment E, *in* Record of Decision final Environmental Impact Statement: Salt Lake City, Utah, U.S. Department of the Interior, Bureau of Reclamation, Upper Colorado Region, National Park Service, Intermountain Region, p. E-1-E-94, http://ltempeis.anl.gov/documents/docs/LTEMP_ROD.pdf.
- U.S. Fish and Wildlife Service, 2024, Biological Opinion for the Near-term Colorado River Operations Final Supplemental Environmental Impact Statement—Glen Canyon Dam Operations: Phoenix, Ariz., U.S. Fish and Wildlife Service, 68 p.
- Valdez, R.A., and Ryel, R.J., 1995, Characterization of the life history and ecology of the humpback chub (*Gila cypha*) in the Colorado River, Grand Canyon, Arizona—final report: Salt Lake City, Utah, BIO/WEST, Inc., submitted to Bureau of Reclamation, contract no. 0-CS-40-09110, technical report no. TR-250-08, 286 p.
- Valdez, R.A., and Ryel, R.J., 1997, Life history and ecology of the humpback chub in the Colorado River, Grand Canyon, Arizona, *in* van Riper, C., III, and Deshler, E.T., eds., Third Biennial Conference on the Colorado Plateau, Flagstaff, Ariz., October 17-20, 1995: National Park Service Transactions and Proceedings Series NPS/NRNAU/NRTP-97/12, 3-31 p.
- Van Haverbeke, D.R., Stone, D.M., Coggins, L.G., and Pillow, M.J., 2013, Long-term monitoring of an endangered desert fish and factors influencing population dynamics: *Journal of Fish and Wildlife Management*, v. 4, no. 1, p. 163-177, <https://doi.org/10.3996/082012-JFWM-071>.
- Van Haverbeke, D.R., Stone, D.M., Dodrill, M.J., Young, K.L., and Pillow, M.J., 2017, Population expansion of humpback chub in western Grand Canyon and hypothesized

- mechanisms: *The Southwestern Naturalist*, v. 62, no. 4, p. 285-292, <https://doi.org/10.1894/0038-4909-62.4.285>.
- Wright, S.A., Anderson, C.R., and Voichick, N., 2009, A simplified water temperature model for the Colorado River below Glen Canyon Dam: *River Research and Applications*, v. 25, no. 6, p. 675-686, <https://doi.org/10.1002/rra.1179>.
- Yackulic, C.B., Dodrill, M.J., Dzul, M.C., Sanderlin, J., and Reid, J., 2020, A need for speed in Bayesian population models—A practical guide to marginalizing and recovering discrete latent states: *Ecological Applications*, v. 30, no. 5, e02112, p. 1-19, <https://doi.org/10.1002/eap.2112>.
- Yackulic, C.B., Korman, J., Yard, M.D., and Dzul, M.C., 2018, Inferring species interactions through joint mark-recapture analysis: *Ecology*, v. 99, no. 4, p. 812-821, <https://doi.org/10.1002/ecy.2166>.
- Yackulic, C.B., Van Haverbeke, D.R., Dzul, M.C., Bair, L.S., and Young, K.L., 2021, Assessing the population impacts and cost-effectiveness of a conservation translocation: *Journal of Applied Ecology*, v. 58, no. 8, p. 1602-1612, <https://doi.org/10.1111/1365-2664.13908>.
- Yackulic, C.B., Yard, M.D., Korman, J., and Van Haverbeke, D.R., 2014, A quantitative life history of endangered humpback chub that spawn in the Little Colorado River—Variation in movement, growth, and survival: *Ecology and Evolution*, v. 4, no. 7, p. 1006-1018, <https://doi.org/10.1002/ece3.990>.

Project H: Salmonid (Trout) Research and Monitoring Project

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Project Summary and Purpose

The LTEMP provides the necessary long-term framework for assessing specific operations at Glen Canyon Dam, including experimental flows as well as other types of management actions conceived during and implemented over the next 20-year period. The LTEMP includes a goal for the rainbow trout (*Oncorhynchus mykiss*) fishery in Glen Canyon: “Achieve a healthy high-quality recreational rainbow trout fishery in GCNRA and reduce or eliminate downstream trout migration consistent with NPS fish management and ESA compliance.” For these reasons, the Salmonid Research and Monitoring Project, which includes research and monitoring of trout, was developed having the long view, to inform stakeholders and management responses to unanticipated and emerging risks (e.g., brown trout *Salmo trutta*; Runge and others, 2018). Rapidly changing conditions related to basin-wide water management and upstream reservoir dynamics have led to occurrences of low dissolved oxygen (DO), warming temperatures, and new warm-water nonnative species expansion in Glen Canyon requiring additional quantitative monitoring sensitive to these changes (Dibble and others, 2021; Bruckerhoff and others, 2022; Eppheimer and others, 2024). Water quality and temperature changes influence populations of salmonids in Glen Canyon (Korman and others, 2021; Korman and others, 2023). Given these new developments, it is unclear whether the new expansion of warm-water nonnative species will disrupt the balance between salmonids and endangered native fishes downstream, and the rainbow trout fishery in Glen Canyon.

In general, the study design described in the previous two work plans (FY 2018-20, FY 2021-23 Triennial Work Plans [TWP]) is still relevant for addressing management questions posed in the LTEMP.

Importantly, recent changes in the water quality and temperatures discharged from Glen Canyon Dam as reservoir storage in the Colorado River basin declines may potentially impact salmonid populations (Bruckerhoff and others, 2022; Healy and others, 2023).

Operational changes designed to respond to warm-water introduced fish expansion and other ecological changes are being proposed by Reclamation (e.g., SEIS, Bureau of Reclamation, 2024). Additionally, GCDAMP stakeholders and Tribes have requested the GCMRC consider and reduce overlapping projects to minimize handling of fish out of respect for cultural values and to address fish health concerns, and to enhance monitoring and research efficiencies and optimize information gathering for decision-making. We recognize monitoring and research related to LTEMP salmonid goals needs to be precise and sensitive to anticipated short- and long-term ecosystem changes. Thus, some modifications to this project are proposed for the 2025-2027 TWP. The type of quantitative approach ongoing and proposed herein is appropriate for understanding large and complex ecosystems, particularly, when quantifying drivers of rainbow trout and brown trout population dynamics. This research project proposes to evaluate 1) the effect of ongoing ecological changes (i.e., temperature, DO, warm-water species expansion) on trout somatic growth, reproduction, recruitment, and survival, 2) effects of spring and fall high flow events (HFEs) on trout recruitment, dispersal, and growth, 3) factors controlling trout recruitment and dispersal into Marble Canyon and Little Colorado River (LCR) reaches, 4) factors controlling the quality of the trout fishery (growth, condition, sexual maturity, and angler catch rates), and 5) factors regulating brown trout and other nonnative fish population dynamics, as well as efficacy of an incentivized brown trout harvest program. Summarized below are monitoring and research elements that address the primary study objectives.

Long-Term Experimental and Management Plan (LTEMP) Resource Goals

Project H addresses the following Long-term Experimental and Management Plan (LTEMP; U.S. Department of the Interior, 2016) goal:

- Achieve a healthy high-quality recreational rainbow trout fishery in GCNRA and reduce or eliminate downstream trout migration consistent with NPS fish management and ESA compliance.

Background

The LTEMP identified potential flow experiments to improve conditions for fishes, benefit the food base, and improve sediment conditions and rebuild sandbars (U.S. Department of the Interior, 2016). Some experiments originally proposed to limit rainbow trout recruitment in Glen Canyon (trout management flows; TMFs) may not occur during the FY 2025-27 TWP; however, new flow actions are being proposed and analyzed to suppress warm-water nonnative species reproduction that may also influence population dynamics of trout and other species (Bureau of Reclamation, 2024).

It is likely that extrinsic factors such as DO and temperature variation are influencing trout growth rates and that more replication across years with and without HFEs is needed to determine if there is a measurable effect related to this type of flow and others.

Brown trout are highly piscivorous and may compete with native fish species (Yard and others, 2011; Whiting and others, 2014), leading to restricted distribution of native fishes in habitats where brown trout achieve high densities (Healy and others, 2020).

The expansion of brown trout in Glen Canyon, which was linked partially to immigration related to fall high flow experiments (HFEs) and a management action (weir installation) to suppress brown trout in Grand Canyon (Healy, and others, 2022a), poses a significant threat to the rainbow trout fishery and likely to endangered humpback chub (*Gila cypha*) populations residing downstream in and near the Little Colorado River (LCR). In the past, management actions have included labor-intensive mechanical removal of rainbow trout and brown trout at the LCR confluence and in Bright Angel Creek (Coggins and others, 2011; Healy and others, 2020). Bright Angel Creek trout suppression is ongoing, however dispersal of brown trout from the mainstem may augment the spawning population (Akland, 2023; Healy and others, 2023).

In 2008, large numbers of young-of-year (YOY) rainbow trout were produced in Glen Canyon following a spring HFE (Korman and others, 2011), some of which may have subsequently dispersed downstream (Korman and others, 2012). Further research has clarified the role of HFEs, nutrient dynamics, prey availability, and competition in driving rainbow trout population dynamics (Korman and others, 2021; Korman and others, 2023; Yard and others, 2023). In the LTEMP Record of Decision (ROD; U.S. Department of the Interior, 2016), a two-year moratorium was imposed due to the possible relationship between spring HFEs and trout recruitment; however, sediment-triggered spring HFEs are now allowed.

Flow variability can influence trout reproductive success – spring floods can decrease survival following emergence via energetic constraints (Cattanéo and others, 2002; Lobón-Cervía, 2004; Healy and others, 2023). While fall HFEs have occurred concurrently with the increase in brown trout recruitment, there has been no causal link established between the two phenomena, suggesting additional research is needed.

Brown trout routinely spawn in late fall to early winter (November-January), as follows, fall HFEs typically implemented in November may function to expurgate fine sediment accrued in spawning bars just prior to spawning. Yet, recent catch data for Glen Canyon would suggest that years with and without fall HFEs have had no measurable effect on the annual spawning success of brown trout recruitment in the following year.

With impending changes in basin-wide water availability and water management decisions, understanding how temperature and flow variability interacts with other extrinsic or intrinsic factors to influence trout population dynamics can help inform management strategies to meet LTEMP goals.

Therefore, it is essential that the influence of different types of flows in the selected ROD alternative be evaluated on behalf of both rainbow trout and brown trout in the FY 2025-27 TWP. The project elements proposed herein are driven by six overarching research questions that apply to both rainbow and brown trout.

What are the Effects of Changing Conditions Downstream of Glen Canyon Dam (i.e., Dissolved Oxygen, Water Temperature) on Trout Recruitment, Growth, and Survival?

Significant changes in water quality and temperature have occurred over the course of the last work plan (Figure 1; U.S. Geological Survey, 2024) that will likely effect population dynamics for trout and other fishes in Glen Canyon and downstream (Korman and others, 2023). Dissolved oxygen (DO) and temperature can have interacting effects on fish metabolic rates that affect survival, growth, and energy allocated to reproduction (Hanks and Secor, 2010; Waldrop and others, 2020).

Current models of DO in GCD releases model bypass DO based on observations during HFEs (Hueftle and Stevens, 2001; Vernieu, 2010). For example, bypass releases of 15,000 cfs during the 2008 high flow experiment resulted in supersaturated DO concentrations (12.6 mg/L; Vernieu, 2010) below the dam.

The degree of supersaturation under lower bypass releases (such as with cold shock flows proposed in supplemental EIS; Bureau of Reclamation, 2024) is a significant uncertainty and would help inform modeling such as that used to inform the NEPA/EIS compliance processes (Yackulic and others, 2024).

Constraints on prey production related to reservoir nutrient dynamics may also influence abundance of rainbow trout in Glen Canyon (Yard and others, 2023). Recent preliminary data collected in 2023 through project H suggests declines in trout body condition, growth rates (Figure 2), and catch rates; additional analysis is needed to understand the drivers of these trends, including interactions between reduced DO, higher water temperatures, and potential food limitations. Continued mark-recapture sampling, along with analysis of reproductive status data (see Crossman and others, 2022), is planned in the FY 2025-27 work plan for project H to further investigate questions related to the effects of changing environmental conditions on trout population dynamics. We will study associations between season-specific estimates of abundance, body condition, growth and survival rates, and spatial and temporal trends in DO, temperature and food (see project F), to understand the strength of drivers of rainbow trout population dynamics. Understanding how these variables interact to influence trout population dynamics and the health of the rainbow trout fishery will help inform decision-making for potential management actions. During the next work plan, management actions designed to influence dissolved oxygen (injection at the penstocks), temperature (use of bypass tubes to release cold water), or stocking may be considered by management agencies.

Basin-wide water management decisions, which influence reservoir levels, temperature, and fish distribution or persistence of salmonids (Dibble and others, 2021; Bruckerhoff and others, 2022; Healy and others, 2023), would also be informed by the results of Project H.

Water Quality Conditions Immediately Below Glen Canyon Dam

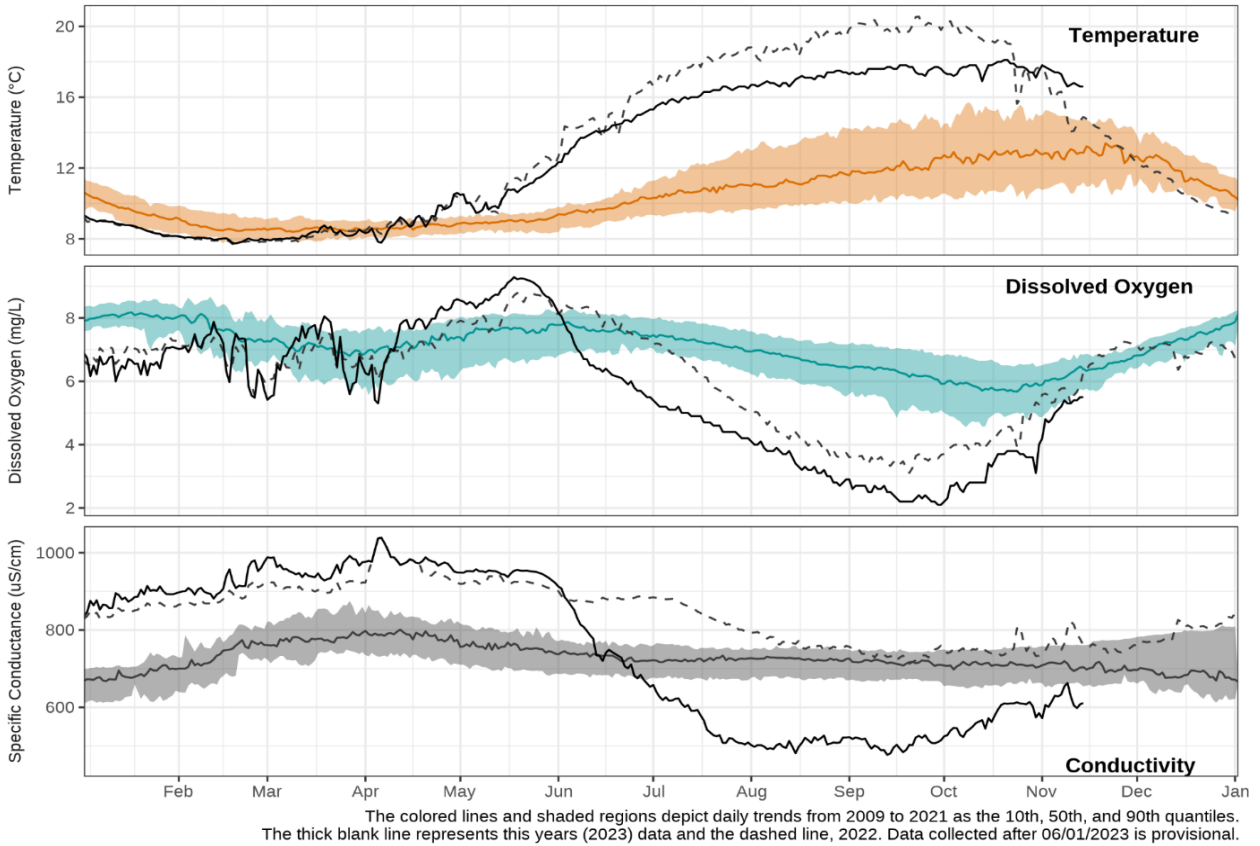


Figure 1. Water quality in Glen Canyon in 2022 and 2023 compared to long-term trends from 2009-2021 (U.S. Geological Survey, 2024).

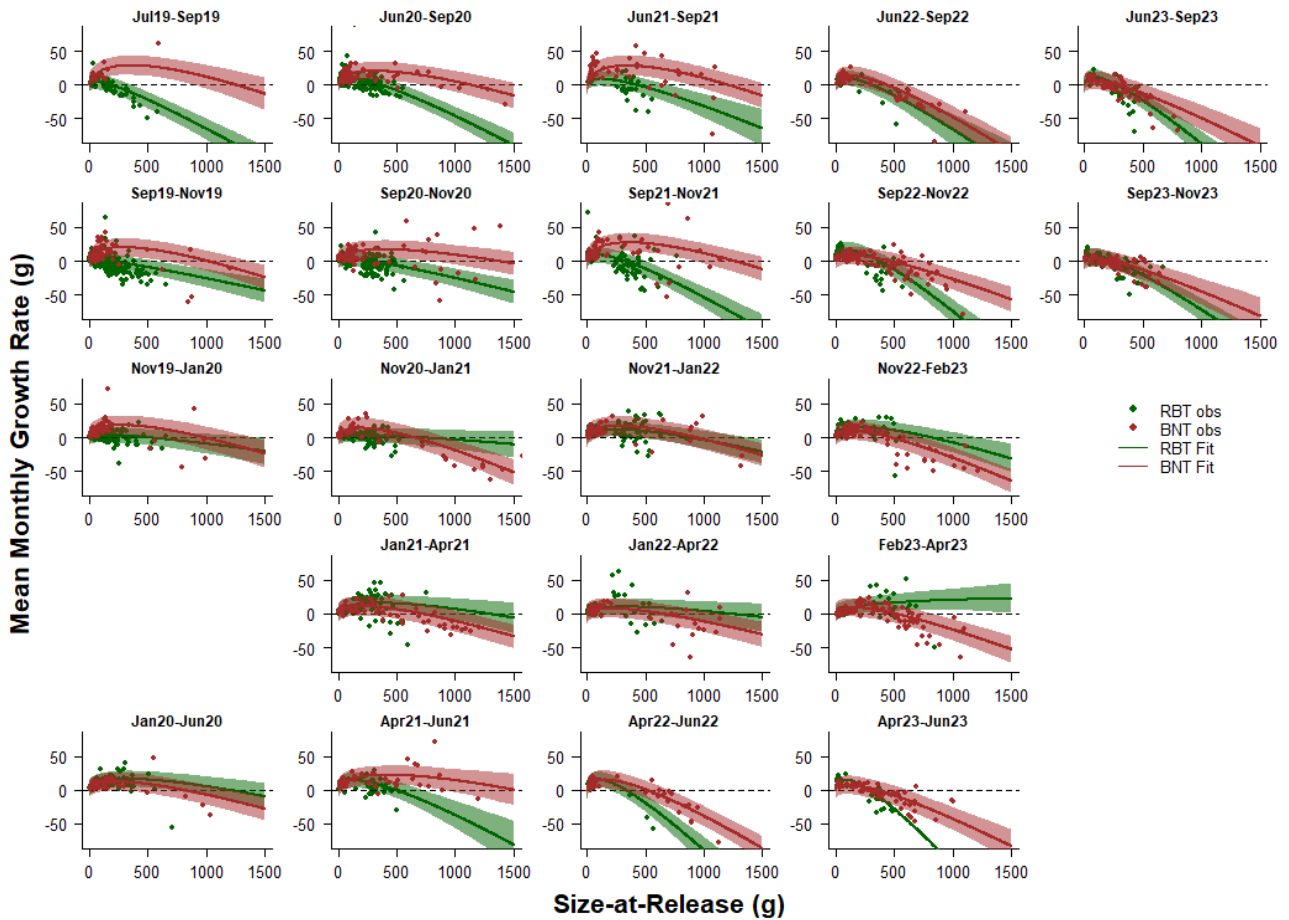


Figure 2. Growth rates in rainbow trout (*Oncorhynchus mykiss*; RBT) and brown trout (*Salmo trutta*; BNT) in Glen Canyon, based on mark-recapture data collected 2019-2023 (Korman and others, 2024).

What are the Effects of Spring and Fall HFEs, Macroinvertebrate Production Flows, and other Management Actions (Flow Management to Suppress Warm-Water Nonnative Species) on Trout Population Dynamics?

LTEMP flow experiments and non-flow fisheries management actions are occurring or planned to occur in Glen Canyon to improve LTEMP resource conditions (U.S. Department of the Interior, 2016), which may influence rainbow trout population dynamics. HFEs in spring or fall, or temperature management strategies using Glen Canyon Dam bypass tubes to limit smallmouth bass (*Micropterus dolomieu*) reproduction, and potentially minimize entrainment through Glen Canyon Dam (Bureau of Reclamation, 2024; also see Project I), among other strategies, may be implemented during the FY 2025-27 work plan.

An increase in rainbow trout growth rates and recruitment in Glen Canyon was observed following the 2008 spring HFE (Korman and others, 2011), but long-term mark-recapture monitoring and analysis of TRD data since 2008 have provided new insights into the role of dam discharge, competition, prey availability, and nutrient dynamics in driving Glen Canyon trout population dynamics, (Korman and others, 2021, 2023). Thus, it is uncertain whether future spring HFEs will produce a similar response because antecedent conditions at the time of the spring HFE may have been unique (e.g., low trout abundance, higher nutrient availability due to large inflows to Lake Powell). Temperature may have an overriding influence on trout growth and survival, but the population-level response in rainbow and brown trout may depend on the temperature management strategy implemented. Therefore, if HFEs or other flow and temperature management actions are implemented, ecosystem responses will be evaluated using the proposed Trout Recruitment and Growth Dynamics (TRGD) sampling program, as described below.

There have been several hypotheses proposed for the expansion of brown trout in Lees Ferry (Runge and others, 2018), one of which is that fall HFEs cleanse spawning gravels immediately prior to brown trout spawning thereby improving egg survival and recruitment. Conversely, high flows during the emergence of brown trout fry may also limit recruitment (Healy and others, 2023). Food availability (e.g., rainbow trout fry) or redd superimposition by rainbow trout, temperature (Healy and others, 2023), or immigration from downstream reaches (Healy and others, 2022a) could also influence brown trout population dynamics. Therefore, additional monitoring and modeling of trout growth and other extrinsic factors may resolve uncertainties about the effects between species and on brown trout population dynamics. Further, evidence indicates fall HFEs reduce adult fish growth during fall and winter, which in turn reduces their rate of sexual maturation and fecundity (Korman and others, 2017). Data analysis is proposed to reassess the hypotheses in Runge and others (2018), and potentially others (e.g., larger rainbow trout competing with smaller brown trout; Blanchet and others, 2007), and examine how adult growth following a fall HFE influences sexual maturation, fecundity, and, by extension, recruitment the following year (Project Element H.3).

What Controls Rainbow and Brown Trout Dispersal from Glen Canyon Downstream into Marble and Grand Canyons?

The LTEMP model used to simulate rainbow trout movement from GCD to the LCR assumed that trout dispersal was a constant proportion of recruitment, and that trout residency in Marble Canyon was constant through time (U.S. Department of the Interior, 2016). Immigration of trout to the LCR reach appears to be a two-step process, initially as a short-duration dispersal of YOYs from Lees Ferry to upper-middle Marble Canyon in their first summer, followed by a downstream dispersal of longer-duration from Marble Canyon to the LCR. Basically, when higher recruitment occurs in Lees Ferry it ultimately leads to higher numbers of young rainbow trout dispersing downstream into Marble Canyon (Korman and others, 2016).

Recent declines in recruitment of trout in Glen Canyon (red bars, Figure 3) and apparent declines in overall catch in 2024 suggest downstream dispersal rates and our assumptions of drivers of dispersal may change.

However, metapopulation modeling (Healy and others, 2023), passive integrated transponder (PIT) tag detections (Schelly and others, 2021), and evidence from brown trout otolith microchemistry (Akland, 2023) suggests the potential for movement of brown trout from mainstem reaches into tributaries that may also be important for native fishes and therefore negatively impact native fish populations (Healy and others, 2020).

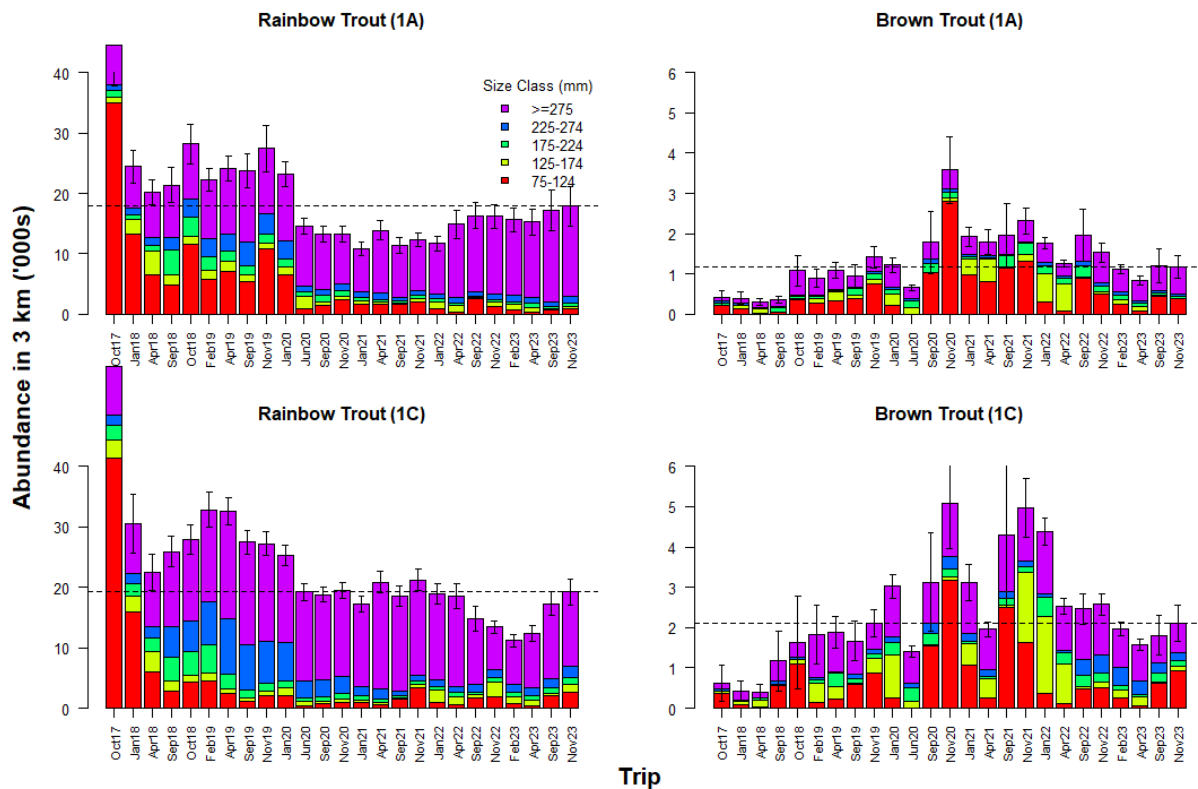


Figure 3. Trends in rainbow trout (*Oncorhynchus mykiss*; left column) and brown trout (*Salmo trutta*; right column) in TRGD sites 1A and 1C (U.S. Geological Survey, 2024).

It remains uncertain whether brown trout will exhibit dispersal downstream similar to dispersal observed for rainbow trout (Korman and others, 2012). Dispersal between reaches of river or tributaries with suitable spawning habitat may serve to maintain a metapopulation of brown trout that is resilient to suppression efforts by the National Park Service (described in Healy and others, 2020). While brown and rainbow trout numbers have been low in Grand Canyon in recent years (Fonkin and others, 2023), depending on the selected dam discharge strategy to suppress spawning of warm-water nonnative fishes such as smallmouth bass, rapid rebounds in salmonids may occur.

We will also integrate data collected for trout in tributaries and the Colorado River to understand movements of trout, including to culturally sensitive areas that are also important for native fishes (LCR inflow).

What Factors Control the Quality of the Rainbow Trout Fishery, and how does Electrofishing Capture Efficiency Influence Interpretation of Rainbow Trout Fishery Metrics?

The quality of the Lees Ferry trout fishery depends on growth rates of rainbow trout and the number of juvenile trout that recruit into the adult population, which are then targeted by anglers. Understanding factors influencing the quality of the rainbow trout fishery is an important endeavor given the economic importance of the fishery to Northern Arizona (Bair and others, 2016). Anecdotal reports from anglers in January-April 2024 suggest the fishery is in steep decline, highlighting the importance of understanding causes of the decline if management responses are to be developed.

There is seasonal variability in rainbow trout condition factor for large-sized fish (> 300 mm FL) (Korman and others, 2017). Water quality, reduced condition, and competition for prey likely led to decreased survival of larger-sized fish and ultimately reduced rainbow trout abundance in the Lees Ferry trout fishery in recent years (Korman and others, 2021; Korman and others, 2023; Yard and others, 2023). Notably, relative condition factor for brown trout has been much higher than rainbow trout (2015-2019), particularly in the fall just prior to the spawning period (November-January), which is the season that rainbow trout appear to be the lowest in condition – nonetheless, condition of both species has declined recently (Figure 4). Comparisons of length-weight relationships between trout species indicate that rainbow trout display a negative-allometric growth relationship, suggesting an overall decrease in growth or an elongation in length without a commensurate increase in weight (Yard and Korman, 2020).

Capture probability biases due to variation in size structure or environmental conditions at the time of sampling (Korman and Yard, 2017; Healy and others, 2022b) can strongly bias interpretation of trends in metrics (catch-per-unit-effort; CPUE) used to assess the quality of the fishery (Glover and others, 2019). For example, index based metrics such as CPUE rely upon the assumption that the index is consistently proportional to actual abundance of trout present in the system (Williams and others, 2002) – this assumption is unlikely to be true based on capture probability estimates for both brown and rainbow trout showing large fish are less likely to be captured than small individuals (Korman and Yard, 2017; Yackulic and others, 2020; Healy and others, 2023). In this work plan, we propose to focus more on mark-recapture based monitoring to track trends in the rainbow trout fishery, to minimize biases associated with CPUE metrics and enhance precision in trends. This will require a crosswalk between past CPUE-based monitoring and mark-recapture data. Depending on the needs and desires of managers, we could recalibrate metrics used to track trends in fishery quality by accounting for biases that could arise in abundance estimation due to individual (size) and temporal heterogeneity in capture probability.

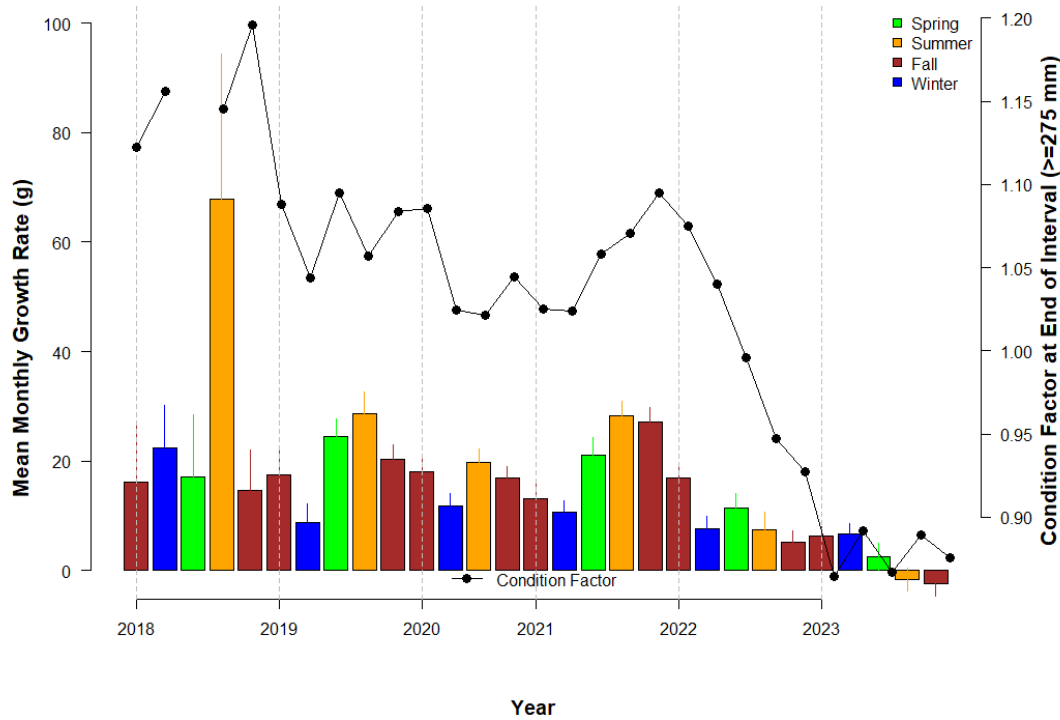
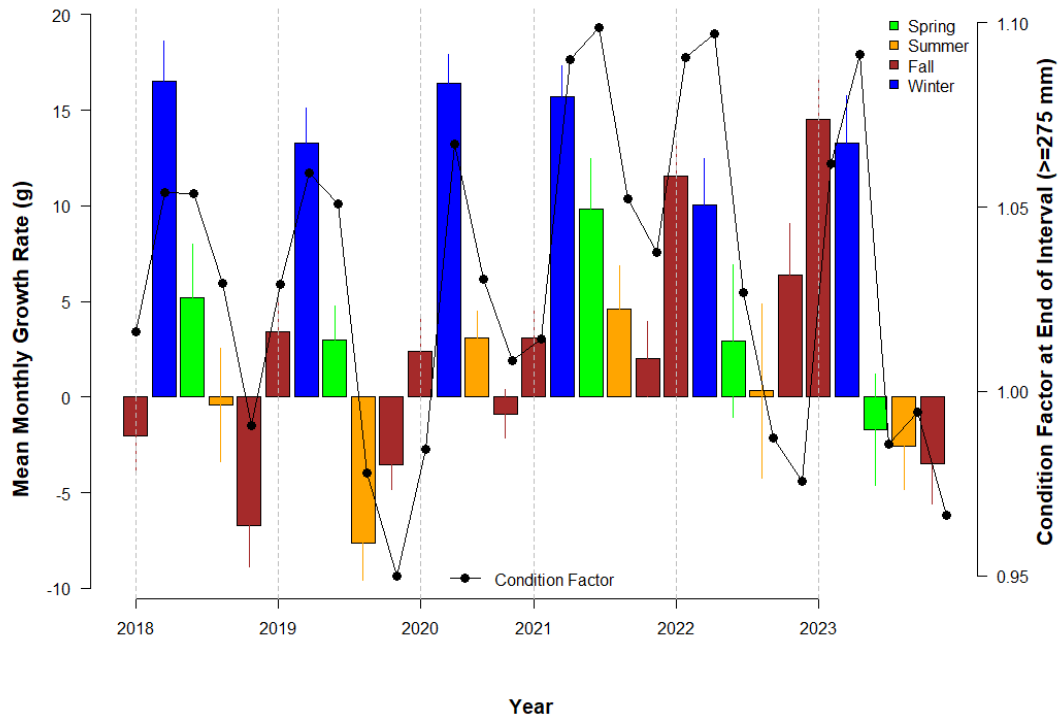


Figure 4. Condition and growth rates of rainbow trout (*Oncorhynchus mykiss*; top) and brown trout (*Salmo trutta*; bottom) in the Glen Canyon reach of the Colorado River, Arizona, USA, from mark-recapture data collected under project H (U.S. Geological Survey, 2024).

How do Expansions of Warm-Water Nonnative Fish Species Influence Trout Population Dynamics?

Several warm-water nonnative fishes have increased in number or have been newly detected downstream of Glen Canyon Dam, including green sunfish (*Lepomis cyanellus*), smallmouth bass, and walleye (*Sander vitreus*). These warm-water species are known to prey upon or potentially compete with native fishes of the Colorado River basin (Rogosch and Olden, 2020) and salmonids elsewhere (Carey and others, 2011), resulting in constrained distribution or abundance of native fishes (Hughes and Herlihy, 2012; Walsworth and Budy, 2015). Outcomes of biotic interactions with salmonids in a food-limited system such as Glen Canyon (Korman and others, 2021) are difficult to predict. We will investigate relationships between abundance or relative abundance (CPUE) of warm-water nonnative fishes and other environmental factors with demographic and somatic growth rates of trout. However, we recognize separating the effects of potential competitors or predators and changing environmental conditions (e.g., temperature) may be difficult. Understanding the relative strength of these potential drivers of trout population dynamics may be nonetheless important to decision-making related to dam operations or nonnative fish suppression efforts, if a goal to maintain a healthy rainbow trout fishery remains a priority for managers.

What Factors Regulate Brown Trout Population Dynamics in Glen Canyon, and what is the Management Efficacy of Control Actions?

As mentioned above, it is unclear what factors led to the increase of brown trout abundance in Lees Ferry (Runge and others, 2018). The efficacy of various flow and non-flow control options for controlling brown trout abundance is also uncertain. Runge and others (2018) examined the likely efficacy of several potential management interventions to reduce brown trout, followed by analyzing the effects of those interventions on other resources of concern. In addition, Healy and others, (2023) used a metapopulation viability model to further simulate the effects of management actions meant to suppress brown trout abundance while accounting for the effects of declining reservoir levels and warming temperatures. They identified some removal strategies that may be effective in moderating population growth in brown trout, including mechanical removal, TMFs, seasonal timing of high flows, and incentivized harvest using anglers to target larger size classes. Currently, the NPS Glen Canyon National Recreation Area (GLCA) is implementing an incentivized harvest of larger-sized brown trout.

Proposed Work

Project Element H.1. Rainbow Trout Fishery Monitoring in Glen Canyon (Modified Study)

Hypotheses and Science Questions

Project Element H.1 is designed to monitor trends in rainbow trout fishing and angler experience, which are directly linked to the LTEMP goal for the rainbow trout fishery.

Methods

In past work plans, the monitoring program used catch-per-unit effort (CPUE) metrics designed to be able to detect population level changes over a five-year or greater time scale; however, for the FY 2025-27 work plan we propose to combine staffing resources and electrofishing-based monitoring trips between H.1 and H.2 to enable more precise and sensitive mark-recapture monitoring, while also adding days onto 2 monitoring trips (June and October) for nonnative fish species surveillance (an additional downstream electrofishing monitoring trip is proposed in lieu of Lees Ferry CPUE trips as well; see Project I). In total (including H.1 and H.2), trout monitoring trips will be reduced from 8 to 4, compared to the previous work plan. Despite these changes, CPUE will continue to be calculated from data collected during the remaining 4 mark-recapture-focused trips.

The Lees Ferry Creel Survey and Arizona Game and Fish Department (AGFD) Citizen Science Project (FY 2021-23 and previous TWPs), which directly evaluates the quality and changes in the recreational experience of angling in the rainbow trout fishery in the Lees Ferry, Glen Canyon National Recreation Area is proposed to continue. AGFD will continue conducting a citizen science project that utilizes fishing guides to collect length data on fishes caught by their clients to reduce bias in angler catch quality (i.e., number of fish $\geq 14''$ and fish $\geq 20''$).

Anticipated Use of Data

Data collected under this element will be used to calculate metrics on the angling experience for both boat and walk-in anglers, including relative angler use, angler catch per unit effort values, and a rating of the fishery. AGFD angler surveys also provide information for Project J.

Outcomes and Products

A presentation will be given at the annual reporting meeting (and potentially at regional meetings) describing the condition of the fishery, with respect to angling satisfaction.

Project Element H.2. Trout Reproductive and Growth Dynamics (Modified Study)

Hypotheses and Science Questions

This is the field data collection component of the research project referred to as the Trout Reproductive and Growth Dynamics (TRGD) project, involving mark-recapture monitoring, as described below, is designed to determine the effects of LTEMP ROD flows on the recruitment of young-of-year (YOY) rainbow and brown trout in Glen Canyon, growth rates of juvenile and adult trout, and dispersal of YOY trout from Glen Canyon to Marble Canyon. Specific hypotheses and science questions addressed using the data collected under Project Element H.2 are listed under Project Element H.3.

Methods

For purposes of study replication, a *multi-reach mark-recapture sampling design* was established having two sub-reaches, each with an assigned 3-km length (sum 24% areal coverage). Trout (and potentially other species) will continue to be marked with a passive-integrated transponder (PIT) tag, and recapture data will be analyzed to estimate demographic parameters using appropriate mark-recapture models (Kendall and others, 1995; Korman and others, 2021). We will also continue to monitor DO levels in the Lees Ferry reach to understand how temporal and spatial variation relates to trout growth, body condition, and demographic rates.

This spatial coverage has allowed for the replication necessary for assessing experimental flow effects (Korman and others, 2021, 2023), and allows the TRGD program to maintain the necessary long-term analysis (comparisons and contrasts) associated with the Natal Origin project (2012-2017; FY 2015-17 TWP, Project Elements 9.1, 9.2, and FY 2018-20 TWP). Data from the TRGD project are used to inform development of a spatially stratified open population model for rainbow trout (Korman and Yard, 2017; Korman and others, 2017), and a population model for brown trout (Runge and others, 2018; Yackulic and others, 2020). These models provide estimates of how experimental flows and riverine conditions influence survival and recruitment of early YOY, as well as other size-classes of trout that may exhibit compensatory response in the population. Trout dispersal out of Glen Canyon will continue to be monitored in Marble Canyon, as well as trout population dynamics near the LCR confluence in conjunction with humpback chub monitoring (Project G).

Lastly, monitoring population dynamics of brown trout provides the means to assess incentivized take harvest measures by National Park Service (Project Element H.3). For the FY 2025-27 work plan, we propose to minimize project costs and avoid monitoring during periods of elevated stress to trout when low DO or high temperatures have been observed to occur (fall), by reducing the total number of electrofishing monitoring trips in Project Elements H.1 and H.2 from 8 to 4 trips (also see Project Element H.1 description above).

Anticipated Use of Data

Data collected on TRGD trips will be used to calculate metrics to assess long-term trends in the rainbow trout fishery such as relative condition, relative density, growth, and recruitment, and the distribution and abundance of nonnative species, and to assess effects and effectiveness of management actions (e.g., brown trout incentivized harvest). These data will be used to address science questions listed under H.3.

Outcomes and Products

We intend to evaluate how experimental flows and management actions influence recruitment, growth, survival, and dispersal of rainbow trout and brown trout in Glen Canyon, for peer-reviewed journal article(s), presentations at GCDAMP annual reporting meetings and at other scientific meetings, as described further under Project Element H.3.

Project Element H.3. Salmonid Modeling (Ongoing Study)

Hypotheses and Science Questions

Project Element H.3 will address science questions identified below, through analysis of data collected in H.1 and H.2. Most of the science questions, as originally proposed in the FY 2021-23 TWP (U.S. Department of the Interior, 2020) and extended in FY 2024, remain relevant now; however, we have revised the list, placing more emphasis on changing environmental conditions and brown trout, because of the relatively unclear risk posed by this species, new expansion of warm-water nonnative species, and recent declines in both trout populations in Glen Canyon.

- 1) What are the effects of changing conditions downstream of Glen Canyon Dam (i.e., DO, water temperature) on trout recruitment, growth, and survival?
- 2) What are the effects of flow experiments and other management actions (removals or flow management to suppress warm-water nonnative species) on trout population dynamics?
- 3) What controls rainbow and brown trout dispersal from Glen Canyon downstream into Marble and Grand canyons?
- 4) What factors control the quality of the rainbow trout fishery, and how does electrofishing capture efficiency influence interpretation of rainbow trout fishery metrics?
- 5) How do expansions of warm-water nonnative fish species influence trout population dynamics?
- 6) What factors regulate brown trout population dynamics in Glen Canyon, and what is the management efficacy of control actions?

- 7) How is climate change likely to impact the rainbow trout fishery below Glen Canyon Dam in short- and longer-term periods?
- 8) Are certain operations of Glen Canyon Dam more conducive than others to allow long-term coexistence of a high-quality rainbow trout fishery with downstream native fish species?

Methods

Modeling approaches will vary by question and will be determined as investigators proceed through the analyses of mark-recapture data.

Anticipated Use of Data

Data collected under Project Element H.2 will be used to address science questions listed above through modeling outlined in Project Element H.3. Salmonid modeling priorities for project H include 1) estimating the efficacy of incentivized harvest on brown trout (by updating brown trout population model and incorporating harvest data) to inform managers and Project J, 2) reassessing the brown trout hypotheses explored in Runge and others (2018) using data collected in recent years and based on natal origins and movement studies (Healy and others, 2022a; Akland, 2023), 3) continuing to develop models to predict recruitment and outmigration of rainbow and brown trout - outmigration of trout and other nonnative fishes from Lees Ferry to downstream areas, including tributaries, can enhance resilience of nonnative salmonid populations (Healy and others, 2023) and is seen as detrimental to native fish conservation goals (Yackulic and others, 2018; Healy and others, 2020), 4) estimating population dynamics of rainbow and brown trout in the Lees Ferry reach in response to experimental flows and environmental conditions, and 5) develop a workflow and analytical approach for calculating LTEMP metric scores for salmonids related to rainbow trout and nonnative fish goals. Finally, we plan to finish the development of manuscripts using reproductive status data collected in the previous work plan, which revealed that condition-effected sexual maturation rate appears to play an important role in regulating annual recruitment. Understanding this relationship may help us develop a more reliable method for forecasting and responding to large recruitment events.

Outcomes and Products

We anticipate completing analyses of data and development of models for salmonid populations to estimate the efficacy of ongoing management actions and improve capacity to predict impacts to fish populations from future management actions and changing environmental conditions, including the introduction and expansion of warm-water nonnative fish populations.

We intend to complete multiple peer reviewed journal article(s), and a presentation of results at the GCDAMP annual reporting meetings and at other scientific meetings related to brown trout population abundance, efficacy of the NPS incentivized brown trout harvest program, rainbow trout population dynamics as related to experimental flows and other drivers, and integrated models to predict rainbow and brown trout recruitment and outmigration.

Personnel and Collaborations

The overall project lead for Project H is Dr. Brian Healy, a Supervisory Research Fish Biologist at GCMRC who specializes in population dynamics, quantitative ecology, and decision science. Dr. Charles Yackulic is a Research Statistician at GCMRC specializing in population dynamics with an emphasis in modeling linkages and vital rates between trout populations. Dr. Kimberly Dibble is a Supervisory Research Fish Biologist at GCMRC with expertise in fish physiology, otolith microstructural analysis, and metadata analysis. Dr. Josh Korman is a Fish Biologist with Ecometric Research, Inc., specializing in analytical models and database development, population dynamics, and modeling capabilities. Dr. Bridget Deemer is a Research Ecologist at GCMRC specializing in water quality and limnology. Dr. David Rogowski is a Fish Biologist with the Arizona Game and Fish Department and is responsible for some of the long-term fish monitoring programs in Glen and Grand canyons and is experienced in statistical models and database management. Dr. Molly Webb is a Research Fishery Biologist with the Bozeman Fish Technology Center, U.S. Fish and Wildlife Service, specializing in the reproductive physiology and ecology of freshwater fishes. Dr. James Crossman is Fish Ecologist with BC Hydro in Vancouver, Canada and will be collaborating with Dr. Webb on the reproductive condition work proposed in Project Element H.3.

Budgets

Project Element Budget

| Fiscal Year 2025 | | | | | | | | | |
|--|------------------|----------------------|-----------------------|-----------------------|---------------------------|--------------------------|-----------------------------|------------------|------------------------------|
| Project H Salmonid Research and Monitoring | Salaries | Travel & Training | Operating Expenses | Logistics Expenses | Cooperative Agreements | To other USGS Centers | AMP Special Burden Rate* | Total | USGS Contributing Funds** |
| | | | | | | | 21.86% | | Est Full Rate = 55% |
| H.1. Rainbow trout fishery monitoring in Glen Canyon | \$7,467 | \$0 | \$0 | \$0 | \$88,000 | \$0 | \$4,272 | \$99,739 | |
| H.2. Experimental flow assessment of trout recruitment (TRGD) | \$78,947 | \$2,500 | \$88,800 | \$111,045 | \$0 | \$0 | \$61,499 | \$342,790 | |
| H.3. Salmonid modeling | \$37,450 | \$0 | \$6,000 | \$0 | \$0 | \$0 | \$9,499 | \$52,949 | |
| Total Project H | \$123,863 | \$2,500 | \$94,800 | \$111,045 | \$88,000 | \$0 | \$75,271 | \$495,479 | \$34,692 |

| Fiscal Year 2026 | | | | | | | | | |
|---|------------------|----------------------|-----------------------|-----------------------|---------------------------|--------------------------|-----------------------------|------------------|------------------------------|
| Project H Salmonid Research and Monitoring | Salaries | Travel & Training | Operating Expenses | Logistics Expenses | Cooperative Agreements | To other USGS Centers | AMP Special Burden Rate* | Total | USGS Contributing Funds** |
| | | | | | | | 22.60% | | Est Full Rate = 55.5% |
| H.1. Rainbow trout fishery monitoring in Glen Canyon | \$7,127 | \$0 | \$0 | \$0 | \$88,000 | \$0 | \$4,251 | \$99,378 | |
| H.2. Experimental flow assessment of trout recruitment (TRGD) | \$78,795 | \$2,500 | \$68,800 | \$115,875 | \$0 | \$0 | \$60,109 | \$326,080 | |
| H.3. Salmonid modeling | \$68,545 | \$0 | \$6,000 | \$0 | \$0 | \$0 | \$16,847 | \$91,392 | |
| Total Project H | \$154,467 | \$2,500 | \$74,800 | \$115,875 | \$88,000 | \$0 | \$81,207 | \$516,850 | \$36,839 |

| Fiscal Year 2027 | | | | | | | | | |
|---|------------------|----------------------|-----------------------|-----------------------|---------------------------|--------------------------|-----------------------------|------------------|------------------------------|
| Project H Salmonid Research and Monitoring | Salaries | Travel & Training | Operating Expenses | Logistics Expenses | Cooperative Agreements | To other USGS Centers | AMP Special Burden Rate* | Total | USGS Contributing Funds** |
| | | | | | | | 23.40% | | Est Full Rate = 56% |
| H.1. Rainbow trout fishery monitoring in Glen Canyon | \$7,484 | \$0 | \$0 | \$0 | \$88,000 | \$0 | \$4,391 | \$99,875 | |
| H.2. Experimental flow assessment of trout recruitment (TRGD) | \$82,735 | \$2,500 | \$68,800 | \$118,998 | \$0 | \$0 | \$63,890 | \$336,923 | |
| H.3. Salmonid modeling | \$71,972 | \$0 | \$6,000 | \$0 | \$0 | \$0 | \$18,245 | \$96,217 | |
| Total Project H | \$162,191 | \$2,500 | \$74,800 | \$118,998 | \$88,000 | \$0 | \$86,526 | \$533,015 | \$38,587 |

*GCDAMP Special Rate: Rate includes current DOI preferred rate (currently 15%; subject to change) and facilities rate (will vary annually). No USGS bureau overhead is charged (unique to the GCDAMP agreement).

**USGS Contributing Funds: The amount of funds required to cover the subsidy created by the reduced burden rate (i.e., GCDAMP Special Rate). As in previous years, SBSC/GCMRC will request these funds of the USGS cost-share program. These funds are not guaranteed.

References Cited

- Akland, M.K., 2023, Brown trout in Grand Canyon—Identifying natal origins and movement patterns using otoliths and eye lenses through established and novel microchemistry techniques: Syracuse, N.Y., State University of New York, M.S. Thesis, 153 p.
- Bair, L.S., Rogowski, D.L., and Neher, C., 2016, Economic value of angling on the Colorado River at Lees Ferry—Using secondary data to estimate the influence of seasonality: North American Journal of Fisheries Management, v. 36, no. 6, p. 1229-1239, <https://doi.org/10.1080/02755947.2016.1204388>.
- Blanchet, S., Loot, G., Grenouillet, G., and Brosse, S., 2007, Competitive interactions between native and exotic salmonids—A combined field and laboratory demonstration: Ecology of Freshwater Fish, v. 16, no. 2, p. 133-143, <https://doi.org/10.1111/j.1600-0633.2006.00205.x>.
- Bruckerhoff, L.A., Wheeler, K., Dibble, K.L., Mihalevich, B.A., Neilson, B.T., Wang, J., Yackulic, C.B., and Schmidt, J.C., 2022, Water storage decisions and consumptive use may constrain ecosystem management under severe sustained drought: Journal of American Water Resources Association, v. 58, no. 5, p. 654-672, <https://doi.org/10.1111/1752-1688.13020>.
- Bureau of Reclamation, 2024, Glen Canyon Dam Long-Term Experimental and Management Plan—Draft Supplemental Environmental Impact Statement: U.S. Department of the Interior, Bureau of Reclamation, Upper Colorado Basins, Interior Region 7,

https://www.usbr.gov/uc/DocLibrary/EnvironmentalImpactStatements/GlenCanyonDamLong-TermExperimentalManagementPlan/20240207-Draft-GCDLTEMP-SEIS_508.pdf.

- Carey, M.P., Sanderson, B.L., Friesen, T.A., Barnas, K.A., and Olden, J.D., 2011, Smallmouth bass in the Pacific Northwest—A threat to native species; a benefit for anglers: *Reviews in Fisheries Science*, v. 19, no. 3, p. 305–315, <https://doi.org/10.1080/10641262.2011.598584>.
- Cattaneo, F., Lamouroux, N., Breil, P., and Capra, H., 2002, The influence of hydrological and biotic processes on brown trout (*Salmo trutta*) population dynamics: *Canadian Journal of Fisheries and Aquatic Sciences*, v. 59, no. 1, p. 12-22, <https://doi.org/10.1139/F01-186>.
- Coggins, L.G., Yard, M.D., and Pine, W.E., 2011, Nonnative fish control in the Colorado River in Grand Canyon, Arizona—An effective program or serendipitous timing?: *Transactions of the American Fisheries Society*, v. 140, no. 2, p. 456-470, <https://doi.org/10.1080/00028487.2011.572009>.
- Crossman, J.A., Webb, M.A.H., Korman, J., and Yard, M.D., 2022, Population reproductive structure of rainbow trout determined by histology and advancing methods to assign sex and assess spawning capability: *Transactions of the American Fisheries Society*, v. 151, no. 4, p. 422-440, <https://doi.org/10.1002/tafs.10356>.
- Dibble, K.L., Yackulic, C.B., Kennedy, T.A., Bestgen, K.R., and Schmidt, J.C., 2021, Water storage decisions will determine the distribution and persistence of imperiled river fishes: *Ecological Applications*, v. 31, no. 2, e02279, p. 1-9, <https://doi.org/10.1002/eap.2279>.
- Eppehimer, D.E., Yackulic, C.B., Bruckerhoff, L.A., Wang, J., Young, K.L., Bestgen, K.R., Mihalevich, B.A., and Schmidt, J.C., 2024, Declining reservoir elevations following a two-decade drought increase water temperatures and non-native fish passage facilitating a downstream invasion: *bioRxiv*, <https://doi.org/10.1101/2024.01.23.576966>.
- Fonken, D., Rogowski, D.L., Fennell, J., and Gardner, T., 2023, Colorado River fish monitoring in the Grand Canyon, Arizona—2022 annual report: Flagstaff, Ariz., Arizona Game and Fish Department, submitted to U.S. Geological Survey, Southwest Biological Science Center, Grand Canyon Monitoring and Research Center, 41 p.
- Glover, R.S., Fryer, R.J., Soulsby, C., and Malcolm, I.A., 2019, These are not the trends you are looking for: poorly calibrated single-pass electrofishing data can bias estimates of trends in fish abundance: *Journal of Fish Biology*, v. 95, no. 5, p. 1223-1235, <https://doi.org/10.1111/jfb.14119>.
- Hanks, D.M., and Secor, D.H., 2011, Bioenergetic responses of Chesapeake Bay white perch (*Morone americana*) to nursery conditions of temperature, dissolved oxygen, and salinity: *Marine Biology*, v. 158, no. 4, p. 805–815, <https://doi.org/10.1007/s00227-010-1608-0>.
- Healy, B., Budy, P., Yackulic, C., Murphy, B.P., Schelly, R.C., and McKinstry, M.C., 2023, Exploring metapopulation-scale suppression alternatives for a global invader in a river network experiencing climate change: *Conservation Biology*, v. 37, no. 1, e13993, p. 1-18, <https://doi.org/10.1111/cobi.13993>.
- Healy, B.D., Yackulic, C.B., and Schelly, R.C., 2022a, Impeding access to tributary spawning habitat and releasing experimental fall-timed floods increases brown trout immigration into a

- dam's tailwater: Canadian Journal of Fisheries and Aquatic Sciences, v. 80, no. 3, p. 614-627, <https://doi.org/10.1139/cjfas-2022-0231>.
- Healy, B.D., Moore, J.F., and Pine III, W.E., 2022b, Monitoring and managing fishes that are invisible and keep moving around—Influences of an invasive species and environmental factors on capture probability: North American Journal of Fisheries Management, v. 42, no. 2, p. 396-409, <https://doi.org/10.1002/nafm.10755>.
- Healy, B.D., Schelly, R.C., Yackulic, C.B., Omana Smith, E.C., and Budy, P., 2020, Remarkable response of native fishes to invasive trout suppression varies with trout density, temperature, and annual hydrology: Canadian Journal of Fisheries and Aquatic Sciences, v. 77, no. 9, p. 1446-1462, <https://doi.org/10.1139/cjfas-2020-0028>.
- Hueftle, S.J., and Stevens, L.E., 2001, Experimental flood effects on the limnology of Lake Powell reservoir, southwestern USA: Ecological Applications, v. 11, no. 3, p. 644-656, <https://doi.org/10.2307/3061107>.
- Hughes, R.M., and Herlihy, A.T., 2012, Patterns in catch per unit effort of native prey fish and alien piscivorous fish in 7 Pacific Northwest USA rivers: Fisheries, v. 37, no. 5, p. 201-211, <https://doi.org/10.1080/03632415.2012.676833>.
- Kendall, W.L., Pollock, K.H., and Brownie, C., 1995, A likelihood-based approach to capture-recapture estimation of demographic parameters under the robust design: Biometrics, v. 51, no. 1, p. 293-308, <https://doi.org/10.2307/2533335>.
- Korman, J., Deemer, B., Yackulic, C.B., Kennedy, T.A., and Giardina, M., 2023, Drought-related changes in water quality surpass effects of experimental flows on trout growth downstream of Lake Powell reservoir: Canadian Journal of Fisheries and Aquatic Sciences, v. 80, no. 3, p. 424-438, <https://doi.org/10.1139/cjfas-2022-0142>.
- Korman, J., Kaplinski, M., and Melis, T.S., 2011, Effects of fluctuating flows and a controlled flood on incubation success and early survival rates and growth of age-0 rainbow trout in a large regulated river: Transactions of the American Fisheries Society, v. 140, no. 2, p. 487-505, <https://doi.org/10.1080/00028487.2011.572015>.
- Korman, J., Martell, S.J.D., Walters, C.J., Makinster, A.S., Coggins, L.G., Yard, M.D., and Persons, W.R., 2012, Estimating recruitment dynamics and movement of rainbow trout (*Oncorhynchus mykiss*) in the Colorado River in Grand Canyon using an integrated assessment model: Canadian Journal of Fisheries and Aquatic Sciences, v. 69, no. 11, p. 1827-1849, <https://doi.org/10.1139/F2012-097>.
- Korman, J., and Yard, M.D., 2017, Effects of environmental covariates and density on the catchability of fish populations and interpretation of catch per unit effort trends: Fisheries Research, v. 189, p. 18-34, <https://doi.org/10.1016/j.fishres.2017.01.005>.
- Korman, J., Yard, M.D., Dzul, M.C., Yackulic, C.B., Dodrill, M.J., Deemer, B.R., and Kennedy, T.A., 2021, Changes in prey, turbidity, and competition reduce somatic growth and cause the collapse of a fish population: Ecological Monographs, v. 91, no. 1, e01427, p. 1-20, <https://doi.org/10.1002/ecm.1427>.

- Korman, J., Yard, M.D., and Kennedy, T.A., 2017, Trends in rainbow trout recruitment, abundance, survival, and growth during a boom-and-bust cycle in a tailwater fishery: Transactions of the American Fisheries Society, v. 146, no. 5, p. 1043-1057, <https://doi.org/10.1080/00028487.2017.1317663>.
- Korman, J., Yard, M.D., and Yackulic, C.B., 2016, Factors controlling the abundance of rainbow trout in the Colorado River in Grand Canyon in a reach utilized by endangered humpback chub: Canadian Journal of Fisheries and Aquatic Sciences, v. 73, no. 1, p. 105-124, <https://doi.org/10.1139/cjfas-2015-0101>.
- Korman, J., Crossman, J. Webb, M., Giardina, M., and Yackulic, C., 2024, Rainbow and brown trout population dynamics in Glen Canyon [presentation]: Phoenix, Ariz., Glen Canyon Dam Adaptive Management Program, Annual Reporting Meeting, January 23-24, 2024, <https://www.usbr.gov/uc/progact/amp/twg/2024-01-25-twg-meeting/20240125-AnnualReportingMeeting-AZGFDLeesFerryMonitoring-508-UCRO.pdf>
- Lobón-Cerviá, J., 2004, Discharge-dependent covariation patterns in the population dynamics of brown trout (*Salmo trutta*) within a Cantabrian river drainage: Canadian Journal of Fisheries and Aquatic Sciences, v. 61, no. 10, p. 1929-1939, <https://doi.org/10.1139/f04-118>.
- Rogosch, J.S., and Olden, J.D., 2020, Invaders induce coordinated isotopic niche shifts in native fish species: Canadian Journal of Fisheries and Aquatic Sciences, v. 77, no. 8, p. 1348-1358, <https://doi.org/10.1139/cjfas-2019-0346>.
- Runge, M.C., Yackulic, C.B., Bair, L.S., Kennedy, T.A., Valdez, R.A., Ellsworth, C., Kershner, J.L., Rogers, R.S., Trammell, M., and Young, K.L., 2018, Brown trout in the Lees Ferry reach of the Colorado River—Evaluation of causal hypotheses and potential interventions: U.S. Geological Survey Open-File Report 2018-1069, 83 p., <https://doi.org/10.3133/ofr20181069>.
- Schelly, R.C., Omana Smith, E., Koller, R., and Healy, B., 2021, Bright Angel Creek comprehensive brown trout control project season report—October 1, 2020-March 1, 2021: Grand Canyon, Ariz., National Park Service, Grand Canyon National Park, prepared for the Upper Colorado Region, Bureau of Reclamation, Interagency agreement no. R17PG00048, 24 p., https://www.researchgate.net/publication/357656772_Bright_Angel_Creek_comprehensive_brown_trout_control_project_October_1_2020_-March_1_2021_season_report.
- U.S. Department of the Interior, 2020, Glen Canyon Dam Adaptive Management Program Triennial Budget and Work Plan—Fiscal years 2021-2023—Final approved by the Secretary of the Interior—December 2, 2020: Flagstaff, Ariz., U.S. Geological Survey, Grand Canyon Monitoring and Research Center, and Salt Lake City, Utah, Bureau of Reclamation, Upper Colorado Region, 384 p., http://gcdamp.com/images_gcdamp_com/5/5d/GCMRC_TWP2021-23_December2_2020_ApprovedBySecretary.pdf.
- U.S. Department of Interior, 2016, Glen Canyon Dam Long-Term Experimental and Management Plan final Environmental Impact Statement (LTEMP EIS): U.S. Department of

- the Interior, Bureau of Reclamation, Upper Colorado Region, National Park Service, Intermountain Region, online, <http://itempeis.anl.gov/documents/final-eis/>.
- U.S. Geological Survey, Grand Canyon Monitoring and Research Center, 2024, U.S. Geological Survey Grand Canyon Monitoring and Research Center Proceedings of the Fiscal Year 2023 Annual Reporting Meeting to the Glen Canyon Dam Adaptive Management Program: Flagstaff, Ariz., prepared by U.S. Geological Survey, Southwest Biological Science Center, submitted to Bureau of Reclamation, Phoenix, Ariz., January 23-25, 2024, 189 p., <https://pubs.usgs.gov/publication/70251210>.
- Vernieu, W.S., 2010, Effects of the 2008 high-flow experiment on water quality in Lake Powell and Glen Canyon Dam releases, Utah-Arizona: U.S. Geological Survey Open-File Report 2010-1159, 25 p., <https://pubs.usgs.gov/of/2010/1159/>.
- Waldrop, T., Summerfelt, S., Mazik, P., Kenney, P.B., and Good, C., 2020, The effects of swimming exercise and dissolved oxygen on growth performance, fin condition and survival of rainbow trout *Oncorhynchus mykiss*: Aquaculture Research, v. 51, no. 6, p. 2582–2589, <https://doi.org/10.1111/are.14600>.
- Walsworth, T.E., and Budy, P., 2015, Integrating nonnative species in niche models to prioritize native fish restoration activity locations along a desert river corridor: Transactions of the American Fisheries Society, v. 144, no. 4, p. 667-681, <https://doi.org/10.1080/00028487.2015.1024333>.
- Whiting, D.P., Paukert, C.P., Healy, B.D., and Spurgeon, J.J., 2014, Macroinvertebrate prey availability and food web dynamics of nonnative trout in a Colorado River tributary, Grand Canyon: Freshwater Science, v. 33, no. 3, p. 872-884, <https://doi.org/10.1086/676915>.
- Williams, B.K., Nichols, J.D., and Conroy, M.J., 2002, Analysis and management of animal populations—Modeling, estimation, and decision making: London, UK, Academic Press.
- Yackulic, C.B., Bair, L.S., Eppheimer, D.E., Salter, G.L., Deemer, B.R., Butterfield, B.J., Kasprak, A., Caster, J.J., Fairley, H.C., Grams, P.E., Mihalevich, B.A., Palmquist, E.C., and Sankey, J.B., 2024, Modeling the impacts of Glen Canyon Dam operations on Colorado River resources: Phoenix, Ariz., U.S. Department of the Interior, Bureau of Reclamation, cooperator publication prepared by U.S. Geological Survey, Southwest Biological Science Center, Grand Canyon Monitoring and Research Center, Flagstaff, Ariz., April 2024, 133 p., <https://pubs.usgs.gov/publication/70252976>.
- Yackulic, C.B., Dodrill, M.J., Dzul, M.C., Sanderlin, J., and Reid, J., 2020, A need for speed in Bayesian population models—A practical guide to marginalizing and recovering discrete latent states: Ecological Applications, v. 30, no. 5, e02112, p. 1-19, <https://doi.org/10.1002/eap.2112>.
- Yackulic, C.B., Korman, J., Yard, M.D., and Dzul, M.C., 2018, Inferring species interactions through joint mark-recapture analysis: Ecology, v. 99, no. 4, p. 812-821, <https://doi.org/10.1002/ecy.2166>.

- Yard, M.D., and Korman, J., 2020, TRGD—Trout recruitment, growth and population dynamics [presentation]: Phoenix, Ariz., Glen Canyon Dam Adaptive Management Program, Annual Reporting Meeting, January 14, 2020.
- Yard, M.D., Coggins, L.G., Baxter, C.V., Bennett, G.E., and Korman, J., 2011, Trout piscivory in the Colorado River, Grand Canyon—Effects of turbidity, temperature, and fish prey availability: Transactions of the American Fisheries Society, v. 140, no. 2, p. 471-486, <https://doi.org/10.1080/00028487.2011.572011>.
- Yard, M.D., Yackulic, C.B., Korman, J., Dodrill, M.J., and Deemer, B.R., 2023, Declines in prey production during the collapse of a tailwater rainbow trout population are associated with changing reservoir conditions: Transactions of the American Fisheries Society, v. 152, no. 1, p. 35-50, <https://doi.org/10.1002/tafs.10381>.

Project I: Nonnative Aquatic Species Monitoring and Research

Investigators

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Project Summary and Purpose

Introductions of nonnative aquatic species into river ecosystems have resulted in increased competition for limited resources, predation on native species, and habitat alterations on a global scale (Gallardo and others, 2016; Havel and others, 2015). In both the upper and lower Colorado River Basins, warm-water predatory fishes are implicated in the lack of recruitment and subsequent population declines of native fishes (Mueller, 2005; Martinez and others, 2014). Outside of preventative measures, control methods are typically most cost effective and successful when nonnative species are detected early (Leung and others, 2002; Dawson and Kolar, 2003). As such, a robust monitoring and research program increases the likelihood that new introductions and expansions of nonnative fish species will be detected early and that managers will have the information needed to control introduced species before distribution expands and abundance increases.

Rapidly changing conditions in the Colorado River downstream from Glen Canyon Dam since 2022, and the immediate threat of smallmouth bass (*Micropterus dolomieu*) and other predatory nonnative species to the Colorado River ecosystem, have necessitated a shift in focus for Project I relative to previous work plans. The focus of Project I will primarily be on nonnative fish/aquatic species, since the category for ‘native fishes’ was moved to Project N [except for Project Element I.1, which focuses on both native and nonnative fishes].

In the Glen Canyon Dam Adaptive Management Program FY 2025-27 Triennial Budget and Work Plan (hereafter, FY 2025-27 TWP), we will focus our research and monitoring efforts on detecting new or notable increases and/or distributional shifts in nonnative species using genomic tools to determine nest origin and potentially the abundance of nonnative fishes being removed from the river, conduct enhanced biosurveillance through additional trips and environmental DNA (eDNA) monitoring, assess entrainment potential for nonnative fishes in Lake Powell, and analyze data from federal and state partners to assess the efficacy of management actions to control nonnative populations of smallmouth bass.

We propose to continue long-term, standardized monitoring conducted by the Arizona Game and Fish Department (AGFD) throughout the Colorado River from Lees Ferry (RM 0) to Pearce Ferry (RM 281) for the combined purposes of tracking the status of native and nonnative fishes (Project Element I.1). These data can be used to determine the distribution of nonnative fishes throughout Glen and Grand Canyons and quantify growth rates through use of length-frequency histograms and modal progression analysis.

We propose to pursue kinship genetic analysis to evaluate the relatedness of individual smallmouth bass and determine whether there is suggestive evidence of local recruitment (i.e., most individuals captured near the slough are siblings or half-siblings) or entrainment (i.e., most individuals captured near the dam represent a greater number of parent-offspring pairs with origins in Lake Powell; Project Element I.2). These data can also be used to conduct close-kin mark-recapture analysis to estimate abundance and other demographic parameters (e.g., survival, population trends) using kinship relationships of individuals identified by genetic sampling.

We would like to re-focus an existing Project Element (I.3) to detect emerging threats in the Colorado River Ecosystem (CRe), which may include nonnative fishes, parasites, disease, crayfish, gastropods, mollusks, and other nonnative aquatic species. One of the primary methods proposed to detect such threats in this work plan will be filtering water to detect eDNA, which is a method that will reduce handling and potential health impacts to fishes.

Last, we propose to add a new Project Element (I.4) that focuses on modeling nonnative fish data from U.S. Geological Survey (USGS), National Park Service (NPS), AGFD, U.S. Fish and Wildlife Service (USFWS), and other cooperators. This modeling will support refinement of existing smallmouth bass models, help determine the effectiveness of Long-Term Experimental and Management Plan (LTEMP) Supplemental Environmental Impact Statement (SEIS) Flow Experiments and smallmouth bass removal actions, and provide science to support discussions of off-ramps laid out in the Invasive Fish Species Strategic Plan (Smallmouth Bass Ad Hoc Group, 2023). We also propose to develop a workflow process for an occupancy model in collaboration with Projects H and G that models native and nonnative fish detection data throughout the CRe for the purpose of reporting out to the Glen Canyon Dam Adaptive Management Program (GCDAMP) on LTEMP Performance Metrics.

If flows to disadvantage smallmouth bass are implemented per the preferred alternative in the LTEMP SEIS, we propose to use the Experimental Fund to analyze monitoring data to evaluate the effectiveness of smallmouth bass flows, collect larval fishes to determine hatch dates, and collect targeted eDNA samples for smallmouth bass prior to and after the experimental flow (Project Elements I.5, I.6).

Long-Term Experimental and Management Plan (LTEMP) Resource Goals

Maintaining self-sustaining native fish populations within the Colorado River and minimizing the presence and expansion of aquatic nonnative species are two resource goals outlined in the LTEMP EIS and associated Biological Opinion for the operation of Glen Canyon Dam (U.S. Department of the Interior, 2016a, b; U.S. Fish and Wildlife Service, 2016). These two resource goals (#5 and #10, respectively) are closely linked together in that nonnative warm water fishes are largely incompatible with Colorado River native fishes including humpback chub, *Gila cypha* (Marsh and Pacey, 2005; Minckley and Marsh, 2009). Nonnative warm-water fishes prey upon juvenile native fishes, and once established, can cause rapid disappearance of native fishes (Moyle and others, 1986). As such, Project I supports the following LTEMP Resource Goals:

- Nonnative Invasive Species
- Other Native Fish Species
- Humpback Chub

Background

Prior to construction of Glen Canyon Dam, water temperature in the Colorado River in Grand Canyon historically fluctuated from near freezing in the winter to almost 30°C in the summer (Vernieu and others, 2005) and was highly turbid during most of the year (Voichick and Topping, 2014). During this period, the Colorado River was dominated by native fishes and introduced warm-water fishes such as channel catfish (*Ictalurus punctatus*) (Haden, 1992; Minckley and Marsh, 2009). These warm-water fishes can handle cold winter water temperatures but need warmer water to grow and complete their life cycles. With the completion of Glen Canyon Dam in 1963 and subsequent filling of Lake Powell, river temperatures in the Lees Ferry reach dropped to 7-10°C by 1973 with little annual variation (Vernieu and others, 2005). This shift in the thermal regime dramatically altered the fish assemblage of the Colorado River in Grand Canyon, changing it from an assemblage dominated by warm-water species to one dominated by cold-water tolerant species including rainbow trout (*Oncorhynchus mykiss*).

This condition persisted for several decades. Until 2022, constant cold water largely kept warm-water nonnative fishes from becoming established throughout Glen Canyon, Marble Canyon, and most of Grand Canyon (Dibble and others, 2021; Eppehimer and others, 2024).

Cold water releases from Lake Powell persisted from 1983, when the reservoir filled completely, until 2005 when drought conditions caused water elevations in Lake Powell to drop. Warmer surface waters (8- 16°C) became entrained in the penstocks and were released downstream, once again creating a varied thermal regime in the Colorado River (Vernieu and others, 2005).

This recent period of warmer water caused by drought conditions, within a system dominated by cold water for a long time, may be one reason for the unique pattern of native fish increases in recent years in the CRe (Van Haverbeke and others, 2017) compared to dramatic declines that have occurred in most other portions of the Colorado River Basin during the same period (Bestgen and others, 2018; Dibble and others, 2021; Johnson and others, 2008). The river flowing through Grand Canyon warmed enough over the past two decades to support growth of native fishes downstream but was cold enough in the tailwater to support a rainbow trout fishery. Lake Powell elevations were high enough during most of this period that penstocks did not draw from the warmer epilimnion, so entrainment of warm-water nonnative fishes was likely low. Essentially, the river was ‘not too hot, not too cold...but just right’ from a perspective of allowing some level of thermal support for the tailwater trout fishery and native fishes including humpback chub downstream. Further, as Lake Mead dropped in elevation below 346 meters above sea level due to declining basin-wide water supply, Pearce Ferry Rapid emerged near the inflow as the river cut through the reservoir delta outside the original channel. Colorado River water now flows over a bedrock ledge at Pearce Ferry Rapid, and this has been hypothesized to be a barrier to upstream movement of nonnative fishes that prey on native fishes (Kegerries and others, 2020).

Nonnative fishes typically have detrimental impacts on the stability of native fish communities (Erős and others, 2020), and these effects can be exacerbated by drought conditions (Rogosch and others, 2019). Identifying sources of warm-water nonnative fishes in the CRe early improves the likelihood that a successful rapid containment/eradication response can be accomplished before negative impacts on endangered populations occur (Martinez and others, 2014). Preventing the introduction and spread of warm-water nonnative fishes is far more environmentally and fiscally desirable than undertaking control or eradication efforts after they become established (Cucherousset and Olden, 2011; Martinez and others, 2014). Introduced fishes may grow rapidly with high reproductive potential, and once established, eradication is often essentially impossible and control typically requires long-term and expensive efforts with often limited efficacy (Pimentel and others, 2000; Simberloff, 2003; Mueller, 2005; Johnson and others, 2009; McIntosh and others, 2010; Martinez and others, 2014).

In 2022, warm-water nonnative fishes including smallmouth bass, green sunfish (*Lepomis cyanellus*), and walleye (*Sander vitreus*) were captured in unprecedented numbers in the Colorado River downstream from Glen Canyon Dam during routine fish monitoring efforts by state and federal agencies (Eppheimer and others, 2024). Reservoir levels in Lake Powell had precipitously declined during a multi-decadal drought, resulting in reservoir releases from the warmer epilimnion (top layer) and upper metalimnion (middle layer) of the lake.

Biological activity and fish production is highest in the warmer epilimnion of Lake Powell, so as reservoir levels dropped more fishes were entrained and moved through the penstocks into the downstream river (Eppheimer and others, 2024). Warmer reservoir releases likely supported higher survival and successful reproduction of warm water entrained fishes including smallmouth bass (Dibble and others, 2021; Bruckerhoff and others, 2022; Eppheimer and others, 2024). Smallmouth bass are a species of high concern to resource managers because they are highly piscivorous and have led to population-level declines in native species in regions where they have been introduced (Loppnow and others, 2013).

Recognizing the threat of smallmouth bass and other warm-water nonnatives to threatened and endangered fish recovery, federal and state agencies mobilized around efforts to characterize the threat, remove warm-water nonnatives from the Colorado River in Glen and Grand Canyons, and develop models to support identification of flow management strategies to disadvantage warm-water nonnative fishes. The Smallmouth Bass Ad Hoc Group (SBAHG), through the Technical Work Group of the GCDAMP in partnership with GCMRC and Reclamation, developed an Invasive Fish Species Strategic Plan (hereafter, 'strategic plan') (Smallmouth Bass Ad Hoc Group, 2023) with a goal to prevent, detect, and respond to the establishment of nonnative fishes in the Colorado River ecosystem. This plan outlines a strategy to prevent fish passage through Glen Canyon Dam while also preventing establishment downstream from the dam.

Management actions taken by federal and state partners from 2022-2024 in accordance with the strategic plan included smallmouth bass and other nonnative fish removals and monitoring in Glen Canyon National Recreation Area (GLCA) and Grand Canyon National Park (GRCA) (Smallmouth Bass Ad Hoc Group, 2023). In addition, GCMRC scientists provided modeling support for the development of the LTEMP SEIS, which evaluated management alternatives to reduce the likelihood of smallmouth bass reproduction and population growth by reducing water temperature (via river outlet release) and/or by spiking flows within powerplant capacity.

As indicated in the 2016 Biological Opinion for the LTEMP EIS, the U.S. Bureau of Reclamation (hereafter, 'Reclamation') will conduct planning and compliance for implementation of rapid response control efforts for newly establishing or existing deleterious nonnative species within and contiguous to the action area (U.S. Fish and Wildlife Service, 2016). Before any management actions can occur, monitoring must first detect the introduced fish and research must be conducted to evaluate risks and inform managers whether control efforts are warranted and feasible. This project provides the information that will allow for effective integration of monitoring, research, and nonnative species management in the Colorado River downstream from Glen Canyon Dam.

Proposed Work

Project Element I.1. System-wide Native Fishes and Nonnative Aquatic Species Monitoring (Ongoing Study)

David Rogowski, John Fennell, Dale Fonken, Trenton Schipper

Arizona Game and Fish Department

The objective of this project element is to provide long-term data on the longitudinal distribution and status of the fish assemblage in the mainstem Colorado River from Lees Ferry (RM 0) to Pierce Ferry Rapid (RM 281). These data could be used with other data sources in the development of LTEMP Performance Metrics focused on detecting the presence and reproduction of aquatic nonnative species downstream from Glen Canyon Dam. Annual AGFD data on adult flannelmouth sucker (*Catostomus latipinnis*) CPUE for three years following LTEMP Experimental Actions is also being used as a surrogate for effects on razorback sucker (*Xyrauchen texanus*) per the 2016 Biological Opinion (U.S. Department of the Interior, 2016a).

In the FY 2021-23 work plan, AGFD conducted one spring sampling trip in FY 2021 and FY 2022 and two spring sampling trips in FY 2023, combined with an annual fall trip that sampled the last 15 miles of river upstream from Pearce Ferry Rapid. In the FY 2025-27 TWP, AGFD will conduct one spring sampling trip annually from Lees Ferry to Pearce Ferry and add a new system-wide sampling trip in fall that is timed when nonnative detections are likely to be highest during the warmest reservoir release months of the year. This trip will replace the existing fall trip that samples the last 15 miles of river, since that area is covered by the new fall system-wide sampling trip. This extended trip is included in the FY 2025-27 TWP to expand upon existing monitoring efforts to detect nonnative fish per the strategic plan (Smallmouth Bass Ad Hoc Group, 2023). Nonnative fish surveillance that was conducted in the Lees Ferry reach during previous AGFD trout monitoring trips will be discontinued during this work plan due to adequate coverage from other agencies sampling the slough at RM -12 and other hot spots, by the dam, and other fixed and random sites throughout the Lees Ferry reach. The cessation of nonnative fish sampling in Lees Ferry associated with this project element is intended to reduce handling of fishes per recreational angler and Tribal concerns.

Hypotheses and Science Questions

- I.1.1: What is the species composition, relative abundance, longitudinal distribution, and population trends of the native and nonnative fish community inhabiting the CRe?
- I.1.2: Can portable submersible PIT tag antennas be used to detect movement of humpback chub in Grand Canyon?

Methods

I.1.1. What is the Species Composition, Relative Abundance, Longitudinal Distribution, and Population Trends of the Native and Nonnative Fish Community Inhabiting the CRe?

The primary objective of Project Element I.1 is to determine the species composition, relative abundance, and trends of fishes in the CRe, which is sampling that has occurred since 2000 (Figure 1). The methods for this project are explained in detail in AGFD Annual Reports (Boyer and Rogowski, 2020; Fonken and others, 2023). For sampling purposes, the Colorado River is divided into 83 reaches 5 miles (8 kilometers, km) in length. Reaches are shorter where rapids define the upstream and downstream extent of a reach. Sample sites within a reach are 250 meters (m) in length on both sides of the river.

Sampling reaches are selected using a spatially stratified random approach (Grafström and Tillé, 2013; Robertson and others, 2013) equivalent to the number of nights sampling in a trip, with the probability of selection weighted according to the percent of available sample sites within that reach relative to the total sampling area (RM 0 - 281.4, n = 3,507 available sites). After a reach is selected, sample sites (20 electrofishing, 16 hoop net) are randomly selected for each reach (36 sites = approximately 56% of sites available in a 5-mile reach). In shorter reaches, sample size is reduced so that 56% of available sites are selected. Sample sites within reaches may be reduced for logistical reasons or due to poor water quality (e.g., turbidity) affecting capture probability. Trip length in the FY 2025-27 TWP will be increased from 13 nights sampling (15 days total) to 15 nights sampling (17 days total) to allow for more nonnative fish sampling and the inclusion of portable PIT tag antennas as an alternate gear type to increase detections of previously tagged fishes (see Project Element G.9). Due to logistical and time constraints with the inclusion of the remote PIT tag antennas to the monitoring program, the number of sites selected for electrofishing and hoop netting within a reach may need to be slightly reduced.

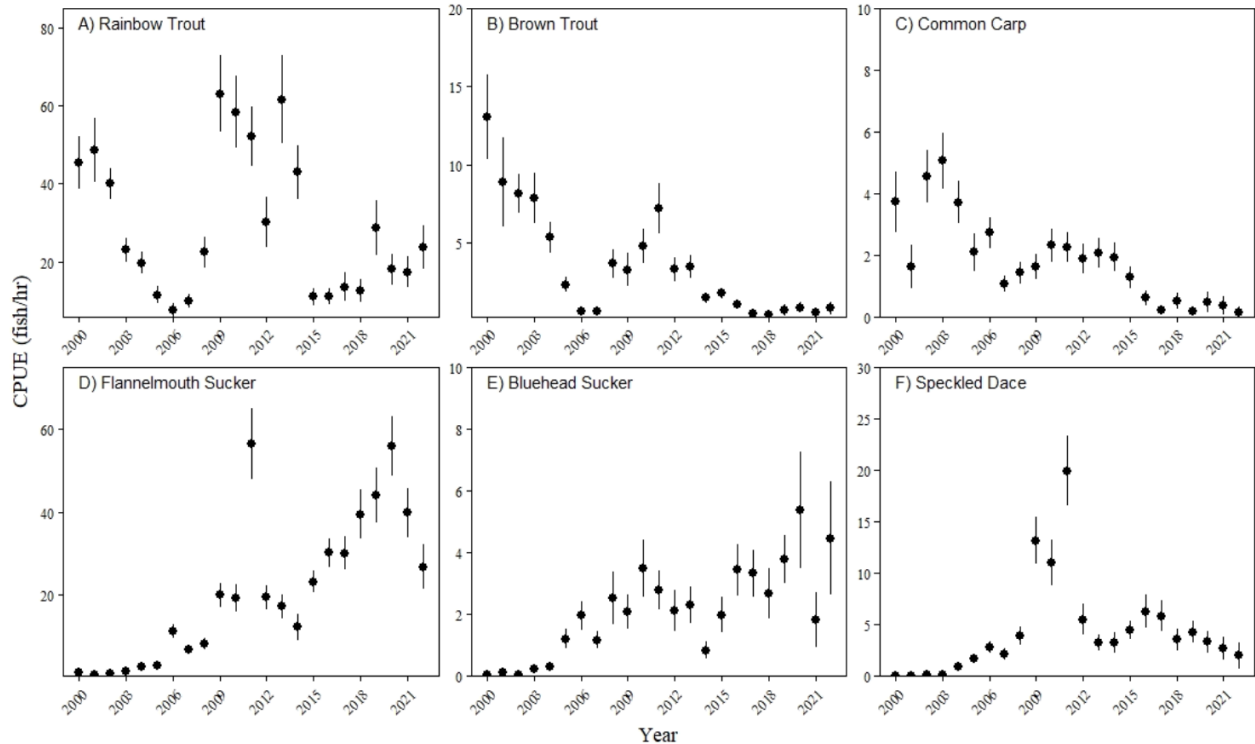


Figure 1. Mean electrofishing CPUE (# fish per hour) for nonnative (A-C) and native (D-F) fish species in the Colorado River from Lees Ferry to Pearce Ferry, from 2000-2022. Error bars indicate 95% confidence intervals (Fonken and others, 2023).

We will use a combination of standardized electrofishing, hoop netting, and angling to sample the native and nonnative fish community. Electrofishing will occur at night (commencing with the appearance of two stars or planets) with two 16-ft Osprey sport boats outfitted with a boat electrofishing system, using an ETS (Electrofishing Systems, LLC) Complex Pulse System (CPS) unit (MBS-1DPQ-CR-AZ) powered by a 6,500W generator (Honda EG6500). The CPS units apply between 195 to 400 volts and 12.0 to 18.5 amperes to one spherical steel anode (25.4 cm diameter) partially submerged (~5 cm exposed) off the front of the boat. The aluminum boat hull acts as the cathode. As sampling moves downstream, voltage is decreased and amperage is increased to maintain similar power output with increasing water conductivity. Seconds spent electrofishing a site are recorded as the unit of effort. Each electrofishing sample consists of a single electrofishing pass along a 250-m shoreline transect. Each boat is crewed by one boatman and one netter.

Hoop nets will be set overnight and measure 1.3 m long and 0.6 m in diameter with 6.35-millimeter (mm) mesh and consist of three hoops and a single 0.1-m throat. All nets will be baited with approximately 117 g of Purina Aquamax fish food. Net set locations within the sample site are based on the ability to effectively secure the net depending on water depth, tie off structures, and river currents.

Each night we will angle for channel catfish using spinning rods baited with Hillshire Farm Lil' Smokies sausages. Angling usually occurs in eddies at camp just before dark, for a minimum of one hour of angling effort per site.

I.1.2. Can Portable Submersible PIT Tag Antennas be used to Detect Movement of Humpback Chub in Grand Canyon?

In the FY 2025-27 work plan we will test submersible portable PIT tag antennas in collaboration with Project Element G.9, with the objective of increasing detections of tagged fish in areas less frequently sampled in Grand Canyon. Increased detections could improve humpback chub population models by providing information on movement of humpback chub out of study reaches (vs. mortality), which may improve our understanding of apparent survival. Since AGFD samples random sites throughout Grand Canyon, deploying antennas will allow us to detect PIT-tagged fishes outside of fixed sites without handling the fish directly. In addition to potentially improving a humpback chub population dynamics model (see Project Element G.1), this information will supplement previous approaches by AGFD to visually assess movement using data from manual sampling and fixed antennas already deployed in Grand Canyon (Figure 2). Antennas will be deployed at a small number of random sites (max 8-10 per reach) and left to passively read data while other sampling is taking place (e.g., hoop netting).

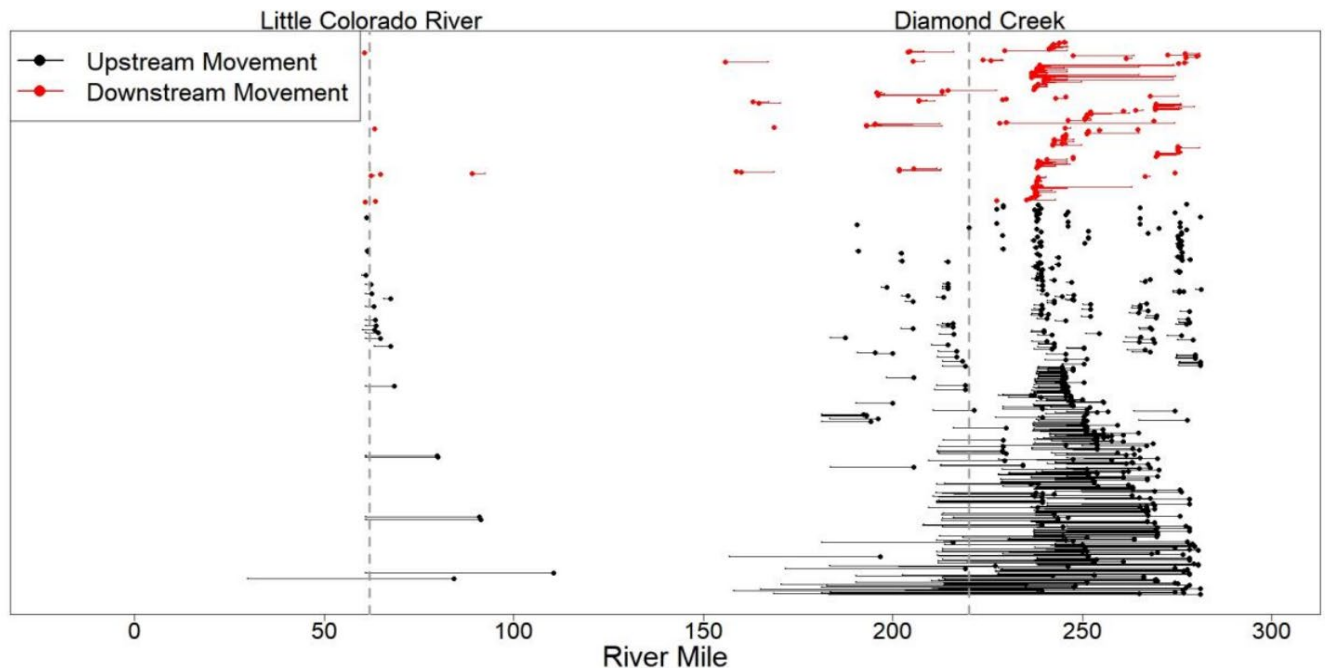


Figure 2. Movement of humpback chub (*Gila cypha*) tagged by AGFD and recaptured from manual sampling and from antennas deployed in Grand Canyon from Lees Ferry (RM 0) to Pearce Ferry (RM 281). Lines represent individual fish that move upstream (black lines) and downstream (red) from the point of origin (dots) (Fonken and others, 2023).

Anticipated Use of Data

These data will be used to monitor the relative abundance (CPUE), condition, distribution, and movement of native and nonnative fishes in the CRe. Annual adult flannelmouth sucker catch rates for three years following LTEMP Experimental Actions is currently being used as a surrogate for the effects of flows on razorback sucker populations.

Outcomes and Products

Outcomes

- Monitoring and detection of trends in distribution and relative abundance of native and nonnative fishes in the Colorado River. Evaluation of fish condition, growth, and movement relative to environmental conditions.

Products

- AGFD trip reports and annual report, AGFD annual presentations at the Grand Canyon Fish Cooperators Meeting and the GCDAMP Annual Reporting Meeting, AGFD conference presentations at the annual Desert Fishes Council Meeting, AGFD conference presentation(s) at the joint annual meeting of the American Fisheries Society and Wildlife Society AZ/NM chapter meetings and Colorado River Aquatic Biologists Annual Meeting, 1-2 peer reviewed journal articles by AGFD during this work plan on issues related to Colorado River fishes.

Project Element I.2. Estimating Kinship and Spawner Abundance of Warm-Water Nonnatives (New Study; Partially Funded)

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The objective of this project element is to use kinship genetic analysis (Diver and others, 2021; Schunter and others, 2014; Städele and Vigilant, 2016) to better understand the ongoing expansion of smallmouth bass into Grand Canyon, determine the extent to which juveniles are locally produced or entrained from the reservoir, and potentially estimate spawner abundance and survival for species being removed from the system via close-kin mark recapture analysis (Bravington and others, 2016; Waples and Feutry, 2022). This work will provide important information on reproductive dynamics and larval sources for a high-risk species that has the potential to prey on ESA-listed species if it disperses downstream.

These data will inform discussions of management strategies, current stage of invasion, and off-ramps as laid out in the Invasive Species Strategic Plan (Smallmouth Bass Ad Hoc Group, 2023). Data on kinship and nest location, when paired with data on hatch dates (see Experimental Fund Project Element I.6) can provide information on the effectiveness of LTEMP SEIS Flow Experiments to disadvantage recruitment and survival of smallmouth bass (U.S. Department of the Interior, 2024).

These data could also be used with other data sources in the development of LTEMP Performance Metrics that are focused on detecting the reproduction of low, medium, high, and very-high risk nonnative aquatic species downstream from Glen Canyon Dam.

Hypotheses and Science Questions

- I.2.1: What proportion of smallmouth bass captured downstream from Glen Canyon Dam are locally produced or entrained from Lake Powell?
- I.2.2: Is there spatial structure in kinship and are there hotspots with higher densities of related individuals?
- I.2.3: How many parents produced the juveniles, and does genetic data indicate monogamous or polygamous mating in smallmouth bass?
- I.2.4: How comparable are population abundance estimates using close-kin mark-recapture and traditional mark-recapture using green sunfish as a surrogate? Can we use close-kin analysis to obtain accurate estimates of abundance for species that must be sacrificed/removed from the CRe?

Methods

I.2.1. What Proportion of Smallmouth Bass Captured Downstream from Glen Canyon Dam are Locally Produced or Entrained from Lake Powell?

Kinship analysis relies on genetic samples from individuals to provide estimates of the number of full siblings (same mother and father), half siblings (same mother or father), and other relationships (e.g., parent offspring pairs) (Diver and others, 2021; Schunter and others, 2014; Städele and Vigilant, 2016). Given large enough sample sizes, kinship analysis can also be used to fit close-kin mark-recapture models that estimate spawner abundance and survival even when individuals are being removed from the population (e.g., for exploited populations, or in the case of smallmouth bass that are being actively removed) (Bravington and others, 2016; Waples and Feutry, 2022). Even if sample sizes are not sufficient for these types of estimates, kinship grouping (e.g., did most individuals derive from a small or large number of parent pairings) and spatial patterns in grouping (e.g., are most larval fish captured near the slough siblings or half-siblings [indicating local recruitment], or do individuals removed from near the dam represent a greater number of parent pairings [suggesting entrainment]) could help to better clarify nest

location for smallmouth bass. Data on kinship grouping and spatial patterns are particularly useful when combined with other ongoing studies using side scan sonar to search for nests and deploying artificial nest substrate to attract nest-building males (pilot work in FY 2024, not anticipated to continue into the FY 2025-27 TWP).

This project element relies on samples collected during smallmouth bass removal efforts conducted by the NPS (GLCA and GRCA) and routine trout monitoring trips by AGFD and the USGS (Project Elements H.1, H.2). Per NPS permit, any smallmouth bass captured downstream from Glen Canyon Dam must be removed, humanely euthanized via Institutional Animal Care and Use Committee (IACUC)-approved methods, and preserved in 95% ethanol for scientific use (e.g., genetic, diet, and hatch date analysis). These specimens provide the source DNA from which kinship can be determined using genetic markers like microsatellites or SNPs. Microsatellite markers for smallmouth bass have been developed (Malloy, Jr and others, 2000), and this is a proven method to identify individual genotypes and their variation among individuals for other species of fish in the Southwest (Turner and others, 2009).

I.2.2. Is There Spatial Structure in Kinship and are there Hotspots with Higher Densities of Related Individuals?

This project aims to understand whether there are hot spots of full siblings that could help managers pinpoint nest locations.

In FY 2023, Reclamation funded pilot work outside the GCDAMP to understand the kinship of smallmouth bass removed in 2022 to determine how related individuals were to each other.

GCMRC sent genetic material from these smallmouth bass to collaborators (Drs. Megan Osborne and Thomas Turner) at the University of New Mexico, Turner Aquatic Conservation Lab to test whether analysis using microsatellite markers could provide estimates of relatedness among individual smallmouth bass. Preliminary analysis indicated there were 94 full sibling pairs, and most of those pairs were dispersed across sampling sites (Figure 3). While there was no broad evidence of spatial structure in kinship, a larger number of sibling pairs were found at RM -14.74 and -14.72 in the Lees Ferry Reach (Figure 3), which may suggest habitat conditions were more suitable in those locations. Most of the 2022 samples analyzed were collected near the dam, so information on parentage would provide greater clarity on whether there were many nests located close to the dam or if juveniles were entrained and survived passage through the penstocks.

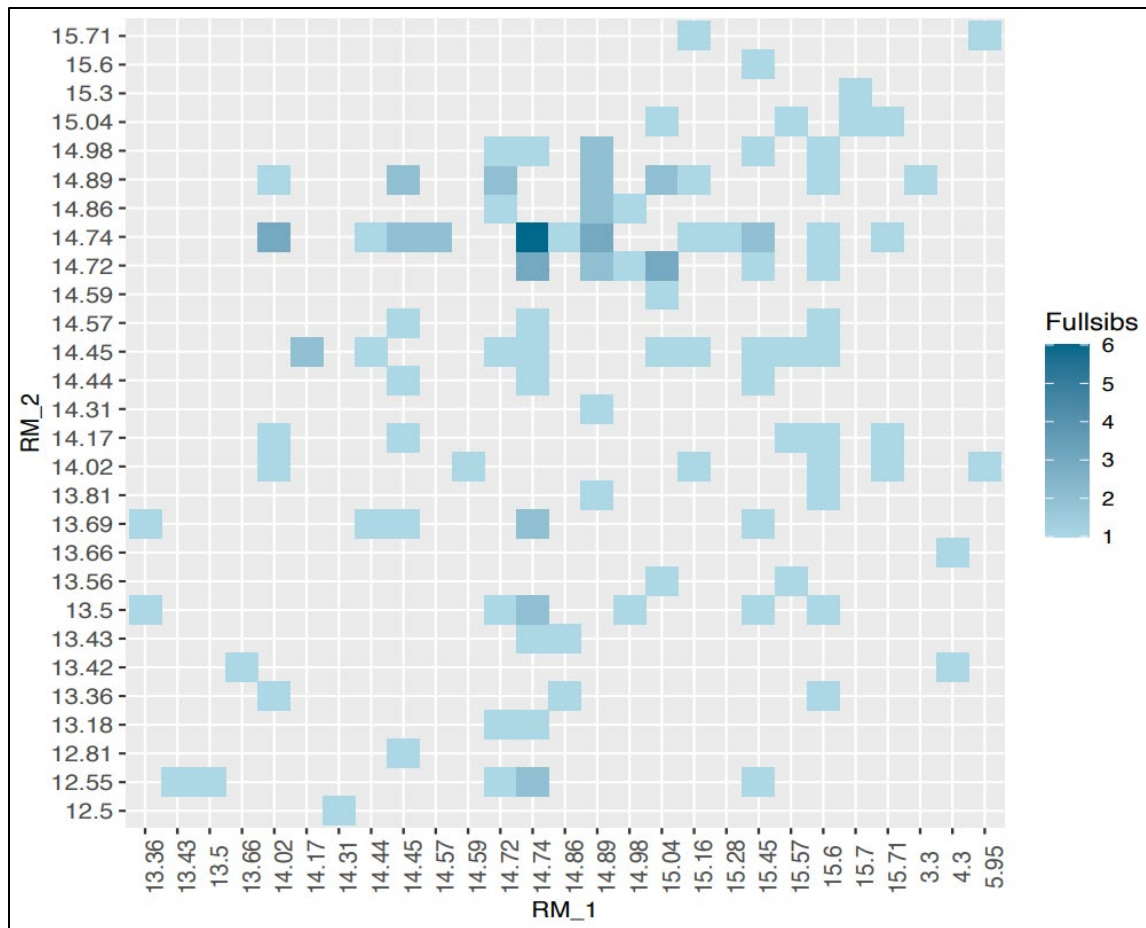


Figure 3. Count of total number of full siblings identified at each river mile pairwise comparison (blue squares). Deeper blue color indicates a larger number of pairs found at that RM comparison. Grey squares reflect zeros. Data do not indicate whether juveniles were produced in Lake Powell or downstream from Glen Canyon Dam (Huachan Liang, Thomas Turner, and Megan Osborne, University of New Mexico, provisional data – do not cite).

1.2.3. How Many Parents Produced the Juveniles, and does Genetic Data Indicate Monogamous or Polygamous Mating in Smallmouth Bass?

Kinship analysis requires sampling of only one generation, whereas parentage analysis requires information from both the offspring and parents. Parentage analysis will provide data on how many pairs of adult smallmouth bass produced the juveniles (i.e., # of nests), and whether mating follows monogamous (Franckowiak and others, 2017) or polygamous relationships (i.e., full siblings vs. half siblings) that can affect the number of nests produced. Preliminary kinship data from 2022 with 192 samples from the Lees Ferry reach and 6 samples from Lake Powell indicated there were 94 unique full sibling clusters (indicating monogamy; Figure 3), and 5 half sibling pairs detected (indicating polygamy), but more data is needed beyond the initial year of sampling.

If guarding males are primarily monogamous and females typically only mate with one male, then there may be fewer opportunities for reproductive adults to build and maintain nests. In 2022 and 2023 very few adults were captured in the Lees Ferry reach, and few samples were taken from the reservoir. This work would benefit from additional genetic sampling of parents and juveniles in Lake Powell near Glen Canyon Dam, as well as continued analysis of individuals captured through monitoring and removal efforts in the CRe. If this genetic technique is successful for smallmouth bass, we could expand this approach to other nonnative fish species with increasing catch rates (e.g., walleye).

I.2.4. How Comparable are Population Abundance Estimates using Close-Kin Mark-Recapture and Traditional Mark-Recapture using Green Sunfish as a Surrogate? Can we use Close-Kin Analysis to Obtain Accurate Estimates of Abundance for Species that must be Sacrificed/Removed from the CRe?

Last, this element will explore the use of close-kin mark-recapture analysis to generate effective population size estimates for nonnative fishes, even when individuals are being removed from the river (Bravington and others, 2016; Waples and Feutry, 2022). We propose to analyze genetic samples from adult and juvenile green sunfish in the Lees Ferry reach using parentage and kinship analysis, and if possible fit close-kin mark-recapture models to compare estimates of abundance and survival to estimates from a concurrent traditional mark-recapture study. We would like to pursue this paired approach because it is difficult to obtain estimates of abundance with removal data, and we may not have enough statistical power to detect population change unless a large decrease in a nonnative fish occurs. Thus, using close-kin information may represent an innovative method to obtain abundances with removal data that could lead to improvements in model precision.

Close-kin mark-recapture models are still fairly novel within the broader field, so this comparison will help us better understand the potential to use this technique for other rarer species (e.g., smallmouth bass, walleye).

The traditional mark-recapture analysis would rely on PIT-tagging and releasing green sunfish captured as part of the AGFD/TRGD trout monitoring project described in Project Elements H.1 and H.2.

Anticipated Use of Data

These data will be used to generate information on the number of smallmouth bass nests in the Lees Ferry reach and an estimate of the proportion of entrainment vs. local reproduction of smallmouth bass. These data will also be used to compare close-kin mark recapture estimates to traditional mark-recapture abundance estimates for a surrogate species (green sunfish) to determine whether genetic tools can be used to estimate abundance for smallmouth bass that are actively being removed from the study reach.

This information can be used by Reclamation and GCDAMP stakeholders to understand (partially) the effectiveness of management actions to control smallmouth bass.

Outcomes and Products

Outcomes

- Estimation of the number of smallmouth bass nests that originated in the Lees Ferry reach or in Lake Powell. Estimation of sibship, parentage, and genetic diversity in juvenile smallmouth bass. Determination of monogamous or polygamous mating, which could affect overall number of nests maintained. Use of close-kin mark-recapture analysis to estimate spawner abundance and effective population size in smallmouth bass and green sunfish (the latter will be compared to traditional mark-recapture analysis).

Products

- Manuscript describing smallmouth bass reproductive dynamics, including information on sibship, parentage, mating strategies, and estimates of spawner abundance and effective population size.
- If genetic data show promise, manuscript describing close-kin mark recapture analysis, paired with traditional mark-recapture analysis to estimate population size/abundance for green sunfish, and its potential use for smallmouth bass and walleye.
- Presentation of results at annual reporting meetings and at a scientific conference.

Project Element I.3. Identifying Emerging Threats to the Colorado River Ecosystem Using Environmental DNA (Modified Study; Partially Funded)

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The objective of this project element is to conduct biosurveillance of nonnative species by collecting water samples and using molecular tools to identify high risk species of fish, crayfish, gastropods, and mollusks that are in the early stages of expansion or those that evade capture by traditional sampling techniques (National Park Service, 2021).

All projects proposed in this element are intended to provide data to resource managers that could inform decision-making as it relates to the NPS Expanded Nonnative Species Management Plan (National Park Service, 2018, 2021) and the Invasive Species Strategic Plan (Smallmouth Bass Ad Hoc Group, 2023). This includes deploying new technology in the Lees Ferry reach to automatically sample water on a set schedule to detect nonnative species, pairing eDNA data with traditional sampling gear to understand its sensitivity as a surrogate for detection and relative abundance data, estimating entrainment potential for high-risk species in Lake Powell, and monitoring Asian tapeworm (*Bothriocephalus acheilognathi*) and *Lernaea* parasites in humpback chub per the 2016 Biological Opinion (U.S. Department of the Interior, 2016a).

These data could also be used with other data sources in the development of LTEMP Performance Metrics that are focused on detecting the presence of low, medium, high, and very-high risk nonnative aquatic species downstream from Glen Canyon Dam. In addition, eDNA data from Experimental Fund Project Element I.5 can provide smallmouth bass detection data to inform planning and evaluation of the effectiveness of LTEMP SEIS Flow Experiments to disadvantage recruitment and survival of smallmouth bass (U.S. Department of the Interior, 2024).

Hypotheses and Science Questions

- I.3.1: What high-risk species in Lake Powell are moving into the Colorado River undetected by traditional sampling techniques?
- I.3.2: How comparable are eDNA methods to traditional sampling gear for detecting sources and dispersal of nonnative fishes in the Colorado River? In comparison to currently used methods, how effective and feasible is using eDNA techniques to monitor changes in distribution and relative abundance of the fish assemblage in relation to dam operations and environmental conditions in the CRe?
- I.3.3: How do the biophysical characteristics of Lake Powell affect the likelihood of nonnative fish entrainment through Glen Canyon Dam?
- I.3.4: What is the incidence of parasite infestation in humpback chub in the Little Colorado River and mainstem Colorado River? Can a molecular assay be developed to reduce handling time and expand the scope of monitoring?

Methods

I.3.1. What High-Risk Species in Lake Powell are Moving into the Colorado River Undetected by Traditional Sampling Techniques? (Unfunded)

Responding quickly to nonnative species introductions before populations become large and established is the least expensive and most effective way to control nonnative species (Leung and others, 2002).

Environmental DNA has become a reliable and cost-effective tool that resource managers can use to detect rare nonnative species prior to population expansion (Pilliod and others, 2013; Klymus and others, 2015; Lacoursiere-Roussel and others, 2016; Pochardt and others, 2020, Spear and others, 2021, Yates and others, 2021; Rourke and others, 2022). In aquatic environments, fish continually shed cellular material into the water via reproduction and feces that can persist in the environment for several weeks. This cellular material can be collected via water sample and DNA can be extracted from cells collected in the environment in which an organism lives, rather than directly from the animals themselves (Pilliod and others, 2013; Klymus and others, 2015; Sepulveda and others, 2019). This reduces or eliminates fish handling and stress (which can be a concern for non-target species) per Tribal concerns and allows managers to detect species at the early stages of expansion, species with low susceptibility to capture, or species residing in habitat areas outside of the range of standard sampling methods.

The warm-water nonnative aquatic species currently present in the Colorado River are only a fraction of the potential nonnative fish species that could establish, especially if fishes continue to be entrained from Lake Powell into warming Colorado River waters. New, emerging threats to the CRe, such as new nonnative species, parasites, or disease, also have the potential to affect fish and wildlife populations and human health. In the FY 2025-27 TWP, we propose to use automatic sampling technology to collect water samples to screen for emerging threats to the CRe. This will include use of an eDNA Autosampler (Smith Root; <https://www.smith-root.com/services/training/environmental-dna-field-sampling-techniques>) to automate sample collection in the field, which will allow us to ‘catch’ aquatic nonnative species year-round every few days even when trips are not on the water. The autosampler will be deployed in an area known to be a hot spot for nonnative species (e.g., downstream from dam, in slough) and/or within the draft tubes of the penstocks, and samples will be retrieved every two to four weeks.

Self-preserved filters will be extracted and analyzed in a controlled laboratory setting using high-throughput quantitative Polymerase Chain Reaction (qPCR), a highly sensitive method that can detect small amounts of DNA in the water column.

Filtered samples will be run using a new Aquatic Invasive Species (AIS) biochip developed by the U.S. Forest Service National Genomics Center for Wildlife and Fish Conservation, which will allow for quantitative, species-specific screening of very high risk or high-risk species that reside in Lake Powell (National Park Service, 2018, 2021). The AIS biochip currently contains 42 of the most problematic nonnative species in the western United States but is planned to be refined in FY 2025 to focus on species most concerning to managers in the Colorado River. Should we detect a new species, or expanded coverage of an existing known species, we will provide this information to GLCA, GRCA, and AGFD for management purposes. These extracted samples can be stored as a library for decades, so as technology develops more data can be obtained from the samples over time (e.g., develop time series, conduct deeper sequencing/kinship analysis, or facilitate collaboration on other questions).

While qPCR will not occur in the field using this instrument, technology is improving rapidly, and this instrument represents a step towards automated sampling and processing in the field. If deployed, this autosampler will be part of a network of pilot eDNA autosamplers that will be installed in the Colorado River basin and in other western US rivers starting in summer 2024 as part of the USGS Rapid environmental (e)DNA Assessment and Deployment Initiative & Network (READI-Net). This pilot initiative aims to accelerate the use of eDNA as a best practice for the early detection of aquatic biological threats and delivery of tools and strategies to resource managers.

I.3.2. How Comparable are eDNA Methods to Traditional Sampling Gear for Detecting Sources and Dispersal of Nonnative Fishes in the Colorado River? (Partially Funded)

A second objective of Project Element I.3 aims to increase biosurveillance of nonnative aquatic species throughout the Colorado River and its tributaries. We also wish to determine the degree to which eDNA sampling is comparable to traditional sampling methods in a large river system and identify its limitations and challenges at this phase in the technology's development. If this method is successful (either now, or with further development), it could be used to target locations for nonnative fish sampling, reduce the amount of hands-on sampling, and/or reduce the number of trips on the water by using automated eDNA sampling to track the distribution and relative abundance of fishes in the CRE.

In 2021 and 2022, GCMRC leveraged outside funding to collect eDNA samples from throughout the river corridor and in major tributaries. We were successful in detecting smallmouth bass, walleye, channel catfish, brown trout (*Salmo trutta*), green sunfish, and razorback sucker and/or razorback/flannemouth sucker hybrids using eDNA. However, key questions remain in how comparable this molecular tool is to detecting fishes captured using traditional sampling gear, and whether estimates of relative abundance (using eDNA copy number [concentration]) are comparable to catch rates for native and nonnative species (Pilliod and others, 2013; Lacoursiere-Roussel and others, 2016; Pochardt and others, 2020; Spear and others, 2021; Yates and others, 2021; Rourke and others, 2022). We seek to provide clarity on that question in the FY 2025-27 TWP by collecting eDNA samples in tandem with regular monitoring trips. The cost of eDNA sampling and processing (salary, logistics, travel, operating expenses, analytical costs) is incurred in Project Element I.3 and not in the projects below.

Seining backwaters in the mainstem Colorado River (funded in FY 2025 only)

Project Element G.5 adds an additional boat and boatman to the Fall USFWS Humpback Chub Aggregations Monitoring trip in September 2025 to search for nonnative species by seining backwaters during the warmest month of the year. Backwaters are often ephemeral and dependent on flows, but previous trips have targeted approximately 200 backwaters throughout Grand Canyon. We will collaborate with the USFWS by collecting up to three eDNA samples from selected backwaters where at least three seining passes are conducted.

This work will allow us to compare a traditional sampling technique with molecular detection tools and determine whether we are fully cataloging nonnative species that utilize warm, isolated habitats throughout the canyon. If these data indicate there are species evading capture through seining, it could lead to a revision of the seining strategy used to capture fishes in backwaters. These samples will also be used to determine whether smallmouth bass are utilizing backwaters in Grand Canyon National Park, which has direct management implications. In addition, if species detection and relative abundance data from eDNA is comparable to seining data, we may be able to use eDNA as a citizen science tool to help monitor backwaters for nonnative fishes in Grand Canyon.

Seining pools in the upper Little Colorado River drainage (Unfunded)

Project Element G.8 monitors spring-fed pools in Piute Canyon in the upper LCR drainage to understand the extent to which humpback chub utilize these pools (and under what hydrologic conditions), as well as identify sources of nonnative species that could be flushed downstream into the LCR during monsoon season. Detection of very high or high-risk nonnative fishes could lead to Tribal management efforts. We plan on collecting three replicate eDNA samples from each paired LCR pool site in FY 2025 in collaboration with seining in Project Element G.8.

Electrofishing in the Lees Ferry Reach (funded in FY 2026 only)

Project Element H.2 monitors the rainbow and brown trout population in the Lees Ferry reach through the Trout Reproductive and Growth Dynamics (TRGD) project, which will be combined with the AGFD trout monitoring project in this work plan (Project Element H.1). The purpose of this sampling is to compare species detection and relative abundance in sites 1A & 1C by deploying a separate eDNA boat to passively collect water samples in each 250-m reach while electrofishing is occurring.

We plan on collecting three replicate samples from 48 reaches split between the upper site (1A) and lower site (1C) during a trip in FY 2026. Species detection and relative abundance data using eDNA copy number will be compared to species catch and abundance data from the combined trout monitoring project to assess the comparability of electrofishing and molecular data collected from sites in Lees Ferry.

Electrofishing, hoop netting, and antennas in the mainstem Colorado River (funded in FY 2027 only)

Project Element I.1 collects system-wide native and nonnative fish monitoring data at random sites located from Lees Ferry to Pearce Ferry. The fish taxa change drastically throughout Grand Canyon, with trout dominating the catch near the dam and native fish dominating in western Grand Canyon. The purpose of this eDNA sampling is to compare fish detection and relative abundance (copy number) data to catch data collected during electrofishing, hoop netting, and passive PIT tag antenna sampling.

We plan on collecting one eDNA sample per reach during electrofishing operations (so as not to delay the pace of sampling) and one sample per reach near deployed hoop nets or portable PIT antenna arrays during a trip in Spring FY 2027 when turbidity is low.

I.3.3. How do the Biophysical Characteristics of Lake Powell Affect the Likelihood of Nonnative Fish Entrainment through Glen Canyon Dam? (Unfunded)

A third objective of I.3 aims to understand how the biological and physical aspects of Lake Powell affect the likelihood that high-risk nonnative fish will be entrained through the penstocks of Glen Canyon Dam and deposited into the river below. This element builds off previous work recently funded by Reclamation that used traditional sampling gear (gill nets, minnow traps, Ichthyoplankton tows, light traps) combined with acoustic telemetry and hydroacoustic surveys to characterize the fish community at depth inhabiting the forebay, confluence, and Wahweap in 2022 and 2023. This sampling occurred over eight sampling events when the reservoir was mixed in winter and stratified from late spring to early fall. Notably, the authors found that 99% of the fish fauna in the Lake Powell forebay was nonnative, and the depth of the epilimnion increased during large inflow years, which increased the probability of occupancy at depth for species such as smallmouth bass. However, smallmouth bass tended to occupy shallower depths in the epilimnion than species such as channel catfish (B. Friesen, Utah State University, unpub. data, 2024).

This study noted that additional avenues of research could comprehensively describe the threat of entrainment by releasing PIT-tagged fish in the forebay at the depth of the penstocks to track survival and movement of tagged fishes downstream. They noted additional hydroacoustic surveys are a cost-effective method to determine fish sizes and densities during years in which the epilimnion is close to the penstocks, which could be used to monitor entrainment-likely fishes.

Last, the authors indicated they did not capture large schools of suspected threadfin shad seen in hydroacoustic surveys via gill nets, but this could be accomplished using trawl nets in the Forebay to confirm species identity (B. Friesen, Utah State University, unpub. data, 2024).

In the FY 2025-27 TWP, we propose to build off this previously funded work by collecting eDNA samples in parcels of water at depth in the Forebay with a Van Dorn sampler, which is used for Lake Powell water quality monitoring and is designed to take samples in lakes and stratified water bodies. We plan to base our depths of sampling on distinct zones of water quality that develop within the reservoir based on reservoir elevation and inflows, including the depth of chlorophyll maximum, dissolved oxygen minimum, the photosynthetically active radiation zone, and water temperature (Figure 4). We will account for movement of water at depth due to the seiche (internal wave) by estimating the period of the seiche and sampling at the same point within each wave. Further, we plan on sampling during the day and at night to capture variation in penstock-entrainable fishes in the lower epilimnion due to diel migration of zooplankton.

During daylight hours zooplankton descend to the bottom of the photic zone to avoid predation by sight-seeing predators, but ascend to the warmer, more nutrient and prey-rich surface layers at night to feed when the predatory threat decreases. We hypothesize that forage fishes and their predators (e.g., smallmouth bass) may descend into deeper depths early in the morning as they follow their food source deeper into the reservoir, potentially leading to higher levels of entrainment at different times of the day.

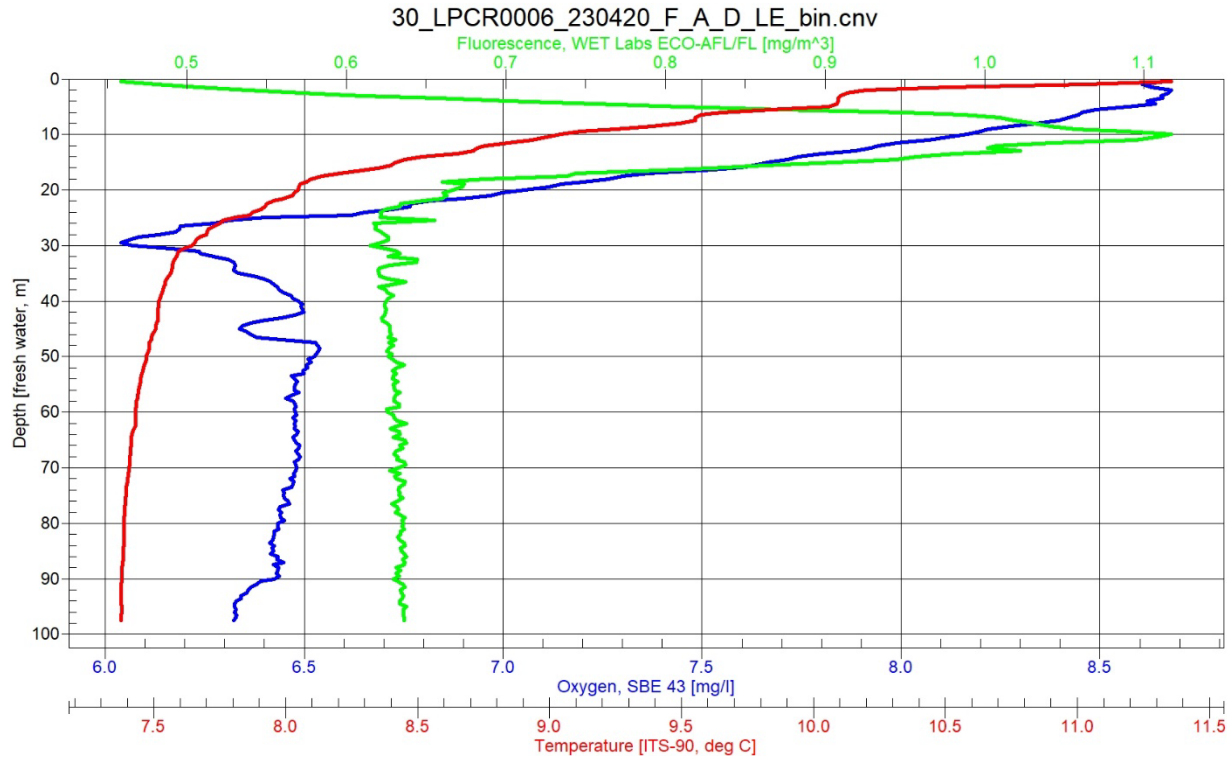


Figure 4. Water quality profile at depth in the forebay of Lake Powell in April 2023 showing chlorophyll (florescence, in green), dissolved oxygen (blue), and water temperature (red) as an example of our approach to selecting sampling depths (Bridget Deemer, USGS, provisional data – do not cite).

We propose to collect water samples for eDNA 1-2 m from the surface, top of chlorophyll maximum (~5 m), bottom of chlorophyll maximum/DO minimum (~15 m), above the penstocks (~25 m, depending on elevation), at the penstocks, and below the penstocks in increments of 10 m (e.g., ~35 and 45 m; Figure 4). We will sample eDNA at depth in both bank (slope) and thalweg locations during spring (April) and fall (August/September) in FY 2026, when the reservoir is stratified. These data will be paired with eDNA samples taken from penstock water and from the base of the dam. Penstock water was sampled in June 2021 and were negative for smallmouth bass and walleye, indicating the DNA of fish upstream was not being transferred downstream (at those reservoir elevations) and being deposited in the Lees Ferry reach. This sampling approach could be paired with ongoing pilot studies to examine entrainment including future hydroacoustic surveys and/or tagged fish released in the Forebay near the penstocks.

This project is intended to provide information to the U.S. Bureau of Reclamation that could inform management actions related to curtailing nonnative fish passage through the penstocks per the 2016 Biological Opinion (e.g., thermal curtain, fish exclusion barrier net in Lake Powell, tucker trawl net downstream from Glen Canyon Dam).

I.3.4: What is the Incidence of Parasite Infestation in Humpback Chub in the Little Colorado River and Mainstem Colorado River? Can a Molecular Assay be Developed to Reduce Handling Time and Expand the Scope of Monitoring? (Partially Funded)

The fourth objective of Project Element I.3 is to continue to monitor Asian fish tapeworm and anchor worm (*Lernaea*) in humpback chub as specified in the Environmental Commitments Section in the LTEMP ROD (and 2016 Biological Opinion) (U.S. Department of the Interior, 2016a, b). To date, Asian fish tapeworm monitoring has been conducted periodically in the Little Colorado River and in the mainstem Colorado River by the USGS in collaboration with USFWS (Figures 5a, 5b), and anchor worm presence has been recorded through exterior visual inspection of fish on USFWS, USGS, NPS, and AGFD fish trips.

In the FY 2025-27 TWP, we propose to continue periodic monitoring to assess infestation of parasites in humpback chub via traditional sampling methodology, but we also propose to develop a molecular assay to detect the presence of those parasites in water and fish fecal matter, and then test the sensitivity and selectivity of the assay using a paired cross-over study.

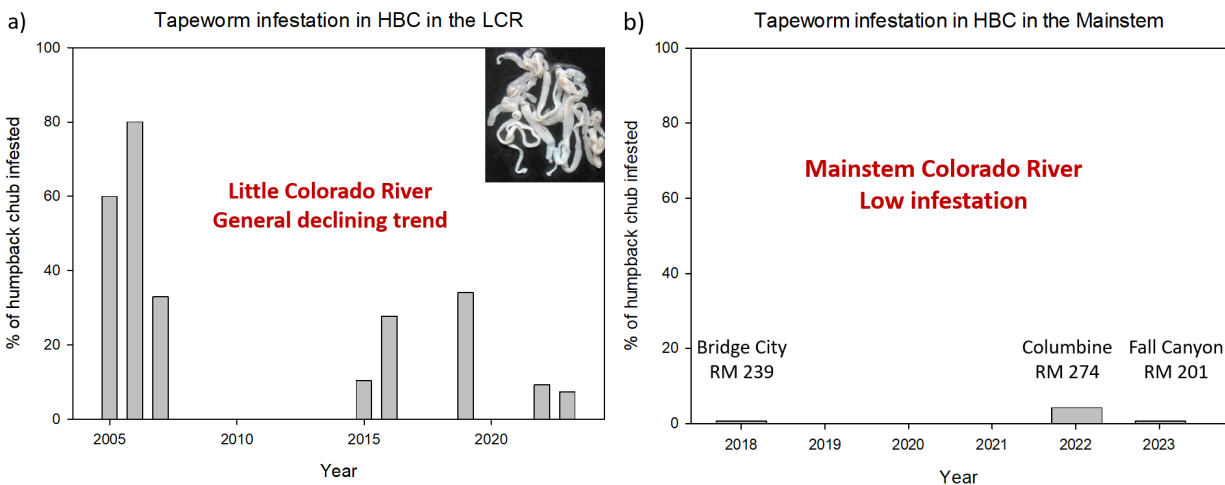


Figure 5. Proportion of humpback chub, *Gila cypha*, infected by Asian tapeworm in the Little Colorado River (a) and in the mainstem Colorado River (b) from 2005-2023 based on periodic sampling (David Ward, U.S. Fish and Wildlife Service, provisional data – do not cite).

Traditional sampling methodology

Asian fish tapeworm is a nonnative species that infests warm-water cyprinid fish. This parasite has been identified as one of six potential threats to the continued existence of endangered humpback chub (U.S. Fish and Wildlife Service, 2002), as it is potentially fatal to new host species (Hoffman and Schubert, 1984). Asian fish tapeworm was first documented in the LCR in Grand Canyon in 1990 (Minckley, 1996) and was hypothesized to be a cause of long-term declines in the condition of adult humpback chub from the LCR (Meretsky and others, 2000), but was rarer in the mainstem due to cold water temperatures (Hoffnagle and others, 2006). Periodic monitoring for the presence of this parasite commenced in 2005 and the proportion of humpback chub infested with tapeworm in the LCR has generally declined over time (Figure 5a).

Infestation of humpback chub in the mainstem CRe has been comparatively low in western Grand Canyon and only recently sampled (Figure 5b). Asian tapeworm and *Lernaea* (anchor worm) monitoring will continue in this work plan as specified in the 2016 Biological Opinion (U.S. Department of the Interior, 2016a) in collaboration with the USFWS (Project Elements G.2, G.5) and on JCM-East trips (Project Element G.3). Briefly, approximately 40 humpback chub will be captured and placed in separate containment baskets on the banks of the river for 48 hours while they are being treated with a bath of Praziquantel, which triggers movement of Asian tapeworm out of their digestive systems. Individual worms will be collected from baskets and proportion of humpback chub infested calculated. Post-treatment, humpback chub (now free of parasitic tapeworms) will be released unharmed into the river from which they were collected. *Lernaea* will be visually identified in the field and recorded by all fish trips, but these parasitic copepods will not be removed from fish as their removal can cause external tissue damage that could lead to secondary infection.

Proposed molecular sampling methodology (Unfunded)

Monitoring parasites and disease in humpback chub in the Little Colorado River and in the mainstem Colorado River is included as a Conservation Measure in the 2016 Biological Opinion (U.S. Fish and Wildlife Service, 2016). At the present time, Asian tapeworm is monitored by collecting and holding humpback chub in a Praziquantel bath for 48 hours on the banks of the Colorado River. *Lernaea* (anchor worm) is monitored through visual inspection of humpback chub and other native and nonnative fishes.

In this work plan, we propose to collaborate with Oregon State University (Dave Lytle and Justin Sanders labs) to develop molecular assays for Asian tapeworm and *Lernaea*, which could eventually replace current detection methods. These assays would be used to test water and fecal eDNA samples from humpback chub to determine presence and infection intensity of parasitized fish. This assay will need to be sensitive enough to detect actual infestation of the definitive host (fish), rather than infestation of the intermediate host (copepods) consumed as a diet item during its procercoid life stage.

We envision a cross-over study using fish treated with Praziquantel but also tested/swabbed for the eDNA of parasites. If successful, moving to molecular-based methods would allow for more efficient sampling, less stress to the fish, more information, and the ability to collect samples throughout the canyon (not just in one or two spots). Asian tapeworm requires a temperature of $>20^{\circ}\text{C}$ to mature, and anchor worm requires $>17^{\circ}\text{C}$, so a good understanding of where parasitic load is highest may help us anticipate changes in parasitism if the Colorado River continues to warm.

Anticipated Use of Data

These data will be used to detect new species introductions into the CRe, or significant increases or expansions of existing nonnative aquatic species in the river. Paired detection and relative abundance data from eDNA and traditional sampling gear will be used to understand whether eDNA monitoring of water samples could be used as a surrogate for hands-on fish sampling in future work plans. Data from Lake Powell at depth will provide information on species that could be entrained through the penstocks relative to biological and physical processes within the lake. Last, we will at minimum gain information on parasite presence and infection intensity in humpback chub in the Little Colorado River and mainstem Colorado River, with the potential for developing a molecular tool that could increase sampling coverage and reduce handling time of humpback chub.

Outcomes and Products

Outcomes

- Refinement of Colorado River-specific high-throughput qPCR biochip for Aquatic Invasive Species known in Lakes Powell and Mead. Expanded detection capability for nonnative species from Glen Canyon Dam to Pearce Ferry, with cross-over comparison of detection and relative abundance data from traditional and eDNA sampling methods. Improved understanding of entrainment potential for nonnative species in Lake Powell. Estimation of parasite infestation and load in humpback chub in the Little Colorado River and mainstem Colorado River.

Products

- Manuscript describing results from pilot project to deploy eDNA autosampler in the Lees Ferry reach. This manuscript will include an analysis of detection data in each month of the year in FY 2025 that will refine future deployment schedules (Unfunded).
- Manuscript describing results from paired eDNA and traditional sampling year study, which includes seining, electrofishing, hoop netting, and PIT tag antennas.

- Manuscript describing detection of aquatic nonnative species DNA at depth in Lake Powell relative to the penstocks, to provide greater understanding of what species might be entrained relative to Lake Powell biophysical characteristics (Unfunded).
- Analysis of Asian tapeworm and *Lernaea* data in humpback chub in the LCR and CRe, for reporting purposes at the annual reporting meeting.
- Manuscript describing parasite detections using Praziquantel vs. detection and quantification of parasites using eDNA and molecular markers (Unfunded).
- Presentation of results at annual reporting meetings and at a scientific conference.

Project Element I.4. Modeling Population Dynamics and Improving Forecasting Tools for Smallmouth Bass and Other Nonnative Fishes (New Study)

Charles Yackulic, Drew Eppheimer, Kimberly Dibble, Maria Dzul, Brian Healy, Eric Frye, Kate Behn, Lindsay Hansen, Ben Miller

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The objective of this project element is to focus on modeling nonnative fish data and reporting results to GCDAMP stakeholders, combined with laboratory work to understand how turbidity influences growth (through energy assimilation) and survival of smallmouth bass.

Modeling projects in this project element include updating forecasting tools for smallmouth bass [if new data or information is available that may improve models], evaluating the effectiveness of management actions including LTEMP Experimental Flows (U.S. Department of the Interior, 2024) on smallmouth bass growth, catch rates, dispersal, and recruitment, and developing a workflow process for an occupancy model that evaluates data from Projects H, G, and I to report on a standard set of LTEMP Performance Metrics that evaluate the presence and reproduction of nonnative aquatic species (Resource Goal #10).

This project element also supports laboratory trials to determine the effects of turbidity and water temperature on the growth and survival of early life stages of smallmouth bass and other nonnative fishes, which is data that may be used to update the smallmouth bass population growth model (Eppheimer and others, 2024). Last, this element allocates time for a GCMRC Research Fish Biologist to lead development of a science plan to support data collection for smallmouth bass sampling and removals in the Lees Ferry reach, and provides time to analyze nonnative fish data, communicate findings to GCDAMP stakeholders, participate in Rapid Response Tech Team and SBAHG calls, and provide science to support discussions of offramps laid out in the Invasive Fish Species Strategic Plan (Smallmouth Bass Ad Hoc Group, 2023).

Hypotheses and Science Questions

- I.4.1: How does turbidity and temperature affect feeding efficiency and survival of early life stages of smallmouth bass? How does adding turbidity as a parameter in the smallmouth bass population growth model affect performance?
- I.4.2: How effective are management actions (LTEMP Flow Experiments, removals) at reducing smallmouth bass population growth, survival, dispersal, and reproduction?
- I.4.3: What are the long-term drivers of distributional changes in native and nonnative fishes in the CRe?

Methods

I.4.1. How does Turbidity and Temperature Affect Feeding Efficiency and Survival of Early Life Stages of Smallmouth Bass? How does Adding Turbidity as a Parameter in the Smallmouth Bass Population Growth Model Affect Performance?

The first objective of Project Element I.4 is to determine how turbidity and temperature affect the growth of early life stages of smallmouth bass in laboratory feeding trials. Smallmouth bass are visual predators that can effectively forage in clear water (Brown and others, 2009), but higher turbidity and total suspended solids in the water column downstream from the Paria and Little Colorado rivers have the potential to decrease foraging efficiency and increase mortality, particularly for early life stages (Sweka and Hartman, 2003; Suedel and others, 2017). High turbidity near the Little Colorado River has been measured at a level known to affect foraging in other sight-feeding predators (Yard and others, 2011), but smallmouth bass in the upper basin have successfully invaded turbid river reaches and have had population-level effects on native species (e.g., Bestgen and others, 2018; Dibble and others, 2021; Martinez and others, 2014).

This laboratory study seeks clarification on whether turbidity in the Colorado River could limit downstream dispersal and growth of this aquatic nonnative species. This work will directly inform any potential updates to the smallmouth bass population growth model developed by Eppehimer and others (2024). The current model assumes water temperature is the key factor influencing the smallmouth bass population growth rate (i.e., through spawning timing/occurrence, age-0 fish growth, and age-0 fish survival) and is currently being used to forecast population-level responses under different reservoir operations and experimental flow regimes (Eppehimer and others, 2024). However, this model could be modified to add turbidity if it has a strong effect on early life stage growth and survival based on laboratory experiments.

I.4.2. How Effective are Management Actions (LTEMP Flow Experiments, Removals) at Reducing Smallmouth Bass Population Growth, Survival, Dispersal, and Reproduction?

The second objective of Project Element I.4 is to evaluate whether management efforts to reduce smallmouth bass catch rates, dispersal, growth, reproduction, and recruitment through removals and/or LTEMP Flow Experiments are working as designed. In the FY 2025-2027 TWP, we will analyze nonnative fish data provided by federal and state partners (e.g., AGFD, NPS, USFWS) that are compiled into a single nonnative fish spreadsheet. Analysis of these data will vary by research question, but we will calculate catch-per unit effort and possibly use close-kin mark-recapture analysis (Project Element I.2) to determine whether the smallmouth bass population is increasing in GLCA and GRCA (Figure 6).

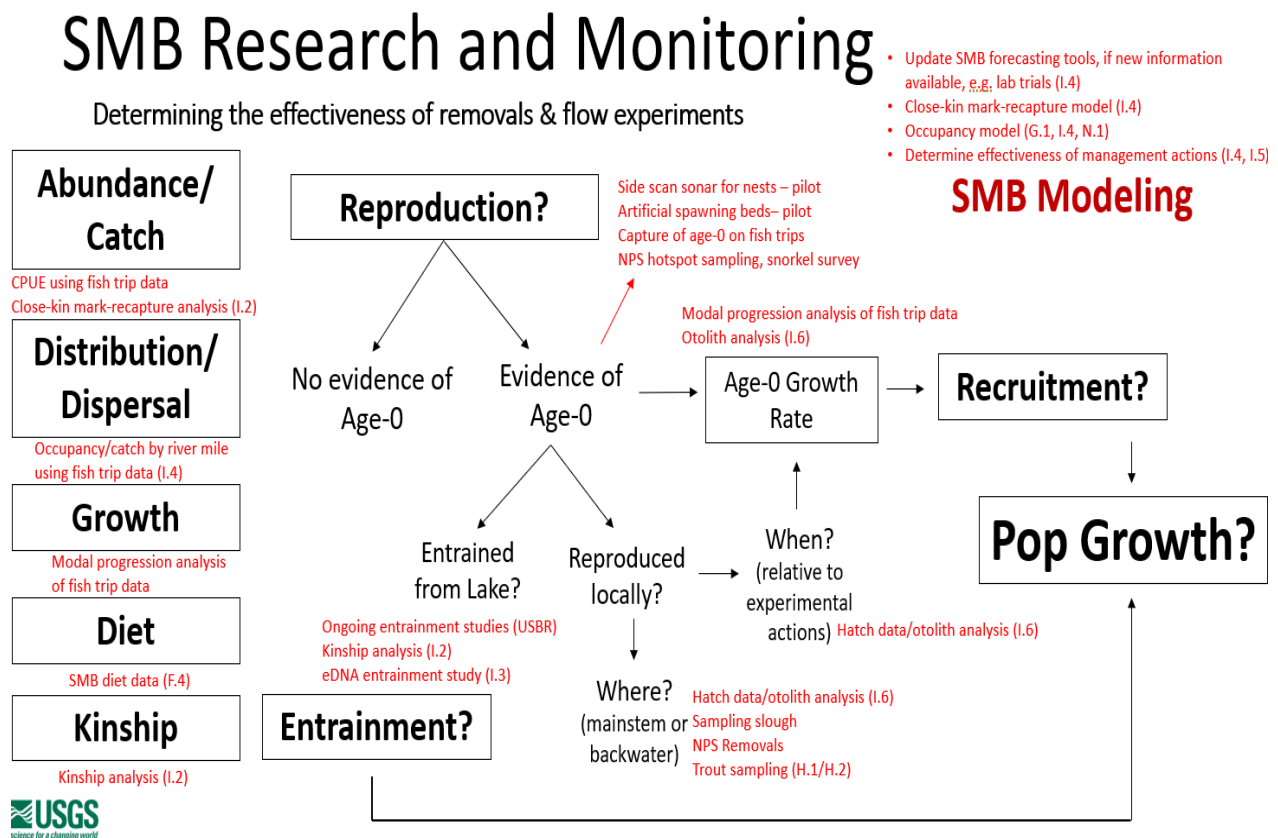


Figure 6. Flow diagram of planned smallmouth bass research and monitoring studies by Project Element (in parentheses) that provide data and analyses for examining the effectiveness of removals and potential LTEMP Flow Experiments on smallmouth bass catch, dispersal, growth, reproduction, and recruitment (Diagram courtesy of Charles Yackulic, Kimberly Dibble, and Drew Eppheimer, USGS).

We will evaluate dispersal using detection data by river mile and/or river reach using traditional sampling gear and eDNA/water sampling (I.3, I.5) and calculate growth rates via modal progression analysis (Figure 7). Smallmouth bass diet data will provide insight into predation pressure at various life stages (F.4). Reproduction and recruitment will be evaluated using pilot data from side scan sonar to search for nests, by deploying artificial substrate to attract nest-building males, capture of larvae or juveniles on fish trips, and using information from NPS snorkel surveys. If we capture age-0 smallmouth bass in the Lees Ferry reach, we will use kinship analysis (I.2), hatch date analysis (I.6), entrainment studies (e.g., I.3), and targeted hot spot sampling to determine whether they were produced locally (and when) or if they were entrained from Lake Powell (Figure 6). Additional LTEMP SEIS-related work is included as options to be funded by the Experimental Fund.

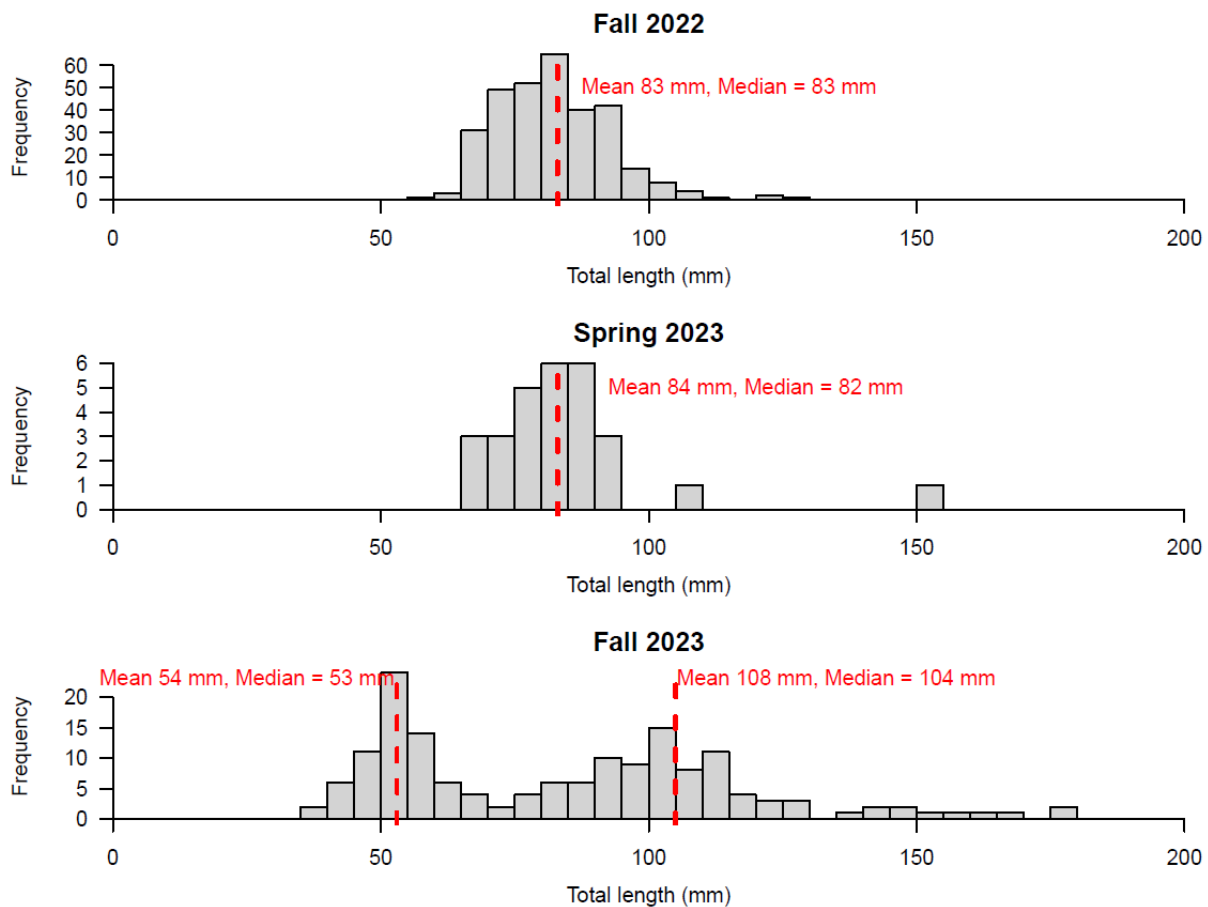


Figure 7. Length-frequency histograms of smallmouth bass that can be analyzed using modal progression analysis to calculate growth rates. Growth from fall 2022 to spring 2023 was minimal due to cold winter temperatures. Temperatures in 2023 were colder overall than in 2022, so age-0 fish did not grow as fast (mean=83 mm in fall 2022; mean=54 mm in fall 2023) (Figure provided by Charles Yackulic using data from NPS, AGFD, and USGS, provisional data – do not cite).

A data visualization dashboard (mapping tool) to view nonnative fish data was piloted in 2023 resulting from a collaboration between Reclamation and GCMRC (Project K). In this work plan, Project K plans to continue to refine that data visualization tool and share it with the public and GCMRC stakeholders. Collectively, the analysis and visualization of multi-agency nonnative fish data described in Project Element I.4 and Project K will be used to support discussions of monitoring effectiveness and offramps laid out in the Invasive Fish Species Strategic Plan (Smallmouth Bass Ad Hoc Group, 2023).

In addition, we propose to use eDNA to evaluate presence and dispersal of smallmouth bass DNA from the dam to the Little Colorado River confluence prior to and after the flow (I.5) and evaluate hatch dates of fish to determine if smallmouth bass were produced before, during, or after LTEMP SEIS flow implementation (I.6). These latter two experiments are proposed for funding by the Experimental Fund.

I.4.3. What are the Long-Term Drivers of Distributional Changes in Native and Nonnative Fishes in the CRe?

The third objective of Project Element I.4 is to develop a workflow process for an occupancy model that evaluates existing monitoring and research data from Projects H, G, and I to report on a standard set of LTEMP Performance Metrics. The proposed metric for Resource Goal #10 includes eight metrics (or two metrics with four sub-categories) based on categories of nonnative species risk (Grand Canyon National Park, 2013; National Park Service, 2018, 2021). Metrics 10.1-10.4 are based on detection (average number) of low, medium, high, and very-high risk nonnative species, while Metrics 10.5-10.8 are based on evidence of nonnative species reproduction for those same risk categories. Data on detections from AGFD, NPS, USFWS, USGS, and other cooperators can be used as data in the development of this occupancy model.

Anticipated Use of Data

These data will be used to estimate the growth and survival of early life stages of smallmouth bass in laboratory experiments, to determine if an update to the smallmouth bass population growth model would improve performance. Data compiled from multi-agency nonnative fish sampling and removal efforts will be analyzed to determine whether management actions are having an effect on catch rates, dispersal, growth, and reproduction.

Outcomes and Products

Outcomes

- Estimation of smallmouth bass early life stage feeding efficiency and survival under turbidity conditions expected in the Colorado River during pre- and post-monsoon

periods. Potential update to forecasting model for smallmouth bass based on results from turbidity trials. Determination of the effectiveness of management actions for reducing smallmouth population growth, survival, dispersal, and reproduction [*if SMB flows occur*].

- Development of workflow for occupancy model for LTEMP Performance Metric for Resource Goal #10.

Products

- Manuscript describing results from laboratory experiments on the effects of turbidity on smallmouth bass growth and survival. Potentially updated forecasting model for smallmouth bass population growth.
- Manuscript describing effectiveness of LTEMP Flow Experiments and removals on smallmouth bass catch rates, growth, dispersal, and reproduction [*if SMB flows occur, see Project Element I.5*].
- Presentation of results at annual reporting meetings and at a scientific conference.

Project Element I.5. Evaluating the Efficacy of Flow Experiments in the LTEMP SEIS to Control Smallmouth Bass (New Study; Experimental Fund)

Drew Eppheimer¹, Charles Yackulic¹, Kimberly Dibble¹, Thomas Franklin²

¹U.S. Geological Survey, Southwest Biological Science Center, Grand Canyon Monitoring and Research Center

²U.S. Forest Service, National Genomics Center for Wildlife and Fish Conservation

The objective of this project element is to use data collected by state and federal agencies and other cooperators (NPS, AZGFD, USFWS, USGS, BioWest/ASIR) to evaluate the efficacy of LTEMP SEIS flow experiments to control the expansion and population growth of smallmouth bass in the Colorado River downstream from Glen Canyon Dam.

Hypotheses and Science Questions

1. How are smallmouth bass distributed throughout the sampled area, and is there evidence of movement/dispersal downstream or hotspots for spawning?
2. Is there in situ recruitment occurring, and if so, where and when did it occur?
3. Is subadult/adult catch stable, increasing, or decreasing, and how does that relate to active removals, experimental flows, and natural mortality?
4. What are the growth rates of subadult and adult smallmouth bass, and how is growth affected by environmental factors vs. experimental flows?

5. What are smallmouth bass eating in GRCA and GLCA and at different life stages, and what are their rates of piscivory?
6. What capture methods are most efficient at various life stages?

Methods

This work will include analyzing data on smallmouth bass abundance/catch, distribution/dispersal, growth, and reproduction. In addition to analyzing multi-agency data to determine the efficacy of LTEMP SEIS Flow Experiments, we propose to use eDNA to evaluate presence and relative abundance of smallmouth bass from the dam to the Little Colorado River confluence prior to smallmouth bass flows. These data can be used as an additional source of data to inform the geographic extent of cooling needed in Marble Canyon resulting from bypass and penstock water released during the flow experiment. Post-flow samples could be used to inform whether the distribution and relative abundance of smallmouth bass has changed following flow implementation.

Outcomes and Products

- **Outcomes**
 - Determination of the effectiveness of management actions for reducing smallmouth population growth, survival, dispersal, and reproduction.
 - Stakeholder update provided during the Annual Reporting meeting and other GCDAMP technical meetings.
- **Products**
 - Manuscript describing results from analysis of multi-agency data to determine the effectiveness of Experimental Flows focused on smallmouth bass.

Anticipated Use of Data

These data will be used to determine whether LTEMP SEIS Experimental Flows and other nonnative fish management actions are having an effect on catch rates, dispersal, growth, and reproduction of smallmouth bass. If eDNA sampling is conducted prior to LTEMP SEIS experimental flows, resource managers can use the data to determine (at a fine scale) the distribution of smallmouth bass in 0.5 or 1-mile increments in Glen and Marble canyons to estimate the extent of river cooling needed to disturb smallmouth bass. Post-flow eDNA sampling can be used by resource managers to determine the effectiveness of experimental flows, as the distribution and relative abundance (eDNA copy number) may have changed following the experiment.

Project Element I.6. Determining Hatch Dates of Larval Smallmouth Bass in Response to LTEMP sEIS Flow Experiments (New Study; Experimental Fund)

Charles Yackulic¹, Kevin Bestgen², Drew Eppehimer¹, Kimberly Dibble¹

¹U.S. Geological Survey, Southwest Biological Science Center, Grand Canyon Monitoring and Research Center

²Colorado State University, Larval Fish Laboratory

The objective of this project element is to evaluate hatch dates of larval smallmouth bass to determine if fish were produced before, during, or after LTEMP SEIS flow implementation. These data can be used to directly determine whether smallmouth bass nests were disturbed and whether eggs were displaced from nests during the flow experiment, which would be evident by larvae not hatching during or shortly after the flow. Further, these data can provide information on whether adult smallmouth bass re-nested at a later date (or in between phases of the flow experiment) and still had a successful hatch that produced juvenile fish.

Methods

Larval fish will be collected throughout the reproductive season in the Lees Ferry reach in a year in which the LTEMP SEIS Flow Experiment occurs using targeted seining and netting efforts to capture fish <40mm. Cobble bars and shallow habitats will be targeted for seining, particularly in areas adjacent to known hot spots for smallmouth bass as identified by multi-agency sampling efforts (see Project Elements I.4, I.5). Larval fish samples will be preserved in 95% ethanol and sent to the Larval Fish Laboratory at Colorado State University for back calculation of hatch dates and calculation of growth rates using otolith microstructural analysis. This project will only occur in years during which LTEMP SEIS Flow Experiments occur.

Outcomes and Products

Outcomes

- Determination of the date age-0 smallmouth bass hatched in relation to the timing of LTEMP SEIS Experimental Flows and other variables (e.g., water temperature). This information can be used to determine whether hatching occurred before, during, or after the experimental flow.

Products

- Inclusion of hatch date data within GCMRC Annual Report and stakeholder update provided during GCDAMP technical meetings.

Anticipated Use of Data

These data will be used to determine the hatch dates of smallmouth bass captured downstream from Glen Canyon Dam. These data can be used to determine whether smallmouth bass hatched prior to, during, or after LTEMP SEIS experimental flows, which can be used to understand nest disturbance and re-nesting of adult smallmouth bass in the Colorado River.

Budgets

Project Element Budget

| Fiscal Year 2025 | | | | | | | | | |
|---|------------------|----------------------|-----------------------|-----------------------|---------------------------|--------------------------|-----------------------------|------------------|------------------------------|
| Project I Non-native Invasive Species Monitoring and Research | Salaries | Travel & Training | Operating Expenses | Logistics Expenses | Cooperative Agreements | To other USGS Centers | AMP Special Burden Rate* | Total | USGS Contributing Funds** |
| | | | | | | | 21.86% | | Est Full Rate = 55% |
| I.1. System-wide native fishes and invasive aquatic species monitoring | \$12,191 | \$0 | \$4,650 | \$105,350 | \$217,550 | \$0 | \$33,241 | \$372,982 | |
| I.2. Estimating kinship and spawner abundance of warm-water non-natives | \$62,715 | \$0 | \$100 | \$0 | \$62,775 | \$0 | \$15,616 | \$141,206 | |
| I.3. Identifying emerging threats to the Colorado River Ecosystem using environmental DNA | \$30,311 | \$300 | \$0 | \$3,215 | \$32,400 | \$0 | \$8,367 | \$74,593 | |
| I.4. Modeling population dynamics and improving forecasting tools for smallmouth bass and other non-native fish | \$160,645 | \$0 | \$1,800 | \$0 | \$0 | \$0 | \$35,515 | \$197,961 | |
| Total Project I | \$265,862 | \$300 | \$6,550 | \$108,565 | \$312,725 | \$0 | \$92,740 | \$786,742 | \$39,817 |

| Fiscal Year 2026 | | | | | | | | | |
|---|------------------|----------------------|-----------------------|-----------------------|---------------------------|--------------------------|-----------------------------|------------------|------------------------------|
| Project I Non-native Invasive Species Monitoring and Research | Salaries | Travel & Training | Operating Expenses | Logistics Expenses | Cooperative Agreements | To other USGS Centers | AMP Special Burden Rate* | Total | USGS Contributing Funds** |
| | | | | | | | 22.60% | | Est Full Rate = 55.5% |
| I.1. System-wide native fishes and invasive aquatic species monitoring | \$10,443 | \$0 | \$4,650 | \$110,692 | \$217,550 | \$0 | \$34,954 | \$378,289 | |
| I.2. Estimating kinship and spawner abundance of warm-water non-natives | \$74,021 | \$0 | \$100 | \$0 | \$47,418 | \$0 | \$18,174 | \$139,713 | |
| I.3. Identifying emerging threats to the Colorado River Ecosystem using environmental DNA | \$24,951 | \$300 | \$0 | \$2,626 | \$23,328 | \$0 | \$7,000 | \$58,205 | |
| I.4. Modeling population dynamics and improving forecasting tools for smallmouth bass and other non-native fish | \$202,107 | \$2,000 | \$1,800 | \$0 | \$0 | \$0 | \$46,535 | \$252,442 | |
| Total Project I | \$311,522 | \$2,300 | \$6,550 | \$113,318 | \$288,296 | \$0 | \$106,663 | \$828,649 | \$45,958 |

| Fiscal Year 2027 | | | | | | | | | |
|---|------------------|----------------------|-----------------------|-----------------------|---------------------------|--------------------------|-----------------------------|------------------|------------------------------|
| Project I Non-native Invasive Species Monitoring and Research | Salaries | Travel & Training | Operating Expenses | Logistics Expenses | Cooperative Agreements | To other USGS Centers | AMP Special Burden Rate* | Total | USGS Contributing Funds** |
| | | | | | | | 23.40% | | Est Full Rate = 56% |
| I.1. System-wide native fishes and invasive aquatic species monitoring | \$10,965 | \$0 | \$4,650 | \$114,150 | \$217,550 | \$0 | \$36,892 | \$384,207 | |
| I.2. Estimating kinship and spawner abundance of warm-water non-natives | \$12,892 | \$2,000 | \$100 | \$0 | \$0 | \$0 | \$3,508 | \$18,500 | |
| I.3. Identifying emerging threats to the Colorado River Ecosystem using environmental DNA | \$33,418 | \$300 | \$0 | \$17,104 | \$48,600 | \$0 | \$13,350 | \$112,772 | |
| I.4. Modeling population dynamics and improving forecasting tools for smallmouth bass and other non-native fish | \$212,212 | \$0 | \$1,800 | \$0 | \$0 | \$0 | \$50,079 | \$264,091 | |
| Total Project I | \$269,487 | \$2,300 | \$6,550 | \$131,254 | \$266,150 | \$0 | \$103,829 | \$779,570 | \$44,088 |

*GCDAMP Special Rate: Rate includes current DOI preferred rate (currently 15%; subject to change) and facilities rate (will vary annually). No USGS bureau overhead is charged (unique to the GCDAMP agreement).

**USGS Contributing Funds: The amount of funds required to cover the subsidy created by the reduced burden rate (i.e., GCDAMP Special Rate). As in previous years, SBSC/GCMRC will request these funds of the USGS cost-share program. These funds are not guaranteed.

Experimental Fund

| Fiscal Year 2025 | | | | | | | | |
|---|-----------------|----------------------|-----------------------|-----------------------|---------------------------|--------------------------|-----------------------------|------------------|
| Project I Experimental Fund Projects | Salaries | Travel & Training | Operating Expenses | Logistics Expenses | Cooperative Agreements | To other USGS Centers | AMP Special Burden Rate* | Total |
| | | | | | | | 21.86% | |
| I.5. Evaluating the efficacy of flow experiments in the LTEMP sEIS to control smallmouth bass | \$30,425 | \$1,800 | \$5,200 | \$38,142 | \$94,478 | \$0 | \$19,356 | \$189,401 |
| I.6. Determining hatch dates of larval smallmouth bass in response to LTEMP sEIS flow experiments | \$19,116 | \$1,500 | \$200 | \$11,102 | \$117,500 | \$0 | \$10,503 | \$159,921 |
| Total Project I | \$49,541 | \$3,300 | \$5,400 | \$49,244 | \$211,978 | \$0 | \$29,859 | \$349,322 |

| Fiscal Year 2026 | | | | | | | | |
|---|-----------------|----------------------|-----------------------|-----------------------|---------------------------|--------------------------|-----------------------------|------------------|
| Project I Experimental Fund Projects | Salaries | Travel & Training | Operating Expenses | Logistics Expenses | Cooperative Agreements | To other USGS Centers | AMP Special Burden Rate* | Total |
| | | | | | | | 22.60% | |
| I.5. Evaluating the efficacy of flow experiments in the LTEMP sEIS to control smallmouth bass | \$31,946 | \$1,800 | \$5,200 | \$39,386 | \$94,478 | \$0 | \$20,537 | \$193,348 |
| I.6. Determining hatch dates of larval smallmouth bass in response to LTEMP sEIS flow experiments | \$20,072 | \$1,500 | \$200 | \$11,316 | \$117,500 | \$0 | \$11,003 | \$161,590 |
| Total Project I | \$52,018 | \$3,300 | \$5,400 | \$50,702 | \$211,978 | \$0 | \$31,540 | \$354,938 |

| Fiscal Year 2027 | | | | | | | | |
|---|-----------------|----------------------|-----------------------|-----------------------|---------------------------|--------------------------|-----------------------------|------------------|
| Project I Experimental Fund Projects | Salaries | Travel & Training | Operating Expenses | Logistics Expenses | Cooperative Agreements | To other USGS Centers | AMP Special Burden Rate* | Total |
| | | | | | | | 23.40% | |
| I.5. Evaluating the efficacy of flow experiments in the LTEMP sEIS to control smallmouth bass | \$33,543 | \$1,800 | \$5,200 | \$40,638 | \$94,478 | \$0 | \$21,831 | \$197,491 |
| I.6. Determining hatch dates of larval smallmouth bass in response to LTEMP sEIS flow experiments | \$21,075 | \$1,500 | \$200 | \$11,536 | \$117,500 | \$0 | \$11,554 | \$163,365 |
| Total Project I | \$54,619 | \$3,300 | \$5,400 | \$52,174 | \$211,978 | \$0 | \$33,385 | \$360,856 |

References Cited

Bestgen, K.R., Walford, C.D., White, G.C., Hawkins, J.A., Jones, M.T., Webber, P.A., Breen, M., Skorupski, J.A., Jr., Howard, J., Creighton, K., Logan, J., Battige, K., and Wright, F.B., 2018, Population status and trends of Colorado pikeminnow in the Green River sub-basin, Utah and Colorado, 2000-2013—final report: Fort Collins, Colo., Colorado State University, Larval Fish Laboratory Contribution 200, submitted to the Colorado River Recovery Implementation Program, Project Number 128, Denver, Colo., 152 p.,

<https://coloradoriverrecovery.org/uc/wp-content/uploads/sites/2/2021/12/TechnicalReport-PROP-Betgen-2018.pdf>.

- Boyer, J.K., and Rogowski, D.L., 2020, Colorado River fish monitoring in the Grand Canyon, Arizona—2019 annual report: Flagstaff, Ariz., Arizona Game and Fish Department, submitted to the U.S. Geological Survey, Southwest Biological Science Center, Grand Canyon Monitoring and Research Center, 32 p., https://www.researchgate.net/publication/353466983_Colorado_River_Fish_Monitoring_in_the_Grand_Canyon_Arizona_2019_Annual_Report. (March 2023; see Supplemental Material, Reference S7).
- Bravington, M.V., and Skaug, H.J., 2016, Close-kin mark-recapture: *Statistical Science*, v. 31, no. 2, p. 259 - 274, <https://doi.org/10.1214/16-STS552>.
- Brown, T.G., Runciman, B., Pollard, S., Grand, A.D.A., and Bradford, M.J., 2009, Biological synopsis of smallmouth bass (*Micropterus dolomieu*): Nanaimo, B.C., Canadian Manuscript Report of Fisheries and Aquatic Science, v. 2887, Fisheries and Oceans Canada, 50 p., https://www.researchgate.net/publication/237358490_Biological_Synopsis_of_Smallmouth_Bass_Micropterus_dolomieu.
- Bruckerhoff, L.A., Wheeler, K., Dibble, K.L., Mihalevich, B.A., Neilson, B.T., Wang, J., Yackulic, C.B., and Schmidt, J.C., 2022, Water storage decisions and consumptive use may constrain ecosystem management under severe sustained drought: *Journal of American Water Resources Association*, v. 58, no. 5, p. 654-672, <https://doi.org/10.1111/1752-1688.13020>.
- Cucherousset, J., and Olden, J.D., 2011, Ecological impacts of nonnative freshwater fishes: *Fisheries*, v. 36, no. 5, p. 215-230, <https://doi.org/10.1080/03632415.2011.574578>.
- Dawson, V.K., and Kolar, C.S., eds., 2003, Integrated management techniques to control nonnative fishes—completion report: La Crosse, Wisc., prepared for the Bureau of Reclamation by U.S. Geological Survey, Upper Midwest Environmental Sciences Center, interagency agreement no. 01-AA-32-0040, 146 p. plus appendices, <https://www.usbr.gov/lc/phoenix/biology/azfish/pdf/TechToControlNNFish.pdf>.
- Dibble, K.L., Yackulic, C.B., Kennedy, T.A., Bestgen, K.R., and Schmidt, J.C., 2021, Water storage decisions will determine the distribution and persistence of imperiled river fishes: *Ecological Applications*, v. 31, no. 2, e02279, p. 1-9, <https://doi.org/10.1002/eap.2279>.
- Diver, T., Mussmann, S., Durst, S., and Franssen, N., 2021, Effective number of breeders and reconstructed sibships reveal low reproductive output by a reintroduced population of endangered fish: *Aquatic Conservation: Marine and Freshwater Ecosystems*, v. 31, <https://doi.org/10.1002/aqc.3722>.
- Eppehimer, D.E., Yackulic, C.B., Bruckerhoff, L.A., Wang, J., Young, K.L., Bestgen, K.R., Mihalevich, B.A., and Schmidt, J.C., 2024, Declining reservoir elevations following a two-decade drought increase water temperatures and non-native fish passage facilitating a downstream invasion: *bioRxiv*, <https://doi.org/10.1101/2024.01.23.576966>.

- Erős, T., Comte, L., Filipe, A.F., Ruhi, A., Tedesco, P.A., Brose, U., Fortin, M.J., Giam, X., Irving, K., Jacquet, C., Larsen, S., Sharma, S., and Olden, J.D., 2020, Effects of nonnative species on the stability of riverine fish communities: *Ecography*, v. 43, no. 8, p. 1156-1166, <https://doi.org/10.1111/ecog.04985>.
- Fonken, D., Rogowski, D.L., Fennell, J., and Gardner, T., 2023, Colorado River fish monitoring in the Grand Canyon, Arizona—2022 annual report: Flagstaff, Ariz., Arizona Game and Fish Department, submitted to U.S. Geological Survey, Southwest Biological Science Center, Grand Canyon Monitoring and Research Center, 41 p.
- Franckowiak, R.P., Ridgway, M.S., and Wilson, C.C., 2017, Genetic mating system and mate selection in smallmouth bass: *Ecology and Evolution*, v. 7, no. 21, p. 8864-8875, <https://doi.org/10.1002/ece3.3423>.
- Gallardo, B., Clavero, M., Sánchez, M.I., and Vilà, M., 2016, Global ecological impacts of invasive species in aquatic ecosystems: *Global Change Biology*, v. 22, no. 1, p. 151-163, <https://onlinelibrary.wiley.com/doi/abs/10.1111/gcb.13004>.
- Grafström, A., and Tillé, Y., 2013, Doubly balanced spatial sampling with spreading and restitution of auxiliary totals: *Environmetrics*, v. 24, no. 2, p. 120-131, <https://doi.org/10.1002/env.2194>.
- Grand Canyon National Park, 2013, Finding of No Significant Impact—Comprehensive Fisheries Management Plan: Coconino County, Ariz., National Park Service, Grand Canyon National Park, Glen Canyon National Recreational Area, 46 p., <https://parkplanning.nps.gov/document.cfm?parkID=65&projectID=35150&documentID=56565>. (March 2023; see Supplemental Material, Reference S12).
- Haden, A., 1992, Nonnative fishes of the Grand Canyon, Arizona—A review with regards to their effects on native fishes: Flagstaff, Ariz., Bureau of Reclamation, Glen Canyon Environmental Studies, 29 p., <http://www.nativefishlab.net/library/textpdf/20950.pdf>.
- Havel, J.E., Kovalenko, K.E., Thomaz, S.M., Amalfitano, S., and Kats, L.B., 2015, Aquatic invasive species: challenges for the future: *Trends in Aquatic Ecology*, v. 750, p. 147–170, <https://doi.org/10.1007/s10750-014-2166-0>.
- Hoffman, G.L., and Schubert, G., 1984, Some parasites of exotic fishes, *in* Courtenay, W.R.J., and Stauffer, J.R.J., eds., *Distribution, biology, and management of exotic fishes*: Baltimore, Md., Johns Hopkins University Press, p. 233-261.
- Hoffnagle, T.L., Choudhury, A., and Cole, R.A., 2006, Parasitism and body condition in humpback chub from the Colorado and Little Colorado Rivers, Grand Canyon, Arizona: *Journal of Aquatic Animal Health*, v. 18, no. 3, p. 184-193, <https://doi.org/10.1577/H05-046.1>.
- Johnson, B.M., Arlinghaus, R., and Martinez, P.J., 2009, Are we doing all we can to stem the tide of illegal fish stocking?: *Fisheries*, v. 34, no. 8, p. 389-394, <https://doi.org/10.1577/1548-8446-34.8.389>.
- Johnson, B.M., Martinez, P.J., Hawkins, J.A., and Bestgen, K.R., 2008, Ranking predatory threats by nonnative fishes in the Yampa River, Colorado, via bioenergetics modeling: *North*

- American Journal of Fisheries Management, v. 28, no. 6, p. 1941-1953,
<https://doi.org/10.1577/M07-199.1>.
- Kegerries, R.B., Albrecht, B., McKinstry, M.C., Rogers, R.J., Valdez, R.A., Barkalow, A.L., Gilbert, E.E., Mohn, H.E., Healy, B., and Omana Smith, E., 2020, Small-bodied fish surveys demonstrate native fish dominance over 300 kilometers of the Colorado River through Grand Canyon, Arizona: *Western North American Naturalist*, v. 80, no. 2, p. 146-156,
<https://doi.org/10.3398/064.080.0202>.
- Klymus, K.E., Richter, C.A., Chapman, D.C., and Paukert, C., 2015, Quantification of eDNA shedding rates from invasive bighead carp *Hypophthalmichthys nobilis* and silver carp *Hypophthalmichthys molitrix*: *Biological Conservation*, v. 183, p. 77-84,
<https://doi.org/10.1016/j.biocon.2014.11.020>.
- Lacoursière-Roussel, A., Rosabal, M., and Bernatchez, L., 2016, Estimating fish abundance and biomass from eDNA concentrations—Variability among capture methods and environmental conditions: *Molecular Ecology Resources*, v. 16, no. 6, p. 1401-1414,
<https://doi.org/10.1111/1755-0998.12522>.
- Leung, B., Lodge, D.M., Finnoff, D., Shogren, J.F., Lewis, M.A., and Lamberti, G., 2002, An ounce of prevention or a pound of cure—Bioeconomic risk analysis of invasive species: *Proceedings of the Royal Society B—Biological Sciences*, v. 269, no. 1508, p. 2407-2413,
<https://doi.org/10.1098/rspb.2002.2179>.
- Loppnow, G.L., Vascotto, K., and Venturelli, P.A., 2013, Invasive smallmouth bass (*Micropterus dolomieu*)—History, impacts, and control: *Management of Biological Invasions*, v. 4, no. 3, p. 191-206, <https://doi.org/10.3391/mbi.2013.4.3.02>.
- Malloy Jr, T.P., Bussche, R.A.V.D., Coughlin, W.D., and Echelle, A.A., 2000, Isolation and characterization of microsatellite loci in smallmouth bass, *Micropterus dolomieu* (Teleostei: Centrarchidae), and cross-species amplification in spotted bass, *M. punctulatus*: *Molecular Ecology*, v. 9, no. 11, p. 1946-1948, <https://doi.org/10.1046/j.1365-294x.2000.01096-16.x>.
- Marsh, P.C., and Pacey, C.A., 2005, Immiscibility of native and non-native fishes, in Brouder, M.J., Springer, C.L., and Leon, C.S., eds., *Restoring natural function within a modified riverine environment—The lower Colorado River, Albuquerque, N.Mex., July 8-9, 1998 and in Restoring native fish to the lower Colorado River—Interactions of native and non-native fishes, Las Vegas, Nev., July 13-14, 1999*, *Proceedings of two symposia*: p. 59-63,
http://www.nativefishlab.net/publications/Symp_Marsh&Pacey.pdf.
- Martinez, P., Wilson, K., Cavalli, P., Crockett, H., Speas, D.W., Trammell, M., Albrecht, B., and Ryden, D., 2014, Upper Colorado River Basin nonnative and invasive aquatic species prevention and control strategy—final report: Lakewood, Colo., U.S. Fish and Wildlife Service Upper Colorado River Endangered Fish Recovery Program, 125 p.,
http://gcdamp.com/images_gcdamp_com/2/25/BASINWIDENNFSTRATEGYFeb2014.pdf.
- McIntosh, C.R., Shogren, J.F., and Finnoff, D.C., 2010, Invasive species and delaying the inevitable—Valuation evidence from a national survey: *Ecological Economics*, v. 69, no. 3, p. 632-640, <https://doi.org/10.1016/j.ecolecon.2009.09.014>.

- Meretsky, V.J., Valdez, R.A., Douglas, M.E., Brouder, M.J., Gorman, O.T., and Marsh, P.C., 2000, Spatiotemporal variation in length-weight relationships of endangered humpback chub—Implications for conservation and management: *Transactions of the American Fisheries Society*, v. 129, no. 2, p. 419-428, [https://doi.org/10.1577/1548-8659\(2000\)129<0419:SVILWR>2.0.CO;2](https://doi.org/10.1577/1548-8659(2000)129<0419:SVILWR>2.0.CO;2).
- Minckley, C.O., 1996, Observations on the biology of the humpback chub in the Colorado River Basin, 1980-1990: Flagstaff, Northern Arizona University, Ph.D. dissertation, 218 p.
- Minckley, W.L., and Marsh, P.C., 2009, *Inland fishes of the greater southwest—Chronicle of a vanishing biota*: Tucson, University of Arizona Press.
- Moyle, P.B., Li, H.W., and Barton, B.A., 1986, The Frankenstein effect—Impact of introduced fishes on native fishes in North America, *in* Stroud, R.H., ed., *Fish culture in fisheries management*: Bethesda, Md., American Fisheries Society, p. 415-426.
- Mueller, G.A., 2005, Predatory fish removal and native fish recovery in the Colorado River mainstem—What have we learned?: *Fisheries*, v. 30, no. 9, p. 10-19, [https://doi.org/10.1577/1548-8446\(2005\)30\[10:PFRANF\]2.0.CO;2](https://doi.org/10.1577/1548-8446(2005)30[10:PFRANF]2.0.CO;2).
- National Park Service, 2018, Environmental Assessment—Expanded non-native aquatic species management plan in Glen Canyon National Recreation Area and Grand Canyon National Park below Glen Canyon Dam: Grand Canyon National Park, National Park Service, Intermountain Region, Glen Canyon National Recreation Area, 154 p., <https://parkplanning.nps.gov/document.cfm?parkID=62&projectID=74515&documentID=90478>.
- National Park Service, 2021, Expanded non-native aquatic species management plan Environmental Assessment—Risk levels of non-native aquatic species in Glen Canyon National Recreation Area and Grand Canyon National Park (2021 update to Appendix F): U.S. Department of the Interior, National Park Service, 11 p., <https://parkplanning.nps.gov/document.cfm?parkID=62&projectID=74515&documentID=117125>.
- Pilliod, D.S., Goldberg, C.S., Arkle, R.S., and Waits, L.P., 2013, Estimating occupancy and abundance of stream amphibians using environmental DNA from filtered water samples: *Canadian Journal of Fisheries and Aquatic Sciences*, v. 70, no. 8, p. 1123-1130, <https://doi.org/10.1139/cjfas-2013-0047>.
- Pimentel, D., Lach, L., Zuniga, R., and Morrison, D., 2000, Environmental and economic costs of nonindigenous species in the United States: *BioScience*, v. 50, no. 1, p. 53-65, [https://doi.org/10.1641/0006-3568\(2000\)050\[0053:EAECON\]2.3.CO;2](https://doi.org/10.1641/0006-3568(2000)050[0053:EAECON]2.3.CO;2).
- Pochardt, M., Allen, J.M., Hart, T., Miller, S.D.L., Yu, D.W., and Levi, T., 2020, Environmental DNA facilitates accurate, inexpensive, and multiyear population estimates of millions of anadromous fish: *Molecular Ecology Resources*, v. 20, no. 2, p. 457-467, <https://doi.org/10.1111/1755-0998.13123>.

- Robertson, B.L., Brown, J.A., McDonald, T., and Jaksons, P., 2013, BAS—Balanced acceptance sampling of natural resources: *Biometrics*, v. 69, no. 3, p. 776-784, <https://doi.org/10.1111/biom.12059>.
- Rogosch, J.S., Tonkin, J.D., Lytle, D.A., Merritt, D.M., Reynolds, L.V., and D., O.J., 2019, Increasing drought favors nonnative fishes in a dryland river—Evidence from a multispecies demographic model: *Ecosphere*, v. 10, no. 4, e02681, p. 1-19, <https://doi.org/10.1002/ecs2.2681>.
- Rourke, M.L., Fowler, A.M., Hughes, J.M., Broadhurst, M.K., DiBattista, J.D., Fielder, S., Wilkes Walburn, J., and Furlan, E.M., 2022, Environmental DNA (eDNA) as a tool for assessing fish biomass—A review of approaches and future considerations for resource surveys: *Environmental DNA*, v. 4, no. 1, p. 9-33, <https://doi.org/10.1002/edn3.185>.
- Schunter, C., Pascual, M., Garza, J.C., Ravento, and Macpherson, E., 2014, Kinship analyses identify fish dispersal events on a temperate coastline: *Proceedings of the Royal Society B—Biological Sciences*, v. 281, no. 1785, p. 1-7, <https://doi.org/10.1098/rspb.2014.0556>.
- Sepulveda, A.J., Schabacker, J., Smith, S., Al-Chokhachy, R., Luikart, G., and Amish, S.J., 2019, Improved detection of rare, endangered and invasive trout in using a new large-volume sampling method for eDNA capture: *Environmental DNA*, v. 1, no. 3, p. 227-237, <https://doi.org/10.1002/edn3.23>.
- Simberloff, D., 2003, How much information on population biology is needed to manage introduced species?: *Conservation Biology*, v. 17, no. 1, p. 83-92, <https://doi.org/10.1046/j.1523-1739.2003.02028.x>.
- Smallmouth Bass Ad Hoc Group, 2023, Invasive fish species below Glen Canyon Dam—A strategic plan to prevent, detect and respond: Phoenix, Ariz., Glen Canyon Dam Adaptive Management Program in partnership with the U.S. Geological Survey Grand Canyon Monitoring and Research Center and the Bureau of Reclamation, presented to the Glen Canyon Dam Adaptive Management Group Technical Work Group, January 26, 2023, 55 p., <https://www.usbr.gov/uc/progact/amp/amwg/2023-02-16-amwg-meeting/20230216-InvasiveFishSpeciesBelowGlenCanyonDam-508-UCRO.pdf>.
- Spear, M.J., Embke, H.S., Krysan, P.J., and Vander Zanden, M.J., 2021, Application of eDNA as a tool for assessing fish population abundance: *Environmental DNA*, v. 3, no. 1, p. 83-91, <https://doi.org/10.1002/edn3.94>.
- Städele, V., and Vigilant, L., 2016, Strategies for determining kinship in wild populations using genetic data: *Ecology and Evolution*, v. 6, no. 17, p. 6107–6120, <https://doi.org/10.1002/ece3.2346>.
- Suedel, B.C., Wilkens, J.L., and Kennedy, A.J., 2017, Effects of suspended sediment on early life stages of smallmouth bass (*Micropterus dolomieu*): *Archives of Environmental Contamination and Toxicology*, v. 72, no. 1, p. 119-131, <https://doi.org/10.1007/s00244-016-0322-4>.

- Sweka, J.A., and Hartman, K.J., 2003, Reduction of reactive distance and foraging success in smallmouth bass, *Micropterus dolomieu*, exposed to elevated turbidity levels: *Environmental Biology of Fishes*, v. 67, no. 4, p. 341–347, <https://doi.org/10.1023/A:1025835031366>.
- Turner, T.F., Dowling, T.E., Osborne, M.J., McPhee, M.V., Broughton, R.E., and Gold, J.R., 2009, Microsatellite markers for the endangered razorback sucker, *Xyrauchen texanus*, are widely applicable to genetic studies of other catostomine fishes: *Conservation Genetics*, v. 10, p. 551–553, <https://doi.org/10.1007/s10592-008-9563-8>.
- U.S. Department of the Interior, 2016a, Glen Canyon Dam Long-term Experimental and Management Plan final Environmental Impact Statement (LTEMP EIS): U.S. Department of the Interior, Bureau of Reclamation, Upper Colorado Region, National Park Service, Intermountain Region, online, <http://ltempeis.anl.gov/documents/final-eis/>.
- U.S. Department of Interior, 2016b, Record of Decision for the Glen Canyon Dam Long-Term Experimental and Management Plan final Environmental Impact Statement (LTEMP ROD): Salt Lake City, Utah, U.S. Department of the Interior, Bureau of Reclamation, Upper Colorado Region, National Park Service, Intermountain Region, 196 p., http://ltempeis.anl.gov/documents/docs/LTEMP_ROD.pdf.
- U.S. Department of the Interior, 2024, Glen Canyon Dam Long-Term Experimental and Management Plan, Draft Supplemental Environmental Impact Statement: U.S. Department of the Interior, Bureau of Reclamation, Upper Colorado River Basins, Interior Region 7, 374 p., https://www.usbr.gov/uc/DocLibrary/EnvironmentalImpactStatements/GlenCanyonDamLong-TermExperimentalManagementPlan/20240207-Draft-GCDLTEMP-SEIS_508.pdf.
- U.S. Fish and Wildlife Service, 2002, Humpback chub (*Gila cypha*) recovery goals—Amendment and supplement to the humpback chub recovery plan: Denver, Colo., U.S. Fish and Wildlife Service, Mountain Prairie Region, 71 p., plus appendices.
- U.S. Fish and Wildlife Service, 2016, Biological Opinion for the Glen Canyon Dam Long-term Experimental Plan, Coconino County, Arizona—Attachment E, *in* Record of Decision final Environmental Impact Statement: Salt Lake City, Utah, U.S. Department of the Interior, Bureau of Reclamation, Upper Colorado Region, National Park Service, Intermountain Region, p. E-1-E-94, http://ltempeis.anl.gov/documents/docs/LTEMP_ROD.pdf.
- Van Haverbeke, D.R., Stone, D.M., Dodrill, M.J., Young, K.L., and Pillow, M.J., 2017, Population expansion of humpback chub in western Grand Canyon and hypothesized mechanisms: *The Southwestern Naturalist*, v. 62, no. 4, p. 285-292, <https://doi.org/10.1894/0038-4909-62.4.285>.
- Vernieu, W.S., Hueftle, S.J., and Gloss, S.P., 2005, Water quality in Lake Powell and the Colorado River, *in* Gloss, S.P., Lovich, J.E., and Melis, T.S., eds., *The state of the Colorado River ecosystem in Grand Canyon—A report of the Grand Canyon Monitoring and Research Center 1991-2004*: U.S. Geological Survey Circular 1282, 69-85 p., <https://pubs.usgs.gov/circ/1282/>.
- Voichick, N., and Topping, D.J., 2014, Extending the turbidity record—Making additional use of continuous data from turbidity, acoustic-Doppler, and laser diffraction instruments and

suspended-sediment samples in the Colorado River in Grand Canyon: U.S. Geological Survey Scientific Investigations Report 2014-5097, 31 p.,
<https://doi.org/10.3133/sir20145097>.

Waples, R.S., and Feutry, P., 2022, Close-kin methods to estimate census size and effective population size: *Fish and Fisheries*, v. 23, no. 2, p. 273-293,
<https://doi.org/10.1111/faf.12615>.

Yard, M.D., Coggins, L.G., Baxter, C.V., Bennett, G.E., and Korman, J., 2011, Trout piscivory in the Colorado River, Grand Canyon—Effects of turbidity, temperature, and fish prey availability: *Transactions of the American Fisheries Society*, v. 140, no. 2, p. 471-486,
<https://doi.org/10.1080/00028487.2011.572011>.

Yates, M.C., Glaser, D.M., Post, J.R., Cristescu, M.E., Fraser, D.J., and Derry, A.M., 2021, The relationship between eDNA particle concentration and organism abundance in nature is strengthened by allometric scaling: *Molecular Ecology*, v. 30, no. 13, p. 3068-3082,
<https://doi.org/10.1111/mec.15543>.

Project J: Socioeconomic Research

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Project Summary and Purpose

In the Long-term Experimental and Management Plan (LTEMP) Environmental Impact Statement (EIS) “recreational experience” and “Tribal resources” are goals that require social and economic monitoring and research to better understand the impacts of management actions related to the operation of Glen Canyon Dam (GCD) on these resources (U.S. Department of Interior, 2016). While there has been some initial monitoring and research related to these goals (Bair and others, 2016; Neher and others, 2017; Neher and others, 2018), there is a need for more comprehensive and long-term social and economic monitoring and research on these topics to adequately inform the Glen Canyon Dam Adaptive Management Program (GCDAMP). This project does not contain elements that establish long-term monitoring but does propose limited research that meet needs for the “recreational experience” and “Tribal resources” goals in the LTEMP EIS:

- 1) **Recreational Experience:** To support research and monitoring needs related to “recreational experience,” this project element will develop a model of whitewater recreation to predict access, participant behavior, and economic value of whitewater rafting under future low flow scenarios, in collaboration with the National Park Service (NPS), and
- 2) **Tribal Resources:** To support research and monitoring needs related to “Tribal resources,” this project element focuses on improving integration of Tribal benefits knowledge, knowledge of benefits and well-being, arising from ecosystems and human engagement with ecosystems, in the GCDAMP (Hoelting and others, 2024a). The effort supports increased recognition and comprehension of the diverse forms of Tribal benefits knowledge that are available to inform GCDAMP decision-making through multiple learning pathways. This effort will be carried out in collaboration and co-production with Tribal led monitoring activities to ground the conversation in Tribal research frameworks.

Related to monitoring and research and in support of goals in the LTEMP EIS, an objective in the LTEMP EIS is to “determine the appropriate experimental framework that allows for a range of programs and actions, including ongoing and necessary research, monitoring, studies, and management actions in keeping with the adaptive management process.” The important phrase in this objective is the identification of necessary monitoring, research and management actions while keeping with the adaptive management process. To further this objective, this project includes an element to conduct research related to integrated modeling of biological, physical, and economic systems, with a focus on value of information (VoI) analysis. The latter effort improves the ability of GCDAMP to evaluate management actions and prioritize monitoring and research.

Background

Project Element J.1. Integrated Models for Adaptive Management

Project Element J.1 will build on the integrated modeling of rainbow trout (*Oncorhynchus mykiss*) and humpback chub (*Gila cypha*) population dynamics, evaluating cost-effective nonnative species management alternatives that meet humpback chub viability goals (Bair and others, 2018; Donovan and others, 2019), and the assessment of the economic costs of experimental flows such as trout management and smallmouth bass (*Micropterus dolomieu*) flows (Yackulic and others, 2024). Continued development of integrated models will provide opportunities to assess the value of information gained from monitoring and research within the GCDAMP.

This element will address the humpback chub and nonnative species goals in the Long-Term Experimental and Management Plan Record of Decision (LTEMP ROD) (U.S. Department of the Interior, 2016), along with compliance activities, by improving our understanding of the dynamics between fisheries resources and impacts of management actions.

These modeling methods reduce uncertainty important to and specify triggers that inform proposed management actions in the LTEMP ROD and other proposed management actions.

The integrated modeling element will focus on integrated modeling of native and nonnative fishes in Glen and Grand Canyon, leveraging ongoing bioeconomic modeling of the system (Bair and others, 2018; Donovan and others, 2019; Yackulic and others, 2024). The project element will emphasize the development of guidance for data collection efforts within the complex bioeconomic system with a continued focus on environmental flow research and hydropower operations at GCD, nonnative species monitoring efforts, management for viable endangered species populations, and other socio-economic stakeholder objectives. This systematic approach for the prioritization of program activities is needed to improve the efficiency of how monitoring and research funding is allocated.

This is consistent with the GCDAMP fisheries review panel’s recommendation that the program, “adopt [a] decision theoretic approach to adaptively manage the rainbow trout fishery and humpback chub population” (Casper and others, 2016). A decision-theoretic approach to adaptive management is when a “predictive model or set of models are created that represent alternative ideas of how the system works” and those priors are evaluated through predicted or actual future resource states (Casper and others, 2016). This approach would allow the GCDAMP to prioritize monitoring and research by identifying the relative efficiency of learning opportunities, while also acquiring information on effectiveness and efficiency of management actions.

Current research includes the exploration of which uncertainties in humpback chub population parameters have the greatest implications for management decisions (i.e., quantitative adaptive management model) and the explicit trade-offs (efficacy and cost) between trout management flows and rainbow trout removals at the Little Colorado River. Continued predictive modeling of multiple resource goals is important as conditions in the Colorado River change, and identifying the relative importance of reducing uncertainties, through application of predictive models, could significantly improve opportunity for cost-effective management. For example, focused effort on one dimension of uncertainty reduction, such as nonnative species monitoring and management, is at the expense of efforts allocated toward other management actions, such as monitoring and research related to experimental dam releases for ecosystem restoration. This work aims to clarify the value of each of these actions and the tradeoffs between them while reducing uncertainty about ecosystem health and function. These efforts also support the analysis expected of short-term rapid response in the Invasive Species Strategic Plan (Smallmouth Bass Ad Hoc Group, 2023), evaluating cost of time and resources dedicated to plan implementation and the costs of management actions into the future.

While we intend to utilize existing predictive models of biological and hydropower resources to evaluate the usefulness of monitoring and research, this project element will also undertake a structured value of information (VoI) analysis with a sub-set of downstream resources.

Limiting the initial VoI assessment to a sub-set of high priority resource goals will provide a tractable first step in the development of a model to demonstrate methods and results to the GCDAMP. In the end, adaptive management is an information problem (Doremus, 2011). Its goal is to improve management through reducing uncertainty in our understanding of the structure of environmental systems. Some uncertainties we can’t reduce, and the benefits of reducing other uncertainties do not outweigh the costs of experimental management actions. It’s the remaining uncertainties that are worth learning more about. Using VoI analysis to characterize uncertainty and identify which uncertainties are worth paying attention to is a way to prioritize monitoring and research when funding is limited (Bolam and others, 2019). We propose the application of VoI analysis in the GCDAMP to assist in the prioritization of monitoring and research.

This type of analysis is an important step in adaptive management, where assessing the value of learning is a necessary component for the effective and efficient management of resource goals (Doremus, 2011; Runge and others, 2011).

Project Element J.2. Recreation Monitoring and Research

Project Element J.2 will develop predictive modeling capability for assessment of whitewater rafting recreational experience under different future management scenarios, as recommended by the GCMRC Socioeconomic Research Review Panel (Hamilton and others, 2010). This research will build on simulation models of recreational activities (Roberts and others, 2002; Gaston and others, 2015; Yackulic and others, 2024) and the estimation of recreational preferences for flows by whitewater rafters in Grand Canyon (Bair and others, 2016; Neher and others, 2017).

The Grand Canyon Protection Act (GCPA) of 1992 states that, “long-term monitoring of GCD shall include any necessary research and studies to determine the effect of the Secretary's actions under section 1804(c) on the...recreational...resources of Grand Canyon National Park and Glen Canyon National Recreation Area” (GCPA, sec. 1805(b)). This research element addresses the language in the GCPA and the “Recreational Experience” resource goal in the LTEMP ROD, by undertaking research to inform management. Ongoing research has established the importance of flow attributes to recreational users in Glen and Grand Canyons (Bair and others, 2016; Duffield and others, 2016; Neher and others, 2017; 2018). This research has also demonstrated the temporal stability of recreational preferences for flow attributes over several decades of dam operations (Bishop and others, 1987; Neher and others, 2017). However, there has been no effort to incorporate these research findings into a dynamic model of recreational use and experience in Glen and Grand Canyons, as recommended by a GCMRC Economics Panel (Hamilton and others, 2010).

This research effort will integrate past research in whitewater recreation trip simulation in the Grand Canyon (Roberts and others, 2002) with the estimates of economic parameters developed in revealed and stated preference research (Bishop and others, 1987; Bair and others, 2016; Duffield and others, 2016; Neher and others, 2018). This approach provides the ability to assess whitewater rafting recreational experience, including the ability to calculate recreational experience as impacted by proposed management actions in the LTEMP ROD and under different future hydrology. Developing a whitewater rafting recreational experience model also allows for the assessment of various aspects of trip behavior that are important to the management of other resources in Grand Canyons, such as archeological resources.

Project Element J.3. Tribal Resources Research (Unfunded)

Project Element J.3 focuses on developing more explicit pathways for recognition and meaningful consideration in the GCDAMP of the values and importance of Tribal resources, and the effects of management on these values.

The “Tribal resources” goal in the LTEMP EIS has not been adequately defined, and comprehensive research and monitoring efforts have not yet been designed to inform the GCDAMP on this goal. This project element is envisioned to address a central issue that will support and enable equitable, accurate research and monitoring around the values of Tribal resources: building understanding within the GCDAMP of diverse forms of knowledge that convey plural values linked to ecosystems and resources, and piloting an ecosystem services-based approach to identify and integrate available Tribal benefits knowledge.

The work combines insights from ecosystem services (ES) theory with place-based, context-specific Tribal understandings of well-being, ecosystems, and human-nature relationships. Specifically, we consider how the Cultural Benefits Learning Framework (CBLF) developed by Hoelting and others (2023) can intersect with Tribal research frameworks to increase recognition and integration of diverse knowledge forms that effectively convey the importance of Tribal resources. We intend to evaluate the application of the CBLF framework within the GCDAMP, in collaboration with interested Tribal partners and ongoing Tribal monitoring programs. This process includes close collaboration with Tribal stakeholders in the evaluation of ongoing monitoring programs, translation of monitoring results into decision relevant knowledge products, and the integration of these knowledge products into broader aspects of the decision process within the GCDAMP. Identifying decision-relevant benefits knowledge, and the knowledge forms that can effectively convey that knowledge, will assist in integration of this information into decision-making in the GCDAMP.

Although there is a growing agreement around the need to elevate diverse knowledges in decision-making, there is less clarity in how to achieve this in practice. The CBLF framework helps operationalize knowledge pluralism by directing attention to diverse forms in which benefits knowledge can be made available, and diverse pathways through which learning from this knowledge can occur in environmental decision-making. For example, the meaningful integration of plural values linked to Tribal resources will require recognition that these values are often best conveyed through knowledge practices and cannot be fully conveyed through quantitative information or written documentation (Hoelting and others, 2024c). The framework is envisioned as a systematic approach to make explicit the diversity of knowledge forms as well as integration pathways most likely to enable learning from these knowledge forms.

Past applications of the CBLF have emphasized improving consideration of plural values associated with cultural benefits specifically. Cultural benefits of ecosystem services remain an important focus of this project element, particularly in the second task on “Retrospective analysis and current assessment of Tribal cultural-benefits-knowledge in the GCDAMP.” This is because of the close association between cultural benefits and value aspects commonly marginalized in ecosystem valuation, e.g., relational values and holistic value perspectives (Hoelting and others, 2024c). However, the theory underlying the CBLF is widely applicable to improving consideration of plural values and value perspectives associated with all ecosystem services and benefits that may be relevant to Tribal resources.

This project task seeks to establish theoretical foundations for research and monitoring of Tribal resources to inform the GCDAMP, grounded in concepts of value pluralism and knowledge pluralism (Hoelting and others, 2024a). This foundation is necessary to ensure pathways are available for integration of plural values – including relational values and holistic value perspectives – alongside the instrumental values typically emphasized in ecosystem valuation and management trade-off analyses (Chan and others, 2018; Himes and Muraca, 2018; Hoelting and others, 2024a). This emphasis attends to recognition justice needs by working to creating space in decision-making for plural values and the associated diversity of knowledge systems and understandings of well-being (Martin and others, 2016; Gould and others, 2020; Hoelting and others, 2024a). This approach is critical for equitable and accurate research and monitoring around Tribal resources specifically but is also relevant to the GCDAMP broadly; lessons learned through this project element will be relevant to consider for monitoring and evaluating success across multiple GCDAMP LTEMP goals.

This effort will be carried out in collaboration and co-production with Tribal led monitoring activities to ground the conversation in Tribal research frameworks. The theoretical foundations developed in this project task are not intended to replace existing Tribal research frameworks, but rather to support and intersect with them to ensure available forms of Tribal cultural-benefits-knowledge, i.e., those forms of knowledge which Tribes choose to offer, are explicitly recognized and legitimized as valid, decision-relevant knowledge to inform the GCDAMP through varied learning pathways. This work will be carried out in collaboration with Tribal stakeholders involved in the GCDAMP. For additional detail about multiple knowledge systems see Hoelting and others (2024a).

There is a broad need for tools and guidance to support knowledge pluralism in applied environmental management (Hoelting and Gould, 2021; Hoelting and others, 2023), and in U.S. Federal decision-making in particular (e.g., White House, 2021, 2022). In addition to the need for adequate theoretical foundations (see first task of Project Element J.3), there is also a need for decision-support tools and guidance to support implementation and iterative learning. The research proposed in this project task aims to pilot and refine a tool both for natural resource management agency personnel and research scientists, as well as Tribal stakeholders, to identify and engage opportunities for more meaningful consideration of plural values linked to Tribal resources. Specifically, this project task will apply the CBLF outlined by Hoelting and others (2023) to carry out both a) retrospective analysis to understand how Tribal cultural-benefits-knowledge has been made available to the GCDAMP historically, and b) current assessment of available knowledge forms and pathways for their integration.

First, *retrospective* analyses of past GCDAMP decision processes can highlight tangible examples as an avenue for reflection around successes and failures, which in turn can feed into guidance to inform current decision processes. In turn, assessment of *current* decision processes can enable systematic identification of available Tribal cultural-benefits-knowledge-forms and opportunities for their meaningful consideration in on-going decision-making processes.

In addition, assessment of current decision-making offers added benefit as a context for social learning among current actors in the GCDAMP, in terms of supporting reflection, systems thinking, negotiation, participation, and integration of diverse knowledges.

Application of the CBLF, as a decision-support tool, can be understood as to offer nested learning opportunities. First, the act of applying the framework offers a social learning process with potential to enhance decision-makers' recognition of the multiple knowledge systems linked to the Grand Canyon ecosystem. Second, the output of CBLF framework application is a list of available forms of benefits knowledge linked to specific opportunities for learning from each knowledge form. These nested outcomes of the CBLF approach can ultimately support diverse learning pathways for meaningful consideration of plural values linked to Tribal resources.

These learning pathways exist across phases of decision-making, including technical and deliberative phases (Brest and Krieger, 2010; Hoelting and others, 2023). Learning opportunities may include 1) technical, single-loop learning to refine understanding of ecosystem function and how management actions affect resources and resource values, 2) deliberative, double-loop learning to integrate values and understandings of ethics and well-being that may have been missed in previous problem definition or goal and objective-setting exercises, and/or 3) triple-loop learning around ways that the existing institutional structures, processes, and paradigms may limit meaningful consideration of particular values or understandings of well-being, i.e., limit consideration of diverse benefits knowledges (Argyris 1977; Pahl-Wostl, 2009; Hoelting and others, 2023, 2024b).

In applying the CBLF to retrospective analysis and current decision assessment of the GCDAMP, this project does not intend to resolve fundamental conflict over natural resource management. Instead, the goal of this tool is to ensure that all forms of cultural-benefits-knowledge relevant to the management of GCD are recognized (Martin and others, 2016; Gould and others, 2020), and identify opportunities, including existing barriers and enabling factors that could be addressed or harnessed, to improve integration of diverse cultural-benefits-knowledge-forms in the GCDAMP (Hoelting and others, 2024b, 2024c).

Proposed Work

Project Element J.1. Integrated Models for Adaptive Management (Modified Study)

This project element will implement interrelated but distinct approaches to further development of dynamic optimization models to improve on predictive modeling of the Colorado River ecosystem and undertake formal VoI analysis.

Hypothesis and Science Questions

- The value of information derived from monitoring and research in a multi-objective adaptive management program is partially dependent on management goals and the relative value of resources, when predicting outcomes of resource states from experimental management actions.

This hypothesis will be tested through continued development of predictive integrated resource models and formal value of information analysis.

Methods

This first approach provides bioeconomic modeling support for the Glen Canyon Adaptive Management Program. The scope of the project involves directing targeted learning efforts within a complicated ecological system encompassing environmental flow research and hydropower operations at Glen Canyon Dam, invasive species monitoring efforts, management for endangered species for ongoing population viability, and other socio-economic stakeholder objectives in the Colorado River ecosystem. A systematic approach for the prioritization of future fieldwork activities that improve our understanding of this system is of particular importance as conditions in the Colorado River change. We recommend assigning priority based on a value of information (VoI) analysis to improve the efficiency of deployed resources.

This project determines the uncertainties which are most important to reduce within the context of our specific management challenge. The proposed model is an extension of current salmonid and humpback chub models (Bair and others, 2018; Donovan and others, 2019) and addresses learning about the expansion dynamics of smallmouth bass while also managing for viable populations of humpback chub. Learning will center around opportunities for reducing uncertainty in system dynamics during smallmouth bass removals and monitoring while also operating GCD for viable humpback chub populations, river temperature management, and minimizing foregone hydropower generation. But explicitly modeling all of these additional model features comes at a high computational cost. The decision model currently applied to the system of interest forces a highly simplified representation of Colorado River ecology in favor of computational feasibility and thus cannot facilitate additional complexity (Donovan and others, 2019).

To address this computational issue and facilitate our VoI analysis, we will use cutting edge dynamic programming methods (Powell 2011; Springborn and Faig, 2019). The Approximate Dynamic Programming (ADP) innovation allows us to optimize systems multiple orders of magnitude more complicated than previously allowed and provides a way to respond to a more accurate representation of the Colorado River ecosystem. This avoids burdensome assumptions about simplified dynamics that only act as a proxy for the underlying system of interest.

ADP also facilitates latent model features that are not directly observed, which is a critical development for cases in which it is infeasible to observe every aspect of a system with 100% certainty. With this modeling paradigm, we can therefore determine the implications of observational focus on one subset of the system versus another.

As part of the integrated modeling task, GCMRC will work with Reclamation, WAPA, and the stakeholders (e.g., Socioeconomic Ad Hoc Group) to present the details on the hydropower models being used in Project J.1 as part of the Hydropower Workshop in the Fall/Winter Fiscal Year 2025 that is being conducted through the Bureau of Reclamation budget item 4.G. The screening tool and integrated models discussed in Project J.1 are directly linked to the scope of 4.G Hydropower Monitoring and Research. The goal of the workshop that is dedicated to Project J.1 hydropower models will be to help stakeholders better understand the benefits and limitations of the hydropower models being proposed within Project J.1. Additionally, this workshop will give stakeholders an opportunity to provide input on the hydropower modeling component in the screening tool and integrated models discussed in Project J.1. GCMRC will coordinate with WAPA and Reclamation in compiling inputs and assumptions (e.g., energy pricing and hourly demand) to be used for the hydropower component of the integrated modeling effort. We anticipate this tool to be helpful in providing rapid assessments of changes in energy amount and value as a limited economic assessment for changing operations at Glen Canyon Dam.

The second approach to further develop formal decision-making methods within the GCDAMP would be by implementing a VoI analysis to guide monitoring and research funding decisions. The use of VoI in conservation settings is at an early stage (Bolam and others, 2019). Applying VoI analysis to a multiple-objective adaptive management program provides an opportunity to advance methods used to identify the value of reducing uncertainty, which may be dependent on the program goals and the relative value of resources states.

We will follow steps in VoI analysis outlined in the literature (Runge and others, 2011; Bolam and others, 2019).

The first step in the analysis will be to document the GCDAMP's objectives and related proposed performance metrics when managing resources downstream from GCD. We will draw on the existing literature specific to the GCDAMP to accomplish this (Runge and others, 2015; U.S Department of the Interior, 2016). We will focus on a sub-set of resource objectives to implement this initial research project. The resources will be selected based on program priorities as defined by legal requirements, priority based on the set of experimental actions, and share of available research funding. A second step will be to identify the complete set of monitoring and research actions that could be used to better meet resource objectives. For each monitoring and research effort, we will consider the observational and structural uncertainty being reduced through monitoring and research. We will then use underlying hypotheses about how the system works, the relative importance, or value, assigned to resources, and the predicted outcomes associated with each monitoring, research, and experimental action to undertake a formal VoI analysis.

Outcomes and Products

- The development of dynamic programming methods for adaptive management in complex bioeconomic systems provides benefits to natural resource economists, ecologists, and program managers operating within and outside of the Grand Canyon setting. Alongside the dissemination of research on the development of the new modeling approach in an academic journal, this project will publish an open-source software package that allows for application to other large, partially observable systems with reduced model development time.

A second scientific article will be published that details methods and result of a VoI analysis related to monitoring and research actions in a multi-objective adaptive management program, intended to assist in prioritizing on monitoring and research funding.

Project Element J.2. Recreation Monitoring and Research (New Study)

This project element will develop a simulation model of annual recreational rafting in Grand Canyon using a sub-daily model. The sub-daily model will be based on forecast hydrology, operational constraints at GCD, energy generation and recreational use as determined by NPS policy and forecast visitor use.

Hypothesis and Science Questions

- The recreational experience and economic value of whitewater rafting on the Colorado River in Grand Canyon will decline over time with projected impacts of aridification in the Colorado River Basin.

Development of a whitewater rafting simulation model that integrates NPS recreation policy, estimated changes in recreational access, behavior, visitation, and impacts to attributes of economic importance will allow us to test these hypotheses under different experimental flow and GCD operational management scenarios.

Methods

The influence of hourly releases at GCD has a significant impact on downstream resources, including river recreation (Bair and others, 2016, Neher and others, 2017). Hourly releases are a function of load following at GCD, where water is released to correspond to electricity demand in the Southwest United States. Demand varies by time of day and season of year, therefore modeling the hourly discharge from GCD over longer time horizons is an important component of evaluating resource outcomes under different scenarios, including low water releases. Therefore, we will integrate a sub-daily energy generation model of flows with recreational use (Yackulic and others, 2024).

Recent research has demonstrated that preferences for recreational attributes (water levels) have remained stable over the last 30 years (Bishop and others, 1987; Neher and others, 2017). However, potential scenarios that deviate from historical conditions present significant uncertainty in recreational access and quality.

We plan to further develop existing modeling frameworks (Gaston and others, 2015; Yackulic and others, 2024) to estimate how Colorado River flow alteration will impact whitewater rafting recreation access, behavior, and quality over a wide range of conditions. Specifically, we aim to address how climate-change related modifications to dam operations, along with general trends in visitation and recreational activity, will influence whitewater rafting recreation outcomes.

To address the uncertainty in whitewater rafting recreational access and behavior under varying river flow conditions we will update the Roberts and others (2002) model that details decision making during recreational trips. The model will simulate whitewater rafting behavior and decision making of individual trips as they seek to optimize their experience. We will utilize federal visitation data, recreational management plans, and qualitative information to update existing models of whitewater rafting recreational use and behavior. The updated models will incorporate the latest visitation trends, updated modeling parameters, and quantitative information about access and visitation in operational scenarios outside of historical norms.

The result of this project sub-element will be an individual trip-based model that uses historic recreational data with expected changes in recreational access and behavior under various conditions, incorporating the revealed and stated preference research that has occurred since 2016 (Bair and others, 2016; Neher and others, 2017, 2018). The objective of the model is to better predict and report on whitewater rafting recreational experience under a variety of operational scenarios and potential impacts to other downstream resources.

Outcomes and Products

Scientific article documenting a whitewater rafting simulation model that is used to estimate recreational use, behavior, and economic value. The publication will integrate past individual trip-based simulation models (Roberts and others, 2002; Gaston and others, 2015) with the revealed and stated preference research specific to whitewater rafting in Grand Canyon (Bair and others 2016; Neher and others 2018).

Project Element J.3. Tribal Resources Research (New Study; Unfunded)

The proposed research outlines a plan for application of the Cultural Benefits Learning Framework (CBLF) detailed in Hoelting and others (2023). This research aims to better understand opportunities for more meaningful consideration of the plural values linked to Tribal resources in the GCDAMP.

Hypothesis and Science Questions

- Adaptive management provides a unique opportunity, through double-loop-learning, to better include diverse benefits knowledge in formal decision-making processes.

To test this hypothesis, we will use a cultural benefit learning framework (Hoelting and others, 2023) to identify opportunities for improved integration of benefits knowledge, within deliberative, double-loop learning (Williams and Brown, 2018). This process will integrate values and understandings of ethics and well-being across Tribes that may have been missed in previous problem definition or goal and objective-setting exercises within the GCDAMP.

Methods

First, the “theoretical foundations for integration of plural values linked to Tribal resources in Adaptive Management” project task seeks to establish appropriate and equitable theoretical foundations for research and monitoring of Tribal resources to inform the GCDAMP, grounded in concepts of value pluralism and knowledge pluralism (Hoelting and others, 2024a). This project task will explore areas of overlap between learning to reduce structural uncertainty of resource dynamics downstream of GCD and learning to reduce uncertainty related to ethics, epistemologies, and institutional structure in the GCDAMP. Drawing on ecosystem services theory in concert with Tribal research frameworks, this research explores how reducing uncertainty through learning in adaptive management is a promising mechanism for more meaningful consideration of the plural values associated with Tribal resources. This research will involve the following elements:

- Tribal review of an existing preliminary literature review and needs assessment of adaptive management as a mechanism to reduce uncertainty related to plural values and benefits knowledges in adaptative management.
 - Explore possible declaration of principles or other principles to guide application of this theory in the context of the GCDAMP.
- Conduct up to 10 interviews with individuals involved with the GCDAMP who have multiple years of involvement and experience with the GCDAMP to understand
 - Details of the GCDAMP decision context, including problem definition, objectives, management alternatives, and processes and protocols for stakeholder engagement and learning,
 - Stakeholders and specifically relevant cultural groups,
 - Knowledge forms through which cultural groups seek to convey cultural-benefits-knowledge (e.g., Hoelting, 2023; Hoelting and others, 2024a, 2024b),
 - Areas of uncertainty, and by extension learning opportunities, with respect to decision making in the GGCDAMP and cultural-benefits-knowledge.

- Collaborate with Tribal stakeholders to finalize the literature review and needs assessment, which is envisioned as documentation of theoretical foundations for integration of plural values linked to Tribal resources in adaptive management, including possible declaration of principles or other guidelines identified as necessary or useful by Tribal collaborators.

Second, the “pilot and refine a decision-support tool for retrospective analysis and current assessment of Tribal cultural-benefits-knowledge in the GCDAMP” project task seeks to pilot the CBLF framework as a tool to understand past successes and limitations around learning from Tribal cultural-benefits-knowledge (retrospective analysis) as well as to identify current opportunities for learning (current decision assessment). Through this process, the research will also support iterative refinement of the CBLF approach. This research will involve the following elements:

- Retrospective Analysis using the CBLF: retrospective case study of cultural-benefits-knowledge integration in the GCDAMP, and
 - Project investigators will conduct interviews and short field visits with a full spectrum of stakeholders.
 - If Tribal participation across the GCDAMP and funding allows, project investigators will support students in initial community meetings and focus groups to launch their fieldwork, and subsequently students will carry out in-person interviews and participant observation, and on-site archival research.
- Current Decision Assessment using the CBLF: assessment of current opportunities for improved integration of cultural-benefits-knowledge in the existing GCDAMP decision-context.
 - Gather GCDAMP actors for a workshop to carry out current decision assessment of Tribal cultural-benefits-knowledge-forms and learning opportunities using the CBLF. Depending on funding, this workshop could be either 1) virtual workshop with a small group of stakeholders representing diverse cultural-benefits-knowledges linked to the Glen Canyon and Grand Canyon areas, 2) an in-person workshop followed by several virtual workshops, or 3) multiple in-person workshops to provide additional support for social learning processes that unfold over time. The latter would allow more adequate space for
 - Reflexivity and mutual understanding around the diverse cultural-benefits-knowledges of workshop participants (i.e., diverse understandings of well-being and human-nature relationship, and aspects of value linked to Tribal resources),

- Reflexivity and understanding around the diverse forms of valid knowledge that have potential to convey these cultural-benefits-knowledges (i.e., understanding knowledge as a practice in addition to documentation of knowledge as a product, and understanding of pathways through which knowledge practices can support learning in the GCDAMP); and
 - Collaborative refinement of the CBLF process.
- As part of current decision assessment, this research will also support continued development of Tribal monitoring programs that have the potential to make cultural-benefits-knowledge available to the GCDAMP in various forms. Workshop(s) will focus in part on supporting understandings of knowledge-as-practice workshops, with potential to gather groups of Tribal members, land managers, and scientists during site visits, ceremonies, or other traditional activities for teaching around plant and mineral collection and other phenomenological, experiential, and social learning as practice. Engagements such as this can allow for better understanding of knowledge practices as valid, decision-relevant forms of cultural-benefits-knowledge, foster development of data collection methods to support learning from these knowledge forms (e.g., participant observation, field notes, surveys, photography, and other activities), and provide insights into data organization, analysis, and review protocols.

As part of current decision assessment, this research will further engage with development and refinement of principles and guidelines for equitable, appropriate engagement with Tribal knowledge forms.

Outcomes and Products

- Scientific article on the potential of adaptive management to support identification and implementation of opportunities for improved consideration of Tribal benefits knowledge, using examples from the GCDAMP to illustrate challenges and areas of potential (Phase 1).
 - The completion of the first deliverable will allow for a detailed proposal for Phases 2 and 3 of this research proposal.
- Scientific article presenting a retrospective case analysis of Tribal benefits knowledge in the GCDAMP, with a detailed assessment of what a current decision assessment would entail (Phase 2). For an example of retrospective application of the CBLF to another decision context see the Elwha River dam removal and ecosystem restoration in Hoelting and others (2023).

- Phase 2 would lead to relationship building to enable Phase 3 of this project proposal.
- Master’s thesis or Ph.D. dissertation focused on the intersections between adaptive management and cultural-benefits-knowledge, using the GCDAMP as a core case study; including the most comprehensive relationship building opportunity to enable Phase 3 of this research agenda.
- Development of guiding principles and ideas for future engagements that could be implemented to carry out current decision assessment (Phase 3).
 - Phase 3 would result in completed current decision assessment, and a scientific article detailing the process of carrying out a current decision assessment.

Development of cross-cultural comprehension among stakeholder groups and opportunities identified through the workshops to considered for implementation.

Budget

Project Element Budget

| Fiscal Year 2025 | | | | | | | | | |
|--|------------------|-------------------|--------------------|--------------------|------------------------|-----------------------|--------------------------|------------------|---------------------------|
| Project J Socioeconomic Research | Salaries | Travel & Training | Operating Expenses | Logistics Expenses | Cooperative Agreements | To other USGS Centers | AMP Special Burden Rate* | Total | USGS Contributing Funds** |
| | | | | | | | 21.86% | | Est Full Rate = 55% |
| J.1. Integrated models for adaptive management | \$92,475 | \$2,000 | \$2,000 | \$0 | \$37,600 | \$0 | \$22,220 | \$156,295 | |
| J.2. Recreation monitoring and research | \$57,732 | \$1,000 | \$1,000 | \$0 | \$0 | \$0 | \$13,059 | \$72,791 | |
| Total Project J | \$150,206 | \$3,000 | \$3,000 | \$0 | \$37,600 | \$0 | \$35,279 | \$229,086 | \$16,313 |

| Fiscal Year 2026 | | | | | | | | | |
|--|------------------|-------------------|--------------------|--------------------|------------------------|-----------------------|--------------------------|------------------|---------------------------|
| Project J Socioeconomic Research | Salaries | Travel & Training | Operating Expenses | Logistics Expenses | Cooperative Agreements | To other USGS Centers | AMP Special Burden Rate* | Total | USGS Contributing Funds** |
| | | | | | | | 22.60% | | Est Full Rate = 55.5% |
| J.1. Integrated models for adaptive management | \$123,809 | \$2,000 | \$2,000 | \$0 | \$0 | \$0 | \$28,885 | \$156,694 | |
| J.2. Recreation monitoring and research | \$90,927 | \$1,000 | \$1,000 | \$0 | \$0 | \$0 | \$21,002 | \$113,929 | |
| Total Project J | \$214,736 | \$3,000 | \$3,000 | \$0 | \$0 | \$0 | \$49,886 | \$270,622 | \$23,391 |

| Fiscal Year 2027 | | | | | | | | | |
|--|------------------|-------------------|--------------------|--------------------|------------------------|-----------------------|--------------------------|------------------|---------------------------|
| Project J Socioeconomic Research | Salaries | Travel & Training | Operating Expenses | Logistics Expenses | Cooperative Agreements | To other USGS Centers | AMP Special Burden Rate* | Total | USGS Contributing Funds** |
| | | | | | | | 23.40% | | Est Full Rate = 56% |
| J.1. Integrated models for adaptive management | \$129,999 | \$2,000 | \$2,000 | \$0 | \$0 | \$0 | \$31,356 | \$165,355 | |
| J.2. Recreation monitoring and research | \$111,386 | \$1,000 | \$1,000 | \$0 | \$0 | \$0 | \$26,532 | \$139,918 | |
| Total Project J | \$241,385 | \$3,000 | \$3,000 | \$0 | \$0 | \$0 | \$57,888 | \$305,273 | \$26,628 |

*GCDAMP Special Rate: Rate includes current DOI preferred rate (currently 15%; subject to change) and facilities rate (will vary annually). No USGS bureau overhead is charged (unique to the GCDAMP agreement).

**USGS Contributing Funds: The amount of funds required to cover the subsidy created by the reduced burden rate (i.e., GCDAMP Special Rate). As in previous years, SBSC/GCMRC will request these funds of the USGS cost-share program. These funds are not guaranteed.

Unfunded Project Element Budget

| Fiscal Year 2025 | | | | | | | | |
|-------------------------------------|-----------------|-------------------|--------------------|--------------------|------------------------|-----------------------|--------------------------|------------------|
| Project J Unfunded Project Elements | Salaries | Travel & Training | Operating Expenses | Logistics Expenses | Cooperative Agreements | To other USGS Centers | AMP Special Burden Rate* | Total |
| J.3. Tribal resources research | \$69,301 | \$7,500 | \$45,000 | \$0 | \$0 | \$0 | 21.86% | \$148,354 |
| Total Project J | \$69,301 | \$7,500 | \$45,000 | \$0 | \$0 | \$0 | \$26,553 | \$148,354 |

| Fiscal Year 2026 | | | | | | | | |
|-------------------------------------|-----------------|-------------------|--------------------|--------------------|------------------------|-----------------------|--------------------------|------------------|
| Project J Unfunded Project Elements | Salaries | Travel & Training | Operating Expenses | Logistics Expenses | Cooperative Agreements | To other USGS Centers | AMP Special Burden Rate* | Total |
| J.3. Tribal resources research | \$89,889 | \$8,500 | \$45,000 | \$0 | \$0 | \$0 | 22.60% | \$175,795 |
| Total Project J | \$89,889 | \$8,500 | \$45,000 | \$0 | \$0 | \$0 | \$32,406 | \$175,795 |

| Fiscal Year 2027 | | | | | | | | |
|-------------------------------------|-----------------|-------------------|--------------------|--------------------|------------------------|-----------------------|--------------------------|------------------|
| Project J Unfunded Project Elements | Salaries | Travel & Training | Operating Expenses | Logistics Expenses | Cooperative Agreements | To other USGS Centers | AMP Special Burden Rate* | Total |
| J.3. Tribal resources research | \$79,343 | \$7,500 | \$45,000 | \$0 | \$0 | \$0 | 23.40% | \$162,694 |
| Total Project J | \$79,343 | \$7,500 | \$45,000 | \$0 | \$0 | \$0 | \$30,851 | \$162,694 |

*GCDAMP Special Rate: Rate includes current DOI preferred rate (currently 15%; subject to change) and facilities rate (will vary annually). No USGS bureau overhead is charged (unique to the GCDAMP agreement).

**USGS Contributing Funds: The amount of funds required to cover the subsidy created by the reduced burden rate (i.e., GCDAMP Special Rate). As in previous years, SBSC/GCMRC will request these funds of the USGS cost-share program. These funds are not guaranteed.

References Cited

- Argyris, C., 1977, Double loop learning in organizations: Harvard Business Review, v. 55, no. 5, p. 115-125, <https://hbr.org/1977/09/double-loop-learning-in-organizations>.
- Bair, L.S., Rogowski, D.L., and Neher, C., 2016, Economic value of angling on the Colorado River at Lees Ferry—Using secondary data to estimate the influence of seasonality: North American Journal of Fisheries Management, v. 36, no. 6, p. 1229-1239, <https://doi.org/10.1080/02755947.2016.1204388>.
- Bair, L.S., Yackulic, C.B., Springborn, M.R., Reimer, M.N., Bond, C.A., and Coggins, L.G., 2018, Identifying cost-effective invasive species control to enhance endangered species populations in the Grand Canyon, USA: Biological Conservation, v. 220, p. 12-20, <https://doi.org/10.1016/j.biocon.2018.01.032>.
- Bishop, R.C., Boyle, K.J., Welsh, M.P., Baumgartner, R.M., and Rathbun, P.R., 1987, Glen Canyon Dam releases and downstream recreation—An analysis of user preferences and economic values: Madison, Wisc., Huberlein-Baumgartner Research Service, submitted to Bureau of Reclamation, National Technical Information Service, Springfield, VA, NTIS report PB88-182546/AS, Glen Canyon Environmental Studies report no. GCES/27/87, 188 p., <https://gcdamphistory.org/key-reading/1987-bishop-et-al-glen-canyon-dam-releases-and-downstream-recreation-an-analysis-of-user-preferences-and-economic-values/1987-bishop->

[et-al-glen-canyon-dam-releases-and-downstream-recreation-an-analysis-of-user-preferences-and-economic-values-3/](#).

- Bolam, F.C., Grainger, M.J., Mengersen, K.L., Stewart, G.B., Sutherland, W.J., Runge, M.C., and McGowan, P.J.K., 2019, Using the value of information to improve conservation decision making: *Biological Reviews*, v. 94, no. 2, p. 629-647, <https://doi.org/10.1111/brv.12471>.
- Brest, P., and Krieger, L.H., 2010, *Problem solving, decision making, and professional judgment—A guide for lawyers and policy makers*: New York, NY, Oxford University Press.
- Casper, A., Jackson, D., Gido, K.B., Peterson, J., and Rahel, F.J., 2016, Final report for the Fisheries Program Review (PEP): Glen Canyon Dam Adaptive Management Program (GCDAMP) Protocols Evaluation Panel, p. 33, http://gcdamp.com/images_gcdamp_com/8/80/Fisheries_PEP_Report_2016-12-27.pdf.
- Chan, K.M., Gould, R.K., and Pascual, U., 2018, Editorial overview—Relational values—what are they, and what’s the fuss about?: *Current Opinion in Environmental Sustainability*, v. 35, p. A1-A7, <https://doi.org/10.1016/j.cosust.2018.11.003>.
- Donovan, P., Bair, L.S., Yackulic, C.B., and Springborn, M.R., 2019, Safety in numbers—Cost-effective endangered species management for viable populations: *Land Economics*, v. 95, no. 3, p. 435-453, <https://doi.org/10.3368/le.95.3.435>.
- Doremus, H., 2011, Adaptive management as an information problem: *North Carolina Law Review*, v. 89, p. 1455-1498, <https://scholarship.law.unc.edu/nclr/vol89/iss5/5/>.
- Duffield, J., Neher, C., and Patterson, D., 2016, Economic analysis of Glen Canyon angler and Grand Canyon whitewater visitor surveys: Missoula, University of Montana, Department of Mathematical Sciences, prepared for U.S. Geological Survey, Grand Canyon Monitoring and Research Center, 160 p., https://www.gcdamp.com/images_gcdamp_com/e/e0/Whitewater_Angler_FINAL_REPORT.pdf.
- Gaston, T.D., Harpman, D.A., Platt, J., and Piper, S., 2015, Recreation economic analysis for the Long-Term Experimental and Management Plan Environmental Impact Statement: Bureau of Reclamation, Upper Colorado River Basin, Technical report EC-2014-03, August 2015.
- Gould, R.K., Bremer, L.L., Pascua, P.a., and Meza-Prado, K., 2020, Frontiers in cultural ecosystem services—Toward greater equity and justice in ecosystem services research and practice: *BioScience*, v. 70, no. 12, p. 1093-1107, <https://doi.org/10.1093/biosci/biaa112>,
- Hamilton, J.R., Hanemann, W.M., Loomis, J.B., and Peters, L., 2010, Final report of the GCMRC socioeconomic research review panel—Report of a workshop held December 2-3, 2009, Phoenix, Ariz.: Flagstaff, Ariz., prepared for the U.S. Geological Survey, Grand Canyon Monitoring and Research Center, February 26, 2010, https://www.usbr.gov/uc/progact/amp/twg/2010-11-15-twg-meeting/Attach_08a.pdf.
- Himes, A., and Muraca, B., 2018, Relational values—The key to pluralistic valuation of ecosystem services: *Current Opinion in Environmental Sustainability*, v. 35, p. 1-7, <https://doi.org/10.1016/j.cosust.2018.09.005>.

- Hoelting, K., and Gould, R., 2021, Federal policy brief—Incorporating cultural ecosystem services into decision-making—Summary Learnings from ACES panel on cultural ecosystem services: ACES—A Community on Ecosystem Services, 3 p., <https://conference.ifas.ufl.edu/aces/prior/aces21/virtual/presentations/Federal-Policy-Brief-on-Cultural-Ecosystem-Services-and-Decision-Making-February-2022.pdf>.
- Hoelting, K.R., Gould, R.K., Cravens, A.E., and Winter, B.D., 2024c, *In Press*, Constraints and enablers for meaningful consideration of plural values through integration of cultural ecosystem services (CES) in decision-making, *in* McElwee, P., Allen, K., Gould, R., He, J., and Hsu, M., eds., *Routledge Handbook of Cultural Ecosystem Services*.
- Hoelting, K.R., Martinez, D.E., Bair, L.S., Schuster, R.M., and Gavin, M.C., 2023, An opportunities framework for improved integration of cultural-benefits-knowledge in environmental decision-making: SocArXiv, <https://osf.io/preprints/socarxiv/v6fxs>.
- Hoelting, K.R., Martinez, D.E., Schuster, R.M., and Gavin, M.C., 2024a, Advancing knowledge pluralism and cultural benefits in ecosystem services theory and application: *Ecosystem Services*, v. 65, article 101583, <https://doi.org/10.1016/j.ecoser.2023.101583>.
- Hoelting, K.R., Morse, J.M., Gould, R.K., Martinez, D.E., Hauptfeld, R.S., Cravens, A.E., Breslow, S.J., Bair, L.S., Schuster, R.M., and Gavin, M.C., 2024b, Opportunities for improved consideration of cultural benefits in environmental decision-making: *Ecosystem Services*, v. 65, article 101587, p. 1-21, <https://doi.org/10.1016/j.ecoser.2023.101587>.
- Martin, A., Coolsaet, B., Corbera, E., Dawson, N.M., Fraser, J.A., Lehmann, I., and Rodriguez, I., 2016, Justice and conservation—The need to incorporate recognition: *Biological Conservation*, v. 197, p. 254-261, <https://doi.org/10.1016/j.biocon.2016.03.021>.
- Neher, C., Bair, L.S., Duffield, J., Patterson, D., and Neher, K., 2018, Convergent validity between willingness to pay elicitation methods—An application to Grand Canyon whitewater boaters: *Journal of Environmental Planning and Management*, v. 62, no. 4, p. 611-625, <https://doi.org/10.1080/09640568.2018.1435411>.
- Neher, C., Duffield, J., Bair, L.S., Patterson, D., and Neher, K., 2017, Testing the limits of temporal stability—Willingness to pay values among Grand Canyon whitewater boaters across decades: *Water Resource Research*, v. 53, no. 12, p. 10108-10120, <https://doi.org/10.1002/2017WR020729>.
- Pahl-Wostl, C., 2009, A conceptual framework for analysing adaptive capacity and multi-level learning processes in resource governance regimes: *Global Environmental Change*, v. 19, no. 3, p. 354-365, <https://doi.org/10.1016/j.gloenvcha.2009.06.001>.
- Powell, W.B., 2011, The challenges of dynamic programming, *in* *Approximate Dynamic Programming—Solving the Curses of Dimensionality*: Hoboken, NJ, Wiley Series in Probability and Statistics, John Wiley & Sons, Inc., p. 1-16, <https://doi.org/10.1002/9780470182963>.
- Roberts, C.A., Stallman, D., and Bieri, J.A., 2002, Modeling complex human-environment interactions—The Grand Canyon river trip simulator: *Ecological Modelling*, v. 153, no. 1-2, p. 181-196, [https://doi.org/10.1016/S0304-3800\(01\)00509-9](https://doi.org/10.1016/S0304-3800(01)00509-9).

- Runge, M.C., Converse, S.J., and Lyons, J.E., 2011, Which uncertainty? Using expert elicitation and expected value of information to design an adaptive program: *Biological Conservation*, v. 144, no. 4, p. 1214-1223, <https://doi.org/10.1016/j.biocon.2010.12.020>.
- Runge, M.C., LaGory, K.E., Russell, Kendra, Balsom, J.R., Butler, R.A., Coggins, L.G., Jr., Gantz, K.A., Hayse, John, Hlohowskyj, Ihor, Korman, Josh, May, J.E., O'Rourke, D.J., Poch, L.A., Prairie, J.R., VanKuiken, J.C., Van Lonkhuyzen, R.A., Varyu, D.R., Verhaaren, B.T., Vesekla, T.D., Williams, N.T., Wuthrich, K.K., Yackulic, C.B., Billerbeck, R.P., and Knowles, G.W., 2015, Decision analysis to support development of the Glen Canyon Dam Long-Term Experimental and Management Plan: U.S. Geological Survey Scientific Investigations Report 2015–5176, 64 p., <https://doi.org/10.3133/sir20155176>.
- Smallmouth Bass Ad Hoc Group, 2023, Invasive fish species below Glen Canyon Dam—A strategic plan to prevent, detect and respond: Phoenix, Ariz., Glen Canyon Dam Adaptive Management Program in partnership with the U.S. Geological Survey Grand Canyon Monitoring and Research Center and the Bureau of Reclamation, presented to the Glen Canyon Dam Adaptive Management Group Technical Work Group, January 26, 2023, 55 p., <https://www.usbr.gov/uc/progact/amp/amwg/2023-02-16-amwg-meeting/20230216-InvasiveFishSpeciesBelowGlenCanyonDam-508-UCRO.pdf>.
- Springborn, M.R., and Faig, A., 2019, Moving forward—A simulation-based approach for solving dynamic resource management problems: *Marine Resource Economics*, v. 34, no. 3, p. 199-224, <https://doi.org/10.1086/704637>.
- U.S. Department of the Interior, 2016, Record of Decision for the Glen Canyon Dam Long-Term Experimental and Management Plan final Environmental Impact Statement (LTEMP ROD): Salt Lake City, Utah, U.S. Department of the Interior, Bureau of Reclamation, Upper Colorado Region, National Park Service, Intermountain Region, 196 p., http://ltempis.anl.gov/documents/docs/LTEMP_ROD.pdf.
- White House, 2021, Memorandum for the heads of departments and agencies—Indigenous Traditional Ecological Knowledge and Federal Decision Making: Washington D.C., Executive Office of the President, Office of Science and Technology Policy, Council on Environmental Quality, November 15, 2021.
- White House, 2022, Memorandum for heads of federal departments and agencies—Guidance for federal departments and agencies on Indigenous knowledge: Washington D.C., Executive Office of the President, Office of Science and Technology Policy, Council on Environmental Quality, November 30, 2024.
- Williams, B.K., and Brown, E.D., 2018, Double-loop learning in adaptive management—The need, the challenge, and the opportunity: *Environmental Management* v. 62, p. 995–1006, <https://doi.org/10.1007/s00267-018-1107-5>.
- Yackulic, C.B., Bair, L.S., Eppheimer, D.E., Salter, G.L., Deemer, B.R., Butterfield, B.J., Kasprak, A., Caster, J.J., Fairley, H.C., Grams, P.E., Mihalevich, B.A., Palmquist, E.C., and Sankey, J.B., 2024, Modeling the impacts of Glen Canyon Dam operations on Colorado River resources: Phoenix, Ariz., U.S. Department of the Interior, Bureau of Reclamation,

cooperator publication prepared by U.S. Geological Survey, Southwest Biological Science Center, Grand Canyon Monitoring and Research Center, Flagstaff, Ariz., April 2024, 133 p., <https://pubs.usgs.gov/publication/70252976>.

Project K: Geospatial Science, Data Management and Technology Project

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Project Summary and Purpose

A crucial component of any long-term adaptive management program is the proper management and accessibility of its data resources necessary for measuring the status, trends, and experimental results related to the program’s objectives. The data collected through the U.S. Geological Survey’s (USGS) Grand Canyon Monitoring and Research Center (GCMRC) are a vital resource used to determine the status of the natural resources identified through the Glen Canyon Dam Adaptive Management Program (GCDAMP) and to make informed and timely decisions on dam operations. Proper data management and data accessibility is critical to protecting the significant GCDAMP investment made over more than two decades in collecting these data. The primary purpose of this project is to provide high-level support to GCDAMP-funded science efforts in the disciplines of geospatial science, data management, database administration, and emerging information technologies. The elements described in this project are a larger part of an organizational approach to the use of geographic information systems (GIS), data management, and information technology in support of the science projects contained within this triennial work plan.

Shifts in the geospatial and information technology industries are pushing the boundaries on how data can be managed and made accessible to outside entities. Much of this change is driven by advances in technology—from improved sensors for monitoring the Earth to increased digital data storage capacity to newer computer systems designed for processing large data sets more efficiently to the greater emphasis of the “Internet of Things (IoT)” where the reliance of web-based technologies has revolutionized our world. This project continues to advance GCMRC’s ability to leverage many of these new technologies for the benefit of the science projects described within this work plan, and the larger adaptive management program. Work performed within this project makes it possible to share important information about trends in resources of the Colorado River ecosystem to the GCDAMP through web-based, interactive tools and mapping products, allowing the GCDAMP to make better informed, time-sensitive decisions on experimental and management actions under the 2016 Long-Term Experimental and Management Plan (LTEMP) and the associated Record of Decision (ROD) (U.S. Department of the Interior, 2016a, b).

Long-Term Experimental and Management Plan (LTEMP) Resource Goals

Project K supports the following LTEMP Resource Goals:

- Archaeological and Cultural Resources
- Natural Processes
- Humpback Chub
- Other Native Fish
- Recreational Experience
- Sediment
- Rainbow Trout
- Nonnative Invasive Species
- Riparian Vegetation

Support is provided to science projects across the three project elements described here. Specifics on the type and level of support within each project element for projects aligning to the LTEMP Goals is outlined in cross-reference tables (Figures 2, 3, and 6) that can be found within each project element section.

Background

Data management in support of research and monitoring has been a part of GCMRC since its inception and was specifically outlined in the 1995 Environmental Impact Statement that clearly defines the GCMRC's responsibilities for managing data in support of the GCDAMP (U.S. Department of the Interior, 1995). The concept of data management encompasses many facets including, but not limited to, data preservation, design, development and maintenance of systems and applications designed to store and serve the data, building systems that provide access to these data, and performing the necessary documentation of data sets. This work was also supported in the 1995 ROD – specifically in GCDAMP Goal 12, to maintain a high-quality monitoring, research and adaptive management program – and in subsequent documents including the most recent LTEMP. Success of LTEMP will rely heavily on the GCMRC's ability to continue to improve on data accessibility for stakeholders, managers, and, when appropriate, the general public (U.S. Department of the Interior, 2016a).

GCMRC has experienced some large shifts in how its data resources have been managed. These shifts have often reflected major organizational changes– both internal and external. Prior to the creation of this project in the first GCMRC Triennial Work Plan for fiscal years 2015 to 2017, the support functions describe within this project were handled through various permutations in organization structure for both GCMRC and the larger Southwest Biological Science Center,

with punctuated events that impacted how the GIS and data support elements functioned. Examples of these events include the creation of the Southwest Biological Science Center in 2005 and eventual disbanding of the original GCMRC Information Technology (IT) Program, and somewhat similarly, the disbanding of GCMRC's program structure altogether in 2013. Such events do have implications on how a long-term monitoring program continues to maintain consistency in its data resources. One challenge that arose from both events was the potential for decentralization of scientific data, and the rise of independent project data management workflows. There are consequences of not standardizing our data management practices, and so Project K has become increasingly involved with these efforts since the first Triennial Work Plan (TWP) of FY 2015-17, and that work continues in this next TWP.

Many reasons exist for an organization to maintain capacity for data management and geospatial expertise within a team dedicated to these tasks. Issues can arise when these responsibilities fall solely on science project staff. At the organizational level, inefficiencies in data storage, data access, and analytical abilities can occur due to decisions made within a specific project as to the data organization, software used to process the data, and separate workflows developed for accessing the data. These inefficiencies translate into higher costs incurred through project staff salary when attempting to perform all data management related tasks while also performing regular project work. These higher costs become magnified when applied across multiple projects or out to an entire science center. Data loss, science project staff turnover, and data siloing are challenges that are indicative of an organization that lacks centralized support for data management functions. For GCMRC, these are challenges that have been experienced in the past, and something that Project K strives to help the organization avoid. Science project staff turnover, in particular, can be expected throughout a 3-year work plan cycle and certainly more likely for longer periods of time within the LTEMP. This may compromise GCMRC's ability to address the larger questions efficiently and effectively, such as the 10-year assessment that is called for in the LTEMP ROD (U.S. Department of the Interior, 2016b; VanderKooi and others, 2017). We acknowledge that staff turnover may also occur within Project K, and indeed it has occurred since the last triennial work plan. The potential for staff turnover helps justify the standardization, as much as is possible, of data management practices, technologies used, and components developed to support GCMRC science projects. Adhering to modern data management concepts and applications while standardizing, where applicable, places these efforts into commonly utilized environments where others with experience in data management principles have much familiarity. Additionally, this work helps support adherence to data management FAIR (Findable, Accessible, Interoperable, and Reusable) principles that have become a major focus of the USGS (Hutchison and others, 2024; Wilkinson and others, 2016).

While this project is still fulfilling to its role as the lead in GIS application to science projects, additional roles have also accumulated as natural extensions to the geospatial science work over the past few years (U.S. Department of the Interior, 2014, 2017, 2020).

The functions of data management were previously addressed within the Data Acquisition, Storage, and Analysis (DASA) program; however, since the reorganization of GCMRC in 2013 away from resource programs and towards the current project-oriented focus, some data management responsibilities have become decentralized for GCMRC. Because of this reorganization, positions that traditionally focused primarily on GIS support have had to expand their roles to include data management oversight, as well as providing computer systems expertise, web server and internet technology leadership, the design, development, and deployment of technologically advanced scientific monitoring equipment and, most recently, the adoption of a hybrid-cloud data storage model and hosted application services. Some of this capacity has existed within GCMRC in previous work plan cycles but is now being described more holistically within the context of a work plan (U.S. Department of the Interior, 2014, 2017, 2020). The project elements presented in Section 5 of this project proposal describe this increased capacity more fully.

In addition to its commitment to the GCDAMP and LTEMP, the GCMRC, as a part of the USGS, must comply with federal guidelines governing many aspects of how geospatial data are collected and maintained by the GCMRC. These aspects range from how specific data are to be collected, to accuracy standards established through federal policy – Federal Geographic Data Committee, National Standard for Spatial Data Accuracy (Federal Geographic Data Committee, 1998), to how data are to be reviewed and released in conjunction with peer-reviewed scientific publications. This last concept is relatively new and is more fully described by the most recent USGS Fundamental Science Practices in response to an Executive Order that redefined the data release policies of federal agencies (U.S. Geological Survey, 2017). As the geospatial science project has increased its role in assisting with proper data management for GCMRC, work performed in this project will continue to lead efforts for adhering to these requirements.

The need for data telemetry has long been identified and addressed by GCMRC over the past two decades through a combination of leveraging commercially available data communication systems and custom engineering initiatives to make those systems actually work in the remote, harsh environments of Grand Canyon. In FY 2017, we became involved with an IoT sensor pilot project to test the feasibility of connecting sensors deployed in the field to the Amazon Web Services (AWS) – USGS Cloud Hosting Solutions cloud environment. This pilot work required the reconfiguration of an existing field sensor system (Vaisala weather station) already deployed at Lees Ferry and development of two-way communication capabilities with the sensor and data logger via cellular transmission to the Amazon cloud. The main objective was to demonstrate the ability to automate the transmission of data from the field to the cloud at some predefined interval, and to allow users to subscribe to “alerts” based on defined data values that would then perform some other action — in this case send a text message regarding extreme air temperature alerts. We successfully achieved this initial goal in 2018 and presented our work at the inaugural USGS Sensor Summit workshop in Denver.

Our IoT efforts have now expanded to include the transmission of IoT data from water quality instruments located at the Lees Ferry Gage Station, River Mile 0 and the outflow of the Glen Canyon Dam, River Mile –15.7, and fish passive integrated transponder (PIT) tag scan data from the Little Colorado River. It is important to add that Project K staff have been successful at applying for and being awarded funding from the USGS Cloud Hosting Solutions team and other USGS entities to offset the costs associated with instituting new or improving upon existing data telemetry systems in Grand Canyon.

Proposed Work

Project Element K.1. Enterprise GIS, Geospatial Analysis and Processing (Ongoing Study)

Work performed within this element will continue to provide the same GIS services and support that have been consistently provided to GCMRC for three previous triennial work plans.

This project is continually striving to improve upon GCMRC's ability to manage its expanding data resources which includes both large, canyon-wide and high-resolution data along with smaller, more site-specific geospatial data representing different sampling locations for a variety of resources. For several years the main focus was on designing, developing, and maintaining consistent and accurate geospatial data sets, workflows and analyses in support of science projects. This work is expected to continue, however, there is also a need to provide more analytical support involving geospatial components than in previous work plans. In fiscal years 2025-27, this project will continue to support research and monitoring projects by providing geospatial expertise to most projects on field mapping methods, development of customized maps, sample site unit definition and selection, GIS layer development and metadata review, Python programming, and GIS tool development and support. GIS staff support also involves the oversight and supervision of resource-specific project staff with GIS-related tasks including spatial analysis, training for staff and cooperators in GIS data entry and database management concepts, data processing techniques, production of printed maps and online map products, error troubleshooting, and other basic GIS methods and techniques. These types of support functions are available to all staff and projects funded by GCDAMP, and this has been the case for many years over successive triennial work plans and even prior to the 3-year work plan cycle which began in fiscal year 2015.

Enterprise GIS Administration

This project maintains an enterprise GIS platform that is built upon Environmental Systems Research Institute (ESRI) ArcGIS Portal and Server applications and used for maintaining existing online data resources (ESRI, 2024a).

Data services developed through this online system can then be shared through multiple endpoints including cloud-based content delivery systems, custom web applications hosted on-premise, and through other applications that can integrate spatial data with resource-specific monitoring data to provide users greater context on the trends and conditions of the Colorado River ecosystem.

During FY 2021-23, a repository of Python scripts was developed to convert and move large batches of data to the on-premise Postgres database, from Postgres to the cloud, and from the cloud to ArcGIS Online. This workflow can significantly improve access to data for stakeholders and the public while reducing the time spent by project staff to work data through this process.

During the FY 2025-27 TWP cycle, we will continue to expand on content that is available through this system and work to improve existing functionality, as well as develop new, web-based analytical tools for interacting with and interpreting GCMRC's data.

This Enterprise GIS consists of an on-premise Postgres database instance that serves as the back-end architecture for storing and serving geospatial data both within the DOI internal network and externally through web-based services.

Online services are published using ESRI's ArcGIS Server web-based application which can then be consumed through a variety of online endpoints. GCMRC staff and DOI cooperators can access the on-premise data resources, while both internal and external clients can connect to the data services managed through ArcGIS Server. Project K also manages an online Geospatial Portal content management system (Figure 1) that packages data services into online maps and web applications.

Additional GIS Administration tasks related to science support include the testing and migration of computer systems to newer versions of industry-standard GIS and remote sensing software, maintaining licensing information, and/or working with IT staff to ensure all licenses, software, extensions, add-ons, and custom applications work properly.

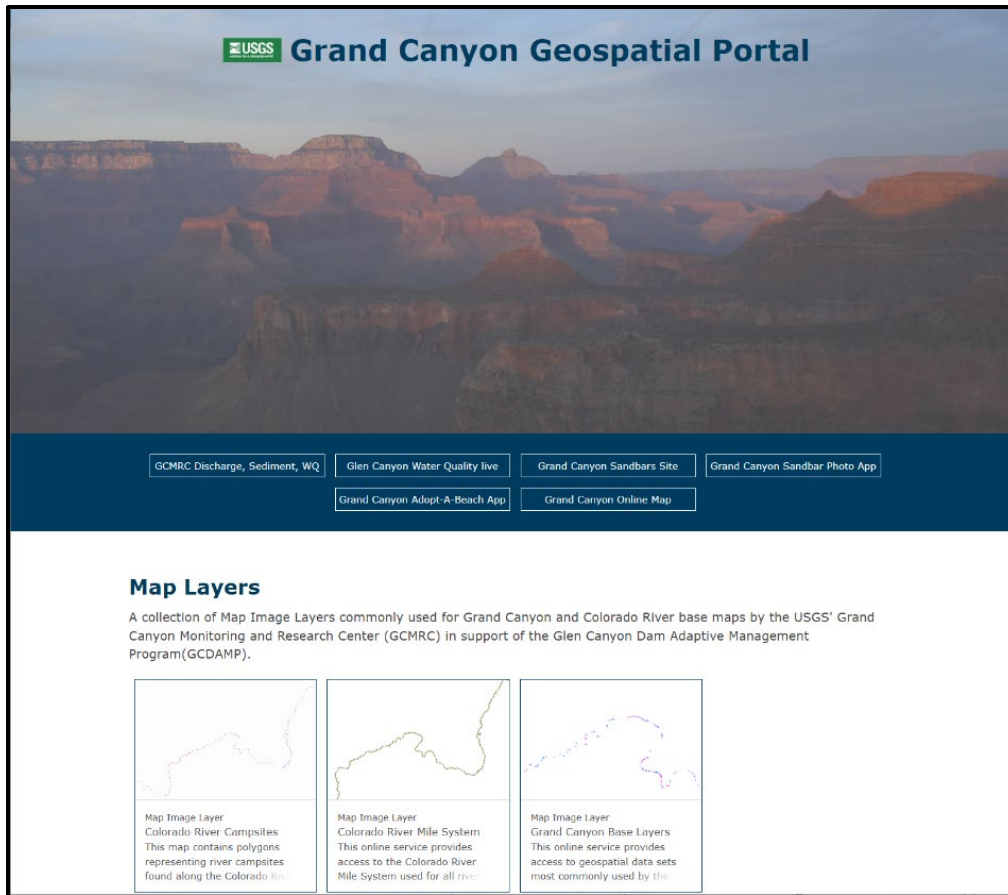


Figure 1. Screen image of GCMRC's Grand Canyon Geospatial Portal that provides access to online geospatial data and contains links to the most used web applications. See URL link in Products subsection.

ArcGIS Online

Another aspect of Project K's GIS administration utilizes external cloud-based platforms for sharing GCMRC geospatial data to a wider audience. The best example of this is the extensive list of services hosted on ESRI's ArcGIS Online content management system (ESRI, 2024b). ArcGIS Online is essentially a cloud-based version of the internal portal that GCMRC maintains on-premise. The power of this system architecture is realized when considering that the data stored and managed on-premise in a PostgreSQL database and served externally through a web server running ArcGIS Server can be simultaneously accessed through numerous maps and applications on both the internal Portal website and the external ArcGIS Online data serving platform.

GIS Support to GCDAMP-funded Projects

Key aspects of the work performed in this element include the processing and analysis of large, complex geospatial data sets that often benefit multiple projects. Examples of this support are evident in past triennial work plans.

One example is the processing and analysis work devoted to the Glen Canyon channel map data set (Kaplinski and others, 2022). Specific tasks performed through GIS staff included processing derived data sets from the 2013 Digital Surface Model to remove vegetation from the surface, thus creating a bare-ground elevation surface to be used in conjunction with field-based topography and bathymetry elevation data to make a composite channel map. In FY 2024, the Python scripts used to create channel mapping datasets were updated for compatibility with ArcGIS Pro and function as custom geoprocessing tools to streamline data analysis. This project element is also responsible for publishing geospatial data sets that support all GCMRC projects, as well as working collaboratively with principal investigators to develop and publish data releases, data series reports and peer-reviewed publications in support of GCDAMP-funded science (Gushue, 2019; Sankey and others, 2024).

While this project element strives to support all projects in all facets of GIS, geospatial analysis and geospatial data processing, the needs of individual projects vary widely between each project and, to a lesser degree, from year to year among projects. This project element attempts to align its support with the work proposed by each project for the next triennial work plan and estimates the anticipated support by drawing from both previous GCMRC work plans and gauging on-going efforts in the current fiscal year. In this way, GIS support provided to projects is meant to have some consistency and continuity from year to year and through each work plan cycle. Examples of this support can include GIS operation, geospatial analysis, data management, relational database tasks, programming and code development for data processing and analysis, and web mapping application development, among other things. The frequency and magnitude of requests ebb and flow during the course of a fiscal year, however, an estimate would be that requests at a base level (i.e., lower-level tasks requiring less than 1 day) occur on a daily or weekly basis. Middle level requests range from two days to approximately one week and can occur from weekly to monthly depending on the time of year, and higher-level tasks involving more than one week occur about every other month. A cross-reference table is provided that highlights general geospatial support components for resource-specific projects (Figure 2).

| Project Supported | Asset / Product | Geospatial analysis | Geospatial data processing | Enterprise GIS | Significance of Support | LTEMP Resource Goals |
|-------------------|---------------------------------------|--|--|--|--|--|
| Project A | Sediment and Water Quality | <ul style="list-style-type: none"> GIS expertise provided to project staff for geospatial analysis Python script programming support, GIS tool development | <ul style="list-style-type: none"> GIS expertise for processing geospatial data Geospatial data documentation support Python script programming support for data processing | <ul style="list-style-type: none"> Publish update geospatial data services used in website Development of online content of sediment monitoring | <ul style="list-style-type: none"> Used by researchers and stakeholders across GCDAMP Used for evaluating HFEs | 7. Sediment |
| Project B | Long-term sediment storage monitoring | <ul style="list-style-type: none"> GIS expertise provided to project staff for geospatial analysis Python script programming support, GIS tool development for analysis | <ul style="list-style-type: none"> GIS expertise for processing geospatial data Geospatial data documentation support Python script programming support for data processing | <ul style="list-style-type: none"> Relational database management for geospatial data Publishing geospatial services Online map / web application development | <ul style="list-style-type: none"> Used by researchers and stakeholders across GCDAMP Used for evaluating HFEs | 7. Sediment |
| Project C | Riparian Vegetation Sampling | <ul style="list-style-type: none"> GIS expertise provided to project staff for geospatial analysis Python script programming support, GIS tool development for analysis Train staff in GIS | <ul style="list-style-type: none"> Random sampling site selection support Modify applications tasks with R scripts for performing analysis Link riparian vegetation sampling database to geospatial framework | <ul style="list-style-type: none"> Link vegetation sampling data to geospatial framework Database management Publishing geospatial services Online map / web application development | <ul style="list-style-type: none"> Used for determining visualizing spatial patterns in long-term trends of terrestrial ecology | 11. Riparian Vegetation |
| Project D | Cultural Resources | <ul style="list-style-type: none"> Support in providing necessary base level data sets and advice on best practices for geospatial analysis Leveraging data sets and tools developed by Project K | <ul style="list-style-type: none"> Support for data processing, GIS operation, python programming beyond the capacity found within project staff (as or if needed) Software maintenance support provided | <ul style="list-style-type: none"> Base-level services available for use by project staff Potential to serve data online with the caveat to safeguard sensitive information | <ul style="list-style-type: none"> Support is at a basic level due to staff within project having proficiency in many areas, and lack of need | 1. Archaeological and Cultural Resources |
| Project E | Nutrients | <ul style="list-style-type: none"> Little to no anticipated support needed for geospatial analysis | <ul style="list-style-type: none"> Occasional need for basic level of GIS support – field maps, data import/export, publication maps | <ul style="list-style-type: none"> No support anticipated | <ul style="list-style-type: none"> Support is at a lower level due to lack of need by project | 2. Natural Processes |
| Project F | Aquatic Invertebrates | <ul style="list-style-type: none"> Little to no anticipated support needed for geospatial analysis | <ul style="list-style-type: none"> Occasional need for basic level of GIS support – field maps, data import/export, publication maps | <ul style="list-style-type: none"> Potential to re-host aquatic invertebrate web applications through new enterprise system | <ul style="list-style-type: none"> Support is at a lower level due to lack of need by project | 2. Natural Processes |
| Project G | Fish Monitoring | <ul style="list-style-type: none"> Maintaining / updating fish monitoring units data sets Nonnative rapid response data curation and analysis Geospatial analysis leveraging fish monitoring database | <ul style="list-style-type: none"> QA/QC support on fish monitoring data for geospatial characteristics Inclusion of fish monitoring data into geospatial framework Field river map books generation and production | <ul style="list-style-type: none"> Relational database management for geospatial data Publishing geospatial services Online map / web application development | <ul style="list-style-type: none"> Used by GCMRC staff and cooperators to put spatial context of fish populations | 3. Humpback Chub |
| Project H | Fish Monitoring | <ul style="list-style-type: none"> Maintaining / updating fish monitoring units data sets Database software installation and updates Database maintenance | <ul style="list-style-type: none"> QA/QC support on fish monitoring data for geospatial characteristics Inclusion of fish monitoring data into geospatial framework Field river map books generation and production | <ul style="list-style-type: none"> Relational database management for geospatial data Publishing geospatial services on nonnative captures Online map / web application development | <ul style="list-style-type: none"> Used by GCMRC staff and cooperators to put spatial context of fish populations | 9. Rainbow Trout Fishery |
| Project I | Fish Monitoring | <ul style="list-style-type: none"> Maintaining / updating fish monitoring units data sets Nonnative rapid response data curation and analysis Database maintenance | <ul style="list-style-type: none"> QA/QC support on fish monitoring data for geospatial characteristics Field river map books generation and production, general field support | <ul style="list-style-type: none"> Publishing geospatial services such as nonnative rapid response presence | <ul style="list-style-type: none"> Used by GCMRC staff and cooperators to put spatial context of fish populations | 3. Humpback Chub 5. Other Native Fishes 10. Nonnative Invasive Species |

Figure 2. Table shows geospatial analysis, processing and enterprise support for resource specific GCMRC science projects funded by GCDAMP. Boxes in white denote unknown assets or areas where Project K is not anticipating providing support.

Science Questions

Project K does not address specific science questions as it is inherently a supportive effort for GCDAMP-funded projects and an important resource for geospatial and data management functions. The purpose of this project is to deliver critical support to science projects funded by GCDAMP by contributing expertise across a wide array of data-centric themes including data processing, data management and documentation, programming and source control, geospatial processing and analysis, and advanced data telemetry applications which are essential to the success of nearly all projects. The following justifications have been used to guide this work plan proposal:

- Data management, including geographic information systems (GIS), has been a part of GCMRC's role in GCDAMP since its inception, and was also supported in the 1995 ROD – specifically in GCDAMP Goal 12, to maintain a high-quality monitoring, research and adaptive management program (U.S. Department of the Interior, 1996).
- Subsequent documents, including the most recent LTEMP, have reaffirmed this important aspect of GCMRC and the adaptive management program (U.S. Department of the Interior, 2016a).
- Project K is designed to support the other proposed science projects that are aligned with resource goals identified in the LTEMP and in more recent Department of the Interior guidance where both documents call for continuity in resource monitoring and consistency in providing high-quality monitoring and research to the Adaptive Management Program (U.S. Department of the Interior, 2016a; Petty, 2019).
- Project K works to share important information about trends in resources of the Colorado River ecosystem through web-based, interactive tools and mapping products (VanderKooi and others, 2017).
- Project K allows for the ability to make better informed, time-sensitive decisions on experimental and management actions under the 2016 LTEMP and the associated ROD (U.S. Department of the Interior, 2016a, b).
- Project K staff participate in USGS-led initiatives, serve on advisory boards and communities of practice, and work to adhere to data-related policies and best management best practices instituted by USGS and the Department of the Interior (Hutchison and others, 2024).

Methods

Project K uses standard geoprocessing and spatial analysis tools and methods for performing much of the requested support to science projects as well as for performing the maintenance of both on-premise, file-base geospatial data sets and the Enterprise GIS relational database environment. Often, however, the level of work requires the development of automated and repeatable methods through scripting which is usually handled through Python code programming to develop reusable and shareable scripts that are commonly run within the ESRI ArcGIS environment. Similarly, advanced geospatial analysis in support of LTEMP-related science projects requires extending the existing desktop software applications through code development. We utilize the USGS-approved GitLab repository for maintaining, updating and sharing these code-base geoprocessing and analysis methods.

Anticipated Use of Data

The data generated and maintained through this Project has a multitude of uses. Internal data resources are made available to science project staff through the internal network and disk storage array that is mapped consistently to each computer. These shared resources are nearly 80TB and extend back to the era of research and monitoring that precedes GCMRC.

Use of online geospatial data resources generated, maintained, and served by this Project is ubiquitous. Services are used in custom web applications, online content stored in ArcGIS Online and when shared publicly can be consumed by the estimated 1.9 million AGOL users, while final web applications served from this platform are available to anyone with an internet connection.

Outcomes and Products

Outcomes

An open environment for sharing geospatial data, geoprocessing tools and spatial analysis scripts in support of science projects and associated LTEMP Goals. Note: Some links listed under Products are not accessible outside of the DOI network, however, sharing outside the DOI may be available upon request.

Products

- Numerous geodatabases contain feature layers, imagery data and other data types representing sampling locations, mapping products such as topographic and bathymetric data sets, derived data sets used for or are results of geoprocessing and spatial analysis.
 - GitLab repository of GCMRC GIS Geospatial Tools: [sbsc / gcmrc / gis / Geospatial Tools · GitLab \(usgs.gov\)](#)
 - GitLab repository of GCMRC GIS Lake Powell Geospatial Tools: [Lake Powell Geospatial · GitLab \(usgs.gov\)](#)
- Online geospatial services representing sampling locations, internally generated map products, derived data sets and canyon-wide base layer information for the Colorado River ecosystem in Grand Canyon.
 - URL link to [GCMRC GIS](#) on ArcGIS Online: <https://www.arcgis.com/home/search.html?q=GCMRC&t=content&restrict=false>
 - GCMRC Portal URL link: <https://grandcanyon.usgs.gov/portal/home/index.html>

- Web mapping applications and other more data-driven applications with spatial components included that are served online and available for use by staff, management, stakeholders, and general public.
 - SBSC GCMRC Website: [GCMRC Data, Web Applications, and Tools | U.S. Geological Survey \(usgs.gov\)](https://www.usgs.gov/data/data-repository/gcmrc)

Project Element K.2. Data Management and Database Administration (Ongoing Study; Partially Funded)

During the last three years this project has worked towards addressing the need to expand concepts developed in GIS to other data resources across GCMRC. This includes the further development of an integrated spatial and tabular relational database environment for GCMRC.

To this end, Project K will continue to incorporate much of the relational database work in support of other science projects defined in this work plan. By building the expertise and capacity in data management, data acquisition, and relational database administration within one group, this project is better aligned to provide more comprehensive support to resource-specific science efforts and to the larger GCDAMP community.

Database Administration

Work proposed within this project element includes the continued maintenance of existing relational databases in support of LTEMP related science efforts, and in some cases, the design and development of new databases for projects or resources. Existing, resource-specific databases that have been developed or managed through this project include Sandbar Area and Volume, Riparian Vegetation Survey, Geodetic Network Control, and Lake Powell Water Quality. For the FY 2025-27 TWP, the primary focus will be on the full documentation, redesign, and re-implementation of the existing fish monitoring database. The fish monitoring database is one of the most important data resources maintained by GCMRC, and this project is now better positioned to greatly improve the entire workflow process for storing, reviewing, analyzing, and accessing fish/aquatic information. The migration of all project data to relational databases will continue throughout this next triennial work plan. A cross-reference table (Figure 3) is provided that highlights the database and application development components that have been identified as being needed to support resource-specific projects in the next triennial work plan. The red text in the table denotes identified work tasks that are necessary to support the respective project and LTEMP resource goal(s) but may not occur in the short-term due to cost increases. The Data Scientist support position in fiscal year 2027 is anticipated to address some of the work tasks shown in red.

The shift in how we approach our data resources will provide a consistent and stable platform for processing and analyzing much of GCMRC’s monitoring and research data within this work plan and beyond.

This work will build upon achievements made in database administration since the last triennial work plan. The largest shift since then was the migration away from Oracle and into PostgreSQL, an open-source relational database platform that has provided a fundamental shift in how we manage and access some high-profile data assets.

| Project Supported | Asset / Product | IT / Database Tasks | Application development | Significance of Support | LTEMP Resource Goals |
|-------------------|--|---|---|---|--|
| Project A | GCMRC Discharge, Sediment and Water Quality website | <ul style="list-style-type: none"> • Coordination with Data Center • Server configuration • Database software installation and updates • Database maintenance | <ul style="list-style-type: none"> • Web application maintenance • Web app development | <ul style="list-style-type: none"> • Used by researchers and stakeholders across GCDAMP • Used for evaluating HFEs | 7. Sediment |
| Project B | Sandbar Area and Volume website and application | <ul style="list-style-type: none"> • Provide some maintenance efforts • Provide updates to online database | <ul style="list-style-type: none"> • Retain code in source control • Troubleshoot issues / errors • Provide | <ul style="list-style-type: none"> • Used by researchers and stakeholders across GCDAMP • Used for evaluating HFEs | 7. Sediment 6. Recreational Experience |
| Project C | Riparian Vegetation Sampling Database | <ul style="list-style-type: none"> • Database maintenance • Train / support staff with accessing database | <ul style="list-style-type: none"> • Maintain data entry application • Modify applications tasks with R scripts for performing analysis | <ul style="list-style-type: none"> • Used for determining long-term trends in terrestrial ecological and responses to operations | 11. Riparian Vegetation |
| Project D | It is unclear what data are being collected within this project that would require support | <ul style="list-style-type: none"> • Geospatial data can be stored in the PostgreSQL environment • Non-sensitive geospatial data can be stored and served | <ul style="list-style-type: none"> • Sensitive nature of data collected in this project limits what can be shared online | <ul style="list-style-type: none"> • Limited to no database support | 1. Archaeological and Cultural Resources |
| Project E | This project utilizes data collected in other projects, in particular, Project A | | | | 2. Natural Processes |
| Project F | Aquatic Invertebrate "Foodbase" database | <ul style="list-style-type: none"> • Currently these data reside in a Microsoft Access database • Automated database backups are pushed to a MS SQL Server backend database | <ul style="list-style-type: none"> • A third-party internal Access application is used for data entry and QA/QC | <ul style="list-style-type: none"> • If K is fully funded, this asset can be upgraded to an enterprise relational database system in the TWP | 2. Natural Processes |
| Project G | Fish Monitoring Database | <ul style="list-style-type: none"> • Server configuration • Database software installation and updates • Database maintenance | <ul style="list-style-type: none"> • Development of processes for data entry, QA/QC and inclusion into master database | <ul style="list-style-type: none"> • Used by GCMRC staff and cooperators to estimate fish populations | 3. Humpback Chub |
| Project H | Fish Monitoring Database | <ul style="list-style-type: none"> • Server configuration • Database software installation and updates • Database maintenance | <ul style="list-style-type: none"> • Development of processes for data entry, QA/QC and inclusion into master database | <ul style="list-style-type: none"> • Used by GCMRC staff and cooperators to estimate fish populations | 9. Rainbow Trout Fishery |
| Project I | Fish Monitoring Database | <ul style="list-style-type: none"> • Server configuration • Database software installation and updates • Database maintenance | <ul style="list-style-type: none"> • Development of processes for data entry, QA/QC and inclusion into master database | <ul style="list-style-type: none"> • Used by GCMRC staff and cooperators to estimate fish populations | 3. Humpback Chub 5. Other Native Fishes 10. Nonnative Invasive Species |

Figure 3. Table showing Grand Canyon Monitoring and Research Center resource-specific projects with relational database support from this project element. Boxes in white denote unknown assets or areas where Project K is not anticipating to provide support. Reasons are given, where applicable. Red text denotes work tasks that are planned to align with the Computer / Data Scientist position in FY 2027.

The fish monitoring database is one of those data assets now currently stored in PostgreSQL. We followed an Infrastructure as Code (IaC) process to generate our PostgreSQL environment which can allow for faster rebuilds due to system failure and aligns these data resources with a hybrid-cloud strategy that will allow for us to replicate these systems in Amazon Web Services (AWS) or other cloud platforms.

In this next triennial work plan, we anticipate that the advance data serving systems developed for geospatial data sets and more recently for the resource-specific data sets, will be better integrated for analysis and informational products. This is illustrated in the GCMRC Enterprise Data System schematic (Figure 4).

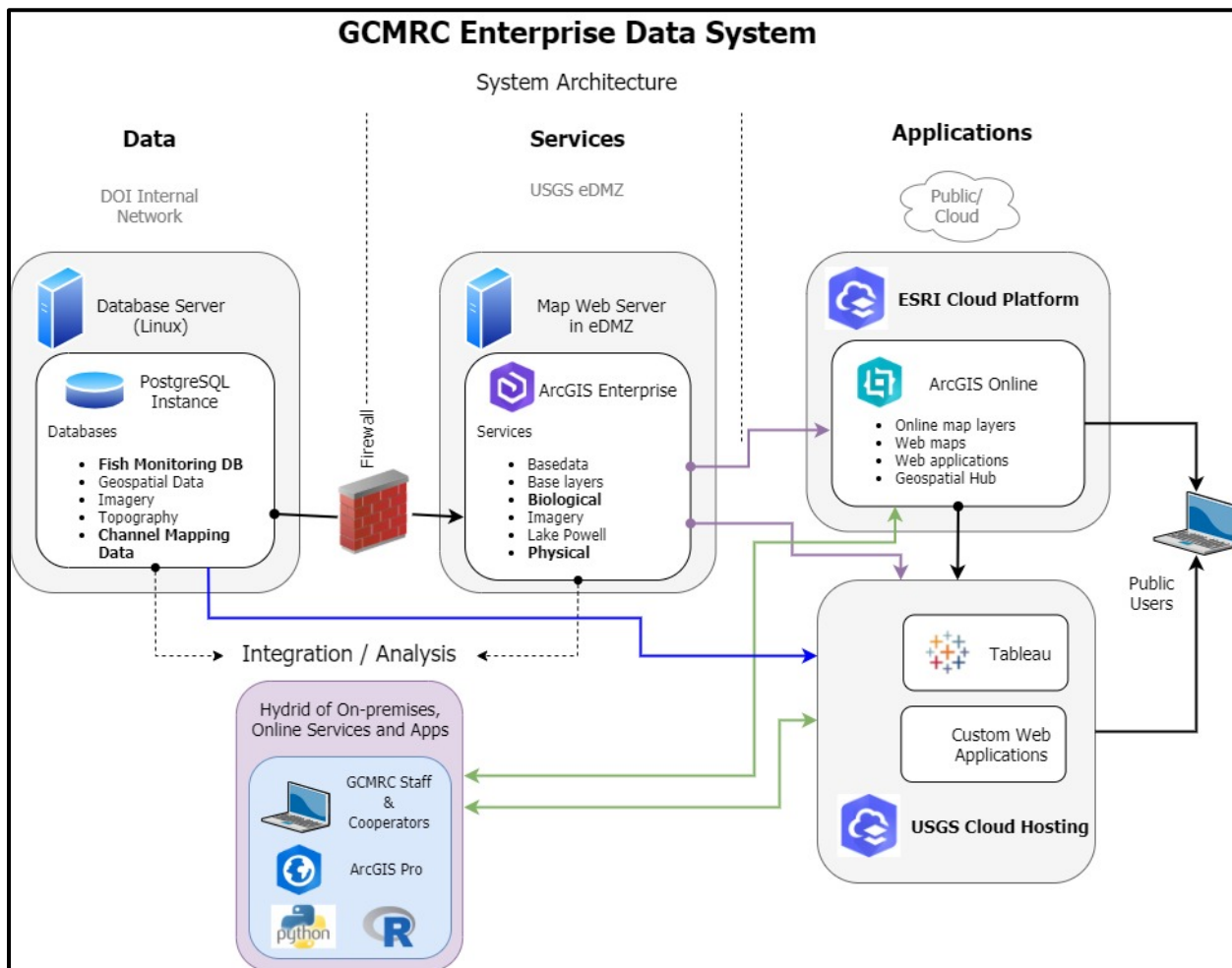


Figure 4. Grand Canyon Monitoring and Research Center enterprise data system architecture schematic, with some example data sets listed within the tiers of this system.

Project K works collaboratively with the SBSC IT staff to address many aspects of data management from a Center-wide perspective, including but not limited to the adoption of hybrid-cloud strategies for data management, the shift for on-premise data backups to leverage Cloudberry and AWS tiered storage, the implementation of Department of the Interior data management policies and data-centric strategies adopted by the USGS for aligning with those policies (Hutchinson and other, 2024).

Cloud-based Data Management

Since 2017, this project has led GCMRC's efforts to adopt and use cloud-based environments for providing better access to its data and applications. By working with the USGS Cloud Hosting Solutions team, the Project K has continued to lead the way for GCMRC in expanding the use of the AWS cloud environment for leveraging cost effective, advanced cloud computing solutions, application development and deployment, and providing access to information through some of the most advanced data serving systems available for natural resource monitoring today.

Modern application of enterprise databases involves standardized source control of all application components, advance system configuration of both local desktop and server environments, and the proper deployment and management of AWS cloud-based components. There are many benefits to leveraging these cloud environments for science applications. They offer scalable resources, many of which only incur costs while the components are being accessed. The cost of server maintenance, security, data/application availability, storage, and redundancy are all managed by AWS, thus reducing the amount of time needed internally for information technology staff to perform these duties. This project will continue to lead GCMRC in adoption of a hybrid-cloud strategy for future data management and application development, as well as continue to investigate other data management solutions being provided through the USGS and the DOI.

Methods

Project element K.2. adheres to commonly used data management principles using standard and, in some cases, advanced relational database management practices, such as employing Infrastructure as Code (IaC) for some systems (Kumara and others, 2021). This means that the actual relational database structure, permissions/privileges, and data all could be re-built from code and database backups in the event of data corruption, database failover or other forms of loss. Guidance is provided through DOI policy and further supported through USGS strategic planning (Hutchison and others, 2024). Some methods used for data management in support of science projects are consistent across all projects, such as file-based data are stored on an internal disk array on the network and receive scheduled backups to the AWS cloud. However, the uniqueness in the types, amounts and frequency of collection for different science project data sets perpetuates the need for the development of custom solutions to acquire, import/upload, manage, maintain, and provide access to these different data resources. Some similarities do exist between different science project data workflows, for instance, data entry and quality accuracy / quality control (QA/QC) software exists for several projects. These Create, Read, Update and Delete (CRUD) applications have been designed to work with data assets maintained in Microsoft SQL Server relational database software, and include the Lake Powell Water Quality database, the Riparian Vegetation database, and the Geodetic Control database.

Conversely, the Fish Monitoring database is hosted in a PostgreSQL database on a Linux environment, and uses a suite of Python, R scripting and SQL programming scripts to perform the data entry, import/upload, QA/QC, and maintenance tasks. As with geoprocessing and spatial analysis scripting methods, these data management scripts are stored using source control methods in the USGS-approved GitLab repository.

Anticipated Use of Data

The data generated and maintained through this Project have a multitude of uses. Internal data resources are made available to science project staff through the internal network and disk storage array that is mapped consistently to each computer. These shared resources are nearly 80TB and extend back to the era of research and monitoring that precedes GCMRC. Here we do not differentiate between geospatial data sets and those that are tabular-based data sets.

Data are used for reporting on the condition of riparian resources of the Colorado River ecosystem. Use of GCMRC data resources can range from real-time consumption of advanced data analytics to internal use by science staff for analyzing data to advanced modeling efforts for determining status and trends of key resources.

Outcomes and Products

Outcomes

An open environment for sharing enterprise relational database information for corresponding science projects and those associated LTEMP goals. Note: Some links listed under Products are not accessible outside of the DOI network, however, sharing outside the DOI may be available upon request.

Database Products

PostgreSQL relational databases:

- Fish Monitoring Database
- Discharge, Sediment, and Water Quality Monitoring Database*
- Base Geospatial Enterprise Geodatabase
- Biological Geospatial Database
- Physical Geospatial Database
- Imagery Geospatial Database
- Topography Geospatial Database

Microsoft SQL Server databases:

- Lake Powell Water Quality database
- Riparian Vegetation Database
- Geodetic Network and Control Database

SQLite Databases:

- Sediment Storage / Sandbar Monitoring Database

Software / Application Products

- Fish Monitoring Data Processing Code Repository: [fish-monitoring · GitLab \(usgs.gov\)](https://github.com/USGS/fish-monitoring)
- Discharge, Sediment, and Water Quality Monitoring Website*: https://www.gcmrc.gov/discharge_qw_sediment/
- Lake Powell Water Quality (Create, Read, Update and Delete) CRUD Application
- Lake Powell Water Quality Data Exploration Website (planned)
- Riparian Vegetation (Create, Read, Update and Delete) CRUD Application

*Asset is maintained through outsourced funds to USGS, Kansas Water Science Center in FY 2025 and FY 2026.

Project Element K.3. Data Telemetry and Field Engineering (Ongoing Study)

The GCMRC conducts much of its work in remote locations throughout Grand Canyon characterized by extreme terrain and harsh environments. This presents challenges in how to monitor important resources at these remote locations. Resource monitoring applications such as remote sensing and data telemetry are a way to leverage newer information technologies to gain better, and more cost-effective, access to monitoring data. While these technologies also have challenges when applied to such remote environments, solutions can and have been developed to greatly reduce the amount of time required for staff to be in the field. Remote monitoring solutions handle these challenges through system design and proper implementation in a way that traditional field site visits cannot. While installing, maintaining, and collecting data from field-based sensors in the Grand Canyon region is difficult, once solutions are applied, those challenges are reduced significantly.

Among the challenges for connecting field-based sensors in this environment are the availability of solar power and adequate connectivity through telecommunications in order to sufficiently supply power to remote monitoring systems and transmit data from these same locations, respectively. Providing power to field sensors in Grand Canyon can be a distinct challenge, especially for field sites located in narrow parts of the canyon.

Commercially available equipment for supplying power and telemetry capabilities is very often not designed for these extreme environments, and so GCMRC has had to design, deploy, and maintain custom solutions that provide consistent, reliable access to valuable data for the Colorado River ecosystem (Figure 5).

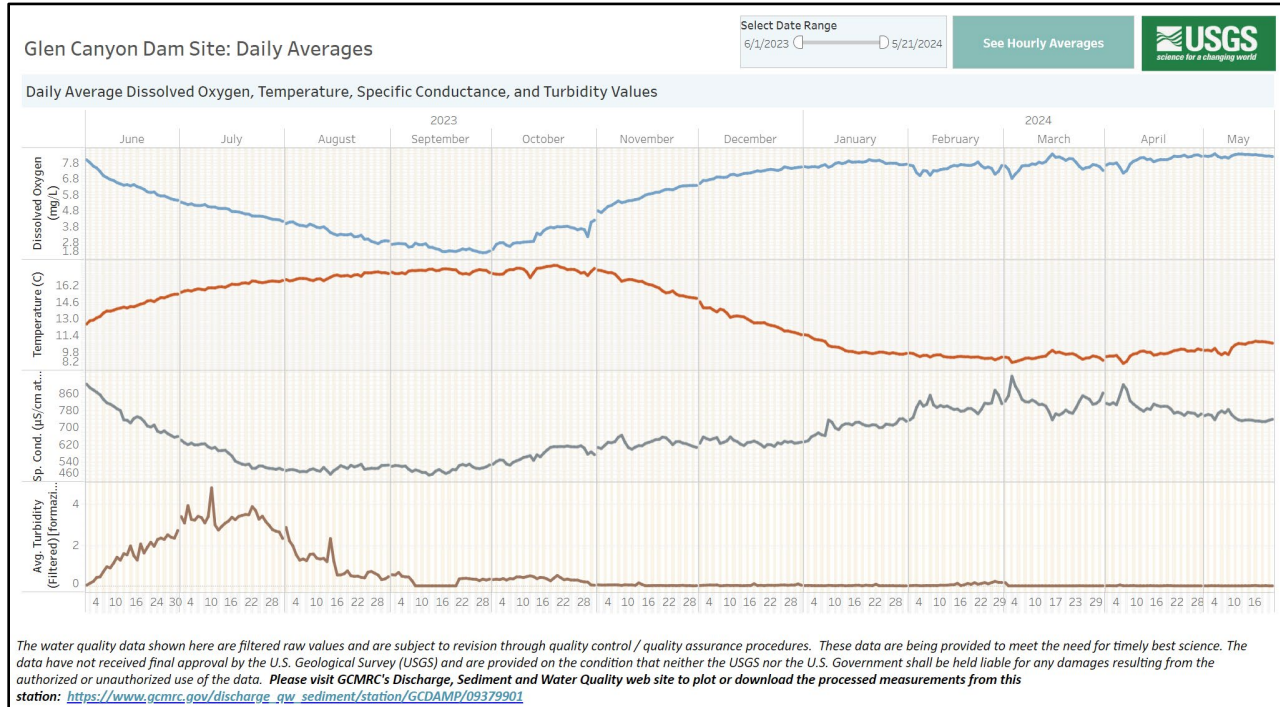


Figure 5. View of online data visualization of Glen Canyon Dam water quality information shared in near real-time, now publicly available with link to quality assessment / quality control version of the data. See URL link in Products subsection.

Expertise in field engineering and direct current system design, installation, and maintenance are required to keep these sites operational. In some instances, these systems have been deployed for more than two decades and underpin some of GCMRC’s most critical data collection efforts. Power management, compute, and telemetry systems for sediment gaging sites, fish PIT tag antennas, water quality monitoring equipment, and weather stations have all been developed, deployed, and maintained by as part of Project K’s field engineering efforts. These systems enable science in Grand Canyon and promote data integrity across the resource-specific projects.

This project element also tracks the technical support and electrical engineering expertise provided to other research projects described in this work plan (Figure 6) which lists specific tasks with individual projects identified, where possible. The type of work performed in this element is varied and must at times adjust to respond to emerging needs within projects or critical responses to system failures.

Some work performed in this element inherently benefits the GCMRC by improving upon the design and development of common components used by most remote monitoring systems deployed by GCMRC.

| Project Supported | Engineering Support | Programming Support | Telemetry | Cloud Services | Significance of Support | LTEMP Resource Goals |
|---|---|--|---|---|---|--|
| Project A Sediment monitoring gauges | <ul style="list-style-type: none"> Sensor installation /maintenance Sensor-to-computer telemetry via radio modems PC power system design / maintenance Power management system maintenance | <ul style="list-style-type: none"> Field computer maintenance and repair Software development and maintenance | <ul style="list-style-type: none"> Starlink dish design, installation, maintenance and repair Telemetry updates and security maintenance | <ul style="list-style-type: none"> Remote access program maintenance Daily check of site uptime and computer health | <ul style="list-style-type: none"> Used by researchers and stakeholders across GCDAMP Used by USBR for water quality modeling | 7. Sediment |
| Project B Sediment storage monitoring | <ul style="list-style-type: none"> Power management system maintenance and repair on boats High-resolution camera (Eyeball) maintenance and repair | <ul style="list-style-type: none"> Eyeball Interactive software maintenance | | | <ul style="list-style-type: none"> Used by researchers and stakeholders across GCDAMP | 7. Sediment |
| Project D Monitoring impacts on Arch. sites | | <ul style="list-style-type: none"> Weather station data collection programming support Weather station telemetry maintenance and design | | | <ul style="list-style-type: none"> Used by researchers in GCMRC | 1. Archaeological and Cultural Resources |
| Project E Controls on ecosystem productivity | <ul style="list-style-type: none"> Sensor and telemetry design, installation, maintenance and repair Maintenance and repair of power system for both field sites DC power design / 120V AC power design | <ul style="list-style-type: none"> Field computer maintenance and repair Software development and maintenance Automatic data downloads from sensor | <ul style="list-style-type: none"> Cellular modem design, installation, maintenance and repair Telemetry updates and security maintenance | <ul style="list-style-type: none"> Remote access program maintenance Daily check of site uptime and computer health Data visualization display via Tableau and email updates | <ul style="list-style-type: none"> Used by researchers and stakeholders across GCDAMP Used by USBR for water quality modeling | 2. Natural Processes |
| Project G Humpback chub population dynamics | <ul style="list-style-type: none"> Sensor and telemetry design, installation, maintenance and repair Maintenance and repair of power system both for fish antenna and rim station. Maintenance and repair of power systems for accessory fish antennas | <ul style="list-style-type: none"> Field computer maintenance Radio antenna maintenance WiFi bridge maintenance Software development and maintenance Automatic data downloads from sensor | <ul style="list-style-type: none"> Cellular modem design, installation, maintenance and repair Telemetry updates and security maintenance | <ul style="list-style-type: none"> Remote access program maintenance Daily check of site uptime and computer health | <ul style="list-style-type: none"> Used by GCMRC staff and cooperators to estimate fish populations | 3. Humpback Chub |
| Project H Salmonid research and monitoring | <ul style="list-style-type: none"> Sensor and telemetry design, installation, maintenance and repair Maintenance and repair of power system for both field sites DC power design / 120V AC power design | <ul style="list-style-type: none"> Field computer maintenance and repair Software development and maintenance Automatic data downloads from sensor | <ul style="list-style-type: none"> Cellular modem design, installation, maintenance and repair Telemetry updates and security maintenance | <ul style="list-style-type: none"> Remote access program maintenance Daily check of site uptime and computer health Data visualization display via Tableau and email updates | <ul style="list-style-type: none"> Used by GCMRC staff and cooperators to estimate fish populations | 9. Rainbow Trout fishery |

Figure 6. Table showing Project Element K.3. support to resource-specific projects, the type and level of effort, and the LTEMP resource goals this work supports. Boxes in white denote unknown assets or areas where Project K is not anticipating providing support.

Project K has recently led efforts to develop and install innovative forms of telemetry at multiple sensor locations throughout Grand Canyon. Efforts to enhance field site telemetry are critical for promoting data integrity and reducing field site maintenance costs. Sensor telemetry enables real-time data analysis from sites which previously could only be accessed periodically through site visits. This improved access to sensors allows for rapid identification of data errors or sensor malfunctions, helping to prevent extended data gaps. By having clear information about potential sensor issues as they occur, staff can strategically respond quickly to repair equipment with increased understanding of the maintenance or repair issue prior to visiting the site.

Advances in Low Earth Orbit (LEO) Satellite Communication

In February of 2023 we tested and deployed the first low earth orbit (LEO) satellite telemetry dish in the USGS at the Grand Canyon gaging station near Phantom Ranch, Grand Canyon National Park. Following the success of this effort, LEO telemetry dishes were installed at the 30-mile and 61-mile sediment gaging stations in August 2023.

These new systems replaced the legacy geostationary satellite dishes first deployed in 2004, significantly increasing both data download speed and dependability. Additionally, the LEO data transmission costs have been reduced by one-third from the geostationary service.

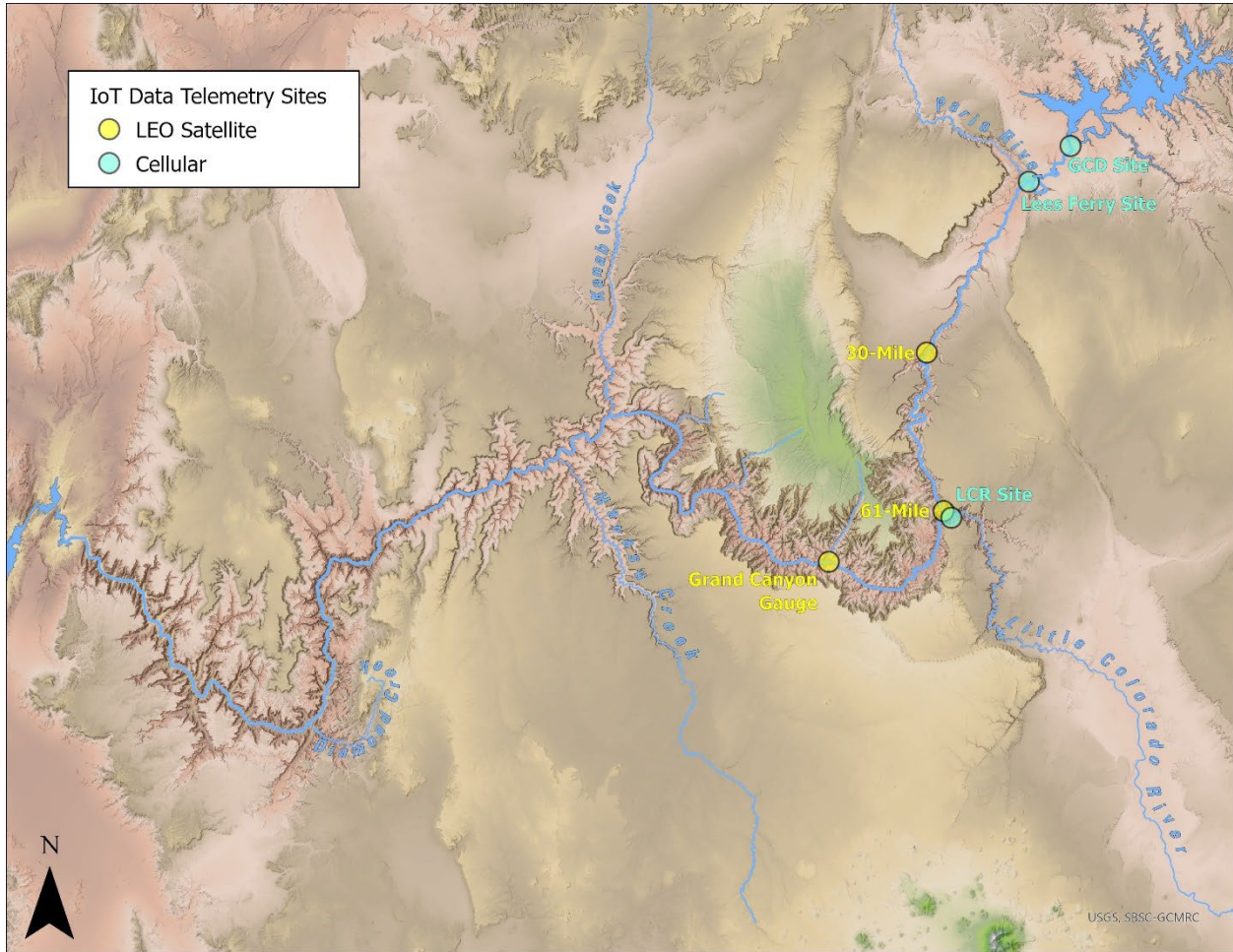


Figure 7. Map of the Grand Canyon region showing the location of IoT Data Telemetry Sites developed, implemented, and maintained by on-going efforts through Project Element K.3 support.

Building upon the success of deploying LEO satellite dishes at the long-term sediment monitoring sites (30-mile, 61-mile, Grand Canyon gauge), Project K has continued to pursue the rapidly emerging trends in this realm of data communication. From Fall 2023 through Spring 2024, we have coordinated with SpaceX and the USGS Associate Chief Information Officer (ACIO) to join a SpaceX beta program for testing direct to cell telemetry systems in the remote environments that GCMRC works. Direct to cell telemetry technology is anticipated to be a transformational breakthrough in remote sensor telemetry, fundamentally changing the way that science is conducted in Grand Canyon and opening the possibility of connecting all sensors to the cloud for real-time data transmission and error identification.

The objective for continuing this work into FY 2025-27 is to leverage this new telemetry technology to develop and install systems for connecting non-telemetered sensors to the cloud and serve as much data as possible in real-time to scientists and stakeholders.

Methods

Project K.3 provides electrical engineering, programming and cloud data ingest services to enable systems ranging from basic power supply to sole data loggers to complex sensor-to-cloud real time data networks and ensure all systems are operating consistently and safely. In order to fulfill these objectives, K.3 utilized a wide range of methods.

Solar and battery bank power systems are required to power a majority of the critical sensor infrastructure deployed by GCMRC. For the design and maintenance of these power systems, Project K.3 conducts site assessments multiple times per year for preventative maintenance, provides power assessment and analysis of new and existing systems, component selection sized in accordance with electrical safety standards, and system sizing to ensure adequate power supply based on the unique nature of field sites. For new sensor deployments, K.3 conducts mounting and orientation of solar panels, charge controllers and battery banks as well as wiring and integration with field computers, data loggers and other field data products. Each of these methods are used at both fixed sensor locations, such as gaging stations, and mobile sensor locations, such as radar enabled boats for survey work. System performance monitoring, documentation, and communication with project leads ensures that sensor system infrastructure is healthy, and that preventative maintenance occurs to reduce potential system downtime.

Project K.3 also provides programming and sensor integration support for other projects as both long-term maintenance and on an as-needed basis. This work is conducted by designing, testing, debugging and maintaining code repositories in USGS-approved software and code management platforms. Programmatic support spans multiple programming languages (R, Python, bash, BASIC, C++) and enables data collection using multiple platforms utilized by GCMRC project leads (data loggers, field computers, cameras). Daily maintenance of telemetered systems through online tools and multiple site visits per year for maintenance of non-telemetered systems ensures that they are operating correctly and that sensor data gaps are prevented.

At field sites where telemetry is available, Project K.3 provides support spanning the entire sensor-to-cloud workflow ensuring access to real time provisional data for project leads and stakeholders. Due to the extreme terrain and remote setting of Grand Canyon, these workflows are challenging to accomplish and require regular preventative maintenance and custom programming of both field data collection devices and cloud hosted data ingest systems. To accomplish the work of real time provisional data, Project K.3 performs daily checks on data workflows, rapid response field site visits for data ingest issues and expertise in all aspects of these systems ranging from basic power supply design to complex cloud-to-database workflows.

Real time provisional data enables Project K.3 staff and project leads to gather data as it is collected and immediately identify critical system issues, preventing or significantly reducing data gaps.

Anticipated Use of Data

Project K supports data generation of multiple other projects through power system design and maintenance, technical support of data loggers, programming support of remote data upload systems and cloud data ingestion and display. Types of data supported by Project K include sediment gaging station sensors (sediment radar, water quality and water samplers), water quality stations at Glen Canyon Dam and Lees Ferry, fish PIT tag data from the Little Colorado River, weather station data from sediment and archeological sites and sandbar cameras.

The Project K.3 Internet of Things (IoT) program supports both real time data access for researchers and real time data transmission for display to resource managers and stakeholders. Real time data access is currently supported for sediment gaging stations, Little Colorado River fish PIT tags, and water quality at Glen Canyon Dam and Lees Ferry. Complex technical field site programming, telemetry and power supply are required to maintain this real time data access. Benefits of real time data access for resource managers includes the ability to check in on and get reports from sensors deployed in the field without an expensive and time-consuming field site visit as well as the ability to know instantly when a sensor malfunctions and conduct remote troubleshooting.

Real time data transmission is currently supported by Project K.3 at water quality sites at Glen Canyon Dam and Lees Ferry and fish PIT tag counts from the Little Colorado River corridor. Sensor to cloud workflows have been developed and maintained that enable resource managers and stakeholders to view in real time provisional data from these sites without the need for a technician to make a field site visit. Recent advances in satellite telemetry are expected to enable real time data transmission from dozens of other field sites in the next few years, dramatically increasing access to provisional data of all kinds including imagery.

An example of how real time data generated by the Project K.3 IoT program has been used by GCDAMP and resource managers / stakeholders is the viewing of water quality parameters at the Glen Canyon Dam outflow during recent cool mix flows. Bureau of Reclamation officials have access to a DOI internal, provisional, data dashboard that shows raw data coming from a water quality sensor located at the outflow of Glen Canyon Dam.

Outcomes and Products

Outcomes

Outcomes from the project element will be measured through custom developed data dashboards that monitor the health of remote IoT data telemetry systems.

Uptime, connectivity, ability to download and responsiveness to planned and unplanned downtime of these systems will be assessed. While the technologies employed through this project element do allow for some measure of acceptable outcomes (i.e., 90% uptime of a system), the remoteness of the IoT sites and the uniqueness of some of the technologies used combine to create scenarios where such outcomes are not always possible. Note: Some links listed under Products are not accessible outside of the DOI network, however, sharing outside the DOI may be available upon request.

Products

While Project K is a support project and does not pursue scientific questions independently, multiple products are expected during the next workplan. These products include codebases stored in Gitlab, operation manuals and documentation for field site telemetry, compute and power supply systems, journal and USGS series publications of innovative system designs, internal to DOI data dashboards for provisional sensor data, and database hosting and maintenance.

- SBSC GCMRC IoT code repository: [sbsc-gcmrc-iot · GitLab \(usgs.gov\)](https://github.com/sbsc-gcmrc-iot)
- Glen Canyon Dam Water Quality Real-Time Data Visualization: <https://tableau.usgs.gov/views/colorado-river-water-quality-gcd/GlenCanyonDamSiteDailyAverages>

Budgets

Project Element Budget

| Fiscal Year 2025 | | | | | | | | | |
|---|------------------|----------------------|-----------------------|-----------------------|---------------------------|--------------------------|-----------------------------|------------------|------------------------------|
| Project K Geospatial Science, Data Management, and Technology | Salaries | Travel & Training | Operating Expenses | Logistics Expenses | Cooperative Agreements | To other USGS Centers | AMP Special Burden Rate* | Total | USGS Contributing Funds** |
| | | | | | | | 21.86% | | Est Full Rate = 55% |
| K.1. Enterprise GIS, geospatial analysis, and processing | \$145,932 | \$4,000 | \$5,600 | \$0 | \$0 | \$0 | \$34,004 | \$189,536 | |
| K.2. Data management and database administration | \$179,715 | \$1,800 | \$3,000 | \$0 | \$0 | \$65,000 | \$40,340 | \$289,855 | |
| K.3. Data telemetry and field engineering | \$107,651 | \$3,000 | \$6,550 | \$0 | \$0 | \$0 | \$25,624 | \$142,824 | |
| Total Project K | \$433,298 | \$8,800 | \$15,150 | \$0 | \$0 | \$65,000 | \$99,968 | \$622,216 | \$47,750 |

| Fiscal Year 2026 | | | | | | | | | |
|---|------------------|----------------------|-----------------------|-----------------------|---------------------------|--------------------------|-----------------------------|------------------|------------------------------|
| Project K Geospatial Science, Data Management, and Technology | Salaries | Travel & Training | Operating Expenses | Logistics Expenses | Cooperative Agreements | To other USGS Centers | AMP Special Burden Rate* | Total | USGS Contributing Funds** |
| | | | | | | | 22.60% | | Est Full Rate = 55.5% |
| K.1. Enterprise GIS, geospatial analysis, and processing | \$153,229 | \$4,000 | \$5,600 | \$0 | \$0 | \$0 | \$36,799 | \$199,628 | |
| K.2. Data management and database administration | \$188,700 | \$1,800 | \$3,000 | \$0 | \$0 | \$65,000 | \$43,731 | \$302,231 | |
| K.3. Data telemetry and field engineering | \$113,033 | \$3,000 | \$6,550 | \$0 | \$0 | \$0 | \$27,704 | \$150,287 | |
| Total Project K | \$454,963 | \$8,800 | \$15,150 | \$0 | \$0 | \$65,000 | \$108,234 | \$652,147 | \$50,750 |

| Fiscal Year 2027 | | | | | | | | | |
|---|------------------|----------------------|-----------------------|-----------------------|---------------------------|--------------------------|-----------------------------|------------------|------------------------------|
| Project K Geospatial Science, Data Management, and Technology | Salaries | Travel & Training | Operating Expenses | Logistics Expenses | Cooperative Agreements | To other USGS Centers | AMP Special Burden Rate* | Total | USGS Contributing Funds** |
| | | | | | | | 23.40% | | Est Full Rate = 56% |
| K.1. Enterprise GIS, geospatial analysis, and processing | \$168,704 | \$4,000 | \$5,600 | \$0 | \$0 | \$0 | \$41,723 | \$220,028 | |
| K.2. Data management and database administration | \$276,796 | \$1,800 | \$3,000 | \$0 | \$0 | \$0 | \$65,893 | \$347,489 | |
| K.3. Data telemetry and field engineering | \$126,499 | \$3,500 | \$6,550 | \$0 | \$0 | \$0 | \$31,952 | \$168,501 | |
| Total Project K | \$571,999 | \$9,300 | \$15,150 | \$0 | \$0 | \$0 | \$139,569 | \$736,018 | \$64,201 |

*GCDAMP Special Rate: Rate includes current DOI preferred rate (currently 15%; subject to change) and facilities rate (will vary annually). No USGS bureau overhead is charged (unique to the GCDAMP agreement).

**USGS Contributing Funds: The amount of funds required to cover the subsidy created by the reduced burden rate (i.e., GCDAMP Special Rate). As in previous years, SBSC/GCMRC will request these funds of the USGS cost-share program. These funds are not guaranteed.

References Cited

- ESRI, 2024a, ArcGIS Enterprise software—copyright 1995–2024: Redlands, Calif., Environmental Systems Research Institute, accessed May 20, 2024, <https://www.esri.com/en-us/legal/copyright-trademarks/>.
- ESRI, 2024b, ArcGIS website—Cloud-based GIS: Redlands, Calif., Environmental Systems Research Institute, accessed May 20, 2024, <https://www.arcgis.com/home/index.html>.
- Federal Geographic Data Committee, 1998, Geospatial positioning accuracy standards, part 3—National standard for spatial data accuracy—Appendix 3-D: Reston, Va., National Spatial Data Infrastructure, Subcommittee for Base Cartographic Data, report no. FGDC-STD-007.3-1998, 26 p., <https://www.fgdc.gov/standards/projects/FGDC-standards-projects/accuracy/part3/chapter3>.
- Gushue, T.M., 2019, Colorado River Mile System, Grand Canyon, Arizona: U.S. Geological Survey data release, <https://doi.org/10.5066/P9IRL3GV>.
- Hutchison, V.B., Burley, T.E., Blasch, K.W., Exter, P.E., Gunther, G.L., Shipman, A.J., Kelley, C.M., and Morris, C.A., 2024, U.S. Geological Survey data strategy 2023–33: U.S. Geological Survey Circular 1517, 7 p., <https://pubs.usgs.gov/publication/cir1517>.
- Kaplinski, M., Hazel, J.E. Jr, Grams, P.E., Gushue, T., Buscombe, D.D., and Kohl, K., 2022, Channel mapping Glen Canyon Dam to Lees Ferry in Glen Canyon National Recreation Area, Arizona – Data: U.S. Geological Survey data release, <https://doi.org/10.5066/P98GFP93>.
- Kumara, I., Garriga, M., Romeu, A.U., Di Nucci, D., Palomba, F., Tamburri, D.A., and van den Heuvel, W.-J., 2021, The do’s and don’ts of infrastructure code: A systematic gray literature review: Information and Software Technology, v. 137, article 106593, p. 1-20, <https://doi.org/10.1016/j.infsof.2021.106593>.

- Petty, T.R., Secretary's Designee, Assistant Secretary for Water and Science, 2019, Glen Canyon Dam Adaptive Management Program guidance—August 14, 2019 memorandum: Washington, D.C., U.S. Department of the Interior, Office of the Secretary, 4 p.
- Sankey, J.B., Bransky, N., Pigue, L., Kohl, K., and Gushue, T.M., 2024, Four band image mosaic of the Colorado River Corridor in Arizona—2021, including accuracy assessment data: U.S. Geological Survey data release, <https://doi.org/10.5066/P9BBGN6G>.
- U.S. Department of the Interior, 1995, Operation of Glen Canyon Dam, Colorado River storage project, Arizona—final Environmental Impact Statement: Salt Lake City, Utah, U.S. Department of the Interior, Bureau of Reclamation, 337 p. plus appendices, <http://www.usbr.gov/uc/envdocs/eis/gc/gcdOpsFEIS.html>.
- U.S. Department of the Interior, 1996, Record of Decision—Operation of Glen Canyon Dam Final Environmental Impact Statement: Salt Lake City, Utah, U.S. Department of the Interior, Bureau of Reclamation, Glen Canyon Dam Adaptive Management Program, 15 p., https://www.usbr.gov/uc/envdocs/rod/Oct1996_OperationGCD_ROD.pdf.
- U.S. Department of the Interior, 2014, Glen Canyon Adaptive Management Program Triennial Budget and Work Plan—Fiscal years 2015-2017: Bureau of Reclamation, Upper Colorado Regional Office and U.S. Geological Survey, Grand Canyon Monitoring and Research Center, prepared in conjunction with the Glen Canyon Dam Adaptive Management Program, 475 p. plus appendices, https://www.usbr.gov/uc/progact/amp/amwg/2017-09-20-amwg-meeting/Attach_04a.pdf.
- U.S. Department of the Interior, 2016a, Glen Canyon Dam Long-term Experimental and Management Plan final Environmental Impact Statement (LTEMP EIS): U.S. Department of the Interior, Bureau of Reclamation, Upper Colorado Region, National Park Service, Intermountain Region, online, <http://ltempeis.anl.gov/documents/final-eis/>.
- U.S. Department of the Interior, 2016b, Record of Decision for the Glen Canyon Dam Long-term Experimental and Management Plan final Environmental Impact Statement (LTEMP ROD): Salt Lake City, Utah, U.S. Department of the Interior, Bureau of Reclamation, Upper Colorado Region, National Park Service, Intermountain Region, 196 p., http://ltempeis.anl.gov/documents/docs/LTEMP_ROD.pdf.
- U.S. Department of the Interior, 2017, Glen Canyon Dam Adaptive Management Program Triennial Budget and Work Plan—Fiscal Years 2018-2020—Final submitted to the Secretary of the Interior: Flagstaff, Ariz., U.S. Geological Survey, Grand Canyon Monitoring and Research Center and Salt Lake City, Utah, Bureau of Reclamation, Upper Colorado Region, 316 p., http://gcdamp.com/index.php/FY18-20_GCMRC_Triennial_Budget_and_Workplan.
- U.S. Department of the Interior, 2020, Glen Canyon Dam Adaptive Management Program Triennial Budget and Work Plan—Fiscal years 2021-2023—Final approved by the Secretary of the Interior—December 2, 2020: Flagstaff, Ariz., U.S. Geological Survey, Grand Canyon Monitoring and Research Center, and Salt Lake City, Utah, Bureau of Reclamation, Upper Colorado Region, 384 p. plus appendices,

http://gcdamp.com/images_gcdamp_com/5/5d/GCMRC_TWP2021-23_December2_2020_ApprovedBySecretary.pdf.

U.S. Geological Survey, 2017, Fundamental science practices (FSP)—Policy directives—online manuals: Washington, D.C., Department of the Interior,

<https://www.usgs.gov/about/organization/science-support/science-quality-and-integrity/policy-directives>.

VanderKooi, S.P., Kennedy, T.A., Topping, D.J., Grams, P.E., Ward, D.L., Fairley, H.C., Bair, L.S., Yackulic, C.B., Schmidt, J.C., and Sankey, J.B., 2017, Scientific monitoring plan in support of the selected alternative of the Glen Canyon Dam Long-Term Experimental and Management Plan: U.S. Geological Survey, Grand Canyon Monitoring and Research Center, U.S. Geological Survey Open-File Report 2017-1006, 18 p.,

<https://doi.org/10.3133/ofr20171006>.

Wilkinson, M.D., Dumontier, M., Aslbersberg, I.J., Appleton, G., Axton, M., Baak, A., Blomberg, N., Boiten, J.-W., Santos, L.B.d.S., Bourne, P.E., Bouwman, J., Brookes, A.J., Clark, T., Crosas, M., Dillo, I., Dumon, O., Edmunds, S., Evelo, C.T., Finkers, R., Gonzalez-Beltran, A., Gray, A.J.G., Groth, P., Goble, C., Grethe, J.S., Heringa, J., Hoen, P.A.C.t., Hooft, R., Kuhn, T., Kok, R., Kok, J., Lusher, S.J., Martone, M.E., Mons, A., Packer, A.L., Persson, B., Rocca-Serra, P., Roos, M., van Schaik, R., Sansone, S.-A., Schultes, E., Sengstag, T., Slater, T., Strawn, G., Swertz, M.A., Thompson, M., van der Lei, J., van Mulligen, E., Velterop, J., Waagmeester, A., Wittenburg, P., Wolstencroft, K., Zhao, J., and Mons, B., 2016, The FAIR Guiding Principles for scientific data management and stewardship: Scientific Data, v. 3, article 160018, <https://doi.org/10.1038/sdata.2016.18>.

Project L: Overflight Remote Sensing in Support of GCDAMP and LTEMP

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Project Summary and Purpose

This project uses remote sensing to monitor the entirety of the terrestrial, riparian, and fluvial ecosystems along the Colorado River from the forebay of Glen Canyon Dam to Lake Mead at Pearce Ferry that are affected by the operation of Glen Canyon Dam. We analyze, interpret, and acquire overflight remote sensing data at a frequency commensurate with significant ecosystem changes along the river, including major tributaries.

During FY 2025-27, Project Element L.1 (partially funded) of this project will analyze and interpret data from the 2021 and previous overflights for collaborative science efforts with other projects (A, B, C, D) to address LTEMP resource goals and calculate important proposed LTEMP performance metrics for assessing sediment and vegetation program objectives. Project Elements L.2 and L.3 (unfunded) of this project will implement the next overflight mission, to acquire imagery and lidar data, respectively. Although significant changes along the Colorado River can happen over days to weeks, we propose limiting data collection to approximately three digital image and topographic datasets per decade in order to maintain a long-term, system-wide, and economically viable remote sensing data record to answer scientific questions about the relationships of dam operations and landcover changes in the Colorado River Ecosystem (CRE) at decadal timescales.

Though unfunded (L.2 and L.3), the overflight mission was proposed to occur in FY 2026, to monitor effects of dam operations on the river ecosystem at the 10-year mark – halfway point – of the LTEMP Environmental Impact Statement (EIS) Record of Decision (ROD) implementation (US Department of the Interior, 2016a, b). Note that the science questions are designed to be addressed in Project Element L.1 using overflight remote sensing datasets that have already been acquired and published by USGS; answering the questions is not contingent on first completing the next overflight mission (i.e., Project Elements L.2 and L.3 in FY 2026).

Long-Term Experimental and Management Plan (LTEMP) Resource Goals

The imagery and derivative data products from overflight remote sensing are used either directly or indirectly by every science project proposed in each Glen Canyon Dam Adaptive Management Program (GCDAMP) Triennial Work Plan (TWP) to address every resource goal

of the Long-Term Experimental and Management Plan (LTEMP; US Department of the Interior, 2016a, b). Table 1 lists the primary datasets from overflight missions and the derivative products used by science projects to address resource goals. Table 1 also provides links to recent examples of these datasets and products.

During FY 2021-24, Project L acquired new remote sensing data from an overflight mission implemented in 2021, and then processed and published those data. The 2021 overflight was the latest in a rich archive of airborne remote sensing in the Grand Canyon in support of the GCDAMP and LTEMP (see <https://www.usgs.gov/centers/southwest-biological-science-center/science/airborne-remote-sensing-grand-canyon>). Importantly, the 2016 Long-Term Experimental and Management (LTEMP) planning relied heavily on these data and derivative products. LTEMP resource goals addressed by this project are:

- Archaeological and Cultural Resources
- Natural Processes
- Humpback Chub
- Hydropower and Energy
- Other Native Fish
- Recreational Experience
- Sediment
- Tribal Resources
- Rainbow Trout Fishery
- Nonnative Invasive Species
- Riparian Vegetation

Background

The imagery and derivative data products from overflight remote sensing are used either directly or indirectly by every science project proposed in this TWP to address every resource goal of the LTEMP. Table 1 lists the primary datasets from overflight missions and the derivative products used by science projects to address resource goals. Table 1 also provides links to recent examples of these datasets and products.

In 2002, 2005, 2009, 2013, and 2021, GCMRC, through the GCDAMP, acquired digital, 4-band multispectral imagery and photogrammetrically derived topography data; similar to the data proposed in Project Element L.2 (unfunded) to be acquired in 2026 (please see Element L.2 narrative below for technical specifications of past and future overflight image acquisitions).

With each of those previous digital image acquisitions, GCMRC remote sensing staff developed and improved upon a methodology for producing a spatially seamless, spectrally consistent, and nearly cloud- and blemish-free image mosaic (Davis, 2012; Durning and others, 2016; Sankey and others, 2024; Table 1). That proven methodology was most recently used to publish an image mosaic from the 2021 acquisition (Sankey and others, 2024).

Table 1. Summary of primary datasets and examples of derived products from overflight missions used by GCDAMP science projects to achieve LTEMP resource goals. Please follow hyperlinks for examples.

| Primary datasets produced from overflight missions | |
|--|--|
| Multispectral imagery [2021, 2013, 2009, 2005, 2002] | Digital Topography |
| Products derived from primary datasets | |
| <ul style="list-style-type: none"> Website content and online maps | <ul style="list-style-type: none"> Cartographic products River map books Publication maps |
| <ul style="list-style-type: none"> Fish sampling unit system for mainstem Colorado River | <ul style="list-style-type: none"> Humpback chub and juvenile monitoring system for Little Colorado River |
| <ul style="list-style-type: none"> Colorado River centerline and river mile system | <ul style="list-style-type: none"> Flowlines Extracted from low-flow water's edge (~8,000 ft³/s) in overflight imagery Modelled from overflight topography and water surface elevation |
| <ul style="list-style-type: none"> Land cover mapping and change detection Water, sand, vegetation, land cover Geomorphic basemap | <ul style="list-style-type: none"> Vegetation species classification Tamarisk vegetation and Tamarisk Beetle impacts |
| <ul style="list-style-type: none"> Campsite delineation Campsite atlas | <ul style="list-style-type: none"> Topographic change detection Hydrologic flow modeling |

Image mosaics from overflight missions have myriad uses that are critical to different aspects of the science implemented through the GCDAMP.

The mosaics are the base map layer for all map books used by science projects to navigate the river and implement field monitoring and research campaigns (Table 1 and Figure 1).

Image mosaics from previous overflights are similarly the base map layer used to produce website content and online GIS maps, such as the GCMRC GIS Base Map Viewer (Table 1). Many science projects develop detailed study designs in a GIS with the imagery.

Examples of such study designs include the fish sampling unit system for the mainstem Colorado River, and the humpback chub monitoring system for the Little Colorado River (Table 1 and Figure 1). Measurements of important river channel characteristics that change over time are periodically updated using the most current overflight imagery.

For example, the river-channel mileage system published by GCMRC and used to navigate and monitor the river, is based on the centerline of the river channel delineated on the published overflight image mosaic data (Table 1; Gushue, 2019).



Figure 1. Remote sensing imagery are the base map layer for all map books used by science projects to navigate the river and implement field monitoring and research campaigns. Top left photo shows a scientist consulting his map book on the river. Top right photo shows a stack of map books printed for a recent field campaign. Bottom panel shows a page of a map book with Fish Sampling Units delineated for a recent monitoring campaign.

As with preceding overflight image mosaics, the 2021 overflight image mosaic was published as a USGS data release (Sankey and others, 2024), and GCMRC is now implementing their workflow for producing landcover classification maps derived from that mosaic. The workflow progresses by first publishing USGS data releases of the most basic, fundamental maps of landcover (Figure 2) including,

- 1) The low-flow river channel at 8,000 ft³/s (e.g., for the 2013 and preceding overflight image mosaics see landcover classification maps of water: Durning and others, 2017a; Sankey and others, 2015b),
- 2) Sand and other river sediment (e.g., for the 2013 and preceding overflight image mosaics see river sand classification maps: Sankey and others, 2018c), and
- 3) Total riparian vegetation (e.g., for the 2013 and preceding overflight image mosaics see landcover classification maps of riparian vegetation: Durning and others, 2017b; Sankey and others, 2015b).



Figure 2. Example landcover classifications of the low-flow river channel at 8,000 ft³/s (water classification); total riparian vegetation (vegetation classification), and sand and other river sediment deposits (sand classification). For each overflight image dataset, GCMRC publishes USGS data releases of these most basic, fundamental maps of landcover covering the entire river corridor from Glen Canyon Dam to Lake Mead at Pearce Ferry. Various GCMRC projects leverage the landcover classifications, with imagery and topographic data for science that relates observations of environmental change in the Colorado River ecosystem to dam operations, other land use and management activities, climate, and other factors.

GCMRC plans to produce and publish the water channel classification map from the 2021 image mosaic during FY 2024 and will produce and publish the 2021 image mosaic riparian vegetation classification map during FY 2025.

Remote sensing staff then collaborate with other GCMRC project staff to leverage the landcover classifications, imagery, and topographic data for science that relates decadal-scale observations of landcover in the Colorado River ecosystem to dam operations, other land use and management activities, climate, and other environmental factors.

One example of this higher-order science leveraged from overflight data is the quantitative assessment of riparian vegetation changes that occurred as a function of dam operations and climate during the first five decades of the operations of Glen Canyon Dam published by Sankey and others (2015a). Other examples are quantitative inventories of tamarisk and tamarisk beetle impacts in Glen Canyon by Sankey and others (2016) and in Grand Canyon by Bedford and others (2018b) and Bransky and others (2021). Hadley and others (2018) used the landcover classifications derived from the imagery to quantify changes in campsites along the river. Sankey and others (2018a, b; 2023) used overflight imagery and topography to assess effects of dam operations on archaeological sites in Grand Canyon. Kasprak and others (2017a, b) used overflight imagery and topography to develop a method for automating the interpretation of repeat survey data in river valleys. Butterfield and others (2020) leveraged the species-level classification of Durning and others (2018) to analyze associations between riparian plant morphological guilds and fluvial sediment dynamics in Grand Canyon.

Durning and others (2021) used the image time series and derived classification maps from 2002-2009-2013 to quantify hydrologic and geomorphic effects on riparian plant species occurrence and encroachment along 360 km of the river corridor in Grand Canyon. Kasprak and others (2017a, b) and Caster and others (2024) used overflight remote sensing digital topographic data and landcover classifications of bare sand and total vegetation to monitor land surface change and its causes at archaeological sites and dunefields. Kasprak and others (2018, 2021) used overflight imagery, topography, and landcover classifications to model the combined influence of changes in river flow and riparian vegetation on the areal extent of sediment available for transport in Grand Canyon from 1921 to 2016, as well as to forecast changes from 2016 to 2036. The work of Kasprak and others (2018, 2021) is currently being implemented to model resource impacts for the LTEMP SEIS (Yackulic and others, 2024).

Proposed Work

Project Element L.1. Analysis and Interpretation of Overflight Remote Sensing Data (Partially Funded; Modified Study)

During FY 2025-27 GCMRC will produce and publish CRe landcover classification maps derived from analysis of the recently published the [orthomosaic of high-resolution multispectral](#)

[imagery](#) acquired during the 2021 overflight (Sankey and others, 2024). GCMRC will use the digital imagery and associated topographic data to conduct analyses of topographic change and changes in landcover derived from the 2021 and prior overflights. This work will support the three major interdisciplinary science efforts outlined in the timeline in Table 2.

Table 2. Timeline of major activities and work effort for Project Element L.1. Analysis and interpretation of overflight remote sensing data during the FY 2025-27 TWP.

| Fiscal Year | Project Element L.1 Science Efforts |
|-------------|--|
| 2025 | <ol style="list-style-type: none"> 1. Produce CRe landcover classification maps derived from analysis of the orthomosaic of high-resolution multispectral imagery (2021 overflight image data release: Sankey and others, 2024). <ul style="list-style-type: none"> • 8,000 ft³/s low-flow river channel (Glen Canyon Dam to Lake Mead; 480 km) • Total riparian vegetation cover (Glen Canyon Dam to Lake Mead; 480 km) • Sand and other river sediment (Marble Canyon; inclusive of Project B.4, C.4, and L.1 high elevation sand analysis extents) |
| 2026 | <ol style="list-style-type: none"> 2. Summarize area of total riparian vegetation cover from 2021 relative to previous overflight classification maps for the entire 480 km of river from Glen Canyon Dam to Lake Mead at Pearce Ferry. <ul style="list-style-type: none"> • Calculate and report on <i>LTEMP Performance Metric 11.1 Total Area of Riparian Vegetation Cover</i> 3. Prepare digital topographic (DSM) datasets from the 2002, 2013, and 2021 overflights for change detection analysis of area and volume of high elevation sand (sand at stage elevations > 25,000 ft³/s deposited by HFEs) for the Marble Canyon reach. |
| 2027 | <ol style="list-style-type: none"> 3. (cont'd) Conduct DSM change detection analysis and interpret results. <ol style="list-style-type: none"> a. Calculate and report on <i>LTEMP Performance Metric 7.3 High Elevation Sand > 25,000 CFS stage that is deposited by HFEs</i> specific to the Marble Canyon reach. |

Science Questions

- L.1 Science Effort 1: How has landcover (i.e., bare sand, total vegetation, and the low flow river channel at 8,000 ft³/s discharge) changed in the Colorado River ecosystem from the current decade relative to preceding decades since the closure of Glen Canyon Dam?
- L.1 Science Effort 1: How are observed landcover changes related to dam operations, other land use and management activities, or climate and other environmental factors in the ecosystem?

- L.1 Science Effort 2: How has the total area of riparian vegetation cover (LTEMP performance Metric 11.1) changed over 480 km of the river from the current decade (2021) relative to preceding decades since the closure of Glen Canyon Dam?
- L.1 Science Effort 3: How have the system-wide area and volume of high elevation sand (sand above the stage of 25,000 ft³/s that is deposited by HFEs; LTEMP performance Metric 7.3) changed during the past two decades in relation to dam operations, HFE protocol, and LTEMP experiments?
- L.1 Science Effort 3: Do remote sensing measurements of high elevation sand provide a more comprehensive metric than the annual monitoring of sandbar sites, periodic topographic-bathymetric channel mapping, or triennial monitoring of archaeological sites in Projects B.1, B.2, and D.1, respectively?

Methods: L.1 Science Effort 1

The first science effort will be to produce landcover change observations for calibration and validation of flow, vegetation, and sediment modeling in support of Projects A, B, C, and D. Classification and change detection methods are detailed in Sankey and others (2015a, 2018c), Durning and others (2021), Bedford and others (2018b), and Bransky and others (2021) and employ machine learning, supervised classification, and unsupervised classification remote sensing image analysis methods.

Anticipated Use of Data: L.1 Science Effort 1

Projects A, B, C, and D include interdisciplinary efforts to better quantify how interactions between river sediment, riparian vegetation, and the flow of water or wind affect long-term changes in the Colorado River channel in response to dam operations (Figure 3). Those efforts require landcover classification maps (Figure 2), and change detection analysis of the maps, derived from overflight remote sensing to either calibrate or validate predictive efforts based on other data acquired by those projects. Project Element L.1 will provide the necessary mapping and change detection analysis for those efforts.

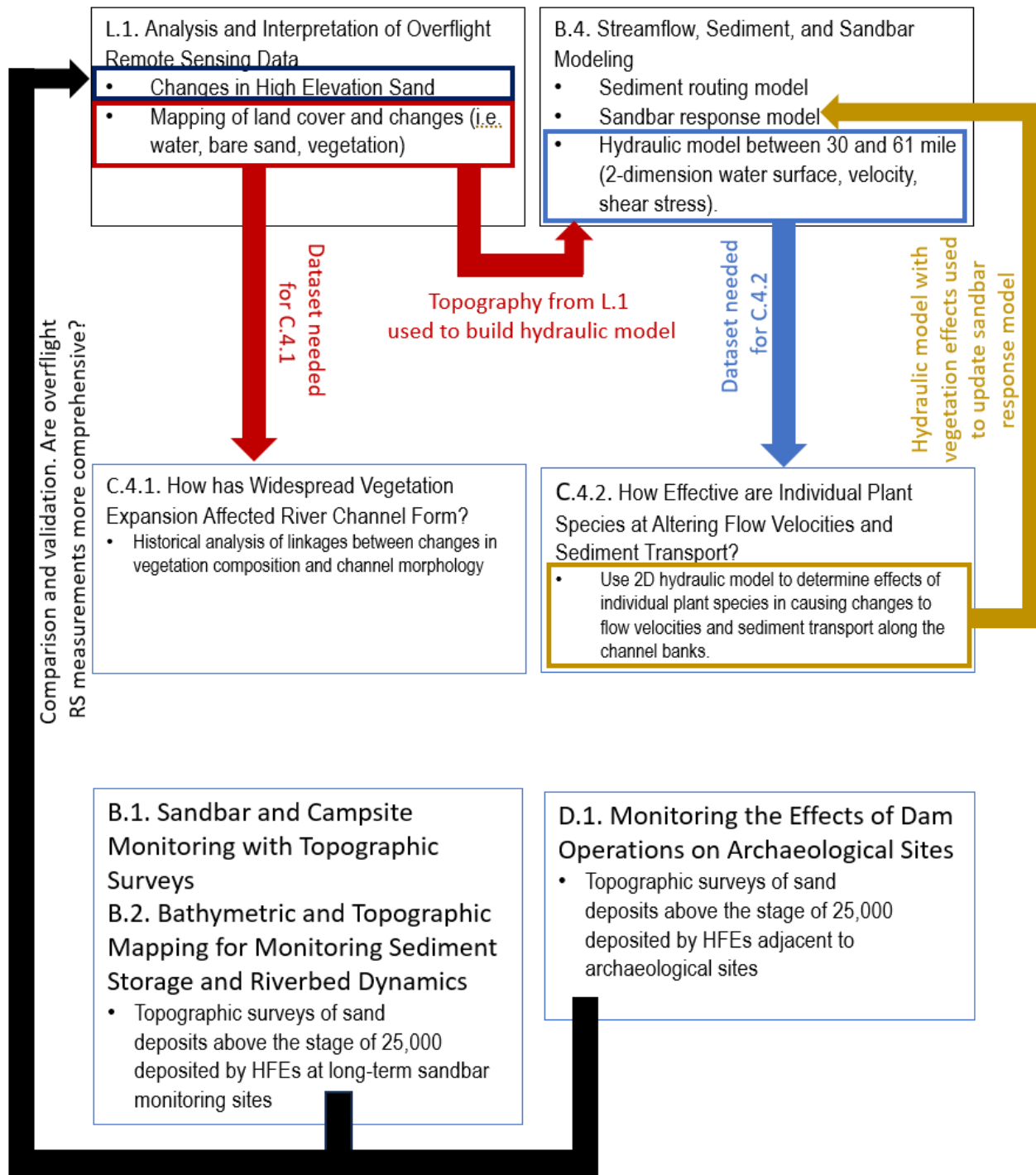


Figure 3. Description of interdisciplinary linkages and dependencies of other project elements on the datasets and analyses provided by Project L.1.

Methods: L.1 Science Effort 2

The second science effort will be to analyze overflight imagery to determine *LTEMP Goal 2, Performance Metric 1.1 Total Vegetation Cover*. This metric is determined periodically after each aerial image acquisition by mapping all of the riparian vegetation in the CRe, and thus provides a census of total vegetation cover throughout the river corridor. At present, the metric has been calculated from aerial image analysis from 1965 to 2013. See Figure 6 in Sankey and others (2015a) and Figure 1 in Durning and others (2021) for examples of the metric spanning 1965-2009 and 1965-2013, respectively. The classification map of total vegetation area will be produced from the 2021 overflight imagery during the first year of this work plan (FY 2025). That classification map of total vegetation within the river channel and riparian area will be interpreted as a function of hydrologic zones based on frequency of inundation: 8,000–25,000 ft³/s (hydrologic zone 1; areas inundated by powerplant operations of GCD; e.g., hydro-peaking), 25,000–45,000 ft³/s (hydrologic zone 2; areas inundated by HFEs) and 45,000 ft³/ and above (hydrologic zone 3; areas not inundated since emergency spillway releases in 1983-84). Methods are detailed in Magirl and others (2008), Sankey and others (2015a), and Durning and others (2021).

Anticipated Use of Data: L.1 Science Effort 2

Total vegetation will be summarized by hydrologic zones to calculate and report *LTEMP Goal 2, Performance Metric 1.1 Total Vegetation Cover*.

Methods: L.1 Science Effort 3

The third major science effort will be to measure system-wide changes in high-elevation sand deposits using the archive of remote sensing digital topography (termed Digital Surface Models; DSMs) acquired by GCMRC and the GCDAMP from remote sensing overflight missions spanning 2002, 2013, and 2021. The volume of high elevation sand (that is, the sand above the stage of 25,000 ft³/s) is a combination of sand deposited by pre-dam floods, sand deposited by HFE's, and river sand transported by wind from lower elevations and deposited at higher elevations. In the absence of HFE's, high-elevation sand can be eroded or deposited by wind, eroded by rainfall-driven runoff from canyon walls and hillslopes, or stabilized by vegetation. However, it is primarily through HFE's that any substantial high-elevation sand may be deposited by the Colorado River. Project B measures the volume of high-elevation sand annually at 45 long-term sandbar monitoring sites. However, it is unknown whether these 45 discrete sandbar sites are diagnostic of system-wide conditions along the 480 km river corridor from Glen Canyon Dam to Lake Mead at Pearce Ferry.

Project D measures the volume of high-elevation sand, including sand at even higher elevations than measured in Project B, during ground-based light detection and ranging (lidar) surveys repeated approximately every three years at sandbars and dunefields associated with approximately 40 archaeological sites. However, it is similarly unknown whether those sites are diagnostic of system-wide conditions. Thus, we will analyze DSMs constructed from photogrammetry of overflight imagery in 2021, 2013, and 2002 to evaluate high-elevation sand conditions within the Marble Canyon reach of the CRe. Interpretation of results will be completed in collaboration with Projects B and D to determine if system-wide high-elevation sand conditions agree with or diverge from high-elevation sand conditions at the discrete long-term sandbar or dunefield monitoring sites in Marble Canyon (Figure 3). Methods for performing geomorphic change detection of the DSMs to determine changes in sand storage are detailed in Kasprak and others (2017a), Sankey and others (2018b), and Caster and others (2024). Figure 4 illustrates examples of geomorphic change detection at Lees Ferry and Soap Creek.

Anticipated Use of Data: L.1 Science Effort 3

This effort will provide needed information on LTEMP Sediment Resource Goals, determine whether additional sites need to be monitored to sufficiently track high-elevation sand volumes, and assess whether system-wide overflight remote sensing derived high-elevation sand measurements provide a more comprehensive understanding of sand resources throughout the CRe.

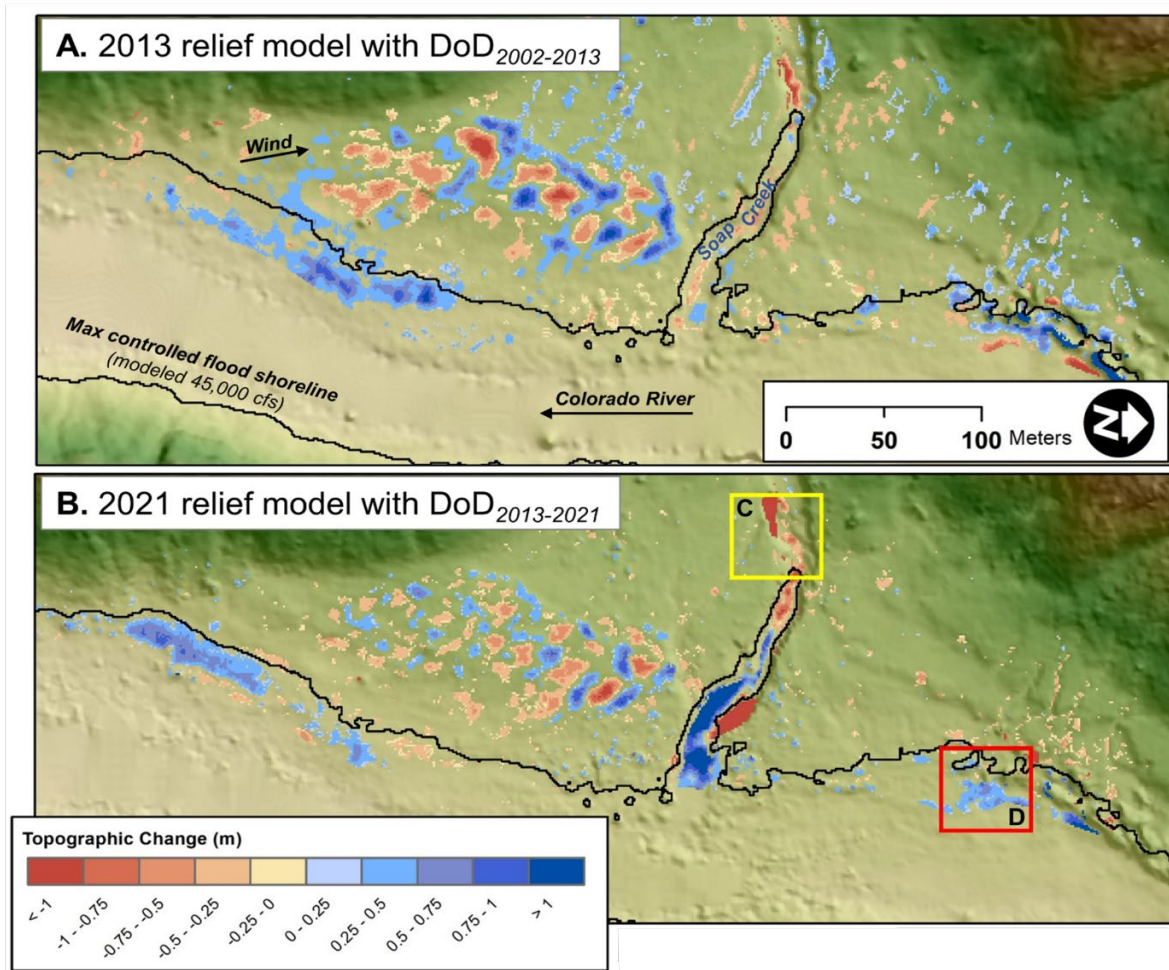


Figure 4. Example of topographic change detection analyses and datasets that will be used for analysis of area and volume of high elevation sand (sand at stage elevations > 25,000 CFS deposited by HFEs) for the entire 480 km of river from Glen Canyon Dam to Lake Mead at Pearce Ferry. Example location is the Soap Creek (approximately 11 miles downstream from Lees Ferry). (A) Basemap with tan-to-green color scale is a topographic is a hillshade of the Digital Surface Model (DSM; relief model) acquired from photogrammetry of 2013 overflight imagery. The basemap is overlaid with a red-to-blue color scale map of topographic changes the occurred between 2002 and 2013 at the Soap Creek camp site, tributary channel, debris fan, dunefield, and sandbar along the Colorado River in Grand Canyon. Topographic changes were determined by differencing the 2013 DSM and 2002 DSM to produce a DSM of Difference (DOD₂₀₀₂₋₂₀₁₃). (B) Basemap with tan-to-green color scale is a topographic hillshade of the Digital Surface Model (DSM; relief model) acquired from photogrammetry of 2021 overflight imagery. The basemap is overlaid with a red-to-blue color scale map of topographic changes that occurred between 2013 and 2021 (DOD₂₀₁₃₋₂₀₂₁). Preliminary results, please don't cite.

Outcomes and Products

- FY 2025; L.1 Science Effort 1: produce and publish, as USGS data releases, classification maps of the (1) low-flow river channel, (2) riparian vegetation, and (3) river sand. Maps 1 and 2 will span the entire 480 km of the river corridor from Glen Canyon Dam to Lake Mead at Pearce Ferry. Map 3 will span the Marble Canyon

reach, which is the common extent of efforts proposed in B.4, C.4, and the L.1 high elevation sand analysis.

- FY 2026; L.1 Science Effort 2: use the derived classification maps as well as the primary imagery from the entire archive of remote sensing overflight missions to evaluate the LTEMP performance metric for total riparian vegetation cover, and to support investigations by other projects in this work plan. The LTEMP performance metric results will be reported in the GCDAMP annual reporting format.
- FY 2027; L.1 Science Effort 3: use the derived classification maps as well as the primary topography data from the archive of remote sensing overflight missions spanning 2002, 2013, to 2021 to evaluate the LTEMP performance metric for high elevation sand specific to the Marble Canyon reach. The results will be reported in the GCDAMP annual reporting format.

Project Element L.2. Acquisition of Overflight Remote Sensing Imagery (New Study; Unfunded)

During FY 2026, GCMRC will implement a remote sensing overflight to collect high-resolution digital, multispectral imagery and topography of the CRe between Glen Canyon Dam and Lake Mead.

Methods

To maintain consistency with previously collected digital, orthorectified aerial imagery acquired in 2002, 2009, 2013, 2021 (Davis, 2012; Durning and others, 2016; Sankey and others, 2024), the mission will be conducted during the same time of year (beginning on Memorial Day weekend in the month of May, and lasting for potentially one week or longer depending on weather) and adhere to the same data collection parameters and significant logistical requirements as used in preceding missions.

Specifications for the data acquisition necessitate that dam releases be held at a steady discharge of 8,000 ft³/s (CFS) for the duration of the overflight mission. As such, the proposed overflight would be within the LTEMP flow regime, and we would request from and work with the Bureau of Reclamation and Western Area Power Administration to maintain the steady 8,000 ft³/s discharge for the duration of the data collection period. This flow adjustment is required to maintain consistency with imagery data sets collected in previous years and maximizes subaerial terrain that is not inundated by the river in the imagery. This will allow for highly accurate image classification of landcover, and for image matching and change detection analysis with previous overflight datasets.

For data collection parameters, we require at least the same 4-band wavelength ranges (red, green, blue, and near infra-red), and same or higher spatial resolution (20-cm pixel resolution), using the same or similar equipment (Leica ADS-100 camera mounted in fixed-wing aircraft), with the option of two cameras and aircraft being made available to increase the rate of data collection and reduce the impact on dam operations. Wavelengths and other technical details will be specified with a Scope of Work (SOW) contract to be written by GCMRC scientists during FY 2025. The overflight mission would occur in FY 2026. Imagery will be acquired using the manned aircraft overflight and Leica sensor, as opposed to other satellite-based platforms and sensors, for example, in order to produce a seamless orthomosaic that is cloud and shadow free, synoptically covers the entire CRe during the low steady dam release of 8,000 CFS beginning on the anniversary date of Memorial Day weekend and meets the image spatial and spectral resolution requirements stated above.

Anticipated Use of Data

The imagery and derivative data products from overflight remote sensing are used either directly or indirectly by every science project to address every resource goal of the LTEMP. Please refer to Table 1 (above) for examples of the primary datasets from overflight missions and the derivative products used by science projects to address resource goals.

Outcomes and Products

- Overflight mission and image acquisition: GCMRC will oversee the contracting, coordinate with all relevant parties, and oversee the mission to acquire multispectral imagery for the river corridor during a low steady dam release of 8,000 CFS beginning at the start of the Memorial Day holiday weekend. Preliminary image mosaic is delivered by the contractor to GCMRC approximately six months after the mission. Staff then perform QA/QC working with the contractor to iteratively review and request further changes as necessary.

Project Element L.3. Acquisition of Airborne Lidar in Conjunction with Overflight Remote Sensing Imagery (New Study; Unfunded)

During the overflight proposed in Project Element L.2, GCMRC proposes to also acquire Quality Level 1 (QLI) or higher resolution airborne lidar data of the CRe.

Anticipated Use of Data

Lidar is a technology that uses a pulsed laser to measure ranges (variable distances) to the Earth. Combined with other data, this generates precise, three-dimensional information about the shape of the Earth's surface and its characteristics.

GCMRC will work through the contracting process for these data to be acquired by the same contractor and in parallel with the high-resolution multispectral imagery (L.2). GCMRC will partner with the USGS 3D Elevation Program (3DEP; <https://www.usgs.gov/3d-elevation-program>) which is leading an ongoing effort to acquire QL1 airborne lidar data coverage of the entire nation.

A large segment of the CRe has yet to be covered by 3DEP. Moreover, 3DEP data acquisition in Grand Canyon to date has not leveraged the low, steady 8,000 CFS flows of the GCDAMP overflight missions.

Acquiring these lidar data during the next GCDAMP overflight would ensure that a low steady dam release of 8,000 CFS is maintained during the data collection, which is critical to maximize the potential for monitoring CRe resources with the data (i.e., when they are subaerially exposed and visible, as opposed to underwater in the river channel).

Anticipated Use of Data

QL1 airborne lidar have a much higher spatial accuracy (latitude, longitude, and elevation) than digital topographic data derived from other airborne remote sensing methods, and thus would be extremely valuable for baseline observations and future monitoring of a variety of resources and applications in the CRe.

Outcomes and Products

- Overflight mission and lidar acquisition: GCMRC will oversee the contracting, coordinate with all relevant parties, and oversee the mission to acquire QL1 lidar for the river corridor in conjunction with the acquisition of multispectral imagery during a low steady dam release of 8,000 CFS beginning at the start of the Memorial Day holiday weekend. Preliminary data is delivered by the contractor to GCMRC at a to-be-determined timeline following the mission. Staff then perform QA/QC working with the contractor to iteratively review and request further changes as necessary.

Budgets

Project Element Budget

| Fiscal Year 2025 | | | | | | | | | |
|--|------------------|----------------------|-----------------------|-----------------------|---------------------------|--------------------------|-----------------------------|------------------|------------------------------|
| Project L Overflight Remote Sensing in Support of GCDAMP and LTEMP | Salaries | Travel & Training | Operating Expenses | Logistics Expenses | Cooperative Agreements | To other USGS Centers | AMP Special Burden Rate* | Total | USGS Contributing Funds** |
| | | | | | | | 21.86% | | Est Full Rate = 55% |
| L.1. Analysis and interpretation of overflight remote sensing data | \$227,154 | \$7,000 | \$5,000 | \$0 | \$0 | \$0 | \$52,286 | \$291,441 | |
| Total Project L | \$227,154 | \$7,000 | \$5,000 | \$0 | \$0 | \$0 | \$52,286 | \$291,441 | \$24,975 |

| Fiscal Year 2026 | | | | | | | | | |
|--|------------------|----------------------|-----------------------|-----------------------|---------------------------|--------------------------|-----------------------------|------------------|------------------------------|
| Project L Overflight Remote Sensing in Support of GCDAMP and LTEMP | Salaries | Travel & Training | Operating Expenses | Logistics Expenses | Cooperative Agreements | To other USGS Centers | AMP Special Burden Rate* | Total | USGS Contributing Funds** |
| | | | | | | | 22.60% | | Est Full Rate = 55.5% |
| L.1. Analysis and interpretation of overflight remote sensing data | \$245,048 | \$7,000 | \$5,000 | \$0 | \$0 | \$0 | \$58,093 | \$315,141 | |
| Total Project L | \$245,048 | \$7,000 | \$5,000 | \$0 | \$0 | \$0 | \$58,093 | \$315,141 | \$27,239 |

| Fiscal Year 2027 | | | | | | | | | |
|--|------------------|----------------------|-----------------------|-----------------------|---------------------------|--------------------------|-----------------------------|------------------|------------------------------|
| Project L Overflight Remote Sensing in Support of GCDAMP and LTEMP | Salaries | Travel & Training | Operating Expenses | Logistics Expenses | Cooperative Agreements | To other USGS Centers | AMP Special Burden Rate* | Total | USGS Contributing Funds** |
| | | | | | | | 23.40% | | Est Full Rate = 56% |
| L.1. Analysis and interpretation of overflight remote sensing data | \$257,300 | \$7,000 | \$5,000 | \$0 | \$0 | \$0 | \$63,016 | \$332,316 | |
| Total Project L | \$257,300 | \$7,000 | \$5,000 | \$0 | \$0 | \$0 | \$63,016 | \$332,316 | \$28,987 |

*GCDAMP Special Rate: Rate includes current DOI preferred rate (currently 15%; subject to change) and facilities rate (will vary annually). No USGS bureau overhead is charged (unique to the GCDAMP agreement).

**USGS Contributing Funds: The amount of funds required to cover the subsidy created by the reduced burden rate (i.e., GCDAMP Special Rate). As in previous years, SBSC/GCMRC will request these funds of the USGS cost-share program. These funds are not guaranteed.

Unfunded Project Element Budget

| Fiscal Year 2025 | | | | | | | | | |
|--|------------|----------------------|-----------------------|-----------------------|---------------------------|--------------------------|-----------------------------|------------|------------|
| Project L Unfunded Project Elements | Salaries | Travel & Training | Operating Expenses | Logistics Expenses | Cooperative Agreements | To other USGS Centers | AMP Special Burden Rate* | Total | |
| | | | | | | | 21.86% | | |
| L.2. Acquisition of overflight remote sensing imagery | \$0 | \$0 | \$0 | \$0 | \$0 | \$0 | \$0 | \$0 | \$0 |
| L.3. Acquisition of airborne lidar in conjunction with overflight remote sensing imagery | \$0 | \$0 | \$0 | \$0 | \$0 | \$0 | \$0 | \$0 | \$0 |
| Total Project L | \$0 | \$0 | \$0 | \$0 | \$0 | \$0 | \$0 | \$0 | \$0 |

| Fiscal Year 2026 | | | | | | | | | |
|--|------------|----------------------|-----------------------|-----------------------|---------------------------|--------------------------|-----------------------------|------------------|--|
| Project L Unfunded Project Elements | Salaries | Travel & Training | Operating Expenses | Logistics Expenses | Cooperative Agreements | To other USGS Centers | AMP Special Burden Rate* | Total | |
| | | | | | | | 22.60% | | |
| L.2. Acquisition of overflight remote sensing imagery | \$0 | \$5,000 | \$30,000 | \$24,727 | \$0 | \$530,000 | \$13,498 | \$603,225 | |
| L.3. Acquisition of airborne lidar in conjunction with overflight remote sensing imagery | \$0 | \$0 | \$0 | \$0 | \$0 | \$265,000 | \$0 | \$265,000 | |
| Total Project L | \$0 | \$5,000 | \$30,000 | \$24,727 | \$0 | \$795,000 | \$13,498 | \$868,225 | |

| Fiscal Year 2027 | | | | | | | | | |
|--|------------|----------------------|-----------------------|-----------------------|---------------------------|--------------------------|-----------------------------|------------|------------|
| Project L Unfunded Project Elements | Salaries | Travel & Training | Operating Expenses | Logistics Expenses | Cooperative Agreements | To other USGS Centers | AMP Special Burden Rate* | Total | |
| | | | | | | | 23.40% | | |
| L.2. Acquisition of overflight remote sensing imagery | \$0 | \$0 | \$0 | \$0 | \$0 | \$0 | \$0 | \$0 | \$0 |
| L.3. Acquisition of airborne lidar in conjunction with overflight remote sensing imagery | \$0 | \$0 | \$0 | \$0 | \$0 | \$0 | \$0 | \$0 | \$0 |
| Total Project L | \$0 | \$0 | \$0 | \$0 | \$0 | \$0 | \$0 | \$0 | \$0 |

References Cited

- Bedford, A., Sankey, T.T., Sankey, J.B., Durning, L.E. and Ralston, B.E., 2018a, Remote sensing derived maps of tamarisk (2009) and beetle impacts (2013) along 412 km of the Colorado River in the Grand Canyon, Arizona: U.S. Geological Survey data release, <https://doi.org/10.5066/F72B8X71>.
- Bedford, A., Sankey, T.T., Sankey, J.B., Durning, L.E., and Ralston, B.E., 2018b, Remote sensing of tamarisk beetle (*Diorhabda carinulata*) impacts along 412 km of the Colorado River in the Grand Canyon, Arizona, USA: *Ecological Indicators*, v. 89, p. 365-375, <https://doi.org/10.1016/j.ecolind.2018.02.026>.
- Bransky, N.D., Sankey, T.T., Sankey, J.B., Johnson, M., and Jamison, L.R., 2021, Monitoring *Tamarix* changes using WorldView-2 satellite imagery in Grand Canyon National Park, Arizona: *Remote Sensing*, v. 13, no. 5, article 958, p. 1-16, <https://doi.org/10.3390/rs13050958>.
- Butterfield, B.J., Grams, P.E., Durning, L.E., Hazel, J.E., Palmquist, E.C., Ralston, B.E., and Sankey, J.B., 2020, Associations between riparian plant morphological guilds and fluvial sediment dynamics along the regulated Colorado River in Grand Canyon: *River Research and Applications*, v. 36, no. 3, p. 410-421, <https://doi.org/10.1002/rra.3589>.
- Caster, J., Sankey, J.B., Sankey, T.T., Kasprak, A., Bowker, M.A., and Joyal, T., 2024, Do topographic changes tell us about variability in aeolian sediment transport and dune mobility? Analysis of monthly to decadal surface changes in a partially vegetated and biocrust covered dunefield: *Geomorphology*, v. 447, article 109021, p. 1-16, <https://doi.org/10.1016/j.geomorph.2023.109021>.
- Davis, P.A., 2012, Airborne digital-image data for monitoring the Colorado River corridor below Glen Canyon Dam, Arizona, 2009—Image-mosaic production and comparison with 2002 and 2005 image mosaics: U.S. Geological Survey Open-File Report 2012–1139, 82 p. <http://pubs.usgs.gov/of/2012/1139/>.
- Durning, L.E., Sankey, J.B., Bedford, A., and Sankey, T.T., 2018, Riparian species vegetation classification data for the Colorado River within Grand Canyon derived from 2013 airborne imagery: U.S. Geological Survey data release, <https://doi.org/10.5066/P9OUB1RS>.
- Durning, L.E., Sankey, J.B., Chain, G.R., and Sankey, T.T., 2017a, Water classification of the Colorado River Corridor, Grand Canyon, Arizona, 2013—Data: U.S. Geological Survey data release, <https://doi.org/10.5066/F7PZ5799>.
- Durning, L.E., Sankey, J.B., Chain, G.R., and Sankey, T.T., 2017b, Riparian vegetation classification of the Colorado River Corridor, Grand Canyon, Arizona, 2013—Data: U.S. Geological Survey data release, <https://doi.org/10.5066/F7K64GJF>.
- Durning, L.E., Sankey, J.B., Davis, P.A., and Sankey, T.T., 2016, Four band image mosaic of the Colorado River corridor in Arizona-2013, including accuracy assessment data: U.S. Geological Survey data release, <http://doi.org/10.5066/F7TX3CHS>.

- Durning, L.E., Sankey, J.B., Yackulic, C.B., Grams, P.E., Butterfield, B.J., and Sankey, T.T., 2021, Hydrologic and geomorphic effects on riparian plant species occurrence and encroachment—Remote sensing of 360 km of the Colorado River in Grand Canyon: *Ecohydrology*, v. 14, no. 8, e2344, <https://doi.org/10.1002/eco.2344>.
- Gushue, T.M., 2019, Colorado River Mile System, Grand Canyon, Arizona: U.S. Geological Survey data release, <https://doi.org/10.5066/P9IRL3GV>.
- Hadley, D.R., Grams, P.E., and Kaplinski, M.A., 2018, Quantifying geomorphic and vegetation change at sandbar campsites in response to flow regulation and controlled floods, Grand Canyon National Park, Arizona: *River Research and Applications*, v. 34, no. 9, p. 1208-1218, <https://doi.org/10.1002/rra.3349>.
- Kasprak, A., Caster, J.J., and Bangen, S.G., 2017a, Geomorphic process from topographic form—Automating the interpretation of repeat survey data in river valleys: *Earth Surface Processes and Landforms*, v. 42, no. 12, p. 1872-1883, <https://doi.org/10.1002/esp.4143>.
- Kasprak, A., Caster, J., Sankey, J.B., and Bangen, S., 2017b, Geomorphic process topographic form, Colorado River, Grand Canyon—Data and models: U.S. Geological Survey data release, <https://doi.org/10.5066/F73776X6>.
- Kasprak, A., Sankey, J.B., Buscombe, D., Caster, J., East, A.E., and Grams, P.E., 2018, Quantifying and forecasting changes in the areal extent of river valley sediment in response to altered hydrology and land cover: *Progress in Physical Geography: Earth and Environment*, v. 42, no. 6, p. 739-764, <https://doi.org/10.1177/0309133318795846>.
- Kasprak, A., Sankey, J.B., and Butterfield, B.J., 2021, Future regulated flows of the Colorado River in Grand Canyon foretell decreased areal extent of sediment and increases in riparian vegetation: *Environmental Research Letters*, v. 16, no. 1, p. 1-15, <https://doi.org/10.1088/1748-9326/abc9e4>.
- Magirl, C.S., Breedlove, M.J., Webb, R.H., and Griffiths, P.G., 2008, Modeling water-surface elevations and virtual shorelines for the Colorado River in Grand Canyon, Arizona: U.S. Geological Survey Scientific Investigations Report 2008-5075, 32 p., <https://pubs.usgs.gov/sir/2008/5075>.
- Sankey, J.B., Bransky, N.D., Pigue, L., Kohl, K., and Gushue, T.M., 2024, Four band image mosaic of the Colorado River corridor in Arizona—2021, including accuracy assessment data: U.S. Geological Survey data release, <https://doi.org/10.5066/P9BBGN6G>.
- Sankey, J.B., Caster, J.J., Kasprak, A., and East, A.E., 2018b, The response of source-bordering aeolian dunefields to sediment-supply changes 2—Controlled floods of the Colorado River in Grand Canyon, Arizona, USA: *Aeolian Research*, v. 32, p. 154-169, <https://doi.org/10.1016/j.aeolia.2018.02.004>.
- Sankey, J.B., Chain, G.R., Solazzo, D., Durning, L.E., Bedford, A., Grams, P.E., and Ross, R.P., 2018c, Sand classifications along the Colorado River in Grand Canyon derived from 2002, 2009, and 2013 high-resolution multispectral airborne imagery: U.S. Geological Survey data release, <https://doi.org/10.5066/P99TN424>.

- Sankey, J.B., East, A., Fairley, H.C., Caster, J., Dierker, J., Brennan, E., Pilkington, L., Bransky, N.D., and Kasprak, A., 2023, Archaeological sites in Grand Canyon National Park along the Colorado River are eroding owing to six decades of Glen Canyon Dam operations: *Journal of Environmental Management*, v. 342, article 118036, p. 1-17, <https://doi.org/10.1016/j.jenvman.2023.118036>.
- Sankey, J.B., Kasprak, A., Caster, J.J., East, A.E., and Fairley, H., 2018a, The response of source-bordering aeolian dunefields to sediment-supply changes 1—Effects of wind variability and river-valley morphodynamics: *Aeolian Research*, v. 32, p. 228-245, <https://doi.org/10.1016/j.aeolia.2018.02.005>.
- Sankey, J.B., Ralston, B.E., Grams, P.E., Schmidt, J.C., and Cagney, L.E., 2015a, Riparian vegetation, Colorado River, and climate—Five decades of spatiotemporal dynamics in the Grand Canyon with river regulation: *Journal of Geophysical Research: Biogeosciences*, v. 120, no. 8, p. 1532-1547, <https://doi.org/10.1002/2015JG002991>.
- Sankey, J.B., Ralston, B.E., Grams, P.E., Schmidt, J.C., and Cagney, L.E., 2015b, Riparian vegetation, Colorado River, and climate—Five decades of spatiotemporal dynamics in the Grand Canyon with river regulation—Data: U.S. Geological Survey data release, <https://doi.org/10.5066/F7J67F0P>.
- Sankey, T.T., Sankey, J.B., Horne, R., and Bedford, A., 2016, Remote sensing of tamarisk biomass, insect herbivory, and defoliation—Novel methods in the Grand Canyon region, Arizona: *Photogrammetric Engineering and Remote Sensing*, v. 82, no. 8, p. 645-652, <https://doi.org/10.14358/PERS.82.8.645>.
- U.S. Department of the Interior, 2016a, Glen Canyon Dam Long-Term Experimental and Management Plan Final Environmental Impact Statement (LTEMP EIS): U.S. Department of the Interior, Bureau of Reclamation, Upper Colorado Region, National Park Service, Intermountain Region, online, <http://ltempeis.anl.gov/documents/final-eis/>.
- U.S. Department of the Interior, 2016b, Record of Decision for the Glen Canyon Dam Long-Term Experimental and Management Plan Final Environmental Impact Statement (LTEMP ROD): Salt Lake City, Utah, U.S. Department of the Interior, Bureau of Reclamation, Upper Colorado Region, National Park Service, Intermountain Region, 196 p., http://ltempeis.anl.gov/documents/docs/LTEMP_ROD.pdf.
- Yackulic, C.B., Bair, L.S., Eppehimer, D.E., Salter, G.L., Deemer, B.R., Butterfield, B.J., Kasprak, A., Caster, J.J., Fairley, H.C., Grams, P.E., Mihalevich, B.A., Palmquist, E.C., and Sankey, J.B., 2024, Modeling the impacts of Glen Canyon Dam operations on Colorado River resources: Phoenix, Ariz., U.S. Department of the Interior, Bureau of Reclamation, cooperator publication prepared by U.S. Geological Survey, Southwest Biological Science Center, Grand Canyon Monitoring and Research Center, Flagstaff, Ariz., April 2024, 133 p., <https://pubs.usgs.gov/publication/70252976>.

Project M: Leadership, Management, and Support

Investigators

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Project Summary and Purpose

Project Element M.1 includes salaries, travel, and operating expenses for the management and operation of Grand Canyon Monitoring and Research Center (GCMRC). For U.S. Geological Survey (USGS) information product release and program administration, M.1 includes partial salaries for the following support staff: Budget Analyst (75%), Public Affairs Specialist (50%), Information Product Data System/Archive Technician (50%), and a Science Data Coordinator (50%). Leadership and management personnel salaries include those for the GCMRC Chief and Deputy Chief. Partial salary for one investigator is included for Tribal coordination and special projects. Project Element M.1 also covers GCMRC travel and training costs (\$16,000 annually), including travel to support activities related to the Glen Canyon Dam Adaptive Management Program. GCMRC operating expenses include General Services Administration (GSA) vehicle costs (e.g., monthly lease fees, mileage costs, and costs for accidents and damage; \$69,000 annually) and DOI vehicle costs (e.g., fuel, maintenance, supplies, and replacement costs; \$45,000 annually). An annual contribution (\$25,000) to the GCMRC equipment and vehicles working capital fund is currently unfunded.

The GCMRC will secure all necessary state and federal permits to conduct projects within this workplan and abide by all permit requirements. As part of operations, GCMRC will summarize and report on the number of native and nonnative fishes that were caught as well as euthanized throughout the mainstem and tributary systems within the Colorado River corridor from Glen Canyon Dam to Pearce Ferry Rapid (CRe) annually for all applicable projects. This summary information will quantify the taking of fish life that occurs within the CRe in response to tribal concerns and will also facilitate identification of range expansions or new nonnative fish species detections that might not be apparent without evaluations across projects conducted by various cooperating agencies. GCMRC will humanely euthanize any animals using Institutional Animal Care and Use Committee (IACUC) approved methods. In line with promoting animal welfare, GCMRC is also undergoing a process to develop a formal protocol for the IACUC to ensure best practices when handling animals.

Project Element M.2 covers salaries for four members of the GCMRC logistics staff (one member was transferred from Project I to logistics in FY 2024 to meet workload demand). The GCMRC logistics section facilitates field efforts of the GCMRC and many cooperator agencies.

Cooperator funding for support of the Partners in Science Program with Grand Canyon Youth is currently unfunded (fiscal year [FY] 2025 = \$40,341; FY 2026 = \$40,551; FY 2027 = \$42,798).

Project Element M.3 covers GCMRC’s information technology equipment and related support costs (FY 2025 = \$59,510; FY 2026 = \$57,250; FY 2027 = \$65,750).

Long-Term Experimental and Management Plan (LTEMP) Resource Goals

Because Project M is related to the full scope of GCMRC operations and support, it touches on all 11 LTEMP Resource Goals. Project M also relates directly to Goal 12 (Maintain a high-quality monitoring, research, and adaptive management program) of the Glen Canyon Dam Adaptive Management Program Strategic Plan.

Project M supports the following 11 LTEMP resource goals:

- Archaeological and Cultural Resources
- Natural Processes
- Humpback Chub
- Hydropower and Energy
- Other Native Fish
- Recreational Experience
- Sediment
- Tribal Resources
- Rainbow Trout Fishery
- Nonnative Invasive Species
- Riparian Vegetation

Budgets

Project Element Budget

| Fiscal Year 2025 | | | | | | | | | |
|---|------------------|-------------------|--------------------|--------------------|------------------------|-----------------------|--------------------------|--------------------|---------------------------|
| Project M Leadership, Management, and Support | Salaries | Travel & Training | Operating Expenses | Logistics Expenses | Cooperative Agreements | To other USGS Centers | AMP Special Burden Rate* | Total | USGS Contributing Funds** |
| | | | | | | | 21.86% | | Est Full Rate = 55% |
| M.1. Leadership, management, and support | \$616,843 | \$16,000 | \$114,000 | \$0 | \$0 | \$0 | \$163,282 | \$910,125 | |
| M.2. Logistics staff | \$364,547 | \$0 | \$0 | \$0 | \$0 | \$0 | \$79,701 | \$444,248 | |
| M.3. IT | \$0 | \$0 | \$59,510 | \$0 | \$0 | \$0 | \$13,011 | \$72,521 | |
| Total Project M | \$981,390 | \$16,000 | \$173,510 | \$0 | \$0 | \$0 | \$255,994 | \$1,426,893 | \$122,277 |

| Fiscal Year 2026 | | | | | | | | | |
|---|--------------------|-------------------|--------------------|--------------------|------------------------|-----------------------|--------------------------|--------------------|---------------------------|
| Project M Leadership, Management, and Support | Salaries | Travel & Training | Operating Expenses | Logistics Expenses | Cooperative Agreements | To other USGS Centers | AMP Special Burden Rate* | Total | USGS Contributing Funds** |
| | | | | | | | 22.60% | | Est Full Rate = 55.5% |
| M.1. Leadership, management, and support | \$661,313 | \$16,000 | \$114,000 | \$0 | \$0 | \$0 | \$178,837 | \$970,150 | |
| M.2. Logistics staff | \$382,774 | \$0 | \$0 | \$0 | \$0 | \$0 | \$86,507 | \$469,281 | |
| M.3. IT | \$0 | \$0 | \$57,250 | \$0 | \$0 | \$0 | \$12,939 | \$70,189 | |
| Total Project M | \$1,044,088 | \$16,000 | \$171,250 | \$0 | \$0 | \$0 | \$278,282 | \$1,509,620 | \$130,483 |

| Fiscal Year 2027 | | | | | | | | | |
|---|--------------------|----------------------|-----------------------|-----------------------|---------------------------|--------------------------|-----------------------------|--------------------|------------------------------|
| Project M Leadership, Management, and Support | Salaries | Travel & Training | Operating Expenses | Logistics Expenses | Cooperative Agreements | To other USGS Centers | AMP Special Burden Rate* | Total | USGS Contributing Funds** |
| | | | | | | | 23.40% | | Est Full Rate = 56% |
| M.1. Leadership, management, and support | \$695,066 | \$16,000 | \$114,000 | \$0 | \$0 | \$0 | \$193,066 | \$1,018,132 | |
| M.2. Logistics staff | \$401,913 | \$0 | \$0 | \$0 | \$0 | \$0 | \$94,048 | \$495,961 | |
| M.3. IT | \$0 | \$0 | \$65,750 | \$0 | \$0 | \$0 | \$15,386 | \$81,136 | |
| Total Project M | \$1,096,980 | \$16,000 | \$179,750 | \$0 | \$0 | \$0 | \$302,499 | \$1,595,228 | \$139,148 |

*GCDAMP Special Rate: Rate includes current DOI preferred rate (currently 15%; subject to change) and facilities rate (will vary annually). No USGS bureau overhead is charged (unique to the GCDAMP agreement).

**USGS Contributing Funds: The amount of funds required to cover the subsidy created by the reduced burden rate (i.e., GCDAMP Special Rate). As in previous years, SBSC/GCMRC will request these funds of the USGS cost-share program. These funds are not guaranteed.

Project N: Native Fish Population Dynamics (New Project)

Note that Project N was previously **Hydropower Monitoring and Research**. That project is no longer continuing, and instead Project N will be a new project, **Native Fish Population Dynamics**.

Investigators

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Project Summary and Purpose

Monitoring of fishes in the Colorado River ecosystem (CRe) has been conducted for >20 years, and yet few analyses of these data have occurred at a system-wide level to understand basic drivers of “Other Native Fishes” (ONF) population dynamics. The purpose of Project N is to complete analysis of (primarily) existing data to understand basic demographic rates, assist managers in decision-making related to ONF conservation, and understand movements and sources of mortality of endangered razorback sucker (*Xyrauchen texanus*) released into the CRe to inform future augmentation strategies.

Long-Term Experimental and Management Plan (LTEMP) Resource Goals

- Maintain self-sustaining native fish species population and their habitats in their natural ranges in the Colorado River and its tributaries.

Project N addresses the following Long-term Experimental and Management Plan (LTEMP; U.S. Department of the Interior 2016) goal of maintaining self-sustaining native fish species population and their habitats in their natural ranges in the Colorado River and its tributaries. The LTEMP goal for ONF) applies to razorback, flannelmouth (*Catostomus latipinnis*), and bluehead (*C. discobolus*) suckers and speckled dace (*Rhinichthys osculus*) in the CRe.

The project was designed to assess how flow and non-flow management actions have affected ONF, and whether progress toward the LTEMP goal is being achieved.

Also, a comprehensive analysis of efficacy of conservation measures in the LTEMP Biological Opinion (U.S. Department of the Interior, 2016) for razorback sucker is needed. Dam-related flow and non-flow management actions related to the LTEMP Biological Opinion, including high flow experiments, suppression of nonnative fishes, or macroinvertebrate flows, may influence fish populations in unexpected ways that are both beneficial (Healy and others, 2020; Deemer, 2022; Hansen and others, 2023) and detrimental (Healy and others, 2022a) to LTEMP resource goals. Further analyses are needed at the population and community level to understand how ONF demographic rates and LTEMP metrics (e.g., proportion of CRe and tributaries occupied by ONF) may respond to management actions.

Background

Significant changes in the CRe are occurring as Glen Canyon Dam discharge becomes warmer in summer due to declining reservoir levels (Dibble and others, 2021), leading to increased reproduction rates and distribution of native fishes (Kegerries and others, 2020; Gilbert and others, 2022; Dzul and others, 2023), potential declines in cold-water nonnative salmonids (Bruckerhoff and others, 2022; Healy and others, 2023), and increased entrainment and expansion of warm-water nonnative fishes from Lake Powell (Epehimer and others, 2024). While abundance or catch rates of humpback chub (*Gila cypha*) and ONF have increased (Van Haverbeke and others, 2017; Rogowski and others, 2018), the expansion in warm-water nonnative sport fishes may threaten the existing native fish community (Epehimer and others, 2024).

Managers are developing novel flow alternatives to respond to new fish introductions and expansions in the Colorado River (U.S. Department of the Interior, 2024) that may also compel managers to consider fish conservation actions focused on tributaries (Bouska and others, 2023). Tributaries have been important to native fish populations when conditions in the mainstem were less conducive to successful native fish reproduction (Yackulic and others, 2014). Tributaries continue to provide reproduction and potential rearing habitat (Healy and others, 2020; Bonjour and others, 2023) but the relative importance of tributary and mainstem habitats in maintaining ONF populations and meeting LTEMP goals is unclear. Thus, a baseline understanding of system-wide ONF population dynamics is needed to understand drivers of survival, recruitment, and basic ecology to inform management.

Endangered Species Act (ESA)-listed razorback sucker was considered extirpated from the Grand Canyon beginning in the mid-1990s, and despite the presence of adults and evidence of spawning after 2012, the species continues to be rare (Kegerries and others, 2017; Gilbert and others, 2022). Conservation measures in the LTEMP (U.S. Department of the Interior, 2016) are focused on understanding the status and life stage- or habitat-specific vulnerabilities of razorback sucker.

Despite documented spawning (Gilbert and others, 2022), no evidence of recruitment from larval to juvenile life stages has been uncovered in the Grand Canyon – it is unclear if the thermal regime, habitat, predation, or other factors limit the razorback sucker in the CRe. Movements of razorback sucker between the Grand Canyon and Lake Mead, where the only population supported by natural reproduction exists (Albrecht and others, 2010; Albrecht and others, 2017), suggests the populations are linked. Adaptive management actions are being considered to both recover razorback sucker in the CRe and Lake Mead and learn about factors limiting their population growth.

For example, a three-year pilot-level razorback sucker augmentation project (Healy and others, 2022b) was initiated with the release of passive-integrated transponder (PIT) tagged age-1 razorback sucker in Havasu Creek by the NPS in 2023. Understanding long-distance movements and dispersal of newly released razorback sucker may require a system-wide approach to identify the fate of both adult and age-1 PIT-tagged fishes (Pennock and others, 2020; Pennock and others, 2024). Glen Canyon Dam Adaptive Management Program (GCDAMP) interagency monitoring programs have and may continue to detect or capture razorback sucker.

Thus, the Grand Canyon presents a unique opportunity to study larval and juvenile survival, growth, and recruitment in a rapidly changing system with fewer existing predators, which are thought to limit natural recruitment and recovery of razorback sucker in other parts of the Colorado River basin (Marsh and Langhorst, 1988; Marsh and Brooks, 1989; Schooley and Marsh, 2007).

Decisions related to fisheries management actions in the CRe are becoming more difficult under uncertain system-wide change and due to the desire to balance a diversity of values held by traditionally associated Tribes and stakeholders, alongside management agency mandates. Department of the Interior utilizes predictive models in many decision-making processes for fish management (Runge and others, 2011a; Runge and others, 2018), but adequate models for ONF have not been developed for the CRe and tributaries. In general, predictions of the outcomes of management actions are also plagued by uncertainty (Runge and others, 2011b). Uncertainty may become increasingly important as new stressors and changes in water quality impact the CRe in ways that may be difficult to predict.

Nonetheless, in systems facing rapid change or newly expanding nonnative fishes, timely decision-making is particularly relevant despite uncertainty in the outcome of taking a management action. The costs and effort needed to maintain important ecological values are often greater once a species that is newly expanding into the CRe has become established, for instance (Healy and others, 2023). Managers may wrestle with different sources of uncertainty during decision-making processes, but not all sources of uncertainty in predictions of management outcomes are important. Value of information (VoI) analytical tools can be used to assist managers in understanding the importance of sources of uncertainty to decision-making and quantifying the tradeoffs between making an immediate management decision under uncertainty or waiting to conduct additional research.

The development of predictive models based on long-term ONF data, and VoI tools can assist managers and researchers in prioritizing future monitoring to improve the outcomes of LTEMP flow and non-flow management for ONF, while also considering diverse values of GCDAMP stakeholders, agencies, and Tribes.

Project N specifically addresses research to understand ONF population dynamics and assess progress toward LTEMP goals for native fishes using a mix of existing data types, and data collected using new technology, to assess management actions.

We will incorporate long-term ONF data collected from throughout the CRe by cooperating agencies (also see projects G and I), including GCMRC, Arizona Game and Fish Department (AZGFD), U.S. Fish and Wildlife Service (USFWS), and National Park Service (NPS), and others in tributaries (Little Colorado River, Bright Angel, Shinumo, and Havasu creeks) and the mainstem Colorado River, to estimate demographic rates (recruitment, survival, population growth), and distribution and abundance using mark-recapture and occupancy models. Mark-recapture modeling has proven to be useful for understanding temporal and spatial variation in survival and growth for humpback chub (Yackulic and others, 2014; Healy and others, 2022c; Dzul and others, 2023), inferring interspecific interactions between native and nonnative species (Yackulic and others, 2018), and understanding the outcomes and influences of management actions on native fish populations (Yackulic and others, 2021; Healy and others, 2022c; Hansen and others, 2023; Healy and Omana Smith, 2024).

The outcome of this project will support the use of monitoring data for key management needs including informing state-dependent decision-making (management decisions in the context of changing environmental conditions), evaluating the effectiveness of both flow and non-flow management actions with respect to LTEMP goals, and provide a structure for future learning through the identification and prioritization of key uncertainties making decision-making difficult (Lyons and others, 2008).

Collaborations between GCMRC, federal and state management agencies, and contract biologists to analyze existing capture, mark-recapture, and PIT tag antenna detection data collected over decades from throughout the CRe and tributaries should allow for unprecedented learning related to the ecology of ONF and improved understanding of influences of LTEMP flow and non-flow management actions. The work proposed in project N is achievable over fiscal years 2025-2027 because most data have already been collected; however, modeling, and other work proposed would provide a baseline for future work plans as more demographic data are collected during monitoring efforts described in projects G, H, and I.

Native fishes, including razorback sucker listed under the ESA, and Colorado River endemic bluehead and flannelmouth suckers, have been sampled through GCDAMP-funded monitoring programs for over 25 years.

Yet, few analyses have been undertaken to understand how and if LTEMP goals are being met for “other native fishes” including speckled dace, bluehead, flannelmouth, and razorback suckers. Meeting this goal may become increasingly difficult as expansion of warm-water sport fishes occur.

Further, understanding the role of tributary and mainstem habitats is needed to develop and implement future management actions meant to maintain populations of these fishes (Bouska and others, 2023). For example, construction of barriers meant to provide refuge from introduced predators may result in isolation and extirpation rather than metapopulation maintenance (Bouska and others, 2023).

Project N is proposed to be minimally funded in the FY 2025-27 TWP. Thus, limited insight will be gained into the above science questions. Nonetheless, we proposed to partially fund a graduate student to answer a subset (to be determined) of the science questions listed below, during this work plan.

Proposed Work

Project Element N.1. Sucker and Dace Distribution and Demographic Modeling (New Study; Partially Funded)

Hypotheses and Science Questions

We propose several basic questions we would attempt to answer (funding limits scope), related to native fish population dynamics, and drivers of recruitment, growth, and survival:

- 1) What are the long-term relationships between variation in temperature, river discharge, and other extrinsic factors in the CRe with recruitment, growth, and survival of native fishes?
- 2) What are the effects of high-flow experiments, other types of designer flows, and other management actions (removals or flow management to suppress warm-water and cold-water nonnative fish species) on native fish population dynamics?
- 3) Which species of warm-water nonnative fish should be prioritized for monitoring and management to minimize negative population-level impacts on native fishes?
- 4) How much mortality in juvenile native fishes is related to predation?

Methods

This project element involves 1) using existing data to estimate the probability of occurrence (occupancy) of all ONF to support the evaluation of trends in LTEMP metrics in development, 2) a synthesis of environmental and mark-recapture data to estimate demographic rates of bluehead and flannelmouth suckers throughout the CRe, and 3) assist NPS and USFWS in estimating

growth and survival of razorback sucker released in Grand Canyon as part of a pilot-level augmentation study (2023-2025). Finally, 4) we plan to collaboratively investigate drivers of early life stage dynamics of ONF using larval and small-bodied fish data (2014-present, ASIR, Inc., and BIO-WEST, Inc.) collected in cooperation with the Bureau of Reclamation. All four emphasis areas in this project depend on the availability of sufficient data for each of the ONF species. It is particularly uncertain whether sufficient recaptures of razorback sucker released as part of the augmentation project will allow for estimation of growth and survival rates (Project Element N.3 may also assist managers in understanding the fate of these fish using acoustic telemetry). Nonetheless, the ultimate objective of Project Element N.1 is to understand and relate long-term variation in environmental (e.g., temperature, watery quality, flow regimes) and biological (e.g., predators, competitor abundance) drivers to ONF demographics (recruitment, survival). Because fishes often use both tributary and mainstem habitats for different life history stages, and tributary occupancy may impart an advantage to native fishes, we intend to incorporate both tributary and mainstem interagency data into these analyses. Movement rates and the use of tributary and mainstem habitats by ONF will also be assessed using a mix of acoustic and PIT-tag telemetry methods (e.g., Bonjour and others, 2023; Dzul and others, 2022). By incorporating system-wide capture or detection data (including through PIT-tag antennas), we may begin to also understand the importance of tributaries and mainstem habitats in supporting native fish populations to inform future management actions.

Anticipated Use of Data

Project Element N.1 involves only the use and analysis of existing data to understand the influence of dam operations, environmental variation, expanding nonnative species, and other management actions on the population dynamics of ONF. This information could be used in models to predict the impacts of future management actions (see Project Element N.2 also), including decisions involving Glen Canyon Dam operations and fish management (e.g., translocations, razorback sucker augmentation). No additional data will be generated by this element; however, data collected during monitoring through other funded Projects G, H, and I may also be incorporated into analyses included in Project N (funding may limit scope). Results from these modeling efforts will be incorporated into predictive models to assess future management alternatives and support DOI decision-making processes.

Outcomes and Products

Funding remaining in this proposed budget is intended to partially fund a PhD student to analyze existing native fish data to answer a subset of priority science questions listed above. However, specific questions are to be determined, in consultation with the project lead, graduate student, and PhD committee members. The project N lead (B. Healy) will serve on the graduate student's advisory committee, to ensure alignment of the student's dissertation work with Project N goals, and to provide guidance, mentoring, and support to the student.

If funded at the level proposed, a presentation may be given at a future annual reporting meeting and potentially at regional scientific meetings, and 1-3 peer-reviewed manuscripts will be produced through the course of the student's graduate work. A dissertation will be finalized in the next work plan, assuming a doctoral degree requires at least 4 years to complete.

If full funding was provided for N.1, based on earlier estimates proposed (i.e., 3-year average of \$136k and \$45k for N.1 and N.2, respectively, to cover salaries in FY 2025-27), the following products may be completed (4-5 manuscripts and multiple presentations):

- Occupancy modeling to calculate LTEMP metrics for ONF.
- Demographic analysis for flannelmouth, bluehead, and razorback suckers, including linking estimates of survival, recruitment, and growth to flow experiments and non-flow actions, and other environmental variables.

Project Element N.2. Predictive Modeling and Decision Support for Native Fishes (New Study; Unfunded)

Hypotheses and Science Questions

- How important is uncertainty to decision-making related to ONF management actions? What are future priorities for research (i.e., critical uncertainties) to meet ONF LTEMP goals?

Methods

For Project Element N.2, we will use the results of demographic modeling in Project Element N.1 to develop predictive models to inform management and decision-making for ONF (e.g., matrix-based models, metapopulation viability model; Runge and others, 2018; Healy and others, 2023). Management decisions may include where and how to focus ONF restoration, augmentation (razorback sucker), or translocation efforts, how to protect native fish populations (e.g., in tributary refuges above barriers), prioritization of uncertain threats including nonnative species, or those related to LTEMP flow actions, among others. GCMRC scientists will work with the U.S. Department of the Interior and state managers, Tribes, and GCDAMP stakeholders, as appropriate, to outline decisions to be made related to GCDAMP recommendations for ONF conservation and LTEMP management objectives, facilitate the development of alternatives, and make predictions about the outcomes of different potential management alternatives. We will also identify and prioritize different sources of uncertainty, and the influence of uncertainty on decision-making for ONF using VoI tools such as expected value of perfect information or expected value of partial perfect information. These tools will allow for the prioritization of future research towards critical uncertainties.

Anticipated Use of Data

No additional data will be generated by this element; however, uses of data are described in the methods above.

However, models developed in association with this element would be used to establish priorities for research in future analysis and work plans (e.g., determine how important future research may be to inform management decisions). The primary purpose of Project Element N.2 is to directly support decision-making for management.

Outcomes and Products

Predictive models to support decision-making will be produced, followed by value of information analysis to understand whether uncertainties in drivers of population dynamics matter for decision-making.

Project Element N.3. Evaluating Dispersal and Sources of Mortality of Razorback Sucker using New Technology (New Study; Unfunded)

Hypotheses and Science Questions

- What are growth, survival, and dispersal rates of juvenile razorback sucker and how do they compare to other Colorado River populations where augmentation programs occur?

Razorback sucker recovery is thought to be plagued by low survival of juvenile or subadult fish due to predation by introduced predators in other Colorado River basin populations (Schooley and Marsh 2007). Populations of razorback sucker have not grown in the Grand Canyon despite the presence of adults and seasonally warming water temperatures, with lower nonnative predator loads in much of the western Grand Canyon where most razorback sucker detections have occurred. Further, rapid and long-distance dispersal of razorback suckers or other native fishes from release sites can complicate interpretation of survival of translocated fish or outcomes of stocking (Spurgeon and others, 2015; Franssen and others, 2021). Predation by nonnative fishes, including by catfishes (Ictaluridae), smallmouth bass (*Micropterus dolomieu*), and others, has been identified as a concern in the existing and future CRe. Quantifying predation rates, and how the likelihood of predation varies spatially, may assist managers in recovering ONF by identifying areas where predation is likely to occur.

Methods

For Project Element N.3, we propose to use acoustic telemetry to quantify dispersal from release sites and predation as a source of mortality of age-1 razorback sucker released in the CRe.

New acoustic tags equipped with predation sensors will be used to assess the probability of predation and location of predation events (e.g., Schultz and others, 2017; Mensinger and others, 2024) following release of razorback sucker, or other fishes, if deemed appropriate. Once consumed by a predator and exposed to stomach acids, predation sensor tags switch emission signals and signify that a predation event has occurred.

A network of acoustic receivers will be temporarily (~4-5 month duration) installed near release sites and in other appropriate areas thought to be important for native fishes, to detect movements and predation event signals from acoustic tagged fish. Multistate mark-recapture models will be used to estimate the likelihood of predation across the network (Mensing and others, 2024). Predation sensor tags are slightly larger than PIT-tags, and the size of fish tagged will be large enough (100-150 mm total length) to avoid predation by humpback chub, but not avoid the gape of most introduced predator sport fishes (e.g., smallmouth bass, trout, walleye [*Sander vitreus*], catfishes). These tags have been tested and used successfully in Atlantic salmon (*Salmo salar*) smolts in rivers of the Northeast United States (Mensing and others, 2024). We will first review testing results for these tags, identify the type of tag most appropriate for our application, and conduct validation trials as needed (e.g., Schultz and others, 2017).

Anticipated Use of Data

Data collected would be used to estimate predation as a source of mortality for released razorback sucker; additional information would provide evidence of where predation may occur (e.g., tributary vs mainstem, Little Colorado inflow reach) to inform management of nonnative predators or mitigate other forms of mortality. Estimates of movement rates of razorback sucker would be used to inform future release strategies for the razorback sucker augmentation program to contribute towards meeting LTEMP conservation measures.

Outcomes and Products

Presentations would be given at the annual reporting meeting and at regional meetings, and a minimum of one manuscript will be produced.

Budgets

Project Element Budget

| Fiscal Year 2025 | | | | | | | | | |
|---|------------|-------------------|--------------------|--------------------|------------------------|-----------------------|--------------------------|-----------------|---------------------------|
| Project N | Salaries | Travel & Training | Operating Expenses | Logistics Expenses | Cooperative Agreements | To other USGS Centers | AMP Special Burden Rate* | Total | USGS Contributing Funds** |
| Native Fish Population Dynamics | | | | | | | 21.86% | | Est Full Rate = 55% |
| N.1. Sucker and dace distribution and demographics (SADDAD) | \$0 | \$0 | \$0 | \$0 | \$22,217 | \$0 | \$667 | \$22,884 | |
| Total Project N | \$0 | \$0 | \$0 | \$0 | \$22,217 | \$0 | \$667 | \$22,884 | \$0 |

| Fiscal Year 2026 | | | | | | | | | |
|---|------------|-------------------|--------------------|--------------------|------------------------|-----------------------|--------------------------|------------|---------------------------|
| Project N Native Fish Population Dynamics | Salaries | Travel & Training | Operating Expenses | Logistics Expenses | Cooperative Agreements | To other USGS Centers | AMP Special Burden Rate* | Total | USGS Contributing Funds** |
| | | | | | | | 22.60% | | Est Full Rate = 55.5% |
| N.1. Sucker and dace distribution and demographics (SADDAD) | \$0 | \$0 | \$0 | \$0 | \$0 | \$0 | \$0 | \$0 | |
| Total Project N | \$0 | \$0 | \$0 | \$0 | \$0 | \$0 | \$0 | \$0 | \$0 |

| Fiscal Year 2027 | | | | | | | | | |
|---|------------|-------------------|--------------------|--------------------|------------------------|-----------------------|--------------------------|------------|---------------------------|
| Project N Native Fish Population Dynamics | Salaries | Travel & Training | Operating Expenses | Logistics Expenses | Cooperative Agreements | To other USGS Centers | AMP Special Burden Rate* | Total | USGS Contributing Funds** |
| | | | | | | | 23.40% | | Est Full Rate = 56% |
| N.1. Sucker and dace distribution and demographics (SADDAD) | \$0 | \$0 | \$0 | \$0 | \$0 | \$0 | \$0 | \$0 | |
| Total Project N | \$0 | \$0 | \$0 | \$0 | \$0 | \$0 | \$0 | \$0 | \$0 |

GCDAMP Special Rate: Rate includes current DOI preferred rate (currently 15%; subject to change) and facilities rate (will vary annually). No USGS bureau overhead is charged (unique to the GCDAMP agreement).

**USGS Contributing Funds: The amount of funds required to cover the subsidy created by the reduced burden rate (i.e., GCDAMP Special Rate). As in previous years, SBSC/GCMRC will request these funds of the USGS cost-share program. These funds are not guaranteed.

Unfunded Project Element Budget

| Fiscal Year 2025 | | | | | | | | | |
|--|------------------|-------------------|--------------------|--------------------|------------------------|-----------------------|--------------------------|------------------|--|
| Project N Unfunded Project Elements | Salaries | Travel & Training | Operating Expenses | Logistics Expenses | Cooperative Agreements | To other USGS Centers | AMP Special Burden Rate* | Total | |
| | | | | | | | 21.86% | | |
| N.1. Sucker and dace distribution and demographics (SADDAD) | \$78,152 | \$2,000 | \$0 | \$0 | \$35,250 | \$0 | \$18,531 | \$133,933 | |
| N.2. Predictive modeling and decision support for native fishes | \$34,587 | \$0 | \$0 | \$0 | \$0 | \$0 | \$7,540 | \$42,127 | |
| N.3. Evaluating dispersal and sources of mortality (Razorback sucker) using new technology | \$28,966 | \$0 | \$143,230 | \$6,702 | \$0 | \$0 | \$39,000 | \$217,898 | |
| Total Project N | \$141,705 | \$2,000 | \$143,230 | \$6,702 | \$35,250 | \$0 | \$65,071 | \$393,958 | |

| Fiscal Year 2026 | | | | | | | | | |
|--|------------------|-------------------|--------------------|--------------------|------------------------|-----------------------|--------------------------|------------------|--|
| Project N Unfunded Project Elements | Salaries | Travel & Training | Operating Expenses | Logistics Expenses | Cooperative Agreements | To other USGS Centers | AMP Special Burden Rate* | Total | |
| | | | | | | | 22.60% | | |
| N.1. Sucker and dace distribution and demographics (SADDAD) | \$105,852 | \$2,000 | \$0 | \$0 | \$0 | \$0 | \$24,375 | \$132,227 | |
| N.2. Predictive modeling and decision support for native fishes | \$37,008 | \$0 | \$0 | \$0 | \$0 | \$0 | \$8,364 | \$45,372 | |
| N.3. Evaluating dispersal and sources of mortality (Razorback sucker) using new technology | \$30,994 | \$0 | \$94,260 | \$6,702 | \$0 | \$0 | \$29,822 | \$161,778 | |
| Total Project N | \$173,854 | \$2,000 | \$94,260 | \$6,702 | \$0 | \$0 | \$62,561 | \$339,377 | |

| Fiscal Year 2027 | | | | | | | | | |
|--|------------------|-------------------|--------------------|--------------------|------------------------|-----------------------|--------------------------|------------------|--|
| Project N Unfunded Project Elements | Salaries | Travel & Training | Operating Expenses | Logistics Expenses | Cooperative Agreements | To other USGS Centers | AMP Special Burden Rate* | Total | |
| | | | | | | | 23.40% | | |
| N.1. Sucker and dace distribution and demographics (SADDAD) | \$113,262 | \$2,000 | \$0 | \$0 | \$0 | \$0 | \$26,971 | \$142,233 | |
| N.2. Predictive modeling and decision support for native fishes | \$39,599 | \$0 | \$0 | \$0 | \$0 | \$0 | \$9,266 | \$48,865 | |
| N.3. Evaluating dispersal and sources of mortality (Razorback sucker) using new technology | \$33,164 | \$0 | \$94,260 | \$6,702 | \$0 | \$0 | \$31,385 | \$165,511 | |
| Total Project N | \$186,025 | \$2,000 | \$94,260 | \$6,702 | \$0 | \$0 | \$67,622 | \$356,609 | |

References Cited

- Albrecht, B.A., Holden, P.B., Kegerries, R.B., and Golden, M.E., 2010, Razorback sucker recruitment in Lake Mead, Nevada–Arizona, why here?: *Lake and Reservoir Management*, v. 26, no. 4, p. 336-344, <https://doi.org/10.1080/07438141.2010.511966>.
- Albrecht, B., Mohn, H.E., Kegerries, R.B., McKinstry, M.C., Rogers, R., Francis, T., Hines, B., Stolberg, J., Ryden, D., Elverud, D., Schleicher, B., Creighton, K., Healy, B., and Senger, B., 2017, Use of inflow areas in two Colorado River Basin reservoirs by the endangered razorback sucker (*Xyrauchen texanus*): *Western North American Naturalist*, v. 77, no. 4, p. 500-514, <https://doi.org/10.3398/064.077.0410>.
- Bonjour, S.M., Gido, K.B., McKinstry, M.C., Cathcart, C.N., Bogaard, M.R., Dzul, M.C., Healy, B.D., Hooley-Underwood, Z.E., Rogowski, D.L., and Yackulic, C.B., 2023, Migration timing and tributary use of spawning flannelmouth sucker (*Catostomus latipinnis*): *Journal of Fish Biology*, v. 103, no. 5, p. 1144-1162, <https://doi.org/10.1111/jfb.15509>.
- Bouska, K.L., Healy, B.D., Moore, M.J., Dunn, C.G., Spurgeon, J.J., and Paukert, C.P., 2023, Diverse portfolios—Investing in tributaries for restoration of large river fishes in the Anthropocene: *Frontiers in Environmental Science*, v. 11, p. 1-18, <https://doi.org/10.3389/fenvs.2023.1151315>.
- Bruckerhoff, L.A., Wheeler, K., Dibble, K.L., Mihalevich, B.A., Neilson, B.T., Wang, J., Yackulic, C.B., and Schmidt, J.C., 2022, Water storage decisions and consumptive use may constrain ecosystem management under severe sustained drought: *Journal of American Water Resources Association*, v. 58, no. 5, p. 654-672, <https://doi.org/10.1111/1752-1688.13020>.
- Deemer, B.R., Yackulic, C.B., Hall, R.O., Jr., Dodrill, M.J., Kennedy, T.A., Muehlbauer, J.D., Topping, D.J., Voichick, N., and Yard, M.D., 2022, Experimental reductions in subdaily flow fluctuations increased gross primary productivity for 425 river kilometers downstream: *PNAS Nexus*, v. 1, no. 3, pgac094, <https://doi.org/10.1093/pnasnexus/pgac094>.
- Dibble, K.L., Yackulic, C.B., Kennedy, T.A., Bestgen, K.R., and Schmidt, J.C., 2021, Water storage decisions will determine the distribution and persistence of imperiled river fishes: *Ecological Applications*, v. 31, no. 2, e02279, p. 1-9, <https://doi.org/10.1002/eap.2279>.
- Dzul, M.C., Yackulic, C.B., Giardina, M., Van Haverbeke, D.R., and Yard, M., 2023, Vital rates of a burgeoning population of humpback chub in western Grand Canyon: *Transactions of the American Fisheries Society*, v. 153, no. 4, p. 443-459, <https://doi.org/10.1002/tafs.10415>.
- Dzul, M.C., Yackulic, C.B., Kendall, W.L., Winkelman, D.L., Conner, M.M., and Yard, M.D., 2022, Incorporating antenna detections into abundance estimates of fish: *Canadian Journal of Fisheries and Aquatic Sciences*, v. 79, no. 3, p. 436–447, <https://doi.org/10.1139/cjfas-2021-0003>.
- Eppehimer, D.E., Yackulic, C.B., Bruckerhoff, L.A., Wang, J., Young, K.L., Bestgen, K.R., Mihalevich, B.A., and Schmidt, J.C., 2024, Declining reservoir elevations following a two-

- decade drought increase water temperatures and non-native fish passage facilitating a downstream invasion: bioRxiv, <https://doi.org/10.1101/2024.01.23.576966>.
- Franssen, N.R., Durst, S.L., Gilbert, E.I., Knight, W.K., and Ulibarri, M., 2021, Flow conditioning of hatchery-reared razorback sucker increases apparent survival in the wild: *North American Journal of Fisheries Management*, v. 41, no. 2, p. 545-555, <https://doi.org/10.1002/nafm.10564>.
- Gilbert, E.I., Brandenburg, W.H., Barkalow, A.L., Kegerries, R.B., Albrecht, B.C., Healy, B.D., Omana Smith, E.C., Stolberg, J.R., McKinstry, M.C., and Platania, S.P., 2022, Systematic larval fish surveys and abiotic correlates characterize extant native fish assemblage reproductive success in the Colorado River, western Grand Canyon, Arizona: *The Southwestern Naturalist*, v. 66, no. 1, p. 67-76, <https://doi.org/10.1894/0038-4909-66.1.67>.
- Hansen, L.E., Yackulic, C.B., Dickson, B.G., Deemer, B.R., and Best, R.J., 2023, Linking ecosystem processes to consumer growth rates—Gross primary productivity as a driver of freshwater fish somatic growth in a resource-limited river: *Canadian Journal of Fisheries and Aquatic Sciences*, v. 80, no. 9, p. 1456-1469, <https://doi.org/10.1139/cjfas-2022-0229>.
- Healy, B.D., Yackulic, C.B., and Schelly, R.C., 2022a, Impeding access to tributary spawning habitat and releasing experimental fall-timed floods increases brown trout immigration into a dam's tailwater: *Canadian Journal of Fisheries and Aquatic Sciences*, v. 80, no. 3, p. 614-627, <https://doi.org/10.1139/cjfas-2022-0231>.
- Healy, B.D., Omana Smith, E.C., Schelly, R.C., Koller, R., Trammell, M., Young, K., Ulibarri, M., Wilson, W.D., and Knight, W., 2022b, Detailed implementation plan—Reversing population declines of endangered razorback sucker through adaptive management in Grand Canyon National Park: Grand Canyon, Ariz., National Park Service, Grand Canyon National Park, 25 p., <https://doi.org/10.13140/RG.2.2.25769.20320>.
- Healy, B.D., Budy, P., Conner, M.M., and Omana Smith, E., 2022c, Life and death in a dynamic environment—Invasive trout, floods, and intraspecific drivers of translocated populations: *Ecological Applications*, v. 32, no. 6, e2635, p. 1-28, <https://doi.org/10.1002/eap.2635>.
- Healy, B., Budy, P., Yackulic, C., Murphy, B.P., Schelly, R.C., and McKinstry, M.C., 2023, Exploring metapopulation-scale suppression alternatives for a global invader in a river network experiencing climate change: *Conservation Biology*, v. 37, no. 1, e13993, p. 1-18, <https://doi.org/10.1111/cobi.13993>.
- Healy, B.D., and Smith, E.O., 2024, Quantifying the contributions of tributaries to large-river fish populations through mark-recapture modeling: *North American Journal of Fisheries Management*, p. 1-20, <https://doi.org/10.1002/nafm.10971>.
- Healy, B.D., Schelly, R.C., Yackulic, C.B., Omana Smith, E.C., and Budy, P., 2020, Remarkable response of native fishes to invasive trout suppression varies with trout density, temperature, and annual hydrology: *Canadian Journal of Fisheries and Aquatic Sciences*, v. 77, no. 9, p. 1446-1462, <https://doi.org/10.1139/cjfas-2020-0028>.
- Kegerries, R.B., Albrecht, B., McKinstry, M.C., Rogers, R.J., Valdez, R.A., Barkalow, A.L., Gilbert, E.E., Mohn, H.E., Healy, B., and Omana Smith, E., 2020, Small-bodied fish surveys

- demonstrate native fish dominance over 300 kilometers of the Colorado River through Grand Canyon, Arizona: *Western North American Naturalist*, v. 80, no. 2, p. 146-156, <https://doi.org/10.3398/064.080.0202>.
- Kegerries, R.B., Albrecht, B.C., Gilbert, E.E., Brandenburg, W.H., Barkalow, A.L., McKinstry, M.C., Mohn, H.E., Healy, B.D., Stolberg, J.R., Omana Smith, E.C., Nelson, C.B., and Rogers, R.J., 2017, Occurrence and reproduction by razorback sucker (*Xyrauchen texanus*) in the Grand Canyon, Arizona: *The Southwestern Naturalist*, v. 62, no. 3, p. 227-232, <https://doi.org/10.1894/0038-4909-62.3.227>.
- Lyons, J.E., Runge, M.C., Laskowski, H.P., and Kendall, W.L., 2008, Monitoring in the context of structured decision-making and adaptive management: *Journal of Wildlife Management*, v. 72, no. 8, p. 1683-1692, <https://doi.org/10.2193/2008-141>.
- Marsh, P.C., and Brooks, J.E., 1989, Predation by ictalurid catfishes as a deterrent to re-establishment of hatchery-reared razorback suckers: *The Southwestern Naturalist*, v. 34, no. 2, p. 188-195, <https://doi.org/10.2307/3671728>.
- Marsh, P.C., and Langhorst, D.R., 1988, Feeding and fate of wild larval razorback sucker: *Environmental Biology of Fishes*, v. 21, no. 59-67, <https://doi.org/10.1007/BF02984443>.
- Mensingher, M.A., Hawkes, J.P., Goulette, G.S., Mortelliti, A., Blomberg, E.J., and Zydlewski, J.D., 2023, Dams facilitate predation during Atlantic salmon (*Salmo salar*) smolt migration: *Canadian Journal of Fisheries and Aquatic Sciences*, v. 81, no. 1, p. 38-51, <https://doi.org/10.1139/cjfas-2023-0175>.
- Pennock, C.A., Healy, B.D., Bogaard, M.R., McKinstry, M.C., Gido, K.B., Cathcart, C.N., and Hines, B., 2024, Translocation in a fragmented river provides demographic benefits for imperiled fishes: *Ecosphere*, v. 15, no. 5, article e4874, <https://doi.org/10.1002/ecs2.4874>.
- Pennock, C.A., McKinstry, M.C., and Gido, K.B., 2020, Razorback sucker movement strategies across a river–reservoir habitat complex: *Transactions of the American Fisheries Society*, v. 149, no. 5, p. 620-634, <https://doi.org/10.1002/tafs.10262>.
- Rogowski, D.L., Osterhoudt, R.J., Mohn, H.E., and Boyer, J.K., 2018, Humpback chub (*Gila cypha*) range expansion in the western Grand Canyon: *Western North American Naturalist*, v. 78, no. 1, article 4, <https://scholarsarchive.byu.edu/wnan/vol78/iss1/4>.
- Runge, M.C., Bean, E., Smith, D.R., and Kokos, S., 2011a, Non-native fish control below Glen Canyon Dam—Report from a structured decision-making project: U.S. Geological Survey Open-File Report 2011-1012, 74 p., <https://pubs.usgs.gov/of/2011/1012/>.
- Runge, M.C., Converse, S.J., and Lyons, J.E., 2011b, Which uncertainty? Using expert elicitation and expected value of information to design an adaptive program: *Biological Conservation*, v. 144, no. 4, p. 1214-1223, <https://doi.org/10.1016/j.biocon.2010.12.020>.
- Runge, M.C., Yackulic, C.B., Bair, L.S., Kennedy, T.A., Valdez, R.A., Ellsworth, C., Kershner, J.L., Rogers, R.S., Trammell, M., and Young, K.L., 2018, Brown trout in the Lees Ferry reach of the Colorado River—Evaluation of causal hypotheses and potential interventions: U.S. Geological Survey Open-File Report 2018-1069, 83 p., <https://doi.org/10.3133/ofr20181069>.

- Schooley, J.D., and Marsh, P.C., 2007, Stocking of endangered razorback suckers in the Lower Colorado River Basin over three decades—1974-2004: *North American Journal of Fisheries Management*, v. 27, no. 1, p. 43-51, <https://doi.org/10.1577/M05-088.1>.
- Schultz, A.A., Afentoulis, V.B., Yip, C.J., and Johnson, M.N., 2017, Efficacy of an acoustic tag with predation detection technology: *North American Journal of Fisheries Management*, v. 37, no. 3, p. 574-581, <https://doi.org/10.1080/02755947.2017.1290720>.
- Spurgeon, J.J., Paukert, C.P., Healy, B.D., Trammel, M., Speas, D.W., and Omana Smith, E., 2015, Translocation of humpback chub into tributary streams of the Colorado River—Implications for conservation of large-river fishes: *Transactions of the American Fisheries Society*, v. 144, no. 3, p. 502-514, <https://doi.org/10.1080/00028487.2015.1007165>.
- U.S. Department of the Interior, 2016, Glen Canyon Dam Long-Term Experimental and Management Plan final Environmental Impact Statement (LTEMP EIS): U.S. Department of the Interior, Bureau of Reclamation, Upper Colorado Region, National Park Service, Intermountain Region, online, <http://ltempeis.anl.gov/documents/final-eis/>.
- U.S. Department of the Interior, 2024, Glen Canyon Dam Long-Term Experimental and Management Plan, Draft Supplemental Environmental Impact Statement: U.S. Department of the Interior, Bureau of Reclamation, Upper Colorado River Basins, Interior Region 7, 374 p., https://www.usbr.gov/uc/DocLibrary/EnvironmentalImpactStatements/GlenCanyonDamLong-TermExperimentalManagementPlan/20240207-Draft-GCDLTEMP-SEIS_508.pdf.
- Van Haverbeke, D.R., Stone, D.M., Dodrill, M.J., Young, K.L., and Pillow, M.J., 2017, Population expansion of humpback chub in western Grand Canyon and hypothesized mechanisms: *The Southwestern Naturalist*, v. 62, no. 4, p. 285-292, <https://doi.org/10.1894/0038-4909-62.4.285>.
- Yackulic, C.B., Korman, J., Yard, M.D., and Dzul, M.C., 2018, Inferring species interactions through joint mark-recapture analysis: *Ecology*, v. 99, no. 4, p. 812-821, <https://doi.org/10.1002/ecy.2166>.
- Yackulic, C.B., Van Haverbeke, D.R., Dzul, M.C., Bair, L.S., and Young, K.L., 2021, Assessing the population impacts and cost-effectiveness of a conservation translocation: *Journal of Applied Ecology*, v. 58, no. 8, p. 1602-1612, <https://doi.org/10.1111/1365-2664.13908>.
- Yackulic, C.B., Yard, M.D., Korman, J., and Van Haverbeke, D.R., 2014, A quantitative life history of endangered humpback chub that spawn in the Little Colorado River—Variation in movement, growth, and survival: *Ecology and Evolution*, v. 4, no. 7, p. 1006-1018, <https://doi.org/10.1002/ece3.990>.

Total TWP Budget by Year

| Total (FY25-27 TWP) | Salaries | Travel & Training | Operating Expenses | Logistics Expenses | Cooperative Agreements | To other USGS Centers | AMP Special Burden Rate* | Total | USGS Contributing Funds** |
|------------------------|---------------------|-------------------|--------------------|--------------------|------------------------|-----------------------|--------------------------|---------------------|---------------------------|
| | | | | | | | | | Est Full Rate = 56% |
| Total (FY25) | \$5,095,360 | \$96,715 | \$535,961 | \$931,414 | \$1,175,910 | \$423,420 | \$1,491,233 | \$9,750,013 | \$694,928 |
| Total (FY26) | \$5,308,839 | \$98,215 | \$488,240 | \$1,090,860 | \$1,129,736 | \$434,160 | \$1,612,763 | \$10,162,813 | \$740,313 |
| Total (FY27) | \$5,512,272 | \$91,215 | \$483,608 | \$1,015,378 | \$1,103,915 | \$382,600 | \$1,695,096 | \$10,284,083 | \$764,502 |
| Total (FY25-27) | \$15,916,471 | \$286,145 | \$1,507,809 | \$3,037,652 | \$3,409,561 | \$1,240,180 | \$4,799,092 | \$30,196,910 | \$2,199,743 |

Appendix 1. Water Quality Monitoring of Lake Powell and Glen Canyon Dam Releases (not GCDAMP funded)

Historical Data Analysis, Water Quality Monitoring, and Drivers of Nutrient Availability in Lake Powell and Glen Canyon Dam Releases

Background

An Interagency Agreement (R23PG00022) between the U.S Department of the Interior, Bureau of Reclamation and U.S. Geological Survey, Southwest Biological Service Center

The Lake Powell water quality monitoring program has been in existence since 1965 and represents a unique and valuable long-term record of water temperatures, dissolved oxygen concentrations, major ions and, more recently (beginning in 1990), nutrients and biological constituents. The monitoring program collects data on water-quality conditions at the forebay of the reservoir (2.4 km from the dam) on a monthly basis and at 25-30 sites reservoir-wide quarterly. Water temperature, specific conductance, dissolved oxygen, pH, redox potential, and turbidity are measured throughout the water column at each site with samples for major ionic constituents, nutrients, dissolved organic carbon, chlorophyll, phytoplankton, and zooplankton being collected at selected sites. Physical and chemical information from this program was published as USGS Data Series Report DS-471 (last revised February 20151) and biological data are contained in USGS Data Series Report DS-959 (last revised October 20152). All information from this program is stored in the Water Quality Database (WQDB) in Microsoft Access, and a subset of the data (through 2013) is available for download with the DS-471 Data Series Report¹. The Lake Powell long term monitoring program has undergone several assessments including a 1996 evaluation by GCMRC that highlighted the role of dam operations in controlling the quality of water in dam releases, a 2001 protocol evaluation panel that recommended a gradual shift in emphasis from Lake Powell to downstream, and a 2017 protocol evaluation panel that recommended improvements in several areas including: 1.) data management and metadata development, 2.) vertical sampling resolution at key sites, and 3.) sensor calibration and QA/QC practices³.

Currently, the Lake Powell water quality monitoring program conducts downstream temperature, dissolved oxygen, and conductivity monitoring directly below the dam, but water quality monitoring at Lees Ferry has been moved to the GCMRC physical science program, and aquatic food base monitoring work has developed and conducts additional downstream monitoring of dissolved oxygen and other water quality parameters like turbidity and conductivity.

Additionally, Reclamation has employed and maintained the CE-QUAL-W2 reservoir model to provide projections of release temperature, dissolved oxygen, and other parameters⁴. Recent low water levels have caused some problems with the historic CE-QUAL-W2 grid, and a

postdoctoral research position was funded in 2022 for additional CE-QUAL-W2 model development that revises grid structure and incorporates new bathymetry and meteorology data. This effort is funded by a separate statement of work (SOW), wherein FY22 funding was appended to the R18PG00108 agreement in the amount of \$57,067 and FY23 funding is planned in the amount of \$92,061. This model development will support investigations of interest to the Glen Canyon Dam Adaptive Management Program (GCDAMP) stakeholders in better understanding and describing how the water quality in dam release water will change with declining water levels, particularly with respect to temperature and dissolved oxygen concentrations.

Since the last agreement was written, some additions have also been made to regular monitoring activities. Reclamation and GCMRC have collaborated to deploy a thermistor string in the reservoir that hangs from the buoy line near the dam. The loggers were first deployed in March of 2018, with temperature loggers at 1m, 5m, 10m, 15m, 20m, 25m, 30m, 35m, 40m, 50m, 55m, 60m and every 10 m until the bottom of the reservoir after that. Conductivity loggers are placed at the penstock depth and at the bottom (periodically readjusted depending on water levels). Units are set to log at least every half hour, providing data describing lake stratification at the sub-daily time scale. A similar thermistor string was placed in the same location in August of 2011. Data from this deployment are available through mid-December of 2014 at which time the thermistor string was lost. USGS has also begun collecting additional nutrient and major ion samples during their monthly and quarterly trips to the reservoir forebay. These samples are collected from 5-10 meters above and below the penstock, with specific depths decided based on existing features of water quality stratification such as chlorophyll maxima or dissolved oxygen minima. USGS also purchased a new Seabird Electronics 19V2 plus CTD profiler in the spring of 2021 (same model as the

Seabird purchased in 2010). Finally, the USGS transferred most of the Lake Powell Water Quality Monitoring data to a Microsoft SQL Server database and designed a custom application that allows for streamlined data import and export. The database is linked to a file transfer protocol (FTP) site for easy sharing of data with Reclamation who uses a subset of the water quality data to run the CE-QUAL-W2 model.

Products that have been developed since the last agreement include the 2017 Protocol Evaluation Panel report³, annual reports written as appendices to the GCMRC work plan annual report, a Lake Powell website (<https://www.usgs.gov/centers/southwest-biologicalscience-center/science/lake-powell-research>), an internally-shared data visualization for water quality directly below GCD that is based on near-real-time data transmitted to USGS GCMRC's Amazon Web Services platform⁵, two peer reviewed papers describing salinity dynamics⁶ and greenhouse gas dynamics⁷ on Lake Powell, and a third peer reviewed paper describing the long term water quality and plankton datasets that is currently in revision⁸. The following updated scope of work aims to outline the role of GCMRC in the ongoing Lake Powell monitoring program.

Description of Products or Services and Milestones

The objectives of this monitoring program have changed since 1965, reflecting changes in scientific interest as the reservoir filled, responsibilities of Reclamation for maintaining salinity levels in the Colorado River, and the monitoring status of Upper Colorado River basin reservoirs.

Objectives have also been responsive to more recent environmental concerns related to the Grand Canyon Protection Act, the establishment of the GCDAMP, and the GCD Long-Term Experimental and Management Plan Environmental Impact Statement and subsequent Record of Decision.

Objectives of this long-term monitoring program include:

1. Determination of water-quality status and trends in Lake Powell and GCD releases.
2. Linking the historical record of Lake Powell water quality to various climatological and hydrological conditions.
3. Documentation of the effects of the structure and operation of GCD on the quality of water in Lake Powell and GCD releases.
4. Integration with GCDAMP information needs and downstream monitoring programs.
5. Documentation of the density structure and associated nutrient distribution in the water column at the GCD forebay and other locations in the reservoir to determine the quality of water available for release from GCD.
6. Assessment of the distribution and patterns of major ionic constituents in Lake Powell and GCD releases.
7. Assessment of the distribution and patterns of nutrient constituents in Lake Powell and GCD releases.
8. Assessment of the structure, status, and trends of the plankton community and its effect on primary and secondary production in Lake Powell.

Task 1: Details of Ongoing Monitoring Program (To Be Continued Under This Agreement)

The ongoing Lake Powell water-quality monitoring program consists of monthly surveys of the reservoir forebay and tailwater (conducted by GCMRC) and quarterly surveys of the entire reservoir, including the Colorado, San Juan, and Escalante arms of the reservoir to the inflow areas (conducted by Reclamation). The GCD forebay station is located approximately 2.4 km upstream from GCD. Two tailwater sites are located immediately downstream from the dam and at Lees Ferry, approximately 25 km downstream. Depending on reservoir elevation, 21-37 established sites, including the forebay and tailwater stations, are sampled for the quarterly surveys.

At each site, initial surface observations (for example, bottom depth, Secchi depth, weather observations) are recorded, after which a depth profile of temperature, specific conductance, dissolved oxygen, pH, redox potential, turbidity, and chlorophyll fluorescence is collected, using the Seabird SBE19plusV2 instrument. These data are downloaded immediately after collection and viewed in the field to determine stratification patterns. Based on stratification patterns, chemical samples for major ionic constituents and nutrient concentrations are collected in the major strata at selected sites. Dissolved organic carbon samples are collected at the forebay tailwater, and tributary inflow sites. Biological samples for chlorophyll concentration, phytoplankton, and zooplankton are also collected at selected sites. Samples are filtered and preserved in the field for subsequent laboratory analysis.

Continuous water quality monitoring is conducted within the GCD draft tubes and directly downstream of GCD via two YSI sondes. The sonde below GCD is setup to transmit near-realtime data to USGS GCMRC's Amazon Web Services platform which is currently shared internally and with BOR via a series of Tableau data visualizations that USGS aims to make publicly available within the next year⁵. Continuous temperature and conductivity monitoring is also conducted directly above GCD in the reservoir forebay via a thermistor string equipped with Hobo temperature and conductivity loggers.

Analysis for major ionic constituent, nutrient, and chlorophyll *concentrations* are performed by Reclamation's Lower Colorado Regional Laboratory in Boulder City, NV. Phytoplankton and zooplankton samples are analyzed under contract by BSA Environmental, Inc. Data processing of the Seabird profile data is performed in the office shortly after the field survey. All field data as well as major ion, nutrient, phytoplankton, and zooplankton data streams are entered into a new Microsoft SQL Server database for statistical and graphical analysis and long-term storage. This database is stored on-premises on a Southwest Biological Science Center (SBSC) Windows Server, with file-based data sets stored on a JBOD (just of bunch of disks) disk array that allows for simple allocation of data to segmented sections of disks. Databases and file-based data that are stored on premises are backup weekly and monthly, with off-site backups being hosted in Amazon Web Services Glacier storage. This will be structured as a dynamic data release that can be regularly updated with new data as it becomes available and undergoes QA/QC and normalization procedures⁹. Details of the monitoring program, a description of the SQL Server database, and physicochemical and biological data from 1964-2021 are contained in a data paper that is currently in revision with the journal *Limnology and Oceanography Letters*⁸.

Task 2: Historical Data Analysis – Towards Improved Predictive Capacity for Reservoir Water Quality Dynamics

Given the large volume of historical nutrient data available for Lake Powell, the river outlet works, and Lees Ferry, and given that there has been relatively little analysis of limnological and plankton data, GCMRC will work to assess and describe spatial and temporal trends in limnological variables and plankton communities.

Preliminary analysis suggests temporal trends in surface water temperature, Secchi depth, and phytoplankton biovolume as well as shifts in plankton communities that include the introduction of *Dreissena bugensis*⁸ (Quagga mussels).

Long-term plankton datasets are particularly limited worldwide due to the technical expertise required for species identification, which increases the cost and time necessary for analysis. Even more rare are the datasets that include both phytoplankton and zooplankton data collected concurrently. Plankton data could become extremely valuable in light of the Quagga mussel invasion. For example, zebra mussels have been implicated in promoting *Microcystis* in noneutrophic waters elsewhere¹⁰. Trends in cyanobacteria composition and abundance are particularly management relevant given that some cyanobacteria can produce harmful toxins (as is the case for *Microcystis*). With warming surface summer water temperatures in Lake Powell⁸ and warmer temperatures downstream of GCD due to lower water levels, it is important to assess changes in plankton assemblages (and their human health and ecological consequences).

Historical limnological data will also be used to improve our predictive power and our mechanistic understanding of low dissolved oxygen events in Lake Powell. This work is largely described in a separate SOW, funding for which is appended to this agreement in the amount of \$92,061 in calendar year 2023. In addition to the CE-QUAL-W2 model development work that falls under that SOW, USGS will use historical profile data to better understand the chances of downstream low dissolved oxygen events based on reservoir elevation, spring inflow, and monsoon events.

Task 3: Characterizing Nutrient Dynamics During Experimental Flows

The potential for experimental flows to modify downstream nutrient regimes will be examined by conducting targeted nutrient sampling before, during, and after experimental flows. Previous sampling efforts have documented the capacity for changes in flow at GCD outlets to affect the chemistry of water below the dam. For example, work by Hueftle and Stevens showed that the 1996 spring high flow event diminished bottom water hypoxia in Lake Powell as far as 100 km uplake while also resulting in high salinity, high oxygen concentrations, and damped DO and pH fluctuations in the dam tailwater¹¹. Nutrient data collected 4 days before and 2 days after the high flow event showed drops in phosphorus concentration at both the penstock and river outlet works, although the magnitude of this drop is difficult to determine given the detection limit of the analyses used at that time. Monitoring during the 2008 high flow experiment also showed elevated dissolved oxygen concentrations downstream of the dam (at maximum 120% of saturation), but relatively minimal effects on the structure of the water column upstream of the dam¹². Still, this study only considered water temperature, specific conductance, and dissolved oxygen, not nutrients. Sampling as part of the last Interagency Agreement captured reservoir water quality profiles and water chemistry at Lees Ferry before, during, and after the 2018 High Flow Experiment. The sampling showed elevated concentrations of phosphorus at Lees Ferry that corresponded to higher phosphorus concentrations at the depth of the bypass intake.

With the current low reservoir water levels, it is possible that new experimental flows (e.g., pertaining to the control of smallmouth bass) will be implemented. Targeted profiling at Wahweap and grab sampling at Lees Ferry will be conducted prior to experimental flow, during, and after experimental flows to better quantify the effects of flow regime on reservoir and outlet chemistry. Thermistor string data will also be examined for changes in physical stratification as well as nutrient dynamics.

Task 4: Improving Access to Historical Dataset

Three key data streams are currently missing from the Science Base data release⁹ and data paper⁸ described above: 1) the thermistor string, 2) filter-based measures of chlorophyll *a* from 1 m depth, and 3) dissolved organic carbon (DOC) data from the reservoir forebay, Lees Ferry, the GCD draft tubes, and the reservoir inlets. The chlorophyll *a* data are described in a previous USGS data series report² which contains publicly available historical data through 2009. All thermistor, DOC, and chlorophyll *a* data is saved on the SBSC local network drive. USGS aims to incorporate these data streams in both the Microsoft SQL Server database and in a future data release. USGS also plans to create data visualization products that can be shared with the public via the USGS Lake Powell website.

References

1. Vernieu, W.S., 2015, Historical physical and chemical data for water in Lake Powell and from Glen Canyon Dam releases, Utah-Arizona, 1964-2013 (ver. 3.0, February 2015): U.S. Geological Survey Data Series 471, 23 p., <https://doi.org/10.3133/ds471>.
2. Vernieu, W.S., 2015, Biological data for water in Lake Powell and from Glen Canyon Dam releases, Utah and Arizona, 1990-2009: U.S. Geological Survey Data Series 959, 12 p., <https://doi.org/10.3133/ds959>.
3. Hamilton, S., Holdren, C., Stets, E., Strock, K., and Tietjen, T., 2018, Grand Canyon Monitoring and Research Center water-quality program review—final report, June 4, 2018: Flagstaff, Ariz., submitted to U.S. Geological Survey, Southwest Biological Science Center, Grand Canyon Monitoring and Research Center, 12 p.
4. Williams, N.T., 2007, Modeling dissolved oxygen in Lake Powell using CE-QUAL-W2 Brigham Young University, M.S. Science thesis, 120 p., https://deercreek.groups.et.byu.net/Papers/Models/W2-MODEL/DO_ModelingPowell_WilliamsThesis.pdf
5. Gushue, T., and Thomas, J.E., 2021, Daily water quality data at Glen Canyon Dam—Online data visualization: U.S. Geological Survey, Southwest Biological Science Center, Grand Canyon Monitoring and Research Center, last modified September 26, 2022, <https://tableau.usgs.gov/t/InternalGuestAccess/views/WaterQualityatGlenCanyonDam/DailyWaterQualityDataatGlenCanyonDam?%3Aembed=y&%3AisGuestRedirectFromVizportal=y#1>.

6. Deemer, B.R., Stets, E.G., and Yackulic, C.B., 2020, Calcite precipitation in Lake Powell reduces alkalinity and total salt loading to the Lower Colorado River Basin: *Limnology and Oceanography*, v. 65, p. 1439-1455, <https://doi.org/10.1002/lno.11399>.
7. Waldo, S., Deemer, B.R., Bair, L.S., and Beaulieu, J.J., 2021, Greenhouse gas emissions from an arid-zone reservoir and their environmental policy significance—Results from existing global models and an exploratory dataset: *Environmental Science and Policy*, v. 120, p. 53-62, <https://doi.org/10.1016/j.envsci.2021.02.006>.
8. Deemer, B.R., Andrews, C.M., Strock, K.E., Voichick, N., Hensleigh, J., Beaver, J.R., and Radtke, R., 2023, Over half a century record of limnology data from Lake Powell, desert southwest United States—From reservoir filling to present day (1964–2021): *Limnology and Oceanography Letters*, v. 8, no. 4, p. 580-594, <https://doi.org/10.1002/lol2.10310>.
9. Andrews, C.M., and Deemer, B.R., 2023, Limnology data from Lake Powell, desert southwest USA (ver. 2.0, Sept. 2023): U.S. Geological Survey data release, <https://doi.org/10.5066/P9ZIKVYW>.
10. Vanderploeg, H.A., Liebig, J.R., Carmichael, W.W., Agy, M.A., Johengen, T.H., Fahnenstiel, G.L., and Nalepa, T.F., 2001, Zebra mussel (*Dreissena polymorpha*) selective filtration promoted toxic *Microcystis* blooms in Saginaw Bay (Lake Huron) and Lake Erie: *Canadian Journal of Fisheries and Aquatic Sciences*, v. 58, no. 6, p. 1208-1221, <https://doi.org/10.1139/f01-066>.
11. Hueftle, S.J., and Stevens, L.E., 2001, Experimental flood effects on the limnology of Lake Powell reservoir, southwestern USA: *Ecological Applications*, v. 11, no. 3, p. 644-656, <https://doi.org/10.2307/3061107>.
12. Vernieu, W.S., 2010, Effects of the 2008 high-flow experiment on water quality in Lake Powell and Glen Canyon Dam releases, Utah-Arizona: U.S. Geological Survey Open-File Report 2010-1159, 25 p., <https://pubs.usgs.gov/of/2010/1159/>.

Agreement Contract

Authority

This Inter/Intra-Agency Agreement (IAA) is entered into between the Department of the Interior, Bureau of Reclamation, hereinafter referred to as “Reclamation,” and Department of the Interior, U.S. Geological Survey, Southwest Biological Service Center, hereinafter referred to as the “USGS” pursuant to the Statutory Authority: Economy Act, 31 U.S.C. 1535.

Period of Performance and Funds Availability

This Agreement becomes effective on the date shown in Block 1a of the Inter/Intra-Agency Agreement (IAA) Form and shall remain in effect until the date shown in Block 1b of the Inter/Intra-Agency Agreement (IAA) Form.

The period of performance for this Agreement may only be modified through written modification of the IAA by a Reclamation Contracting Officer (CO). The total estimated amount of Reclamation funding for this IAA is \$1,686,216.63, of which the initial amount of federal funds available is limited to \$383,068.24 as indicated by Block 16 of the Inter/Intra-Agency Agreement (IAA) Form. Subject to the availability of Congressional appropriations, and as necessary, subsequent funds will be made available for payment through written modifications to this IAA by a Reclamation CO. No legal liability on the part of Reclamation for any payment may arise until funds are made available, through written modification of the IAA by a Reclamation CO.

Description of Products or Services and Milestones

The Lake Powell water quality monitoring program has been in existence since 1965 and represents a unique and valuable long-term record of water temperatures, dissolved oxygen concentrations, major ions and, more recently nutrients and biological constituents. The monitoring program collects data on water-quality conditions at the forebay of the reservoir on a monthly basis and at 25-30 sites reservoir-wide quarterly. Water temperature, specific conductance, dissolved oxygen, pH, redox potential, and turbidity are measured throughout the water column at each site with samples for major ionic constituents, nutrients, dissolved organic carbon, chlorophyll, phytoplankton, and zooplankton being collected at selected sites. Physical and chemical information from this program is published in Reports and stored in the Water Quality Database.

The objectives of this monitoring program have changed since 1965, reflecting changes in scientific interest as the reservoir filled, responsibilities of Reclamation for maintaining salinity levels in the Colorado River, and the monitoring status of Upper Colorado River basin reservoirs. Objectives have also been responsive to more recent environmental concerns related to the Grand Canyon Protection Act, the establishment of the GCDAMP, and the GCD Long-Term Experimental and Management Plan Environmental Impact Statement and subsequent Record of Decision.

Objectives of this long-term monitoring program include:

1. Determination of water-quality status and trends in Lake Powell and GCD releases.
2. Linking the historical record of Lake Powell water quality to various climatological and hydrological conditions.
3. Documentation of the effects of the structure and operation of GCD on the quality of water in Lake Powell and GCD releases.
4. Integration with GCDAMP information needs and downstream monitoring programs.

5. Documentation of the density structure and associated nutrient distribution in the water column at the GCD forebay and other locations in the reservoir to determine the quality of water available for release from GCD.
6. Assessment of the distribution and patterns of major ionic constituents in Lake Powell and GCD releases.
7. Assessment of the distribution and patterns of nutrient constituents in Lake Powell and GCD releases.
8. Assessment of the structure, status, and trends of the plankton community and its effect on primary and secondary production in Lake Powell.

Three Primary Tasks comprise the work of this agreement and are fully outlined in the attached statement of work.

- Task 1: Details of Ongoing Monitoring Program (to be continued under this agreement)
- Task 2: Historical Data Analysis- Towards Improved Predictive Capacity for Reservoir Water

Quality Dynamics

- Task 3: Characterizing Nutrient Dynamics During Experimental Flows Task 4- Improving Access to Historical Dataset
- Task 4: Improving Access to Historical Dataset

Roles and Responsibilities

A. Requesting Agency Responsibilities

Reclamation will serve as the program manager for this study on behalf of the Department of the Interior and will be responsible for the following:

Reclamation will:

1. Provide funds for field support for quarterly sampling events and laboratory analysis of samples collected.
2. Provide technical support to assist in planning.
3. Collaborate with GCMRC to select and determine the type and amount of water quality samples required.
4. Oversee monitoring activities and associated needs.
5. Provide logistics, supplies and crew members and funding for additional agency personnel as requested by the servicing agency.

6. Review results.
7. Fund route monitoring lab and sample analytical costs.
8. Publish annual data as required.

B. Servicing Agency Responsibilities

USGS shall serve as the program manager for this study on behalf of the Department of the Interior and will be responsible for the following items.

USGS shall:

1. Provide analysis of historic data.
2. Store and serve water quality data.
3. Design and implement targeted nutrient sampling and profiling,
4. Conduct monthly sampling events.
5. Provide crew members for quarterly sampling events as needed.
6. Collaborated with Reclamation to select and determine the type and amount of water quality samples required.
7. Provide supplies, logistics and samples for forebay (monthly) monitoring activities and additional targeted nutrient sampling.
8. Provide crew members for quarterly lake wide water quality monitoring sampling as needed.
9. Send all water quality samples to the lab for analysis.
10. Fund all additional lab sample costs not covered by routine monitoring costs.
11. Provide an annual report which provides insights into how climatological, meteorological and hydrodynamic processes (including the operation of GCD) affect the reservoir and the water quality of releases from GCD 12. Store and manage water quality data in a database.
12. Serve water quality data online.

Budget

- A. Budget Estimate.** The following is the estimated budget for this IAA. As interagency agreements are cost-reimbursable, the budget provided is for estimation purposes only. Final costs incurred under the budget categories listed may be either higher or lower than the estimated costs. Final determination of the allowability, allocability, or reasonableness of costs incurred under this agreement is the responsibility of the Contracting Officer. Servicing Agencies are encouraged to direct any questions regarding allowability, allocability or reasonableness of costs to the Requesting Agency's Contracting Officer for review prior to incurrence of the costs in question.

| FY 23 | FY 24 | FY 25 | FY 26 | FY 27 | Total |
|--------------|--------------|--------------|--------------|--------------|----------------|
| \$383,068.24 | \$305,992.59 | \$319,941.84 | \$329,418.29 | \$347,795.67 | \$1,686,216.63 |

B. Funding Schedule. Funds in the amount of \$383,068.24 have been pre-validated on PR 40602908 and will be obligated to the interagency agreement at time of award.

| FY 23 | FY 24 | FY 25 | FY 26 | FY 27 | Total |
|--------------|--------------|--------------|--------------|--------------|----------------|
| \$383,068.24 | \$305,992.59 | \$319,941.84 | \$329,418.29 | \$347,795.67 | \$1,686,216.63 |

C. Pre-Award Incurrence of Costs. The Servicing Agency shall be entitled to reimbursement for costs incurred on or after 1/1/2023 which if had been incurred after this IAA was entered into, would have been allowable, allocable, and reasonable under the terms and conditions of this IAA.

Billing and Payment

The Servicing Agency will be reimbursed by Reclamation using the Intergovernmental Payment and Collection (IPAC) method. The Servicing Agency shall bill via IPAC to Reclamation’s Agency Location Code (ALC) identified in Block 7a. Bill on a QUARTERLY basis for costs incurred as authorized by the agreement. Include the IAA number R23PG00022, account numbers from Block 7a of the first page of this agreement, the billing time period, and the Servicing agency’s point of contact and the telephone number for billing information. Billing may not begin until after the date of the award.

Description block of IPAC billing must include the following:

1. Reclamation accounting information;
2. IAA number;
3. billing period;
4. accountable contact; and
5. telephone number.

The Servicing Agency shall submit the appropriate IPAC payment support documentation to Reclamation’s Financial POC and to the IATR as identified on the cover-page of this IAA.

Servicing Agency Funding Information

The Servicing Agency shall complete the table below.

| | |
|--|--------------|
| Basic appropriation symbol (Treasury account symbol) | 14X0804 |
| Fund citation (line of accounting) | |
| Business event type code | COLL |
| Agency location code (8-digit) for IPAC | 14-08-0001 |
| DUNS/BPN number (Business Partner Network of BPN #) | NJQMLNG5L8A5 |

Property

Title to all property acquired with funds provided under this interagency agreement shall be vested in Reclamation and is subject to the condition that the property shall be used for authorized purposes of the project. Should the Servicing Agency wish to take unrestricted title to any property acquired or to change the use of the facilities or real property so acquired, such transactions shall be governed by Federal Property Management Regulations. All procurements of property using interagency agreements funds shall be approved in writing by Reclamation prior to the transaction being initiated. The Servicing Agency shall not make disposition of any property except as directed in writing by Reclamation.

Modifications

Authority to modify an IAA on behalf of Reclamation is expressly limited to the Contracting Officer. Authority of the IATR is subject to limitations that do not include the authority to modify an IAA. This IAA may be modified through bilateral agreement between the parties. Any modification made to this IAA shall be confirmed in writing prior to performance of the change. The Servicing Agency assumes all risks, liabilities, and consequences of performing additional work outside the specified scope of work without prior written approval from the Contracting Officer. Any modifications to the terms and conditions shall be made in writing and signed by both the Servicing Agency and the Requesting Agency. The Servicing Agency will notify Reclamation as soon as practicable if it appears that additional funding may be needed to perform the work. If additional funding is needed, Reclamation may provide it or agree to adjust the work to fit within available funding, as appropriate.

Termination

This IAA may be terminated upon thirty (30) calendar days of written notice by either party. If this agreement is cancelled, any implementing contract/order may also be cancelled. If the IAA is terminated, the agencies shall agree on the terms of the termination, including costs attributable to each party and the disposition of awarded and pending actions.

If the Servicing Agency incurs costs due to the Requesting Agency's failure to give the requisite notice of its intent to terminate the IAA, the Requesting Agency shall pay any actual costs incurred by the Servicing Agency as a result of the delay in notification, provided such costs are directly attributable to the failure to give notice.

Interpretation of the IAA

If the Servicing Agency and Requesting Agency are unable to agree about a material aspect of the terms and conditions of this IAA, the parties agree to engage in an effort to reach mutual agreement in the proper interpretation of this IAA, including amendment of this IAA, as necessary, by escalating the dispute within their respective organizations. If a dispute related to funding remains unresolved for more than sixty (60) calendar days after the parties have engaged in an escalation of the dispute, the parties agree to refer the matter to their respective Agency Chief Financial Officers with a recommendation that the parties submit the dispute to the CFO Council Intragovernmental Dispute Resolution Committee for review in accordance with the Treasury Financial Manual, Volume 1, Appendix 10, Chapter 4700, entitled Intergovernmental Transaction Guide, or subsequent guidance.

Appendix 2. Budget Allocation by Year

FY 2025

| Fiscal Year 2025 | | | | | | | | | |
|------------------|---|------------------|-------------------|--------------------|--------------------|------------------------|-----------------------|------------------|--------------------|
| Project | Project Description | Salaries | Travel & Training | Operating Expenses | Logistics Expenses | Cooperative Agreements | To other USGS Centers | Burden | Total |
| | | | | | | | | 21.86% | |
| A | Streamflow, Water Quality, and Sediment Transport and Budgeting in the Colorado River Ecosystem | | | | | | | | |
| A.1 | Stream gaging and hydrologic analyses | \$181,311 | \$6,000 | \$6,000 | \$27,000 | \$0 | \$174,800 | \$48,166 | \$443,277 |
| A.2 | Continuous water quality parameters | \$114,492 | \$1,000 | \$12,000 | \$26,000 | \$0 | \$29,260 | \$33,558 | \$216,310 |
| A.3 | Sediment transport and budgeting | \$333,120 | \$6,000 | \$38,000 | \$26,200 | \$0 | \$154,360 | \$88,178 | \$645,857 |
| | Total A | \$628,922 | \$13,000 | \$56,000 | \$79,200 | \$0 | \$358,420 | \$169,902 | \$1,305,444 |
| B | Sandbar and Sediment Storage Monitoring and Research | | | | | | | | |
| B.1 | Sandbar and campsite monitoring with topographic surveys and remote cameras | \$224,543 | \$2,000 | \$4,000 | \$27,162 | \$23,500 | \$0 | \$57,047 | \$338,252 |
| B.2 | Bathymetric and topographic mapping for monitoring sediment storage and riverbed dynamics | \$348,531 | \$2,000 | \$2,000 | \$0 | \$18,800 | \$0 | \$77,638 | \$448,969 |
| B.3 | Control network and survey support | \$0 | \$0 | \$0 | \$0 | \$0 | \$0 | \$0 | \$0 |
| B.4 | Streamflow, sediment, and sandbar modeling | \$0 | \$0 | \$0 | \$0 | \$0 | \$0 | \$0 | \$0 |
| | Total B | \$573,074 | \$4,000 | \$6,000 | \$27,162 | \$42,300 | \$0 | \$134,685 | \$787,221 |
| C | Riparian Vegetation Monitoring and Research | | | | | | | | |
| C.1 | Ground-based riparian vegetation monitoring | \$124,827 | \$3,615 | \$3,500 | \$81,775 | \$12,542 | \$0 | \$47,101 | \$273,360 |
| C.2 | Determining hydrological tolerances and management tools for plant species of interest | \$13,445 | \$0 | \$0 | \$0 | \$45,452 | \$0 | \$4,303 | \$63,201 |
| C.3 | Predictive models and synthesis | \$9,845 | \$0 | \$0 | \$0 | \$20,007 | \$0 | \$2,753 | \$32,604 |
| C.4 | Biogeomorphic Linkages between streamflow, sediment transport, and vegetation composition | \$105,222 | \$0 | \$0 | \$0 | \$0 | \$0 | \$23,005 | \$128,226 |
| C.5 | Vegetation management decision support | \$0 | \$0 | \$0 | \$0 | \$0 | \$0 | \$0 | \$0 |
| | Total C | \$253,339 | \$3,615 | \$3,500 | \$81,775 | \$78,001 | \$0 | \$77,162 | \$497,392 |
| D | Effects of Dam Operations and Vegetation Management for Archaeological Sites | | | | | | | | |
| D.1 | Monitoring the effects of dam operations on archaeological sites | \$166,947 | \$10,000 | \$20,000 | \$34,015 | \$0 | \$0 | \$50,495 | \$281,457 |
| D.2 | Monitoring landscape-scale ecosystem change with repeat photography | \$70,048 | \$3,000 | \$2,500 | \$0 | \$0 | \$0 | \$16,517 | \$92,065 |
| D.3 | Evaluating effects of LTEMP non-flow actions and other experimental vegetation management on archaeological sites | \$19,605 | \$9,000 | \$6,000 | \$0 | \$0 | \$0 | \$7,566 | \$42,170 |
| D.4 | Pilot study to evaluate potential to extract cultural and ecological information from Colorado River deposits using eDNA and pollen | \$0 | \$0 | \$0 | \$0 | \$0 | \$0 | \$0 | \$0 |
| D.5 | Monitoring petroglyphs and pictographs with photogrammetry and lidar | \$0 | \$0 | \$0 | \$0 | \$0 | \$0 | \$0 | \$0 |
| | Total D | \$256,599 | \$22,000 | \$28,500 | \$34,015 | \$0 | \$0 | \$74,578 | \$415,692 |

| | | | | | | | | | |
|----------|--|------------------|-----------------|-----------------|------------------|------------------|------------|------------------|--------------------|
| E | Controls on Ecosystem Productivity: Nutrients, Flow, and Temperature | | | | | | | | |
| E.1 | Phosphorus budgeting in the Colorado River | \$19,449 | \$0 | \$3,500 | \$1,225 | \$0 | \$0 | \$5,285 | \$29,459 |
| E.2 | Rates and composition of primary producers in the Colorado River | \$217,868 | \$4,000 | \$24,637 | \$1,750 | \$0 | \$0 | \$54,276 | \$302,531 |
| E.3 | Understanding the energetic basis of the food web in Western Grand Canyon | \$0 | \$0 | \$0 | \$0 | \$0 | \$0 | \$0 | \$0 |
| E.4 | Productivity at higher trophic levels | \$96,625 | \$0 | \$3,000 | \$0 | \$0 | \$0 | \$21,781 | \$121,406 |
| | Total E | \$333,942 | \$4,000 | \$31,137 | \$2,975 | \$0 | \$0 | \$81,342 | \$453,396 |
| F | Aquatic Invertebrate Ecology | | | | | | | | |
| F.1 | Invertebrate and bat monitoring in Marble and Grand Canyons | \$228,969 | \$1,000 | \$17,000 | \$0 | \$0 | \$0 | \$53,995 | \$300,964 |
| F.2 | Aquatic invertebrate monitoring in Glen Canyon | \$170,121 | \$500 | \$1,500 | \$875 | \$0 | \$0 | \$37,822 | \$210,818 |
| F.3 | Aquatic invertebrate monitoring of Grand Canyon tributaries | \$22,756 | \$500 | \$0 | \$0 | \$45,825 | \$0 | \$6,459 | \$75,540 |
| F.4 | Invertebrate and fish diet studies | \$108,428 | \$500 | \$1,500 | \$0 | \$0 | \$0 | \$24,143 | \$134,571 |
| | Total F | \$530,274 | \$2,500 | \$20,000 | \$875 | \$45,825 | \$0 | \$122,419 | \$721,893 |
| G | Humpback Chub Population Dynamics throughout the Colorado River Ecosystem | | | | | | | | |
| G.1 | Humpback chub population modeling | \$145,751 | \$8,000 | \$12,000 | \$0 | \$0 | \$0 | \$36,238 | \$201,989 |
| G.2 | Annual spring/fall HBC abundance estimates in the lower 13.6 km of the LCR | \$4,719 | \$0 | \$20,222 | \$102,192 | \$406,258 | \$0 | \$39,983 | \$573,373 |
| G.3 | Juvenile chub monitoring near the LCR confluence (JCM-East) | \$150,152 | \$2,000 | \$31,156 | \$301,244 | \$0 | \$0 | \$105,938 | \$590,489 |
| G.4 | Remote PIT-tag array monitoring in the LCR | \$22,136 | \$0 | \$5,000 | \$3,000 | \$0 | \$0 | \$6,589 | \$36,724 |
| G.5 | Monitoring humpback chub aggregation relative abundance and distribution | \$3,938 | \$0 | \$12,436 | \$79,366 | \$142,984 | \$0 | \$25,221 | \$263,945 |
| G.6 | Juvenile chub monitoring - Western Grand Canyon (JCM-West) | \$0 | \$0 | \$0 | \$0 | \$0 | \$0 | \$0 | \$0 |
| G.7 | Chute Falls translocations | \$0 | \$0 | \$0 | \$0 | \$0 | \$0 | \$0 | \$0 |
| G.8 | Sampling of springs in the upper LCR | \$0 | \$0 | \$0 | \$0 | \$0 | \$0 | \$0 | \$0 |
| G.9 | Movement in western Grand Canyon from system-wide antenna monitoring | \$10,742 | \$0 | \$12,000 | \$0 | \$0 | \$0 | \$4,972 | \$27,713 |
| | Total G | \$337,437 | \$10,000 | \$92,814 | \$485,802 | \$549,242 | \$0 | \$218,940 | \$1,694,235 |
| H | Salmonid Research and Monitoring | | | | | | | | |
| H.1 | Rainbow trout fishery monitoring in Glen Canyon | \$7,467 | \$0 | \$0 | \$0 | \$88,000 | \$0 | \$4,272 | \$99,739 |
| H.2 | Experimental flow assessment of trout recruitment (TRGD) | \$78,947 | \$2,500 | \$88,800 | \$111,045 | \$0 | \$0 | \$61,499 | \$342,790 |
| H.3 | Salmonid modeling | \$37,450 | \$0 | \$6,000 | \$0 | \$0 | \$0 | \$9,499 | \$52,949 |
| | Total H | \$123,863 | \$2,500 | \$94,800 | \$111,045 | \$88,000 | \$0 | \$75,271 | \$495,479 |

| | | | | | | | | | |
|----------|--|------------------|----------------|-----------------|------------------|------------------|-----------------|-----------------|------------------|
| I | Non-native Invasive Species Monitoring and Research | | | | | | | | |
| I.1 | System-wide native fishes and invasive aquatic species monitoring | \$12,191 | \$0 | \$4,650 | \$105,350 | \$217,550 | \$0 | \$33,241 | \$372,982 |
| I.2 | Estimating kinship and spawner abundance of warm water non-natives | \$62,715 | \$0 | \$100 | \$0 | \$62,775 | \$0 | \$15,616 | \$141,206 |
| I.3 | Identifying emerging threats to the Colorado River Ecosystem using environmental DNA | \$30,311 | \$300 | \$0 | \$3,215 | \$32,400 | \$0 | \$8,367 | \$74,593 |
| I.4 | Modeling population dynamics and improving forecasting tools for smallmouth bass and other non-native fish | \$160,645 | \$0 | \$1,800 | \$0 | \$0 | \$0 | \$35,515 | \$197,961 |
| | Total I | \$265,862 | \$300 | \$6,550 | \$108,565 | \$312,725 | \$0 | \$92,740 | \$786,742 |
| J | Socioeconomic Research | | | | | | | | |
| J.1 | Integrated models for adaptive management | \$92,475 | \$2,000 | \$2,000 | \$0 | \$37,600 | \$0 | \$22,220 | \$156,295 |
| J.2 | Recreation monitoring and research | \$57,732 | \$1,000 | \$1,000 | \$0 | \$0 | \$0 | \$13,059 | \$72,791 |
| J.3 | Tribal resources research | \$0 | \$0 | \$0 | \$0 | \$0 | \$0 | \$0 | \$0 |
| | Total J | \$150,206 | \$3,000 | \$3,000 | \$0 | \$37,600 | \$0 | \$35,279 | \$229,086 |
| K | Geospatial Science, Data Management, and Technology | | | | | | | | |
| K.1 | Enterprise GIS, geospatial analysis, and processing | \$145,932 | \$4,000 | \$5,600 | \$0 | \$0 | \$0 | \$34,004 | \$189,536 |
| K.2 | Data management and database administration | \$179,715 | \$1,800 | \$3,000 | \$0 | \$0 | \$65,000 | \$40,340 | \$289,855 |
| K.3 | Data telemetry and field engineering | \$107,651 | \$3,000 | \$6,550 | \$0 | \$0 | \$0 | \$25,624 | \$142,824 |
| | Total K | \$433,298 | \$8,800 | \$15,150 | \$0 | \$0 | \$65,000 | \$99,968 | \$622,216 |
| L | Overflight Remote Sensing in Support of GCDAMP and LTEMP | | | | | | | | |
| L.1 | Analysis and interpretation of overflight remote sensing data | \$227,154 | \$7,000 | \$5,000 | \$0 | \$0 | \$0 | \$52,286 | \$291,441 |
| L.2 | Acquisition of overflight remote sensing imagery | \$0 | \$0 | \$0 | \$0 | \$0 | \$0 | \$0 | \$0 |
| L.3 | Acquisition of airborne lidar in conjunction with overflight remote sensing imagery | \$0 | \$0 | \$0 | \$0 | \$0 | \$0 | \$0 | \$0 |
| | Total L | \$227,154 | \$7,000 | \$5,000 | \$0 | \$0 | \$0 | \$52,286 | \$291,441 |

| | | | | | | | | | |
|----------|---|--------------------|-----------------|------------------|------------------|--------------------|------------------|--------------------|--------------------|
| M | Leadership, Management, and Support | | | | | | | | |
| M.1 | Leadership, management, and support | \$616,843 | \$16,000 | \$114,000 | \$0 | \$0 | \$0 | \$163,282 | \$910,125 |
| M.2 | Logistics staff | \$364,547 | \$0 | \$0 | \$0 | \$0 | \$0 | \$79,701 | \$444,248 |
| M.3 | IT | \$0 | \$0 | \$59,510 | \$0 | \$0 | \$0 | \$13,011 | \$72,521 |
| | Total M | \$981,390 | \$16,000 | \$173,510 | \$0 | \$0 | \$0 | \$255,994 | \$1,426,893 |
| N | Native Fish Population Dynamics | | | | | | | | |
| N.1 | Sucker and dace distribution and demographics (SADDAD) | \$0 | \$0 | \$0 | \$0 | \$22,217 | \$0 | \$667 | \$22,884 |
| N.2 | Predictive modeling and Decision support for native fishes | \$0 | \$0 | \$0 | \$0 | \$0 | \$0 | \$0 | \$0 |
| N.3 | Evaluating dispersal and sources of mortality (Razorback sucker) using new technology | \$0 | \$0 | \$0 | \$0 | \$0 | \$0 | \$0 | \$0 |
| | Total N | \$0 | \$0 | \$0 | \$0 | \$22,217 | \$0 | \$667 | \$22,884 |
| | Total (FY25) | \$5,095,360 | \$96,715 | \$535,961 | \$931,414 | \$1,175,910 | \$423,420 | \$1,491,233 | \$9,750,013 |
| | Anticipated AMP Funding Available (80.0% and 0% CPI) | | | | | | | | \$10,000,000 |
| | AMP Over/Under Budget | | | | | | | | \$249,987 |
| | GCMRC Grand Total | \$5,095,360 | \$96,715 | \$535,961 | \$931,414 | \$1,175,910 | \$423,420 | \$1,491,233 | \$9,750,013 |

FY 2026

| Fiscal Year 2026 | | | | | | | | | |
|------------------|--|------------------|-------------------|--------------------|--------------------|------------------------|-----------------------|------------------|--------------------|
| Project | Project Description | Salaries | Travel & Training | Operating Expenses | Logistics Expenses | Cooperative Agreements | To other USGS Centers | Burden | Total |
| A | Streamflow, Water Quality, and Sediment Transport and Budgeting in the Colorado River Ecosystem | | | | | | | 22.60% | |
| A.1 | Stream gaging and hydrologic analyses | \$160,807 | \$6,000 | \$6,000 | \$28,000 | \$0 | \$180,120 | \$45,382 | \$426,309 |
| A.2 | Continuous water quality parameters | \$120,217 | \$1,000 | \$12,000 | \$26,800 | \$0 | \$30,080 | \$36,164 | \$226,260 |
| A.3 | Sediment transport and budgeting | \$339,919 | \$6,000 | \$38,000 | \$27,000 | \$0 | \$158,960 | \$92,868 | \$662,747 |
| | Total A | \$620,943 | \$13,000 | \$56,000 | \$81,800 | \$0 | \$369,160 | \$174,414 | \$1,315,317 |
| B | Sandbar and Sediment Storage Monitoring and Research | | | | | | | | |
| B.1 | Sandbar and campsite monitoring with topographic surveys and remote cameras | \$252,959 | \$2,000 | \$4,000 | \$31,921 | \$23,500 | \$0 | \$66,444 | \$380,824 |
| B.2 | Bathymetric and topographic mapping for monitoring sediment storage and riverbed dynamics | \$399,706 | \$2,000 | \$5,000 | \$86,680 | \$37,600 | \$0 | \$112,633 | \$643,619 |
| B.3 | Control network and survey support | \$0 | \$0 | \$0 | \$0 | \$0 | \$0 | \$0 | \$0 |
| B.4 | Streamflow, sediment, and sandbar modeling | \$0 | \$0 | \$0 | \$0 | \$0 | \$0 | \$0 | \$0 |
| | Total B | \$652,664 | \$4,000 | \$9,000 | \$118,601 | \$61,100 | \$0 | \$179,077 | \$1,024,443 |
| C | Riparian Vegetation Monitoring and Research | | | | | | | | |
| C.1 | Ground-based riparian vegetation monitoring | \$131,069 | \$3,615 | \$3,500 | \$85,182 | \$13,327 | \$0 | \$50,880 | \$287,573 |
| C.2 | Determining hydrological tolerances and management tools for plant species of interest | \$14,634 | \$0 | \$0 | \$0 | \$48,681 | \$0 | \$4,768 | \$68,083 |
| C.3 | Predictive models and synthesis | \$10,337 | \$0 | \$0 | \$0 | \$24,423 | \$0 | \$3,069 | \$37,829 |
| C.4 | Biogeomorphic Linkages between streamflow, sediment transport, and vegetation composition | \$41,431 | \$0 | \$0 | \$0 | \$0 | \$0 | \$9,363 | \$50,794 |
| C.5 | Vegetation management decision support | \$0 | \$0 | \$0 | \$0 | \$0 | \$0 | \$0 | \$0 |
| | Total C | \$197,471 | \$3,615 | \$3,500 | \$85,182 | \$86,431 | \$0 | \$68,081 | \$444,279 |

| | | | | | | | | | |
|----------|---|------------------|-----------------|-----------------|------------------|------------------|------------|------------------|--------------------|
| D | Effects of Dam Operations and Vegetation Management for Archaeological Sites | | | | | | | | |
| D.1 | Monitoring the effects of dam operations on archaeological sites | \$236,012 | \$9,000 | \$15,000 | \$35,129 | \$0 | \$0 | \$66,702 | \$361,843 |
| D.2 | Monitoring landscape-scale ecosystem change with repeat photography | \$58,840 | \$3,000 | \$1,000 | \$0 | \$0 | \$0 | \$14,202 | \$77,042 |
| D.3 | Evaluating effects of LTEMP non-flow actions and other experimental vegetation management on archaeological sites | \$77,940 | \$8,000 | \$2,500 | \$12,699 | \$0 | \$0 | \$22,857 | \$123,996 |
| D.4 | Pilot study to evaluate potential to extract cultural and ecological information from Colorado River deposits using eDNA and pollen | \$0 | \$0 | \$0 | \$0 | \$0 | \$0 | \$0 | \$0 |
| D.5 | Monitoring petroglyphs and pictographs with photogrammetry and lidar | \$0 | \$0 | \$0 | \$0 | \$0 | \$0 | \$0 | \$0 |
| | Total D | \$372,792 | \$20,000 | \$18,500 | \$47,828 | \$0 | \$0 | \$103,761 | \$562,881 |
| E | Controls on Ecosystem Productivity: Nutrients, Flow, and Temperature | | | | | | | | |
| E.1 | Phosphorus budgeting in the Colorado River | \$20,358 | \$0 | \$3,500 | \$700 | \$0 | \$0 | \$5,550 | \$30,109 |
| E.2 | Rates and composition of primary producers in the Colorado River | \$100,742 | \$4,000 | \$28,676 | \$1,750 | \$0 | \$0 | \$30,548 | \$165,717 |
| E.3 | Understanding the energetic basis of the food web in Western Grand Canyon | \$0 | \$0 | \$0 | \$0 | \$0 | \$0 | \$0 | \$0 |
| E.4 | Productivity at higher trophic levels | \$69,525 | \$0 | \$3,000 | \$0 | \$0 | \$0 | \$16,391 | \$88,915 |
| | Total E | \$190,625 | \$4,000 | \$35,176 | \$2,450 | \$0 | \$0 | \$52,489 | \$284,740 |
| F | Aquatic Invertebrate Ecology | | | | | | | | |
| F.1 | Invertebrate and bat monitoring in Marble and Grand Canyons | \$234,793 | \$1,000 | \$17,000 | \$0 | \$0 | \$0 | \$57,131 | \$309,924 |
| F.2 | Aquatic invertebrate monitoring in Glen Canyon | \$122,673 | \$1,000 | \$1,500 | \$875 | \$0 | \$0 | \$28,487 | \$154,535 |
| F.3 | Aquatic invertebrate monitoring of Grand Canyon tributaries | \$36,189 | \$1,000 | \$0 | \$8,669 | \$48,175 | \$0 | \$11,809 | \$105,843 |
| F.4 | Invertebrate and fish diet studies | \$89,472 | \$1,000 | \$1,500 | \$8,669 | \$0 | \$0 | \$22,745 | \$123,386 |
| | Total F | \$483,127 | \$4,000 | \$20,000 | \$18,213 | \$48,175 | \$0 | \$120,172 | \$693,687 |
| G | Humpback Chub Population Dynamics throughout the Colorado River Ecosystem | | | | | | | | |
| G.1 | Humpback chub population modeling | \$185,221 | \$8,000 | \$5,000 | \$0 | \$0 | \$0 | \$44,798 | \$243,019 |
| G.2 | Annual spring/fall HBC abundance estimates in the lower 13.6 km of the LCR | \$2,905 | \$0 | \$20,222 | \$105,220 | \$412,700 | \$0 | \$41,388 | \$582,435 |
| G.3 | Juvenile chub monitoring near the LCR confluence (JCM-East) | \$152,119 | \$2,000 | \$29,656 | \$314,734 | \$0 | \$0 | \$112,663 | \$611,172 |
| G.4 | Remote PIT-tag array monitoring in the LCR | \$23,242 | \$0 | \$3,000 | \$3,000 | \$0 | \$0 | \$6,609 | \$35,851 |
| G.5 | Monitoring humpback chub aggregation relative abundance and distribution | \$2,905 | \$0 | \$12,436 | \$84,639 | \$145,034 | \$0 | \$26,947 | \$271,960 |
| G.6 | Juvenile chub monitoring - Western Grand Canyon (JCM-West) | \$0 | \$0 | \$0 | \$0 | \$0 | \$0 | \$0 | \$0 |
| G.7 | Chute Falls translocations | \$0 | \$0 | \$0 | \$0 | \$0 | \$0 | \$0 | \$0 |
| G.8 | Sampling of springs in the upper LCR | \$0 | \$0 | \$0 | \$0 | \$0 | \$0 | \$0 | \$0 |
| G.9 | Movement in western Grand Canyon from system-wide antenna monitoring | \$0 | \$0 | \$0 | \$0 | \$0 | \$0 | \$0 | \$0 |
| | Total G | \$366,393 | \$10,000 | \$70,314 | \$507,593 | \$557,734 | \$0 | \$232,404 | \$1,744,438 |

| | | | | | | | | | |
|----------|--|------------------|----------------|-----------------|------------------|------------------|-----------------|------------------|------------------|
| H | Salmonid Research and Monitoring | | | | | | | | |
| H.1 | Rainbow trout fishery monitoring in Glen Canyon | \$7,127 | \$0 | \$0 | \$0 | \$88,000 | \$0 | \$4,251 | \$99,378 |
| H.2 | Experimental flow assessment of trout recruitment (TRGD) | \$78,795 | \$2,500 | \$68,800 | \$115,875 | \$0 | \$0 | \$60,109 | \$326,080 |
| H.3 | Salmonid modeling | \$68,545 | \$0 | \$6,000 | \$0 | \$0 | \$0 | \$16,847 | \$91,392 |
| | Total H | \$154,467 | \$2,500 | \$74,800 | \$115,875 | \$88,000 | \$0 | \$81,207 | \$516,850 |
| I | Non-native Invasive Species Monitoring and Research | | | | | | | | |
| I.1 | System-wide native fishes and invasive aquatic species monitoring | \$10,443 | \$0 | \$4,650 | \$110,692 | \$217,550 | \$0 | \$34,954 | \$378,289 |
| I.2 | Estimating kinship and spawner abundance of warm water non-natives | \$74,021 | \$0 | \$100 | \$0 | \$47,418 | \$0 | \$18,174 | \$139,713 |
| I.3 | Identifying emerging threats to the Colorado River Ecosystem using environmental DNA | \$24,951 | \$300 | \$0 | \$2,626 | \$23,328 | \$0 | \$7,000 | \$58,205 |
| I.4 | Modeling population dynamics and improving forecasting tools for smallmouth bass and other non-native fish | \$202,107 | \$2,000 | \$1,800 | \$0 | \$0 | \$0 | \$46,535 | \$252,442 |
| | Total I | \$311,522 | \$2,300 | \$6,550 | \$113,318 | \$288,296 | \$0 | \$106,663 | \$828,649 |
| J | Socioeconomic Research | | | | | | | | |
| J.1 | Integrated models for adaptive management | \$123,809 | \$2,000 | \$2,000 | \$0 | \$0 | \$0 | \$28,885 | \$156,694 |
| J.2 | Recreation monitoring and research | \$90,927 | \$1,000 | \$1,000 | \$0 | \$0 | \$0 | \$21,002 | \$113,929 |
| J.3 | Tribal resources research | \$0 | \$0 | \$0 | \$0 | \$0 | \$0 | \$0 | \$0 |
| | Total J | \$214,736 | \$3,000 | \$3,000 | \$0 | \$0 | \$0 | \$49,886 | \$270,622 |
| K | Geospatial Science, Data Management, and Technology | | | | | | | | |
| K.1 | Enterprise GIS, geospatial analysis, and processing | \$153,229 | \$4,000 | \$5,600 | \$0 | \$0 | \$0 | \$36,799 | \$199,628 |
| K.2 | Data management and database administration | \$188,700 | \$1,800 | \$3,000 | \$0 | \$0 | \$65,000 | \$43,731 | \$302,231 |
| K.3 | Data telemetry and field engineering | \$113,033 | \$3,000 | \$6,550 | \$0 | \$0 | \$0 | \$27,704 | \$150,287 |
| | Total K | \$454,963 | \$8,800 | \$15,150 | \$0 | \$0 | \$65,000 | \$108,234 | \$652,147 |
| L | Overflight Remote Sensing in Support of GCDAMP and LTEMP | | | | | | | | |
| L.1 | Analysis and interpretation of overflight remote sensing data | \$245,048 | \$7,000 | \$5,000 | \$0 | \$0 | \$0 | \$58,093 | \$315,141 |
| L.2 | Acquisition of overflight remote sensing imagery | \$0 | \$0 | \$0 | \$0 | \$0 | \$0 | \$0 | \$0 |
| L.3 | Acquisition of airborne lidar in conjunction with overflight remote sensing imagery | \$0 | \$0 | \$0 | \$0 | \$0 | \$0 | \$0 | \$0 |
| | Total L | \$245,048 | \$7,000 | \$5,000 | \$0 | \$0 | \$0 | \$58,093 | \$315,141 |

| | | | | | | | | | |
|----------|---|--------------------|-----------------|------------------|--------------------|--------------------|------------------|--------------------|---------------------|
| M | Leadership, Management, and Support | | | | | | | | |
| M.1 | Leadership, management, and support | \$661,313 | \$16,000 | \$114,000 | \$0 | \$0 | \$0 | \$178,837 | \$970,150 |
| M.2 | Logistics staff | \$382,774 | \$0 | \$0 | \$0 | \$0 | \$0 | \$86,507 | \$469,281 |
| M.3 | IT | \$0 | \$0 | \$57,250 | \$0 | \$0 | \$0 | \$12,939 | \$70,189 |
| | Total M | \$1,044,088 | \$16,000 | \$171,250 | \$0 | \$0 | \$0 | \$278,282 | \$1,509,620 |
| N | Native Fish Population Dynamics | | | | | | | | |
| N.1 | Sucker and dace distribution and demographics (SADDAD) | \$0 | \$0 | \$0 | \$0 | \$0 | \$0 | \$0 | \$0 |
| N.2 | Predictive modeling and Decision support for native fishes | \$0 | \$0 | \$0 | \$0 | \$0 | \$0 | \$0 | \$0 |
| N.3 | Evaluating dispersal and sources of mortality (Razorback sucker) using new technology | \$0 | \$0 | \$0 | \$0 | \$0 | \$0 | \$0 | \$0 |
| | Total N | \$0 | \$0 | \$0 | \$0 | \$0 | \$0 | \$0 | \$0 |
| | Total (FY26) | \$5,308,839 | \$98,215 | \$488,240 | \$1,090,860 | \$1,129,736 | \$434,160 | \$1,612,763 | \$10,162,813 |
| | Anticipated AMP Funding Available (80.0% and 0% CPI) | | | | | | | | \$10,000,000 |
| | AMP Over/Under Budget | | | | | | | | (\$162,813) |
| | GCMRC Grand Total | \$5,308,839 | \$98,215 | \$488,240 | \$1,090,860 | \$1,129,736 | \$434,160 | \$1,612,763 | \$10,162,813 |

FY 2027

| Fiscal Year 2027 | | | | | | | | | |
|------------------|--|------------------|-------------------|--------------------|--------------------|------------------------|-----------------------|------------------|--------------------|
| Project | Project Description | Salaries | Travel & Training | Operating Expenses | Logistics Expenses | Cooperative Agreements | To other USGS Centers | Burden | Total |
| | | | | | | | | 23.40% | |
| A | Streamflow, Water Quality, and Sediment Transport and Budgeting in the Colorado River Ecosystem | | | | | | | | |
| A.1 | Stream gaging and hydrologic analyses | \$168,847 | \$6,000 | \$6,000 | \$29,000 | \$0 | \$185,800 | \$49,104 | \$444,752 |
| A.2 | Continuous water quality parameters | \$130,384 | \$1,000 | \$12,000 | \$27,400 | \$0 | \$30,900 | \$39,963 | \$241,647 |
| A.3 | Sediment transport and budgeting | \$356,915 | \$6,000 | \$38,000 | \$28,000 | \$0 | \$165,900 | \$100,366 | \$695,181 |
| | Total A | \$656,146 | \$13,000 | \$56,000 | \$84,400 | \$0 | \$382,600 | \$189,434 | \$1,381,580 |
| B | Sandbar and Sediment Storage Monitoring and Research | | | | | | | | |
| B.1 | Sandbar and campsite monitoring with topographic surveys and remote cameras | \$260,273 | \$2,000 | \$4,000 | \$32,895 | \$23,500 | \$0 | \$70,710 | \$393,378 |
| B.2 | Bathymetric and topographic mapping for monitoring sediment storage and riverbed dynamics | \$431,301 | \$2,000 | \$2,000 | \$0 | \$18,800 | \$0 | \$102,424 | \$556,526 |
| B.3 | Control network and survey support | \$0 | \$0 | \$0 | \$0 | \$0 | \$0 | \$0 | \$0 |
| B.4 | Streamflow, sediment, and sandbar modeling | \$0 | \$0 | \$0 | \$0 | \$0 | \$0 | \$0 | \$0 |
| | Total B | \$691,574 | \$4,000 | \$6,000 | \$32,895 | \$42,300 | \$0 | \$173,135 | \$949,904 |
| C | Riparian Vegetation Monitoring and Research | | | | | | | | |
| C.1 | Ground-based riparian vegetation monitoring | \$141,592 | \$1,615 | \$3,500 | \$88,103 | \$14,112 | \$0 | \$55,369 | \$304,291 |
| C.2 | Determining hydrological tolerances and management tools for plant species of interest | \$11,397 | \$0 | \$0 | \$0 | \$48,820 | \$0 | \$4,131 | \$64,348 |
| C.3 | Predictive models and synthesis | \$10,854 | \$0 | \$0 | \$0 | \$25,156 | \$0 | \$3,294 | \$39,304 |
| C.4 | Biogeomorphic Linkages between streamflow, sediment transport, and vegetation composition | \$43,503 | \$0 | \$0 | \$0 | \$0 | \$0 | \$10,180 | \$53,682 |
| C.5 | Vegetation management decision support | \$0 | \$0 | \$0 | \$0 | \$0 | \$0 | \$0 | \$0 |
| | Total C | \$207,345 | \$1,615 | \$3,500 | \$88,103 | \$88,088 | \$0 | \$72,974 | \$461,625 |

| | | | | | | | | | |
|----------|---|------------------|-----------------|-----------------|-----------------|-----------------|------------|------------------|------------------|
| D | Effects of Dam Operations and Vegetation Management for Archaeological Sites | | | | | | | | |
| D.1 | Monitoring the effects of dam operations on archaeological sites | \$247,813 | \$9,000 | \$15,000 | \$36,257 | \$0 | \$0 | \$72,088 | \$380,158 |
| D.2 | Monitoring landscape-scale ecosystem change with repeat photography | \$61,782 | \$2,000 | \$1,000 | \$0 | \$0 | \$0 | \$15,159 | \$79,941 |
| D.3 | Evaluating effects of LTEMP non-flow actions and other experimental vegetation management on archaeological sites | \$81,837 | \$7,500 | \$2,500 | \$0 | \$0 | \$0 | \$21,490 | \$113,326 |
| D.4 | Pilot study to evaluate potential to extract cultural and ecological information from Colorado River deposits using eDNA and pollen | \$0 | \$0 | \$0 | \$0 | \$0 | \$0 | \$0 | \$0 |
| D.5 | Monitoring petroglyphs and pictographs with photogrammetry and lidar | \$0 | \$0 | \$0 | \$0 | \$0 | \$0 | \$0 | \$0 |
| | Total D | \$391,431 | \$18,500 | \$18,500 | \$36,257 | \$0 | \$0 | \$108,737 | \$573,425 |
| E | Controls on Ecosystem Productivity: Nutrients, Flow, and Temperature | | | | | | | | |
| E.1 | Phosphorus budgeting in the Colorado River | \$44,632 | \$0 | \$3,500 | \$0 | \$0 | \$0 | \$11,263 | \$59,395 |
| E.2 | Rates and composition of primary producers in the Colorado River | \$63,290 | \$0 | \$18,044 | \$1,750 | \$0 | \$0 | \$19,442 | \$102,526 |
| E.3 | Understanding the energetic basis of the food web in Western Grand Canyon | \$0 | \$0 | \$0 | \$0 | \$0 | \$0 | \$0 | \$0 |
| E.4 | Productivity at higher trophic levels | \$50,485 | \$0 | \$3,000 | \$0 | \$0 | \$0 | \$12,516 | \$66,001 |
| | Total E | \$158,408 | \$0 | \$24,544 | \$1,750 | \$0 | \$0 | \$43,220 | \$227,922 |
| F | Aquatic Invertebrate Ecology | | | | | | | | |
| F.1 | Invertebrate and bat monitoring in Marble and Grand Canyons | \$171,354 | \$1,000 | \$17,000 | \$0 | \$0 | \$0 | \$44,309 | \$233,663 |
| F.2 | Aquatic invertebrate monitoring in Glen Canyon | \$139,043 | \$1,000 | \$1,500 | \$875 | \$0 | \$0 | \$33,326 | \$175,744 |
| F.3 | Aquatic invertebrate monitoring of Grand Canyon tributaries | \$42,302 | \$1,000 | \$0 | \$8,961 | \$49,350 | \$0 | \$13,710 | \$115,324 |
| F.4 | Invertebrate and fish diet studies | \$76,487 | \$1,000 | \$1,500 | \$8,961 | \$0 | \$0 | \$20,580 | \$108,527 |
| | Total F | \$429,186 | \$4,000 | \$20,000 | \$18,797 | \$49,350 | \$0 | \$111,924 | \$633,257 |

| | | | | | | | | | |
|----------|--|------------------|-----------------|-----------------|------------------|------------------|------------|------------------|--------------------|
| G | Humpback Chub Population Dynamics throughout the Colorado River Ecosystem | | | | | | | | |
| G.1 | Humpback chub population modeling | \$194,482 | \$8,000 | \$5,000 | \$0 | \$0 | \$0 | \$48,551 | \$256,033 |
| G.2 | Annual spring/fall HBC abundance estimates in the lower 13.6 km of the LCR | \$3,051 | \$0 | \$20,222 | \$108,272 | \$422,944 | \$0 | \$43,470 | \$597,958 |
| G.3 | Juvenile chub monitoring near the LCR confluence (JCM-East) | \$153,851 | \$2,000 | \$30,156 | \$324,740 | \$0 | \$0 | \$119,515 | \$630,262 |
| G.4 | Remote PIT-tag array monitoring in the LCR | \$24,405 | \$0 | \$3,000 | \$3,000 | \$0 | \$0 | \$7,115 | \$37,519 |
| G.5 | Monitoring humpback chub aggregation relative abundance and distribution | \$3,051 | \$0 | \$12,436 | \$66,912 | \$147,083 | \$0 | \$23,694 | \$253,176 |
| G.6 | Juvenile chub monitoring - Western Grand Canyon (JCM-West) | \$0 | \$0 | \$0 | \$0 | \$0 | \$0 | \$0 | \$0 |
| G.7 | Chute Falls translocations | \$0 | \$0 | \$0 | \$0 | \$0 | \$0 | \$0 | \$0 |
| G.8 | Sampling of springs in the upper LCR | \$0 | \$0 | \$0 | \$0 | \$0 | \$0 | \$0 | \$0 |
| G.9 | Movement in western Grand Canyon from system-wide antenna monitoring | \$0 | \$0 | \$0 | \$0 | \$0 | \$0 | \$0 | \$0 |
| | Total G | \$378,839 | \$10,000 | \$70,814 | \$502,924 | \$570,027 | \$0 | \$242,344 | \$1,774,948 |
| H | Salmonid Research and Monitoring | | | | | | | | |
| H.1 | Rainbow trout fishery monitoring in Glen Canyon | \$7,484 | \$0 | \$0 | \$0 | \$88,000 | \$0 | \$4,391 | \$99,875 |
| H.2 | Experimental flow assessment of trout recruitment (TRGD) | \$82,735 | \$2,500 | \$68,800 | \$118,998 | \$0 | \$0 | \$63,890 | \$336,923 |
| H.3 | Salmonid modeling | \$71,972 | \$0 | \$6,000 | \$0 | \$0 | \$0 | \$18,245 | \$96,217 |
| | Total H | \$162,191 | \$2,500 | \$74,800 | \$118,998 | \$88,000 | \$0 | \$86,526 | \$533,015 |
| I | Non-native Invasive Species Monitoring and Research | | | | | | | | |
| I.1 | System-wide native fishes and invasive aquatic species monitoring | \$10,965 | \$0 | \$4,650 | \$114,150 | \$217,550 | \$0 | \$36,892 | \$384,207 |
| I.2 | Estimating kinship and spawner abundance of warm water non-natives | \$12,892 | \$2,000 | \$100 | \$0 | \$0 | \$0 | \$3,508 | \$18,500 |
| I.3 | Identifying emerging threats to the Colorado River Ecosystem using environmental DNA | \$33,418 | \$300 | \$0 | \$17,104 | \$48,600 | \$0 | \$13,350 | \$112,772 |
| I.4 | Modeling population dynamics and improving forecasting tools for smallmouth bass and other non-native fish | \$212,212 | \$0 | \$1,800 | \$0 | \$0 | \$0 | \$50,079 | \$264,091 |
| | Total I | \$269,487 | \$2,300 | \$6,550 | \$131,254 | \$266,150 | \$0 | \$103,829 | \$779,570 |

| | | | | | | | | | |
|----------|---|--------------------|-----------------|------------------|--------------------|--------------------|------------------|--------------------|---------------------|
| J | Socioeconomic Research | | | | | | | | |
| J.1 | Integrated models for adaptive management | \$129,999 | \$2,000 | \$2,000 | \$0 | \$0 | \$0 | \$31,356 | \$165,355 |
| J.2 | Recreation monitoring and research | \$111,386 | \$1,000 | \$1,000 | \$0 | \$0 | \$0 | \$26,532 | \$139,918 |
| J.3 | Tribal resources research | \$0 | \$0 | \$0 | \$0 | \$0 | \$0 | \$0 | \$0 |
| | Total J | \$241,385 | \$3,000 | \$3,000 | \$0 | \$0 | \$0 | \$57,888 | \$305,273 |
| K | Geospatial Science, Data Management, and Technology | | | | | | | | |
| K.1 | Enterprise GIS, geospatial analysis, and processing | \$168,704 | \$4,000 | \$5,600 | \$0 | \$0 | \$0 | \$41,723 | \$220,028 |
| K.2 | Data management and database administration | \$276,796 | \$1,800 | \$3,000 | \$0 | \$0 | \$0 | \$65,893 | \$347,489 |
| K.3 | Data telemetry and field engineering | \$126,499 | \$3,500 | \$6,550 | \$0 | \$0 | \$0 | \$31,952 | \$168,501 |
| | Total K | \$571,999 | \$9,300 | \$15,150 | \$0 | \$0 | \$0 | \$139,569 | \$736,018 |
| L | Overflight Remote Sensing in Support of GCDAMP and LTEMP | | | | | | | | |
| L.1 | Analysis and interpretation of overflight remote sensing data | \$257,300 | \$7,000 | \$5,000 | \$0 | \$0 | \$0 | \$63,016 | \$332,316 |
| L.2 | Acquisition of overflight remote sensing imagery | \$0 | \$0 | \$0 | \$0 | \$0 | \$0 | \$0 | \$0 |
| L.3 | Acquisition of airborne lidar in conjunction with overflight remote sensing imagery | \$0 | \$0 | \$0 | \$0 | \$0 | \$0 | \$0 | \$0 |
| | Total L | \$257,300 | \$7,000 | \$5,000 | \$0 | \$0 | \$0 | \$63,016 | \$332,316 |
| M | Leadership, Management, and Support | | | | | | | | |
| M.1 | Leadership, management, and support | \$695,066 | \$16,000 | \$114,000 | \$0 | \$0 | \$0 | \$193,066 | \$1,018,132 |
| M.2 | Logistics staff | \$401,913 | \$0 | \$0 | \$0 | \$0 | \$0 | \$94,048 | \$495,961 |
| M.3 | IT | \$0 | \$0 | \$65,750 | \$0 | \$0 | \$0 | \$15,386 | \$81,136 |
| | Total M | \$1,096,980 | \$16,000 | \$179,750 | \$0 | \$0 | \$0 | \$302,499 | \$1,595,228 |
| N | Native Fish Population Dynamics | | | | | | | | |
| N.1 | Sucker and dace distribution and demographics (SADDAD) | \$0 | \$0 | \$0 | \$0 | \$0 | \$0 | \$0 | \$0 |
| N.2 | Predictive modeling and Decision support for native fishes | \$0 | \$0 | \$0 | \$0 | \$0 | \$0 | \$0 | \$0 |
| N.3 | Evaluating dispersal and sources of mortality (Razorback sucker) using new technology | \$0 | \$0 | \$0 | \$0 | \$0 | \$0 | \$0 | \$0 |
| | Total N | \$0 | \$0 | \$0 | \$0 | \$0 | \$0 | \$0 | \$0 |
| | GCMRC AMP Total (FY27) | \$5,512,272 | \$91,215 | \$483,608 | \$1,015,378 | \$1,103,915 | \$382,600 | \$1,695,096 | \$10,284,083 |
| | Anticipated AMP Funding Available (80.0% and 0% CPI) | | | | | | | | \$10,000,000 |
| | AMP Over/Under Budget | | | | | | | | (\$284,083) |
| | GCMRC Grand Total | \$5,512,272 | \$91,215 | \$483,608 | \$1,015,378 | \$1,103,915 | \$382,600 | \$1,695,096 | \$10,284,083 |

Appendix 3. Experimental Fund Summaries by Year

FY 2025

| Fiscal Year 2025 | | | | | | | | |
|---|------------------|-------------------|--------------------|--------------------|------------------------|-----------------------|------------------|------------------|
| Experimental Fund Projects | Salaries | Travel & Training | Operating Expenses | Logistics Expenses | Cooperative Agreements | To other USGS Centers | Burden | Total |
| | | | | | | | 21.86% | |
| A.4. HFE Experimental Fund | \$117,666 | \$2,000 | \$20,000 | \$45,200 | \$0 | \$0 | \$40,417 | \$225,283 |
| A.5. Real-time water temperature in Marble Canyon | \$0 | \$0 | \$0 | \$0 | \$0 | \$20,000 | \$0 | \$20,000 |
| Total A | \$117,666 | \$2,000 | \$20,000 | \$45,200 | \$0 | \$20,000 | \$40,417 | \$245,283 |
| B.5.1. Extended duration Fall HFE (daily surveys during HFE + 1 set sandbar surveys w/o bathymetry) | \$39,793 | \$3,000 | \$1,000 | \$96,787 | \$69,325 | \$0 | \$32,815 | \$242,719 |
| B.5.2. Proactive Spring HFE (1 set of sandbar surveys w/o bathymetry) | \$36,569 | \$1,500 | \$1,000 | \$55,144 | \$36,425 | \$0 | \$21,691 | \$152,329 |
| B.5.3. Proactive Spring HFE (1 set of sandbar surveys with bathymetry) | \$36,569 | \$1,500 | \$1,000 | \$61,858 | \$69,325 | \$0 | \$24,145 | \$194,398 |
| B.5.4. Variation in HFE downramp rate (2 sets of sandbar surveys w/o bathymetry) | \$43,016 | \$3,000 | \$1,000 | \$117,611 | \$69,325 | \$0 | \$38,072 | \$272,025 |
| B.5.5. Channel response to HFE in western Grand Canyon (4 surveys around fall HFE) | \$33,346 | \$1,000 | \$1,000 | \$4,587 | \$22,325 | \$0 | \$9,400 | \$71,658 |
| Total B | \$189,294 | \$10,000 | \$5,000 | \$335,986 | \$266,725 | \$0 | \$126,123 | \$933,128 |
| C.5. Experimental Vegetation Treatment Support | \$9,845 | \$0 | \$0 | \$0 | \$4,001 | \$0 | \$2,272 | \$16,119 |
| C.6. Plant physiological responses to experimental flows | \$0 | \$450 | \$400 | \$9,362 | \$0 | \$0 | \$2,233 | \$12,445 |
| C.7. Effects of plants on flow velocity and sand deposition | \$55,869 | \$0 | \$0 | \$14,932 | \$0 | \$0 | \$15,479 | \$86,280 |
| Total C | \$65,714 | \$450 | \$400 | \$24,294 | \$4,001 | \$0 | \$19,984 | \$114,843 |
| D.6. Post-HFE surveys | \$29,541 | \$250 | \$0 | \$0 | \$0 | \$0 | \$6,513 | \$36,304 |
| Total D | \$29,541 | \$250 | \$0 | \$0 | \$0 | \$0 | \$6,513 | \$36,304 |
| I.5. Evaluating the efficacy of flow experiments in the LTEMP sEIS to control smallmouth bass | \$30,425 | \$1,800 | \$5,200 | \$38,142 | \$94,478 | \$0 | \$19,356 | \$189,401 |
| I.6. Determining hatch dates of larval smallmouth bass in response to LTEMP sEIS flow experiments | \$19,116 | \$1,500 | \$200 | \$11,102 | \$117,500 | \$0 | \$10,503 | \$159,921 |
| Total I | \$49,541 | \$3,300 | \$5,400 | \$49,244 | \$211,978 | \$0 | \$29,859 | \$349,322 |

FY 2026

| Fiscal Year 2026 | | | | | | | | |
|---|------------------|-------------------|--------------------|--------------------|------------------------|-----------------------|------------------|--------------------|
| Experimental Fund Projects | Salaries | Travel & Training | Operating Expenses | Logistics Expenses | Cooperative Agreements | To other USGS Centers | Burden | Total |
| | | | | | | | 22.60% | |
| A.4. HFE Experimental Fund | \$123,549 | \$2,050 | \$20,000 | \$46,700 | \$0 | \$0 | \$43,460 | \$235,758 |
| A.5. Real-time water temperature in Marble Canyon | \$0 | \$0 | \$0 | \$0 | \$0 | \$20,000 | \$0 | \$20,000 |
| Total A | \$123,549 | \$2,050 | \$20,000 | \$46,700 | \$0 | \$20,000 | \$43,460 | \$255,758 |
| B.5.1. Extended duration Fall HFE (daily surveys during HFE + 1 set sandbar surveys w/o bathymetry) | \$41,783 | \$3,000 | \$1,000 | \$99,980 | \$69,325 | \$0 | \$35,022 | \$250,109 |
| B.5.2. Proactive Spring HFE (1 set of sandbar surveys w/o bathymetry) | \$38,398 | \$1,500 | \$1,000 | \$56,966 | \$36,425 | \$0 | \$23,210 | \$157,499 |
| B.5.3. Proactive Spring HFE (1 set of sandbar surveys with bathymetry) | \$38,398 | \$1,500 | \$1,000 | \$63,843 | \$69,325 | \$0 | \$25,751 | \$199,817 |
| B.5.4. Variation in HFE downramp rate (2 sets of sandbar surveys w/o bathymetry) | \$45,167 | \$3,000 | \$1,000 | \$113,932 | \$69,325 | \$0 | \$38,940 | \$271,365 |
| B.5.5. Channel response to HFE in western Grand Canyon (4 surveys around fall HFE) | \$35,013 | \$1,000 | \$1,000 | \$4,667 | \$22,325 | \$0 | \$10,089 | \$74,095 |
| Total B | \$198,759 | \$10,000 | \$5,000 | \$339,388 | \$266,725 | \$0 | \$133,013 | \$952,885 |
| C.5. Experimental Vegetation Treatment Support | \$10,337 | \$0 | \$0 | \$0 | \$4,884 | \$0 | \$2,483 | \$17,704 |
| C.6. Plant physiological responses to experimental flows | \$0 | \$450 | \$400 | \$9,774 | \$0 | \$0 | \$2,401 | \$13,025 |
| C.7. Effects of plants on flow velocity and sand deposition | \$58,663 | \$0 | \$0 | \$15,432 | \$0 | \$0 | \$16,745 | \$90,840 |
| Total C | \$69,000 | \$450 | \$400 | \$25,206 | \$4,884 | \$0 | \$21,629 | \$121,568 |
| D.6. Post-HFE surveys | \$31,018 | \$250 | \$0 | \$0 | \$0 | \$0 | \$7,067 | \$38,335 |
| Total D | \$31,018 | \$250 | \$0 | \$0 | \$0 | \$0 | \$7,067 | \$38,335 |
| I.5. Evaluating the efficacy of flow experiments in the LTEMP sEIS to control smallmouth bass | \$31,946 | \$1,800 | \$5,200 | \$39,386 | \$94,478 | \$0 | \$20,537 | \$193,348 |
| I.6. Determining hatch dates of larval smallmouth bass in response to LTEMP sEIS flow experiments | \$20,072 | \$1,500 | \$200 | \$11,316 | \$117,500 | \$0 | \$11,003 | \$161,590 |
| Total I | \$52,018 | \$3,300 | \$5,400 | \$50,702 | \$211,978 | \$0 | \$31,540 | \$354,938 |
| Total Experimental Fund (FY26) | \$474,343 | \$16,050 | \$30,800 | \$461,996 | \$483,587 | \$20,000 | \$236,708 | \$1,723,485 |

FY 2027

| Fiscal Year 2027 | | | | | | | | |
|---|------------------|-------------------|--------------------|--------------------|------------------------|-----------------------|------------------|--------------------|
| Experimental Fund Projects | Salaries | Travel & Training | Operating Expenses | Logistics Expenses | Cooperative Agreements | To other USGS Centers | Burden | Total |
| | | | | | | | 23.40% | |
| A.4. HFE Experimental Fund | \$129,726 | \$2,100 | \$20,000 | \$48,100 | \$0 | \$0 | \$46,783 | \$246,709 |
| A.5. Real-time water temperature in Marble Canyon | \$0 | \$0 | \$0 | \$0 | \$0 | \$20,000 | \$0 | \$20,000 |
| Total A | \$129,726 | \$2,100 | \$20,000 | \$48,100 | \$0 | \$20,000 | \$46,783 | \$266,709 |
| B.5.1. Extended duration Fall HFE (daily surveys during HFE + 1 set sandbar surveys w/o bathymetry) | \$43,872 | \$3,000 | \$1,000 | \$103,201 | \$69,325 | \$0 | \$37,431 | \$257,829 |
| B.5.2. Proactive Spring HFE (1 set of sandbar surveys w/o bathymetry) | \$40,318 | \$1,500 | \$1,000 | \$58,806 | \$36,425 | \$0 | \$24,873 | \$162,921 |
| B.5.3. Proactive Spring HFE (1 set of sandbar surveys with bathymetry) | \$40,318 | \$1,500 | \$1,000 | \$65,259 | \$69,325 | \$0 | \$27,370 | \$204,772 |
| B.5.4. Variation in HFE downramp rate (2 sets of sandbar surveys w/o bathymetry) | \$47,426 | \$3,000 | \$1,000 | \$117,611 | \$69,325 | \$0 | \$41,634 | \$279,996 |
| B.5.5. Channel response to HFE in western Grand Canyon (4 surveys around fall HFE) | \$36,764 | \$1,000 | \$1,000 | \$4,749 | \$22,325 | \$0 | \$10,852 | \$76,689 |
| Total B | \$208,697 | \$10,000 | \$5,000 | \$349,626 | \$266,725 | \$0 | \$142,159 | \$982,206 |
| C.5. Experimental Vegetation Treatment Support | \$10,854 | \$0 | \$0 | \$0 | \$5,032 | \$0 | \$2,691 | \$18,576 |
| C.6. Plant physiological responses to experimental flows | \$0 | \$450 | \$400 | \$10,118 | \$0 | \$0 | \$2,567 | \$13,535 |
| C.7. Effects of plants on flow velocity and sand deposition | \$61,596 | \$0 | \$0 | \$15,933 | \$0 | \$0 | \$18,142 | \$95,671 |
| Total C | \$72,450 | \$450 | \$400 | \$26,051 | \$5,032 | \$0 | \$23,399 | \$127,782 |
| D.6. Post-HFE surveys | \$32,569 | \$250 | \$0 | \$0 | \$0 | \$0 | \$7,680 | \$40,499 |
| Total D | \$32,569 | \$250 | \$0 | \$0 | \$0 | \$0 | \$7,680 | \$40,499 |
| I.5. Evaluating the efficacy of flow experiments in the LTEMP sEIS to control smallmouth bass | \$33,543 | \$1,800 | \$5,200 | \$40,638 | \$94,478 | \$0 | \$21,831 | \$197,491 |
| I.6. Determining hatch dates of larval smallmouth bass in response to LTEMP sEIS flow experiments | \$21,075 | \$1,500 | \$200 | \$11,536 | \$117,500 | \$0 | \$11,554 | \$163,365 |
| Total I | \$54,619 | \$3,300 | \$5,400 | \$52,174 | \$211,978 | \$0 | \$33,385 | \$360,856 |
| Total Experimental Fund (FY27) | \$498,060 | \$16,100 | \$30,800 | \$475,951 | \$483,735 | \$20,000 | \$253,405 | \$1,778,052 |